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In situ growth of CuS NPs on 3D porous cellulose macrospheres as recyclable biocatalysts for organic dye degradation

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Aiming at recyclable catalyst carriers, porous cellulose macrospheres from wood pulp dissolved in an alkaline urea system were regenerated by simple injection regeneration. After solvent exchange, porous cellulose macrospheres (CMs) with a high specific surface area of $325.3 \text{ m}^2 \text{ g}^{-1}$ were obtained by lyophilization, and CuS nanoparticles (CuS NPs) were coated on CMs by *in situ* growth in the liquid phase to achieve CuS-supported CM macrospheres (CuS@CM). The results indicated that the CuS@CM biocatalyst was successfully prepared with an average diameter of approximately 1.2 mm. In addition, CuS@CM was further used as a heterogeneous catalyst for the catalytic degradation of methylene blue (MB) and methyl orange (MO) model dyes during the oxidation of hydrogen peroxide (H_2O_2). In the presence of low doses of H_2O_2 , the degradation rate of MB reached 94.8% within 10 min, showing high catalytic activity under neutral and alkaline conditions. After five cycles, 90.1% of the original catalytic activity was still retained, indicating that the prepared CuS@CM composite possessed excellent catalytic activity and reusability.

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1. Introduction

With the rapid development of textile industries accompanied by high energy consumption and increased waste discharge, industrial wastewater containing a great deal of organic dye pollutants has an adverse effect on the environment, aquatic organisms and human health. Because of its non-biodegradability, toxicity, possible carcinogenicity and mutagenicity,¹ waste discharge poses a serious threat to the aquatic ecosystem. In particular, the treatment of dye wastewater containing a large number of persistent and toxic substances with various chemical structures is a difficult and challenging task.^{2,3} To this end, various wastewater treatment technologies of adsorption,⁴ photocatalytic degradation,⁵ ozonization,⁶ electrochemistry,⁷ biological methods,⁸ and others,^{9,10} have been used to treat dye wastewater.^{11–13} However, most of these methods have a lower removal efficiency and high processing costs. Fenton-like reagents, which used transition metal oxides or sulfides as catalysts to release hydroxyl groups, present

extremely high reactivity and have proved to be effective materials for converting some organic dye macromolecules into non-toxic carbon dioxide and water,^{14–16} and hence, they are widely used in the field of wastewater treatment. However, in practical applications, the catalyst is usually dispersed in a solution in the form of fine particles, which makes it more difficult to separate and recover the catalyst in the form of fine powder.

The effective solution to this problem is to find a suitable supporting substrate for the catalyst. Cellulose, as the most abundant natural polymer material in nature,^{17,18} usually exists in the form of a variety of biomass such as cotton, wood pulp and flax.^{19–21} It is environmentally friendly, non-toxic and biodegradable as a supporting substrate material. Therefore, many prepared cellulose composites such as Pickering emulsions,²² cellulose composite membranes,²³ functionalized nanocrystals,²⁴ cellulose fibers and fabrics,^{25,26} hydrogels and aerogels^{27–29} have been reported. Cellulose gel is a three-dimensional mesh structure material, which has a high specific surface area,³⁰ and abundant active sites that could increase the contact opportunity between contaminants and catalysts, and the pores in the structure also facilitated the mass transport of reactants and products.³¹ In general, spherical porous materials possess better fluidity and a higher specific surface area, making cellulose into a spherical form, and can not only increase the catalyst load, but also more easily disperse evenly in the wastewater solution during the treatment process.

It has been reported that copper sulfide (CuS), as a typical p-type semiconductor, has been widely used in various fields such

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as sensing,³² catalysis,³³ solar cell fabrication,³⁴ and biomedicine³⁵ because of its excellent optical, electronic and other physical and chemical properties.^{36,37} In recent years, some researchers have shown that the one-dimensional or three-dimensional structure of copper sulfide is desirable for the photocatalytic activity, while all its structures are suitable for the catalytic activity.^{38–41} Therefore, the combination of copper sulfide and porous cellulose gel spheres was conducive to reusability and maximum utilization of the catalyst.

In this work, we used an alkaline urea system to dissolve wood pulp cellulose^{42,43} and form gel spheres by simple injection regeneration. After solvent exchange, porous cellulose microspheres (CM) with a high specific surface area were obtained by lyophilization, and CuS was loaded on CM by *in situ* growth (CuS@CM). The prepared CuS immobilized on CM microspheres in the presence of H₂O₂ showed excellent catalytic activity for the degradation of methylene blue (MB) and methyl orange (MO) due to its wide application and toxicity. In addition, CuS@CM has a wide range of pH applicability under environmental conditions, and still maintains more than 90% of the catalytic efficiency after five cycles. Therefore, the CuS@CM composite could be used as a standard candidate for recoverable Fenton reagents.

2. Materials and methods

2.1 Materials

Cellulose (wood pulp board, DP500, purity of 92.8%) was supplied by Baoding Swan New Fiber Manufacturing Co., Ltd (China). Lithium hydroxide (LiOH), sodium hydroxide (NaOH), nitric acid (HNO₃), urea, *tert*-butanol (TBA), sodium bicarbonate solution (NaHCO₃), 30% hydrogen peroxide (H₂O₂), and thioacetamide (CH₃CSNH₂) were provided by Shanghai Sinopharm Chemical Reagents Co., Ltd (China). Copper(II) nitrate trihydrate (Cu(NO₃)₂·3H₂O) was purchased from Shanghai Macklin Biochemical Co., Ltd (China). All reagents were of analytical grade and used as received without further purification. Methylene blue (C₁₆H₁₈ClN₃S·3H₂O) and methyl orange (C₁₄H₁₄N₃·NaO₃S) dyes were purchased from Shanghai Sinopharm Chemical Reagents Co., Ltd.

2.2 Preparation of cellulose microsphere (CM) and CuS-supported CM (CuS@CM) biocatalysts

The preparation process of cellulose microspheres is as follows: 2.5 g cellulose was put into a LiOH/urea/water solution with a mass ratio of 4.5 : 15 : 80.5, vigorously stirred at room temperature, and uniformly dispersed. This mixture was refrigerated at –80 °C until completely frozen, thawed at room temperature and then stirred constantly. Cellulose was completely dissolved after two freeze–thaw cycles. The cellulose solution was injected into 5% sodium bicarbonate solution (NaHCO₃) with constant current pump at a certain speed and stirred at 200 rpm for regeneration. It was filtered and rinsed with a large amount of deionized water until neutral pH. Then, the beads were put into *tert*-butanol (TBA) for solvent exchange at 120 rpm for 24 h, and after filtration and lyophilization,

cellulose microspheres with a high specific surface area were obtained.

A CuS-supported cellulose microsphere biocatalyst was prepared by *in situ* growth in the liquid phase.⁴⁴ Briefly, 300 mg of CM was immersed in 12.5–100 mM Cu(NO₃)₂·3H₂O solution. After stirring at 200 rpm for 1 h to reach adsorption equilibrium, the pH was further adjusted to 3 using dilute nitric acid solution. Then 25 mL constant concentration of CH₃CSNH₂ solution was added drop by drop, and the color of the solution gradually changed from blue to yellow. After 1 h heating in an oil bath at 90 °C, the solution turned dark green, indicating the formation of different CuS-supported CM (CuS@CM-12.5, CuS@CM-25, CuS@CM-50, and CuS@CM-100). After filtration, cleaning with deionized water, and lyophilization with *tert*-butanol solvent replacement, the CuS@CM catalyst was obtained. The pure copper sulfide sample was prepared by the reaction of Cu(NO₃)₂·3H₂O with CH₃CSNH₂.

2.3 Characterizations

The morphology and structure of CM and CuS@CM were studied using a field emission scanning electron microscope (FESEM, Hitachi S-4800, Japan). The BET test was performed using a nitrogen adsorption desorption analyzer (TriStar II 3020M). Transmission electron microscopy (TEM, JEM-2100, Japan) was performed to observe CuS nanoparticles. The crystallographic structures of copper sulfide and supported copper sulfide were analyzed by X-ray (DX-2700BH, HAOYUAN). The content of CuS loaded on cellulose microspheres was measured using a plasma spectrometer (Prodigy-ICP, America). Fourier transform infrared spectroscopy (FTIR, Spectrum Two, PerkinElmer, America) and microscopic confocal Raman spectroscopy (Raman, DXR2xi, Thermo Fisher Scientific, America) were performed to analyze the product ingredients. The CuS@CM microsphere was characterized by X-ray photoelectron spectroscopy (XPS, ESCALAB 250Xi). Thermogravimetric analysis was performed in a nitrogen atmosphere at a heating rate of 10 °C min^{–1} to 900 °C (TG 209 F3 Tarsus, NETZSCH, Germany). The catalytic activity was tested with ultraviolet-visible spectra (UV-1800, Shimadzu, Japan).

2.4 Evaluation of dye degradation by CuS@CM biocatalysts

MB and MO were commonly used as probe molecules in heterogeneous catalytic reactions and as indicators for various reactions, respectively. Therefore, CuS@CM was used to degrade these two dyes to evaluate the catalytic performance of biocatalysts. The specific catalytic process is as follows: 100 mg of different loads of CuS@CM were added to the dye solution (100 mL, 20 mg L^{–1}). The catalytic reaction was then started by the addition of H₂O₂ (500 μL, 30 wt%) and oscillating at 120 rpm, 50 °C. Then, 1 mL of residual solution was taken out at a standard time to measure the dye concentration immediately using a UV-visible light spectrophotometer. After the reaction, the solution was filtered to recover the catalyst, and then rinsed with DI water, and the cyclic usability of CuS@CM was measured under the same conditions. The pseudo-first-order kinetics model was used to calculate the initial reaction rate.



The degradation efficiency formula and the pseudo-first-order equation of the dye are expressed as follows:

$$\text{Degradation efficiency (\%)} = (C_0 - C_t)/C_0 - \ln(C_t/C_0) = kt \quad (1)$$

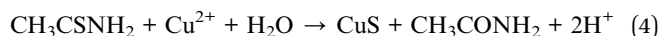
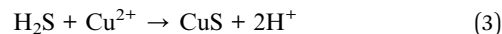
where C_0 (g L^{-1}) represents the initial dye concentration, C_t (g L^{-1}) the dye concentration after a certain time of catalysis, k (min^{-1}) the rate constant of the pseudo-first-order reaction, and t (min) the reaction time.

In addition, the effects of different reaction conditions on the degradation efficiency were studied. Catalytic experiments were carried out at different pH values (3–11), H_2O_2 amounts (0–1 mL), temperatures (30–60 °C) and dye concentrations (10–40 mg L^{-1}). Other experimental conditions were the same as above.

3. Result and discussions

3.1 Formation mechanism of CuS particles on porous cellulose microspheres

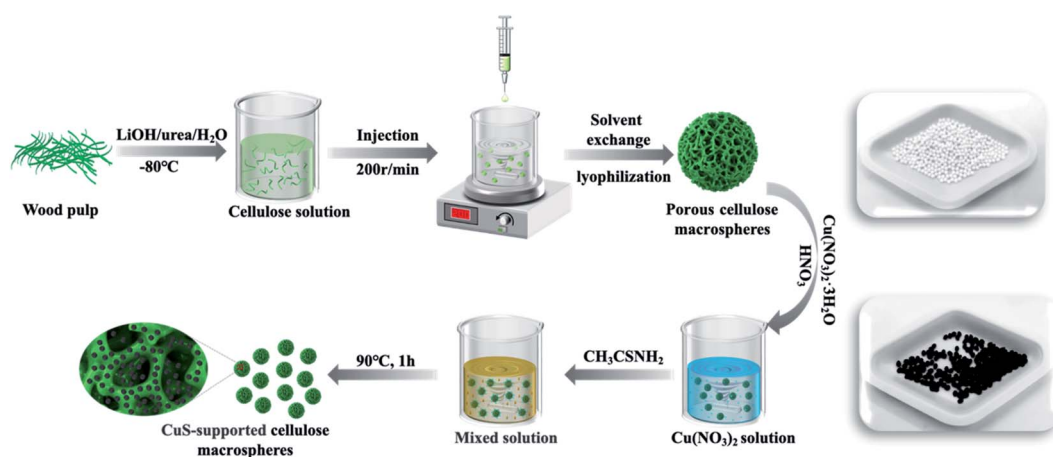
The synthesis schematic diagram of cellulose microspheres and CuS-supported CM catalysts is shown in Scheme 1. In this work, CuS was grown *in situ* on porous CM with a large specific surface area of $325.3 \text{ m}^2 \text{ g}^{-1}$ in an aqueous solution at 90 °C. Compared with the method of preparing copper sulfide in a high-temperature autoclave,⁴⁵ our method is simpler and faster. It is reported that the specific mechanism of CuS nanostructure formation is still being studied, which is quite complex.^{32,46} In our experiment, we inferred from the following reaction equation (eqn (2)–(4)). The CuNO_3 aqueous solution was acidic due to the hydrolysis of Cu^{2+} . In the reaction, thioacetamide could be complexed with Cu^{2+} and hydrolyzed in an acidic solution to produce H_2S , acting as a precipitator to deposit CuS particles on CM. Therefore, a dilute HNO_3 solution was used to adjust the reaction system to the optimal pH value, so that CuS could grow rapidly on CM. The entire formation process of CuS particles can be expressed by the following equation:



As the porous structure of cellulose microsphere is conducive to the effective adsorption of Cu^{2+} , CuS particles could be formed *in situ* on the surface and channels of the three-dimensional cellulose microsphere.

3.2 Morphological and chemical composition analysis of CuS-supported CM biocatalysts

Fig. 1 shows the morphology of CM and a series of CuS@CM catalysts. As illustrated in Fig. 1A and B, CM presented a regular spherical shape with a diameter of about 1.2 mm and a porous network structure on the surface, which was beneficial to the loading of catalyst. Fig. 1C–F show the morphology of CuS-supported CM at different initial Cu^{2+} concentrations, which demonstrate that CuS nanoparticles successfully aggregate and grow on the surface and channels of CM. The TEM images (Fig. 1G and H) reveal that CuS nanoparticles have similar sizes and structures to CuS NPs grown on CM, with a diameter of about 20 nm, indicating that this *in situ* growth method does not damage the original structure and properties of CuS. Meanwhile, the elemental mapping images (Fig. 2A–D) showed that sulfur (S) and copper (Cu) were uniformly distributed in CuS@CM-50 biocatalysts, and EDS analysis revealed the chemical composition of copper and sulfur. The atomic percentage of Cu and S is 32.77 and 30.36 with an atomic ratio of $\text{Cu} : \text{S} = 1 : 1$. Fig. 2F further shows the increase in the amount of copper sulfide by increasing the concentration of copper salt from 12.5 mM to 100 mM. However, with the increase in copper source, the amount of CuS gradually reached equilibrium or even slightly decreased, and the maximum load was 135.2 mg g^{-1} , which may be due to the excessive aggregation of copper sulfide nanoparticles leading to some separation. The above-mentioned results indicate that CuS nanoparticles are uniformly and densely anchored on the cellulose microspheres with three-dimensional porous structures, which not



Scheme 1 Schematic diagram of the preparation process of 3D porous CM and CuS-supported CM composite biocatalysts.



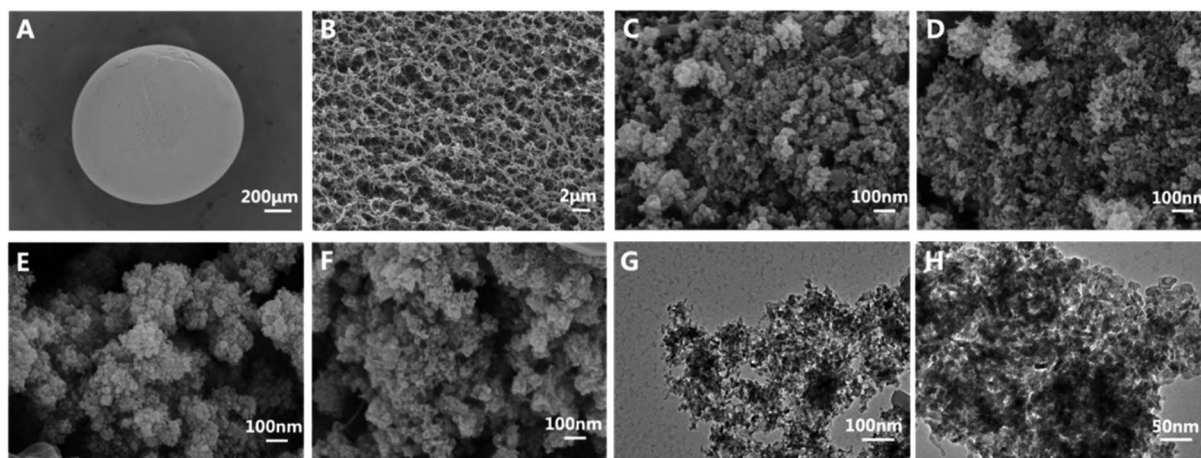


Fig. 1 Apparent SEM images of (A and B) cellulose macrospheres, (C) CuS@CM-12.5, (D) CuS@CM-25, (E) CuS@CM-50 and (F) CuS@CM-100. (G and H) TEM images of CuS nanoparticles.

only make the dye more easily adsorbed to contact CuS, but also provide more contact sites for catalysis.

To confirm the preparation of crystalline copper sulfide on the macrosphere surface of cellulose, XRD analysis was performed, as shown in Fig. 3A. It can be seen from the XRD pattern that all diffraction peaks of pure CuS particles and CuS@CM catalyst can be assigned to cobalt-phase copper sulfide, corresponding to the standard data (PDF card no. 00-001-1281). The crystal peaks at 29.04° , 31.96° , 33.22° , 47.86° , and 59.34° were attributed to the planar structures of (102),

(103), (006), (110), and (116) respectively.⁴⁴ The sharp peak at 20.7° and the low peak at 12.1° belonged to the characteristic diffraction peaks of cellulose II, corresponding to the crystal planes of (110) and (200), respectively.⁴⁷ This indicated that CuS has good crystallinity and stability on CM.

TGA measurement was used to study the thermal stability of CuS, CM, and CuS@CM in a N_2 atmosphere. As shown in Fig. 3B, the TGA curve of the CuS sample presented several obvious steps of mass loss. At a temperature of $200^\circ C$, the

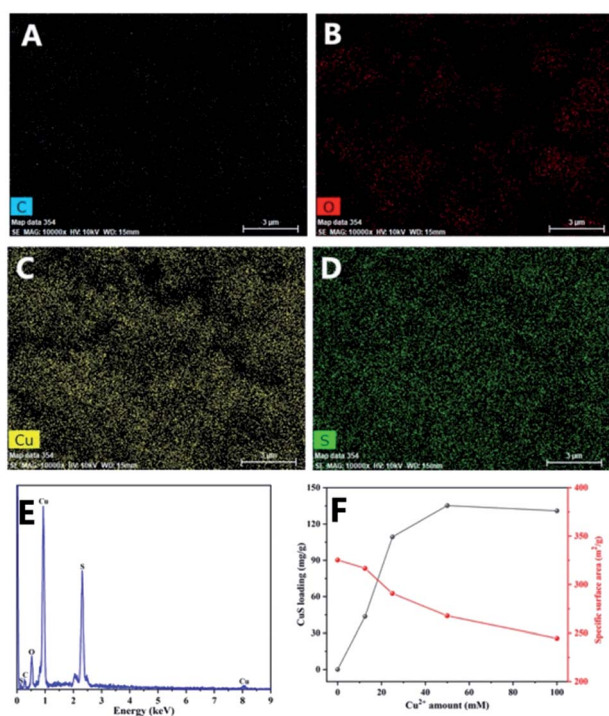


Fig. 2 (A–D) Corresponding elemental mapping image and (E) EDS spectrum of CuS@CM-50. (F) Amount of CuS in CuS@CM composites as a function of $CuNO_3$ concentration and changes of specific surface area of different samples.

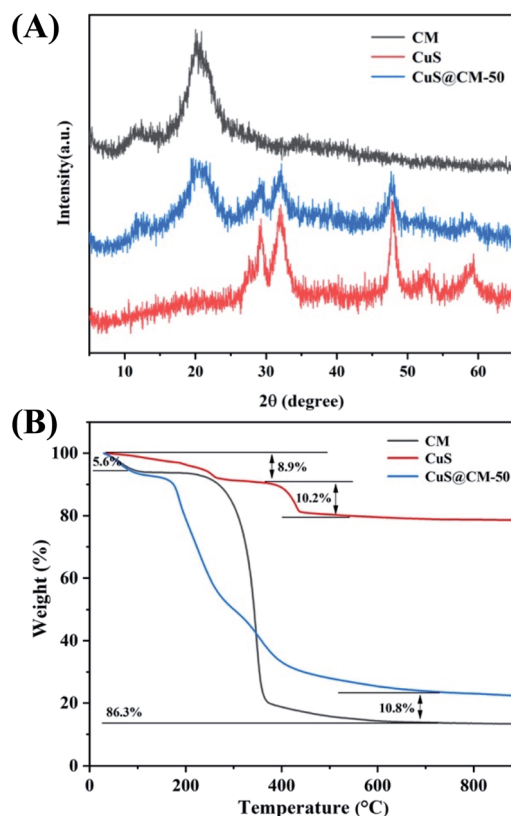


Fig. 3 (A) XRD patterns and (B) TGA curve of CM, CuS, and CuS@CM-50.

thermal weight loss (8.9%) was mainly attributed to the loss of residual solvent and sulfur source in the sample.⁴⁸ This indicated that the sample was extremely unstable at this temperature, and CuS decomposed into various crystalline Cu_xS . The weight loss (10.2%) from 330 °C to 450 °C was classified as CuS decomposition into Cu_2S and phase transition of sulfur elements.⁴⁹ For CM and CuS@CM samples, there was an initial peak at 50–60 °C, which was the weightlessness (5.6%) peak of water adsorption.⁵⁰ The main decomposition temperature of CM was approximately 350–360 °C, which was mainly the decomposition and oxidative degradation of cellulose (80.7%).⁵¹ Obviously, CuS@CM integrated the two above-mentioned decomposition steps and decomposed gradually in the range of 200–400 °C, which was sufficient to prove that the CuS-supported CM biocatalyst was successfully prepared.

FTIR and Raman spectra were used to further determine the binding of CuS on cellulose macrospheres, and the results are shown in Fig. 4. As shown in Fig. 4A, the FTIR spectrum of CM showed characteristic peaks at 890 and 1430 cm^{-1} , and the absorption peak was wider at 3440 cm^{-1} , which was considered to be the typical energy band peak of regenerated cellulose.^{47,52,53} In the CM sample deposited by CuS, the Cu–S tensile vibration band appears at the peak of 620 cm^{-1} , and the peak at 3440 cm^{-1} moved to the high wave digit, indicating the attraction between CuS and –OH groups on cellulose.^{54,55} As shown in

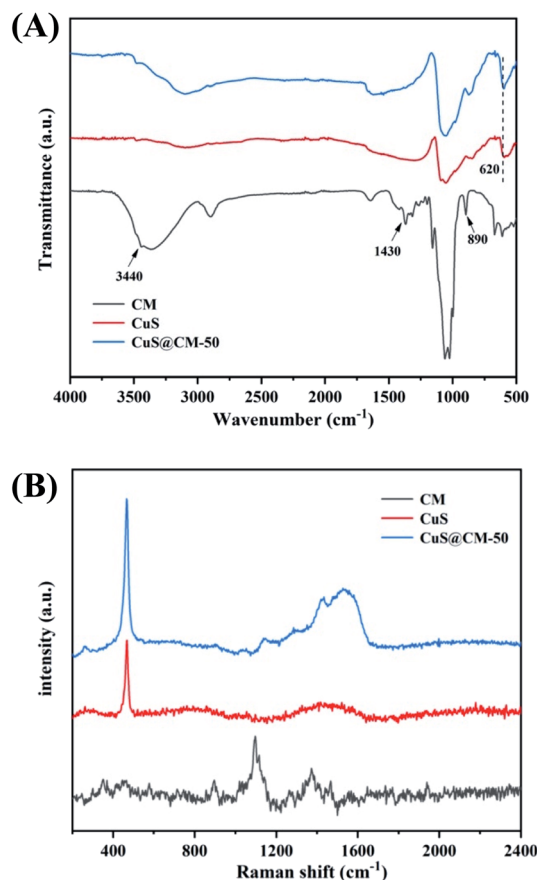


Fig. 4 (A) FTIR and (B) Raman spectra of the as-prepared CM, CuS and CuS@CM-50.

Fig. 4B, Raman spectra also showed characteristic peaks of cellulose and CuS, with 345–390 cm^{-1} , 978–1178 cm^{-1} , and 1280–1411 cm^{-1} respectively belonging to the β -D-glucose ring, glycosidic bond (C–O–C), and subunit methyl (–CH₂) absorption bands.^{56–58} The sharp peak at 471 cm^{-1} belonged to the CuS Raman characteristic absorption peak.⁵⁹ The above-mentioned results indicated that CuS was stably deposited on CM after complexing with CH_3CSNH_2 in the liquid phase.

To further investigate the chemical structure of CuS@CM, X-ray photoelectron spectroscopy was performed to analyze the sample surface. Fig. 5A presents the wide-scan XPS spectra of the prepared CuS@CM microspheres. It can be seen that C, O, Cu, and S elements are significantly detected, indicating their

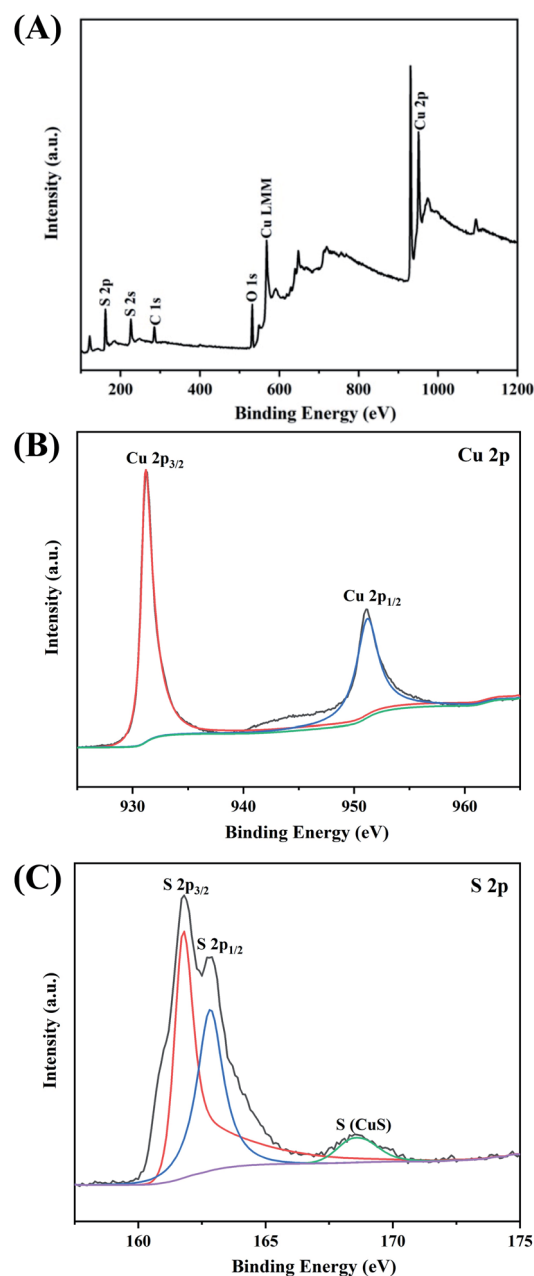


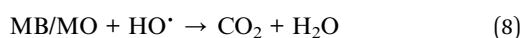
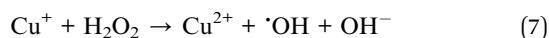
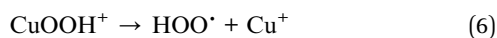
Fig. 5 XPS spectra of CuS@CM microsphere: (A) full spectrum, (B) Cu 2p spectrum, and (C) S 2p spectrum.



presence in the sample. The Auger line of Cu (Cu LMM) appears at 569.0 eV, which is the typical binding energy value for CuS, indicating that the Cu element is present in the form of bivalent states.³² The peaks of Cu 2p (Fig. 5B) and S 2p (Fig. 5C) were further identified by high-resolution XPS spectra. In Fig. 5B, the two main peaks at 932.4 and 952.3 eV were attributed to Cu 2p_{3/2} and Cu 2p_{1/2}, respectively.⁵⁴ Fig. 5C shows that the binding energies of S 2p_{3/2} and S 2p_{1/2} were 161.7 and 163.3, respectively, 168.1 eV corresponding to the S²⁻ state in CuS.⁶⁰ All measurements confirmed that CuS and not Cu₂S or other substances were bound to cellulose macrospheres.

3.3 Catalytic activity of the prepared CuS@CM composite macrospheres

A series of prepared CuS@CM (12.5, 25, 50, and 100 mM) composite macrospheres were used to degrade MB and MO in H₂O₂. During the reaction process, CuS initiated hydrogen peroxide to release hydroxyl radicals of high reaction, which could convert organic dyes into non-toxic small molecules.^{61,62} The mechanism of catalysis can be expressed by the following equations (eqn (5)–(8)):



The evolution of UV-vis spectra with respect to time is shown in Fig. 6A and B. The strong absorption peaks located at 660 nm

and 609 nm belonged to the maximum absorption of the monomer and dimer of MB in aqueous solutions, and the absorption peak at 463 nm belonged to the characteristic absorption peak of MO. It is obvious that the original peaks of MB and MO decreased gradually within 90 min, and no new peaks can be observed after degradation reaction. The results indicated that CuS@CM composite spheres can successfully degrade MB and MO in H₂O₂.

Fig. 6C and D show the influence of different catalytic systems on the catalytic effects of MB and MO. All of the catalytic reactions were performed under the set experimental conditions. When CuS@CM alone was added, the removal rates of MB and MO were 3.57% (Fig. 6C, curve a) and 3.67% (Fig. 6D, curve a) within 1 h, which was due to the effect of porous macrospheres on dye adsorption, while only H₂O₂ was added, the degradation rates of MB and MO were 5.65% (Fig. 6C, curve b) and 3.67% (Fig. 6D, curve b), indicating that the higher temperature promoted the generation of free radicals by H₂O₂ leading to the oxidative degradation of dyes. However, in the presence of CuS@CM macrospheres and H₂O₂, the degradation of MB and MO in the prepared four samples within 1 h reached more than 95% (Fig. 6C, curve c–f; Fig. 6D, curve c–f), which fully indicates that the as-prepared catalysts has excellent catalytic efficiency. In addition, the initial reaction rate constants of the four catalysts were calculated using the pseudo-first-order equation, and the obtained fitting curve conformed to this equation (Fig. 6E and F). It can be seen that the CuS@CM-50 composite material has the fastest reactive rate *k* (MB: 0.28729 min⁻¹; MO: 0.101 min⁻¹), and the results were because this sample has the largest loading capacity and a relatively high specific surface area, which made the dye

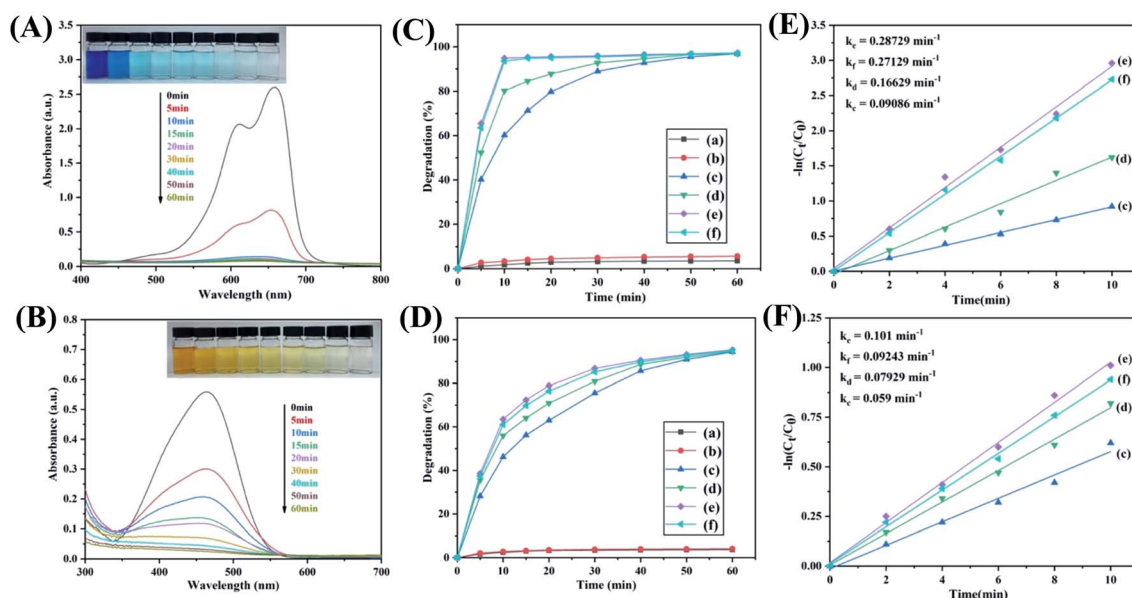


Fig. 6 Changes in the UV-vis absorbance spectra of (A) MB and (B) MO aqueous solutions in the presence of CuS@CM-50 composites and H₂O₂. The degradation rates of (C) MB and (D) MO in different catalytic systems for 60 min, described by the decrease in the intensity of peak absorption of MB and MO solutions: (a) only CM; (b) only H₂O₂; (c) CuS@CM-12.5 + H₂O₂; (d) CuS@CM-25 + H₂O₂; (e) CuS@CM-50 + H₂O₂; and (f) CuS@CM-100 + H₂O₂. (E and F) First-order kinetic plot of $\ln(C_t/C_0)$ vs. time (0–10 min) degradation of MB and MO by different biocatalyst samples.

molecules more receptive to contact with CuS, thus improving the reaction probability.

3.4 Effects of pH, temperature, H₂O₂ and dye concentration on catalytic efficiency

The CuS@CM-50 biocatalyst with the best degradation effect was used for the reaction condition influence determination, with MB as the test dye. As shown in Fig. 7A, degradation rates within 20 min were 75.4%, 82.8%, 95.6%, 96.6% and 96.7%, corresponding to pH values of 3, 5, 7, 9, and 11, respectively. Obviously, the catalytic effect was better than that of acid under alkaline conditions, and the degradation efficiency was maintained highly at pH ≥ 7 , this is because the solution in this pH range has an isoelectric point higher than that of copper sulfide, the cationic MB dye has better affinity for the surface catalyst, and H₂O₂ is dominant under acidic conditions, while HO²⁻ or O₂²⁻ is more under alkaline conditions. The results were consistent with the test results of Wang *et al.*,⁵⁴ indicating that this catalyst has a wider range of pH applicability.

Fig. 7B shows the catalytic efficiency of CuS@CM composites at different temperatures. The results illustrated that the degradation efficiency increased significantly with the increase in reaction temperature. When the reaction temperature was higher than 45 °C, the degradation rate of MB reached more than 95% within 10 min, while when the reaction temperature was 25 °C or 35 °C, only 62.6% and 84.5% of MB were degraded within 20 min. The results indicate that higher temperatures are conducive to rapid dye removal, which is consistent with the report by Zhan *et al.*⁶³

The influence of H₂O₂ concentration on the catalytic efficiency was also investigated, because it is directly related to the number of free radicals generated in the reaction, and has an significant influence on the reaction rate, especially the initial reaction rate. As can be seen from Fig. 7C, degradation efficiency gradually increased with the dosage of H₂O₂ from 65 μ L to 500 μ L. When the dosage was further increased to 1 mL, the catalytic efficiency did not significantly increase, indicating that the reaction rate reached the maximum when 500 μ L of H₂O₂ was added, and the hydroxyl free radical in the dye solution had reached saturation. On the contrary, the dosage was further increased, and the reaction rate would decrease, which may be due to its scavenging effect on the hydroxyl radical caused by low reaction.⁶⁴ Therefore, in order to maintain a high level of reaction rate, a dosage of 500 μ L of H₂O₂ was selected in this experiment.

At the same time, different concentrations of MB were degraded under the above-mentioned optimal conditions. As shown in Fig. 7D, the degradation rates of dyes within 20 min were 96.8%, 95.6%, 92.6% and 90.3%, which were attributed to MB concentrations of 10, 20, 30 and 40 g L⁻¹ respectively. For a series of dye concentrations, the degradation rates were all over 90% in a short time. The results indicated that at a given amount of catalyst, it also had excellent catalytic activity for dyes with high initial concentrations.

3.5 Reusability of CuS@CM biocatalysts

Catalyst reusability and stability are critical in industrial applications because it is an important way to save costs. In

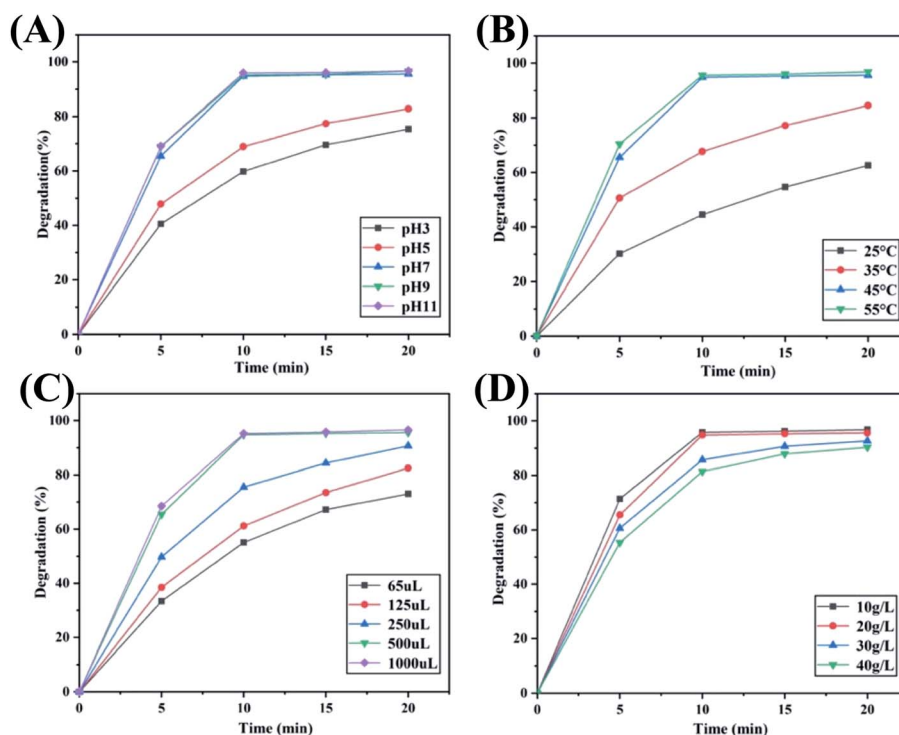


Fig. 7 Effect of (A) initial pH, (B) temperature and (C) H₂O₂ dosage on the catalytic degradation of MB using CuS@CM-50 biocatalyst and the degradation efficiency of (D) different concentrations of MB solution.



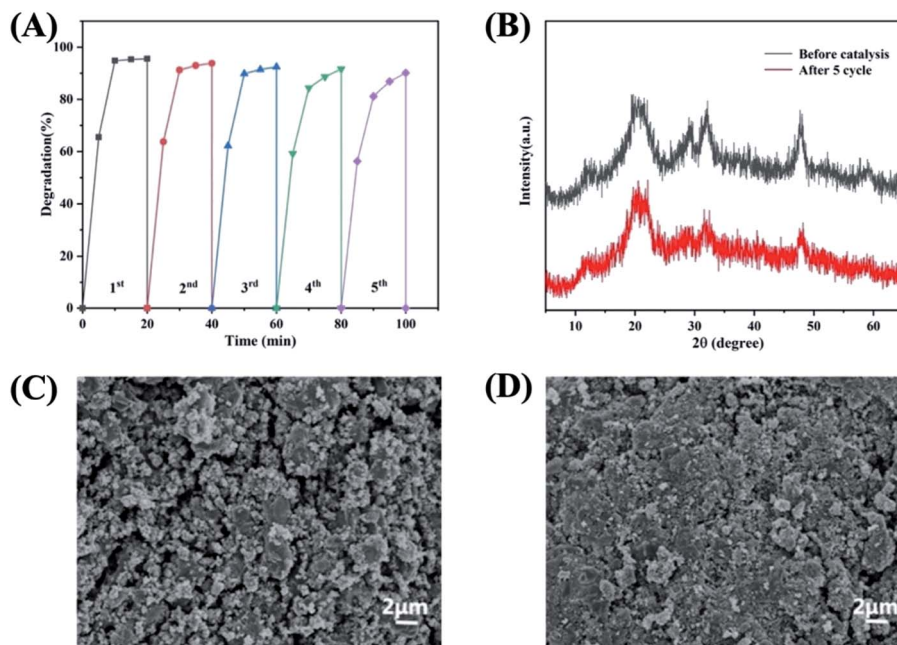


Fig. 8 (A) Catalytic cycle of the prepared CuS@CM-50 biocatalyst for the MB degradation under optimal conditions for 20 min. (B) XRD patterns of the CuS@CM-50 composite after consecutive recycling for five times. SEM images (C) before and (D) after five cycles of degradation.

Table 1 Degradation of different dyes by photocatalysts and peroxide recently published compared to this work

Materials	Synthesis method	Dye conc. (mg L ⁻¹)	Means	Degradation (%)	Time (min)	Ref.
CuS/cellulose aerogel	Ion exchange	20 of MB	3 mL H ₂ O ₂	94.1	60	60
CuS/PAN	Hydrothermal approach	10 of MB	1 mL H ₂ O ₂	93.6	45	65
CuS-CdS	Hydrothermal approach	10 of MB	2 mL H ₂ O ₂ and visible light	94.7	10	66
CuS/rGO	Solvothermal	15 of MB	UV-vis. light	80	140	67
CNT/CuS	Co-precipitation method	5 of RhB	Sun light	120	92.4	68
GO-FA-CuS	Hydrothermal approach	10 of MB	NIR	94.3	20	69
CuS/cellulose microspheres	Liquid phase deposition	20 of MB	0.5 mL H ₂ O ₂	94.8	10	This work

order to confirm the reusability of the synthetic CuS@CM, five cycles were carried out in the process of MB degradation under the same catalytic conditions. As illustrated in Fig. 8A, the initial degradation rate of MB was 95.6%. Although the catalytic activity decreased to a certain extent during use, the degradation degree of CuS@CM was still 90.1% after five cycles, indicating that this material is a potential candidate for recoverable Fenton-like reagents.

Furthermore, the morphology and structural stability of the recovered CuS@CM were characterized by SEM and XRD respectively. The SEM image after cycling (Fig. 8D) showed that the porosity of cellulose was significantly lower than that of cellulose before catalytic degradation (Fig. 8C), which may be due to the partial shedding or collapse of cellulose during the degradation process, resulting in a decrease in the catalytic activity, and the XRD spectra (Fig. 8B) also proved that no new diffraction peaks were generated after

five cycles compared with the fresh catalyst, both of which showed CuS characteristic peaks at 29.04°, 31.96° and 47.86°, indicating that CuS supported on cellulose microspheres remained stable after catalysis. However, the intensity of characteristic peaks decreases after cycling, which may be due to the lower crystallinity of CuS caused by a higher catalytic temperature. It is worth noting that CuS itself did not suffer any significant damage and showed excellent cyclic catalytic stability. In addition, the results of this study were compared with those of previous studies (Table 1), and the activity of peroxide oxidation and photocatalytic oxidation of organic dyes by copper sulfide is summarized. Briefly, the CuS-supported CM composite prepared in this work had higher catalytic activity in the presence of less H₂O₂ and was a kind of green and promising material with excellent catalytic activity, stability and reusability.



4. Conclusion

In summary, a 3D porous cellulose macrosphere with high specific surface area has been successfully prepared and CuS was stably immobilized on the porous cellulose macrosphere with high capacity. The prepared CuS@CM exhibited excellent catalytic activity in the presence of a small amount of H₂O₂ and a wide range of pH applicability (5–11) for the degradation of MB and MO dyes. After several cycles of catalysis, more than 90% of the initial activity still remained and the loaded CuS nanoparticles did not suffer any serious damage, showing good recycling stability. The results indicated that CuS@CM is promising to solve the recycling problem of Fenton reagents and has great application potential in the removal of dyes from wastewater. In this study, only laboratory-simulated wastewater was treated. Future research should use more complex industrial wastewater to further develop practical water treatment technologies.

Conflicts of interest

Authors declare that they have no conflict of interest.

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References

- 1 M. Visa, *Appl. Surf. Sci.*, 2012, **263**, 753–762.
- 2 A. Rodríguez, G. Ovejero, J. L. Sotelo, M. Mestanza and J. García, *Ind. Eng. Chem. Res.*, 2010, **49**, 498–505.
- 3 K. Pakshirajan and S. Singh, *Ind. Eng. Chem. Res.*, 2010, **49**, 7484–7487.
- 4 U.-J. Kim, D. Kim, J. You, J. W. Choi, S. Kimura and M. Wada, *Cellulose*, 2018, **25**, 2615–2628.
- 5 H. E. Emam and H. B. Ahmed, *Int. J. Biol. Macromol.*, 2019, **138**, 450–461.
- 6 G. Ciardelli and N. Ranieri, *Water Res.*, 2001, **35**, 567–572.
- 7 M. F. Elahmadi, N. Bensalah and A. Gadri, *J. Hazard. Mater.*, 2009, **168**, 1163–1169.
- 8 A. Baban, A. Yediler, N. Ciliz and A. Kettrup, *Chemosphere*, 2004, **57**, 731–738.
- 9 A. Khan, S. J. Shah, K. Mehmood, Awais, N. Ali and H. Khan, *J. Mater. Sci.: Mater. Electron.*, 2018, **30**, 406–414.
- 10 M. S. Mia, P. Yao, X. Zhu, X. Lei, T. Xing and G. Chen, *RSC Adv.*, 2021, **11**, 8290–8305.
- 11 J. Lunagariya, A. Dhar and R. L. Vekariya, *RSC Adv.*, 2017, **7**, 5412–5420.
- 12 A. Dhar, N. S. Kumar, M. Asif and R. L. Vekariya, *New J. Chem.*, 2018, **42**, 6990–6996.
- 13 A. Dhar, N. Siva Kumar, M. Khimani, A. S. Al-Fatesh, A. A. Ibrahim, A. H. Fakeeha, H. Patel and R. L. Vekariya, *RSC Adv.*, 2020, **10**, 15282–15292.
- 14 G. Liu, Q. Deng, H. Wang, D. H. L. Ng, M. Kong, W. Cai and G. Wang, *J. Mater. Chem.*, 2012, **22**, 9704–9713.
- 15 J. H. Ramirez, C. A. Costa, L. M. Madeira, G. Mata, M. A. Vicente, M. L. Rojas-Cervantes, A. J. López-Peinado and R. M. Martín-Aranda, *Appl. Catal., B*, 2007, **71**, 44–56.
- 16 H. R. Pouretedal, A. Norozi, M. H. Keshavarz and A. Semnani, *J. Hazard. Mater.*, 2009, **162**, 674–681.
- 17 A. D. French, *Cellulose*, 2017, **24**, 4605–4609.
- 18 H. Shaghaleh, X. Xu and S. Wang, *RSC Adv.*, 2018, **8**, 825–842.
- 19 N.-M. Park, S. Choi, J. E. Oh and D. Y. Hwang, *Carbohydr. Polym.*, 2019, **223**, 115114.
- 20 A. Farooq, M. K. Patoary, M. Zhang, H. Mussana, M. Li, M. A. Naeem, M. Mushtaq, A. Farooq and L. Liu, *Int. J. Biol. Macromol.*, 2020, **154**, 1050–1073.
- 21 J. L. Slavin, P. M. Brauer and J. A. Marlett, *J. Nutr.*, 1981, **111**, 287–297.
- 22 H. Dong, Q. Ding, Y. Jiang, X. Li and W. Han, *Carbohydr. Polym.*, 2021, **265**, 118101.
- 23 D. Georgouvelas, H. N. Abdelhamid, J. Li, U. Edlund and A. P. Mathew, *Carbohydr. Polym.*, 2021, **264**, 118044.
- 24 N. Li, W. Lu, J. Yu, Y. Xiao, S. Liu, L. Gan and J. Huang, *Mater. Sci. Eng., C*, 2018, **91**, 179–189.
- 25 A. Batool and S. Valiyaveetil, *J. Hazard. Mater.*, 2021, **413**, 125301.
- 26 P. Zhou, L. Zhang, X. Sui, Y. Zhong, B. Wang, Z. Chen, X. Feng, H. Xu and Z. Mao, *Cellulose*, 2020, **27**, 3517–3530.
- 27 R. M. Barajas-Ledesma, L. Hossain, V. N. L. Wong, A. F. Patti and G. Garnier, *J. Colloid Interface Sci.*, 2021, **599**, 140–148.
- 28 X. He, T. Chen, T. Jiang, C. Wang, Y. Luan, P. Liu and Z. Liu, *Carbohydr. Polym.*, 2021, **260**, 117790.
- 29 D. Li, Y. Wang, F. Long, L. Gan and J. Huang, *ACS Appl. Mater. Interfaces*, 2020, **12**, 1549–1557.
- 30 Y. Zhao, Y. Zhang, A. Liu, Z. Wei and S. Liu, *ACS Appl. Mater. Interfaces*, 2017, **9**, 4006–4014.
- 31 W. Lyu, J. Li, L. Zheng, H. Liu, J. Chen, W. Zhang and Y. Liao, *Chem. Eng. J.*, 2021, **414**, 128931.
- 32 L. Qian, J. Mao, X. Tian, H. Yuan and D. Xiao, *Sens. Actuators, B*, 2013, **176**, 952–959.
- 33 G. Nie, Z. Li, X. Lu, J. Lei, C. Zhang and C. Wang, *Appl. Surf. Sci.*, 2013, **284**, 595–600.
- 34 L. Zheng, F. Teng, X. Ye, H. Zheng and X. Fang, *Adv. Energy Mater.*, 2020, **10**, 1902355.
- 35 S. Goel, F. Chen and W. Cai, *Small*, 2014, **10**, 631–645.
- 36 W. He, H. Jia, X. Li, Y. Lei, J. Li, H. Zhao, L. Mi, L. Zhang and Z. Zheng, *Nanoscale*, 2012, **4**, 3501–3506.
- 37 T. Zhu, B. Xia, L. Zhou and X. Wen Lou, *J. Mater. Chem.*, 2012, **22**, 7851–7855.
- 38 A. Manzi, T. Simon, C. Sonnleitner, M. Döblinger, R. Wyrwich, O. Stern, J. K. Stolarczyk and J. Feldmann, *J. Am. Chem. Soc.*, 2015, **137**, 14007–14010.
- 39 J. Kundu, S. Khilari and D. Pradhan, *ACS Appl. Mater. Interfaces*, 2017, **9**, 9669–9680.
- 40 X.-j. Yang, X.-m. Xu, J. Xu and Y.-f. Han, *J. Am. Chem. Soc.*, 2013, **135**, 16058–16061.
- 41 C. Zhou, J. Luo, Q. Chen, Y. Jiang, X. Dong and F. Cui, *Chem. Commun.*, 2015, **51**, 10847–10849.
- 42 J. Cai and L. Zhang, *Macromol. Biosci.*, 2005, **5**, 539–548.



- 43 J. Yang, J. Duan, L. Zhang, B. Lindman, H. Edlund and M. Norgren, *Cellulose*, 2016, **23**, 3105–3115.
- 44 J. Huang, J. Zhou, J. Zhuang, H. Gao, D. Huang, L. Wang, W. Wu, Q. Li, D. P. Yang and M. Y. Han, *ACS Appl. Mater. Interfaces*, 2017, **9**, 36606–36614.
- 45 V. K. Gupta, D. Pathania, S. Agarwal and P. Singh, *J. Hazard. Mater.*, 2012, **243**, 179–186.
- 46 Q. Y. Lu, F. Gao and D. Y. Zhao, *Nano Lett.*, 2002, **2**, 725–728.
- 47 X. Luo and L. Zhang, *Biomacromolecules*, 2010, **11**, 2896–2903.
- 48 J. Madarász, M. Krunk, L. Niinistö and G. Pokol, *J. Therm. Anal. Calorim.*, 2015, **120**, 189–199.
- 49 K. R. Nemade and S. A. Waghuley, *Mater. Sci. Semicond. Process.*, 2015, **39**, 781–785.
- 50 A. M. A. Nada and M. L. Hassan, *Polym. Degrad. Stab.*, 2000, **67**, 111–115.
- 51 L. Zhou, J. He, J. Zhang, Z. He, Y. Hu, C. Zhang and H. He, *J. Phys. Chem. C*, 2011, **115**, 16873–16878.
- 52 M. L. Nelson and R. T. O'Connor, *J. Appl. Polym. Sci.*, 1964, **8**, 1311–1324.
- 53 P. Langan, Y. Nishiyama and H. Chanzy, *J. Am. Chem. Soc.*, 1999, **121**, 9940–9946.
- 54 J. Qian, K. Wang, Q. Guan, H. Li, H. Xu, Q. Liu, W. Liu and B. Qiu, *Appl. Surf. Sci.*, 2014, **288**, 633–640.
- 55 J. Guo, H. Tian and J. He, *Chem. Eng. J.*, 2021, **408**, 127336.
- 56 R. H. Atalla and B. E. Dimick, *Carbohydr. Res.*, 1975, **39**, C1–C3.
- 57 K. Schenzel, H. Almlöf and U. Germgård, *Cellulose*, 2009, **16**, 407–415.
- 58 A. Jahn, M. W. Schroder, M. Futing, K. Schenzel and W. Diepenbrock, *Spectrochim. Acta, Part A*, 2002, **58**, 2271–2279.
- 59 Y. Zou, L. Jiang, T. Zhai, T. You, X. Jing, R. Liu, F. Li, W. Zhou and S. Jin, *J. Alloys Compd.*, 2021, **865**, 158919.
- 60 R. M. Y. Saeed, Z. Bano, J. Sun, F. Wang, N. Ullah and Q. Wang, *J. Appl. Polym. Sci.*, 2019, **136**, 47404.
- 61 A. Y. Satoh, J. E. Trosko and S. J. Masten, *Environ. Sci. Technol.*, 2007, **41**, 2881–2887.
- 62 Z. K. Yang, L. X. Song, Y. Teng and J. Xia, *J. Mater. Chem. A*, 2014, **2**, 20004–20009.
- 63 Y. Zhan, H. Li and Y. Chen, *J. Hazard. Mater.*, 2010, **180**, 481–485.
- 64 N. N. Fathima, R. Aravindhan, J. R. Rao and B. U. Nair, *Chemosphere*, 2008, **70**, 1146–1151.
- 65 Y. Lu, Y. Wang, S. Cui, W. Chen and L. Mi, *RSC Adv.*, 2018, **8**, 40589–40594.
- 66 M. Mahanthappa, N. Kottam and S. Yellappa, *Appl. Surf. Sci.*, 2019, **475**, 828–838.
- 67 X.-S. Hu, Y. Shen, Y.-T. Zhang and J.-J. Nie, *J. Phys. Chem. Solids*, 2017, **103**, 201–208.
- 68 Y. Wang, F. Jiang, J. Chen, X. Sun, T. Xian and H. Yang, *Nanomaterials*, 2020, **10**, 178.
- 69 G. M. Neelgund and A. Oki, *Appl. Surf. Sci.*, 2021, **566**, 150648.

