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Influence of Ag nanoparticles anchored on protonated g-C₃N₄–Bi₂MoO₆ nanocomposites for effective antibiotic and organic pollutant degradation

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The development of noble metal-anchored semiconductors for photocatalytic processes is now garnering interest for potential application to toxic pollutants as well as antibiotic degradation. Herein, we report novel Ag@p-g-C₃N₄–Bi₂MoO₆ nanocomposites synthesized by facile hydrothermal and calcination methods with a size of about 50 nm, exhibiting superior photocatalytic activity for charge separation. The resulting nanocomposites were evaluated by various physiochemical techniques such as X-ray diffraction, X-ray photoelectron spectroscopy, Fourier-transform infrared spectroscopy, scanning electron microscopy, and high-resolution transmission electron microscopy. The charge transfer photogenerated carriers were confirmed by photoluminescence spectra and electrochemical impedance spectroscopy. The anchoring of Ag nanoparticles over p-g-C₃N₄/Bi₂MoO₆ decreased the band gap energy from 2.67 to 2.48 eV, to exhibit an abnormal increase in absorption of light towards the visible light region. The degradation performance of the nanocomposites in terms of antibiotic ciprofloxacin and rhodamine B degradation efficiency was measured 85 and 99.7% respectively. The superoxide radical anion 'O₂^{•−} played a significant role throughout the entire degradation process. Focusing on the probable mechanism based on the desirable results, the present work follows the heterostructure mechanism. Moreover, this work features the feasible applications of Ag@p-g-C₃N₄–Bi₂MoO₆ as a modified photocatalyst in the treatment of both domestic and industrial waste water.

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1. Introduction

Currently, there is a great demand for natural resources worldwide mainly due to environmental degradation and pollution.¹ To resolve this issue, the provision of alternative renewable energy sources is required. Solar energy harvesting has received extensive attention in the fields of photocatalysis and energy storage applications.^{2,3} Due to the rapid development of industrial and human activities most of the carcinogenic dyes from tannery industries get passed directly into water, thereby resulting in environmental pollution.^{4–6} Among them, ciprofloxacin (CIP), C₁₇H₁₈N₃O₃, belongs to the second-generation fluoroquinolone antibiotic family mainly composed of quinolone structure and a piperazine moiety.⁷

CIP is extensively used in various fields, such as medicine, livestock and agriculture. Due to the metabolic activities, CIP has a negative impact on human health and the environment. To overcome these issues, antibiotics are being replaced by environmentally friendly and renewable technology. Photocatalysis offers an alternative solution for the complete degradation of antibiotics by boosting reaction conditions.^{8–14}

Most of the traditional methods used for wastewater treatment have recently suffered due to them receiving less attention in environmental remediation, particularly anaerobic oxidation layer partition and adsorption.^{15,16} To overcome this issue, the use of a noble metal-based semiconductor photocatalyst has been found to be the most efficient and viable technique to degrade macromolecular contaminants into less harmful and non-toxic compounds. In general, an enormous number of semiconductor metal oxide and sulfide photocatalysts, including TiO₂, ZnO, WO₃, CdS, In₂S₃, and Ag₃PO₄, have been bench-marked for photocatalytic activity through the efficient degradation of organic pollutants.^{17,18}

It is well known that polymeric graphitic carbon nitride (g-C₃N₄) with a layered 2D structure has a broad range of

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applications in materials chemistry due to its flexible physical and chemical properties and tunable electronic structure, showing outstanding potential for visible-light active photocatalysis.^{19,20} Besides this, g-C₃N₄ contains abundant functional groups such as carbon (C) and nitrogen (N), which augment the material properties to boost its broad specific area housing active absorption sites for organic pollutants.^{21,22} Meanwhile, there have been few attempts on investigating noble metals anchored on heterojunction semiconductors to reduce the band gap energy. Indeed, g-C₃N₄ has certain limitations in realistic photocatalysis applications, such as inadequate real surface area, less separation of photoexcited electron-hole pairs, limited electron transfer, poor quantum yield, and low light-harvesting ability. Promoting the efficiency of g-C₃N₄ by doping with transition metals (*e.g.*, Cu, Co, and Ag) and non-metal components (*e.g.* P, S, I, O, and B) not only improves the density of both donor and acceptor, but also facilitates the transport of electrons and increases the visible-light absorption at an appropriate wavelength.^{23–26} In recent reports the aurivillius metal oxide group is generally represented by the formula Bi₂X_{n-1}Y_nO_{3n+3} (X = Ca, Sr, Ba, Pb, Na and Y = Ti, Nb, Ta, Mo, W) and is considered to be the most popular and active photocatalyst owing to its layered structures and outstanding photoelectrical properties. Bismuth molybdate (Bi₂MoO₆) consists of interlaced [Bi₂O₂]²⁺ and perovskite group (MoO₄²⁻) layers.²⁷ The foremost virtues of pure Bi₂MoO₆ are those such as lower band gap energy nearly equal to 2.4–2.8 eV, nontoxicity and low cost, and it showed an improved light response compared with wide-band-gap window photocatalysts such as TiO₂ and ZnO. Moreover, pristine Bi₂MoO₆ suffers from a wide range of problems in large-scale practical applications, like poor separation of charge transfer, low quantum efficiency and scarce energetic sites.^{28–30} To resolve these disadvantages, numerous endeavors have been adopted involving modifying doping techniques.³¹

Considering the above key points, we are interested in synthesizing composites of Ag nanoparticles anchored on p-g-C₃N₄/Bi₂MoO₆ surface *via* a facile and straightforward hydrothermal method. As is evident, Ag nanoparticles affixed on the surface layer of p-g-C₃N₄/Bi₂MoO₆ can further improve the photoinduced charge carriers mainly due to surface plasmon resonance.³² Furthermore in comparison with bare g-C₃N₄/Bi₂MoO₆, decorating Ag nanoparticles on p-g-C₃N₄/Bi₂MoO₆ could enhance the photocatalytic degradation of various contaminants such as CIP and rhodamine B (RhB) under visible light irradiation, measured at around 85% and 99.7%, respectively, within 60 min.

2. Experimental

2.1 Materials

All reagents used in the present work were of analytical grade and used without further purification. Ultrapure water was used in all the experiments. Melamine (99%), bismuth(III) nitrate pentahydrate, ammonium molybdate tetrahydrate, CIP, and RhB were purchased from Sigma-Aldrich, India. HCl (37 wt%)

and ethanol used as solvents for synthesis were obtained from Avra Chemicals, India.

2.2 Synthesis of g-C₃N₄

For pristine g-C₃N₄, a modified procedure was adopted from a previous publication.³³ Firstly, melamine was ground into fine powder with a mortar and pestle. The fine powder was transferred into an alumina boat crucible and kept for calcination at 550 °C at a ramp rate of 5 °C min⁻¹ for 4 h. After the reaction was cooled to ambient temperature, the final yellow product was crushed and named as g-C₃N₄.

2.3 Synthesis of protonated g-C₃N₄ or p-g-C₃N₄

Typically, 1 g of bulk g-C₃N₄ was dissolved in 25 mL of 3 M HCl and stirred for 12 h at room temperature. After the completion of the reaction, the product suspension was washed with deionized water until it was neutral. The mixture was placed in a hot-air oven for drying at 80 °C for 12 h. Finally the dried powder was obtained and labelled as p-g-C₃N₄.³³

2.4 Synthesis of Bi₂MoO₆

Bi₂MoO₆ was synthesized by a facile hydrothermal method with slight modification.³⁴ About 0.485 g of Bi(NO₃)₃·5H₂O and 0.088 g of Na₂MoO₄·2H₂O were dissolved in 40 mL MilliQ-pore water under vigorous stirring for 5 h, followed by sonication for 15 min. The mixture was then transferred into a Teflon-lined stainless steel autoclave with a filling volume of 50 mL and heated up to 150 °C for 15 h. After being dried at room temperature the obtained product was washed with deionized water and absolute ethanol about three times and dried in a hot-air oven for 6 h at 80 °C. It was further annealed in a muffle furnace at 300 °C for 3 h (temperature ramp rate: 5 °C min⁻¹).

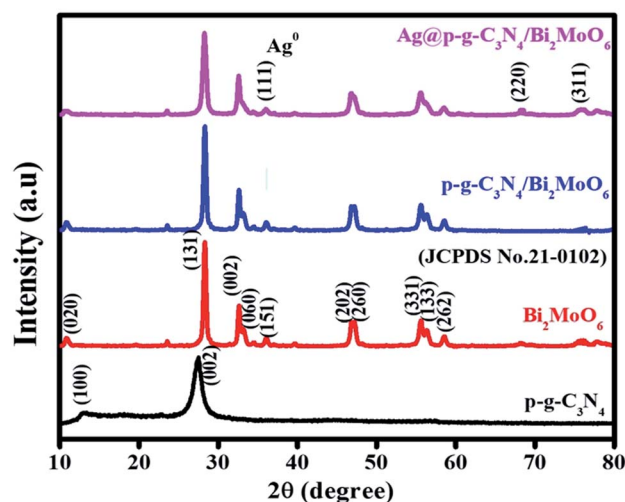


Fig. 1 XRD patterns of the prepared p-g-C₃N₄, Bi₂MoO₆, p-g-C₃N₄/Bi₂MoO₆, and Ag@p-g-C₃N₄/Bi₂MoO₆ photocatalysts.



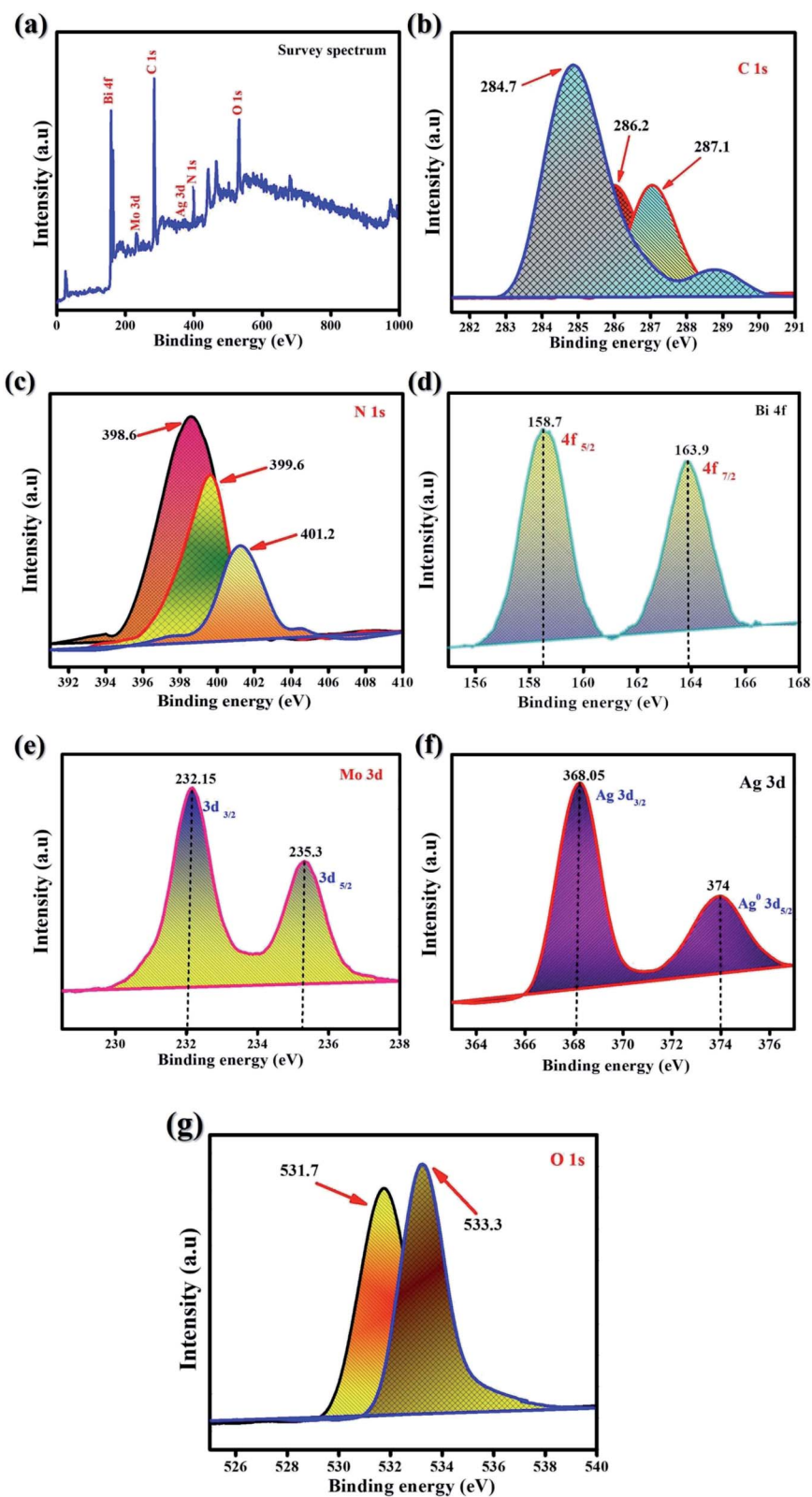


Fig. 2 XPS spectra of the as-synthesized Ag@p-g-C₃N₄/Bi₂MoO₆. (a) Survey scan, (b) C 1s, (c) N 1s, (d) Bi 4f, (e) Mo 3d, (f) Ag 3d, and (g) O 1s.

2.5 Synthesis of p-g-C₃N₄/Bi₂MoO₆ and Ag@p-g-C₃N₄/Bi₂MoO₆

To fabricate the Ag-loaded nanocomposites, 100 mg of p-g-C₃N₄ was dissolved in 40 mL of deionized water and stirred for 30 min followed by the dropwise addition of 1 : 2 stoichiometric ratio of 0.485 g of Bi(NO₃)₃·5H₂O and 0.088 g of Na₂MoO₄·2H₂O. The solution was kept under vigorous magnetic stirring for 30 min. After that, for loading Ag into the bulk solution, prior as-weighed AgNO₃ (0.0424 g) was added to above the mixture. Then, the solution was ultrasonically treated for 15 min under vigorous magnetic stirring for 1 h. It was then transferred into a Teflon-lined stainless steel autoclave of 50 mL capacity and heated up to 150 °C for 15 h. The obtained final product was centrifuged and washed with distilled water and absolute ethanol to remove the impurities and dried at 80 °C for 6 h. After drying, the samples were further annealed in a muffle furnace at 300 °C for 3 h (temperature ramp rate: 5 °C min⁻¹). The obtained powder sample was denoted as Ag@p-g-C₃N₄/Bi₂MoO₆. For comparison, the same synthetic procedure was followed for p-g-C₃N₄/Bi₂MoO₆ but without addition of AgNO₃.

2.6 Materials characterization

The crystallinity of the samples was probed using the Cu K alpha line ($\lambda = 0.154$ nm) of an X-ray diffractometer (XRD, Panalytical Xpert Pro). XRD data were taken from $2\theta = 10$ to 80° with a step size of 0.025° . The obtained diffraction patterns were compared to a reference pattern.

Morphological studies were conducted to identify the structure of Ag@p-g-C₃N₄/Bi₂MoO₆ by scanning electron microscopy (SEM). The microscopic images were obtained using a Bruker FE-SEM operating at 15 000 V with a built-in EDX setup. High-resolution transmission electron microscopic (HRTEM) imaging and EDS elemental mapping were carried out with a JEOL 2010F TEM with an accelerating voltage of 200 kV. To confirm the chemical composition and oxidation state of the materials, X-ray photoelectron spectroscopy (XPS) was performed with a Physical Electronics System India, using Al K α radiation as the excitation source. The obtained resultant binding energy values from XPS were calibrated with carbon peak as background reference. Fourier transform infrared spectroscopy (FTIR) was conducted with a Perkin Elmer instrument (USA) in the range of 4000 to 400 cm⁻¹. The optical band gap and charge separation efficiency of the synthesized catalyst were analyzed by UV-visible DRS with BaSO₄ as the background (Evolution 220 PC spectrophotometer). Photoluminescence (PL) analysis was performed with a Fluorolog (Horiba Yvon) spectrophotometer. A lamp source was used in the photocatalytic performance investigation (Xe lamp, 300 W). The transient photocurrent response of all of the synthesized samples was investigated on an electrochemical workstation using a general three-electrode setup (Shanghai Chenhua CHI-660D).

2.7 Photocatalytic experiments

The photodegradation of RhB dye and CIP antibiotic as model pollutants was investigated using a 300 W Xe lamp, with a UV cut-off filter ($\lambda > 400$ nm). Typically, 50 mg of Ag@p-g-C₃N₄/

Bi₂MoO₆ photocatalyst was dispersed in 50 mL of pollutant solution (concentration of 10 ppm) and placed in a Pyrex reactor. For 30 min, the solution was stirred under dark condition to reach the adsorption-desorption equilibrium state for the solution. At regular intervals of 10 min, an aliquot of 4 mL was extracted at a particular time after visible light irradiation and the concentration of CIP antibiotic and RhB dye was determined by measuring the absorbance corresponding to their respective wavelengths (λ_{\max}) of 278 and 554 nm that was analyzed by a UV-visible spectrophotometer (Specord-200 plus UV-visible spectrophotometer, Germany). The photocatalyst and the remaining pollutant were separated using centrifugation. Furthermore, for recycling experiments, the photocatalyst cycling ability was investigated by analyzing the removal of CIP after the sequential running of four cycles. To identify the mineralization rate in the pollutant, total organic carbon (TOC) was measured using an analyzer (Shimadzu TOC-L instrument, Japan). Additionally, a trapping experiment was used to determine the significant active species involved in the degradation of the antibiotic. Different scavengers (concentration of 5 mM), namely benzoquinone (BQ), ethylenediaminetetraacetic acid disodium (EDTA-2Na), and isopropyl alcohol (IPA), were added into the CIP solution as scavengers of superoxide anion radicals, holes, and hydroxyl radicals, respectively.

3. Results and discussion

3.1 XRD analysis

Powder XRD studies were used to investigate the crystalline structure of the as-prepared samples (p-g-C₃N₄, Bi₂MoO₆, p-g-C₃N₄/Bi₂MoO₆ and Ag@p-g-C₃N₄/Bi₂MoO₆). As seen in Fig. 1 the pristine p-g-C₃N₄ displays two distinct peaks at $2\theta = 12.08^\circ$ and 27.8° , which are due to the interplanar stacking corresponding to the tri-triazine groups [100] and [002] of the aromatic laminate plane, well matched to the standard JCPDS (87-1526)

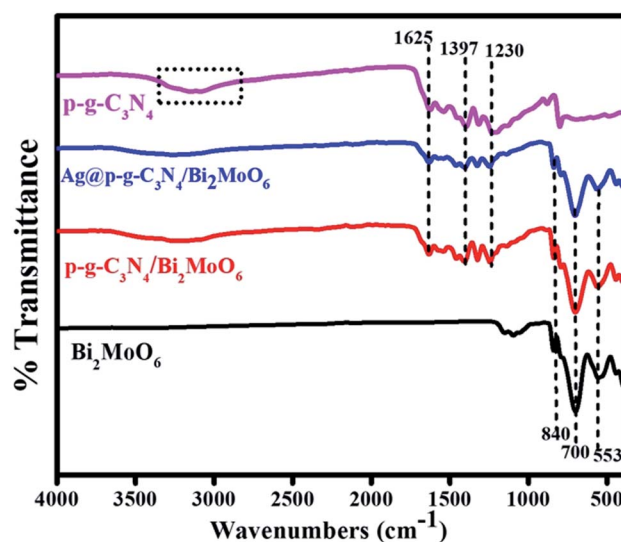


Fig. 3 FTIR spectra of p-g-C₃N₄, Bi₂MoO₆, p-g-C₃N₄/Bi₂MoO₆, and Ag@p-g-C₃N₄/Bi₂MoO₆.



values.^{35,36} The XRD patterns are clearly defined to validate the orthorhombic crystalline structure of γ -BMO (JCPDS no. 21-0102). The diffraction peaks of Bi_2MoO_6 were observed in the pattern of $\text{p-g-C}_3\text{N}_4/\text{Bi}_2\text{MoO}_6$ composites, suggesting that heterojunctions were successfully established. The peaks of $\text{p-g-C}_3\text{N}_4$ of (100) and (002) phases were found to be invisible and overlapped with the (131) peak of Bi_2MoO_6 . The characteristic diffraction peaks of Ag at $2\theta = 32.7, 46.8, 55.7$, and 67.9° suggest the (111), (200), (220), and (311) planes were observed for Ag-doped $\text{p-g-C}_3\text{N}_4/\text{Bi}_2\text{MoO}_6$ composites perfectly matching JCPDS no. 02-1067.³⁷ Simultaneously, the rest of the major diffraction peaks of Bi_2MoO_6 show that Ag nanoparticles are decorated on the $\text{p-g-C}_3\text{N}_4/\text{Bi}_2\text{MoO}_6$ heterojunction.

3.2 XPS analysis

XPS was performed to analyze the valency states and all chemical species of the synthesized samples. Fig. 2 presents the survey spectrum of all of the characteristic elements of carbon, nitrogen, bismuth, molybdenum, oxygen, and silver. Fig. 2a shows the XPS survey spectrum of as-synthesized samples of $\text{Ag@p-g-C}_3\text{N}_4/\text{Bi}_2\text{MoO}_6$. Fig. 2b shows the two distinct XPS peaks of C 1s, which could be attributed to amorphous carbon of graphite C-C bonds appearing at 284.7 and 287.1 eV. For $\text{p-g-C}_3\text{N}_4$ composite, the binding energies at 398.6 eV, 399.9 eV and 401.2 eV are observed for the N 1s spectrum as shown in Fig. 2c and fine characteristic peaks are attributed to C-N-C coordination. Bi 4f is detected *via* the signal as shown in Fig. 2d at consistent binding energies of 158.7 eV ($\text{Bi } 4f_{5/2}$) and 163.9 eV

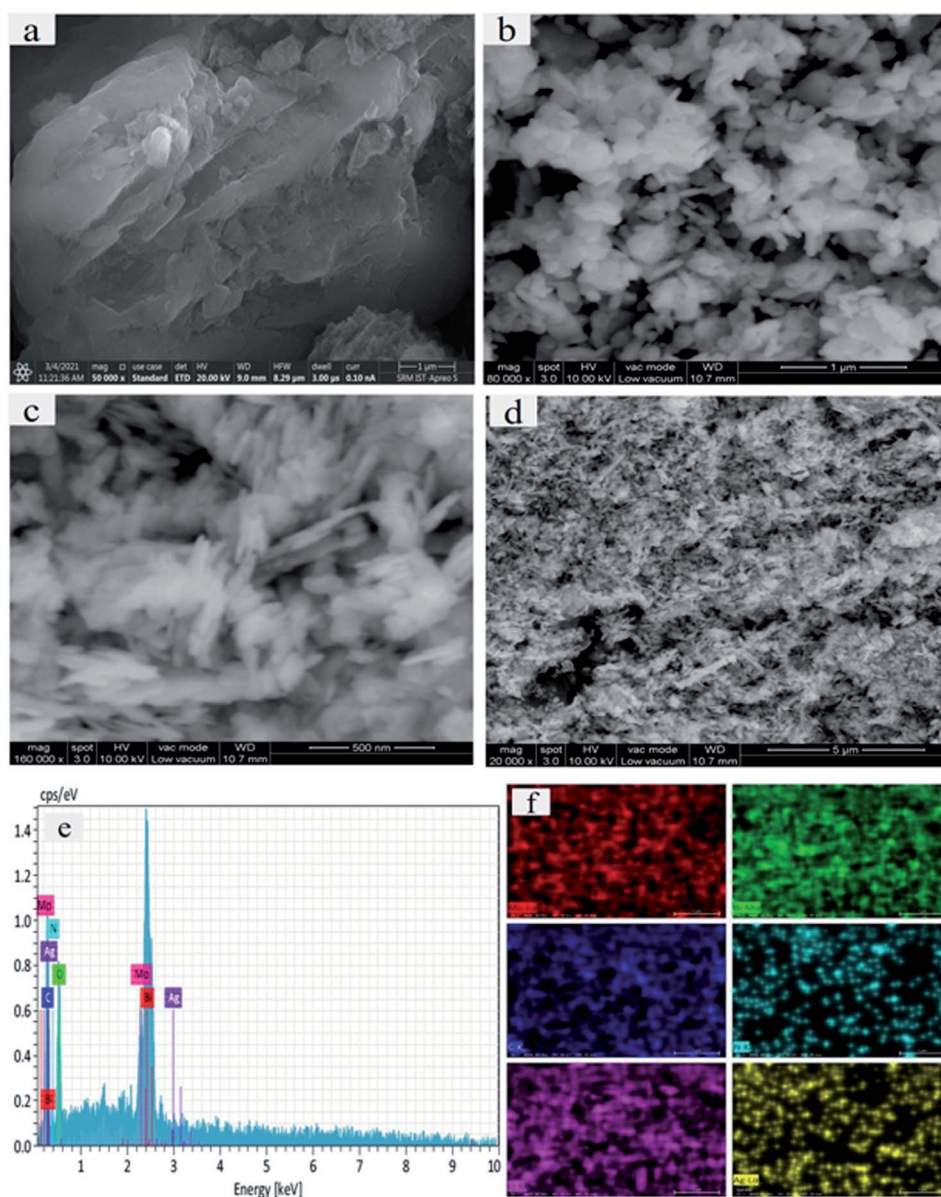


Fig. 4 (a–d) SEM images of as-synthesized samples: (a) $\text{p-g-C}_3\text{N}_4$, (b) Bi_2MoO_6 , (c) $\text{p-g-ssC}_3\text{N}_4/\text{Bi}_2\text{MoO}_6$ and (d) $\text{Ag@p-g-C}_3\text{N}_4/\text{Bi}_2\text{MoO}_6$. (e) Elemental composition of $\text{Ag@p-g-C}_3\text{N}_4/\text{Bi}_2\text{MoO}_6$. (f) Elemental maps of $\text{Ag@p-g-C}_3\text{N}_4/\text{Bi}_2\text{MoO}_6$.

(Bi 4f_{7/2}), attributed to Bi metallic compound. The characteristic spin-orbital splitting of photoelectrons for Mo⁶⁺ oxidation state in the 3d orbital, particularly in d_{5/2} and d_{3/2}, corresponds to the binding energy peaks in the range of 232.1 and 235.3 eV corresponding to individual Mo atoms that co-exist in different chemical states as shown in Fig. 2e. The standard peaks of Ag 3d can be found in the elemental silver metal state as shown in Fig. 2f, the peaks at 367.8 and 374 eV corresponding to the Ag 3d_{3/2} and Ag 3d_{5/2} orbital regions, respectively. In comparison, the influential O 1s peaks can be seen in Fig. 2g at 531.7 and 533.3 eV, which is compatible with the MoO₆ functionalization of Ag@p-g-C₃N₄/Bi₂MoO₆ nanocomposites.^{38,39}

3.3 FTIR spectroscopy analysis

The presence of functional groups for as-prepared samples was further analyzed by FTIR. Fig. 3 shows the two pure sharp peaks at 800 cm⁻¹ confirming the presence of bending mode of S-

triazine in CN heterocycles. The peaks at 1230–1625 cm⁻¹ were assigned to the stretching vibrations of C≡N. Moreover, the large bands at 1230, 1317 and 1397 cm⁻¹ could be assigned to stretching vibration mode of C–N bond in the aromatic moiety. The peak at 1625 cm⁻¹ corresponds to the stretching vibration mode of heptazine group. Additionally, the broad band at 3000–3500 cm⁻¹ was ascribed to stretching mode of uncondensed amine N–H groups in g-C₃N₄. The pure Bi₂MoO₆ composite characteristic peaks at 553 and 700 cm⁻¹ show the asymmetric and stretching vibration mode of the oxygen atoms in MoO₆.⁴⁰ The sharp bands at 700 and 840 cm⁻¹ were attributed to the Mo–O stretching vibration mode in an octahedral structure.

3.4 SEM analysis

The surface morphologies and microstructural elemental composition of the as-prepared samples were further examined

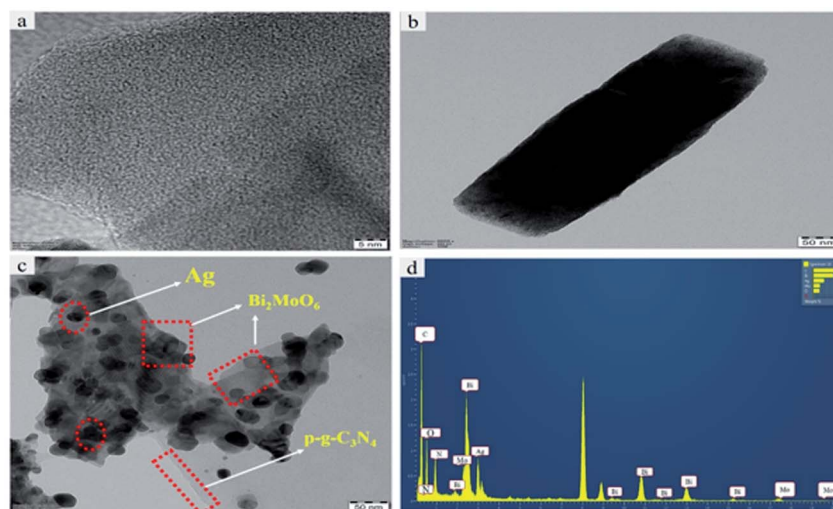


Fig. 5 Typical TEM images of (a) pristine p-g-C₃N₄, (b) Bi₂MoO₆, and (c) Ag@p-g-C₃N₄/Bi₂MoO₆. (d) EDX spectrum of Ag@p-g-C₃N₄/Bi₂MoO₆.

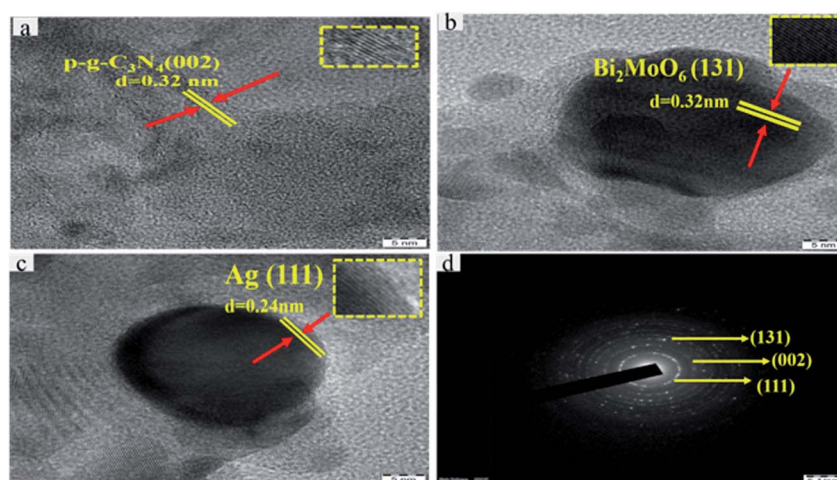


Fig. 6 HRTEM images of (a) p-g-C₃N₄, (b) Bi₂MoO₆, and (c) Ag. (d) Selected area electron diffraction pattern for Ag@p-g-C₃N₄/Bi₂MoO₆.



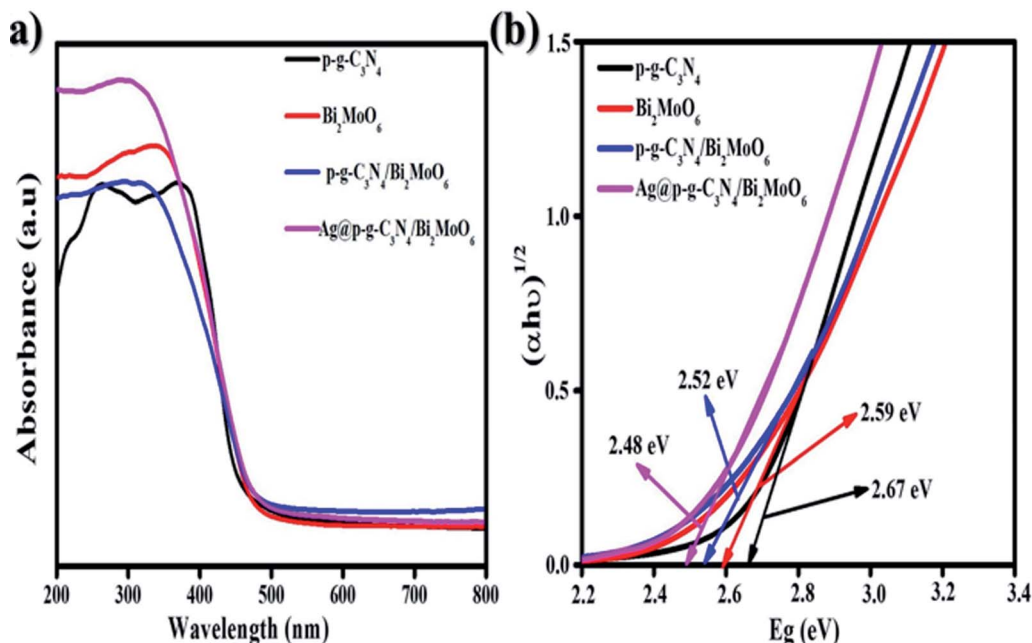


Fig. 7 (a) UV diffuse reflectance spectra and (b) graph of $(\alpha h\nu)^{1/2}$ versus energy (eV) of the as-synthesized Ag@p-g-C₃N₄/Bi₂MoO₆ photocatalyst.

by SEM. Fig. 4a–d displays a large surface area of p-g-C₃N₄ wrinkled 2d morphology doping with p-g-C₃N₄/Bi₂MoO₆. Fig. 4a depicts the ultrathin uniform bulk layer of pristine p-g-C₃N₄. Notably, the pure Bi₂MoO₆ micro-petals with an irregular plate-like microstructure are shown in Fig. 4b. Moreover, the microplate combination of both p-g-C₃N₄ and Bi₂MoO₆ nanocomposites is shown in Fig. 4c. Interestingly, Fig. 4d shows the final composite of Ag nanoparticles anchored on p-g-C₃N₄ and Bi₂MoO₆ thereby confirming that some amount of Ag nanoparticles were loaded into the heterostructure, sufficiently enhancing the photocatalytic activity. Interestingly the final catalyst Ag@p-g-C₃N₄/Bi₂MoO₆ shows a rice husk-like

morphology. Fig. 4e shows the EDS spectrum of the final compositions of the prepared nanocomposites. This also indicates that all of the elements C, N, Mo, Bi and Ag are present in the final composition, and the heterostructure of Ag@p-g-C₃N₄/Bi₂MoO₆ had been effectively fabricated. Fig. 4f shows the elemental mapping of Ag@p-g-C₃N₄/Bi₂MoO₆, with all the elements being present in the final composition.

3.5 TEM analysis

To further investigate the final heterostructured surface morphology, TEM was carried out for synthesized p-g-C₃N₄, Bi₂MoO₆ and Ag@p-g-C₃N₄/Bi₂MoO₆ nanocomposites as

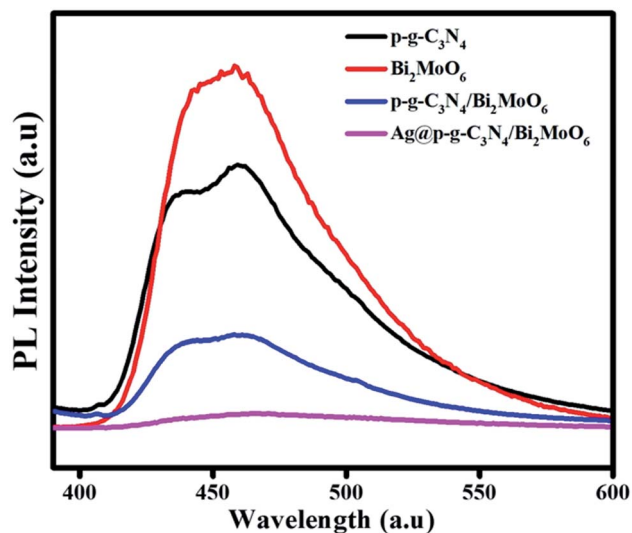


Fig. 8 PL spectra of p-g-C₃N₄, Bi₂MoO₆, p-g-C₃N₄/Bi₂MoO₆, and Ag@p-g-C₃N₄/Bi₂MoO₆.

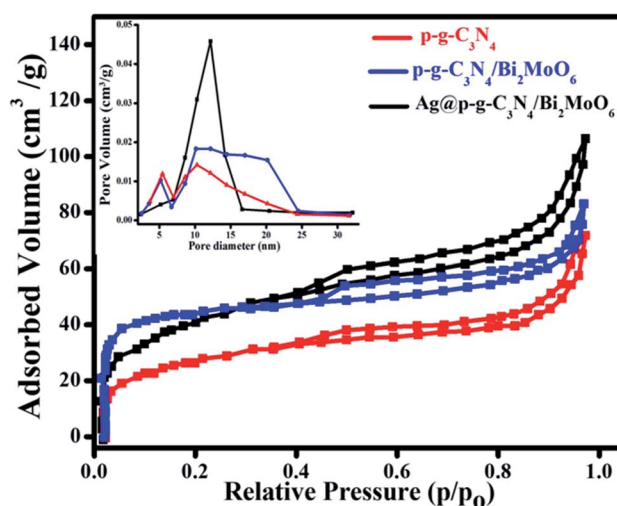


Fig. 9 N₂ adsorption-desorption isotherms and the corresponding (inset) pore-size distribution curves for as-prepared nanocomposites.

depicted in Fig. 5. As illustrated in Fig. 5a the TEM image of p-g-C₃N₄ shows a layered nanosheet, with an average thickness of 6–10 nm. Fig. 5b reveals a lower magnification image of Bi₂MoO₆ nanopetal. It is worth noting that Fig. 5c shows a close-up view of dark spot highlighted in red of Ag wrapped with nanoplates of Bi₂MoO₆ and p-g-C₃N₄ nanosheet, which strongly confirms the formation of Ag@p-g-C₃N₄/Bi₂MoO₆. Fig. 5d confirms the EDX spectrum for Ag@p-g-C₃N₄/Bi₂MoO₆. This clearly shows that all the chemical components of Ag@p-g-C₃N₄/Bi₂MoO₆ derived rice husk composites contain Ag, Bi, C, N, O, and Mo elements. Furthermore, HRTEM images of pure p-g-C₃N₄, Bi₂MoO₆, and Ag@p-g-C₃N₄/Bi₂MoO₆ are displayed in Fig. 6. It is clear from Fig. 6a that the boxed area shows the interplanar spacing lattice fringes of about 0.32 nm corresponding to the (002) facets of p-g-C₃N₄ and is in good agreement with the XRD pattern. Similarly Fig. 6b displays the selected part of Bi₂MoO₆ highlighted in red color. It also shows the interplanar spacing lattice fringes of about 0.32 nm corresponding to the (131) plane of Bi₂MoO₆. Likewise, Fig. 6c shows the interplanar spacing lattice fringes of about 0.240 nm corresponding to the (111) plane of Ag. The selected area electron diffraction pattern in Fig. 6d reveals selected discrete spots indexed to (131), (002) and (111) of the planes corresponding to Ag@p-g-C₃N₄/Bi₂MoO₆.

3.6 UV-visible diffuse reflectance spectroscopy analysis

To examine the absorption ranges and optimum energy band gaps of the synthesized nanocomposites, the optical absorption was examined from 200 to 800 nm, as shown in Fig. 7a. The p-g-C₃N₄ exhibited a moderate absorption edge around 435 nm while the bare Bi₂MoO₆ showed a steep absorption edge that is found to be 470 nm. When p-g-C₃N₄ was intercalated with Bi₂MoO₆, outstanding peaks were observed in the region of 550 nm and the absorption edges became steeper due to p-g-C₃N₄ intercalation properties. After loading Ag nanoparticles onto the surface of p-g-C₃N₄/Bi₂MoO₆, the absorption edge exhibited higher absorption intensities than that for bare p-g-C₃N₄, Bi₂MoO₆, and p-g-C₃N₄/Bi₂MoO₆. From the DRS studies it can be clearly understood that incorporating Ag onto p-g-C₃N₄/

Bi₂MoO₆ extends the light absorption into the visible region thus contributing to the great potential for photocatalytic degradation. The energy gaps were determined using Tauc plots of $\alpha h\nu = A(h\nu - E_g)^{1/2}$ versus photon energy as shown in Fig. 7b, where α = absorption coefficient, h = Planck's constant, ν = light frequency and E_g = band gap energy. With the excellent agreement of the curve the obtained band gap energy (E_g) value was evaluated at 2.48 eV. The final catalyst was found to possess a direct band gap based on a literature survey.⁴¹

3.7 Photoluminescence spectra analysis

The PL emission spectra of pure p-g-C₃N₄, Bi₂MoO₆ and combinations of p-g-C₃N₄/Bi₂MoO₆ and Ag@p-g-C₃N₄/Bi₂MoO₆ catalysts were recorded under an excitation irradiation wavelength of 400 nm as shown in Fig. 8. PL is an important technique to investigate the interfacial charge transfer and isolation of photogenerated electrons and holes in photocatalysts. A lower PL intensity of Ag@p-g-C₃N₄/Bi₂MoO₆ than pure p-g-C₃N₄ and Bi₂MoO₆ can be observed from the PL spectra. This may be due to conduction band levels of bare p-g-C₃N₄ and Ag@p-g-C₃N₄ that stand at equal positions, quickly suppressing the rate of charge carrier recombination. Besides, it is interesting to note when Ag nanoparticles were decorated on the surface of p-g-C₃N₄/Bi₂MoO₆, this reduced the recombination rate of photogenerated electron-hole pairs and exhibited the lowest emission intensity, which greatly enhanced the photocatalytic activity due to the surface plasmon resonance effect of Ag.

3.8 BET analysis

As seen in Fig. 9, the Brunauer–Emmett–Teller (BET) technique was used to quantify the precise surface area and pore size distributions of the prepared nanocomposites under N₂ adsorption–desorption equilibrium conditions.⁴² The surface area of p-g-C₃N₄, p-g-C₃N₄/Bi₂MoO₆, and Ag@p-g-C₃N₄/Bi₂MoO₆ nanocomposites was found to be 43.5, 64.2, and 98.4 cm² g^{−1}, according to the BET analysis results. The broad surface area of the Ag@p-g-C₃N₄/Bi₂MoO₆ photocatalyst shows that the recombination rate of photogenerated charge carriers has been reduced. Furthermore, the increased pore volume and

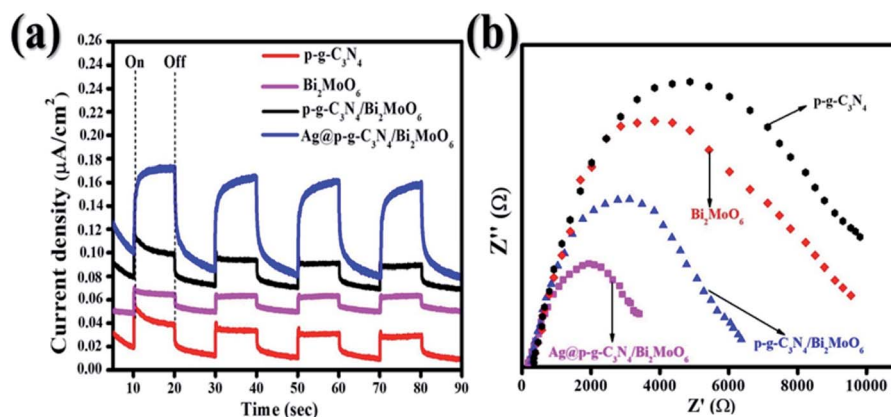


Fig. 10 (a) Transient photocurrent response of all synthesized samples and (b) EIS curves of all synthesized samples.



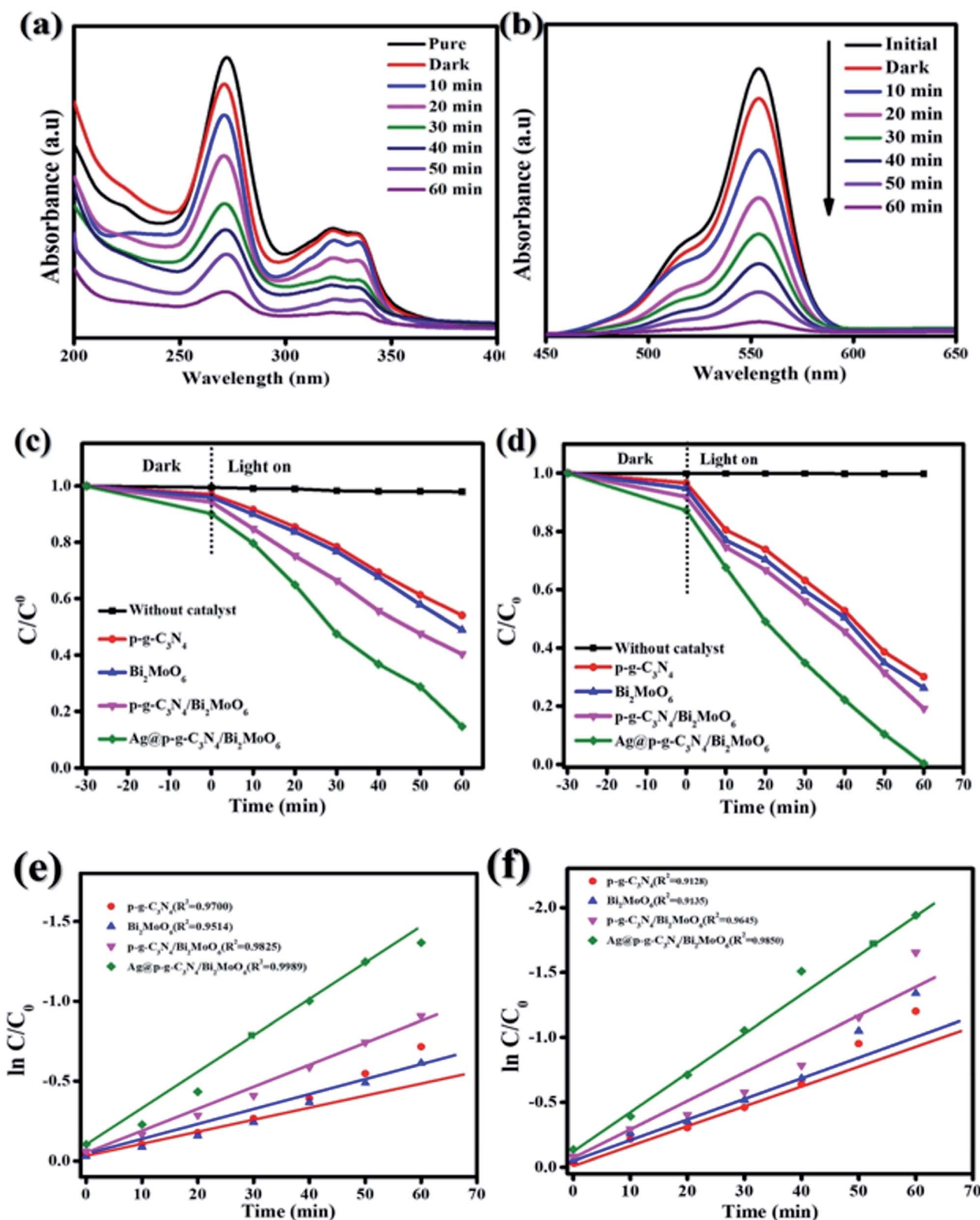


Fig. 11 (a and b) Degradation of CIP and RhB over $\text{Ag}@p\text{-}g\text{-C}_3\text{N}_4/\text{Bi}_2\text{MoO}_6$ nanocomposite. (c and d) Photocatalytic CIP and RhB degradation for all synthesized nanocomposites. (e and f) Corresponding kinetic plots (k_{obs}) of the as-synthesized nanocomposites.

diameter, as well as better isolation and migration of photo-generated charges, contributes to increased photocatalytic activity. As a consequence of the above results, the main $\text{Ag}@p\text{-}g\text{-C}_3\text{N}_4/\text{Bi}_2\text{MoO}_6$ composite is useful for improved adsorption

and also provides a greater number of reactive sites for photocatalytic processes, with a major effect on photocatalytic

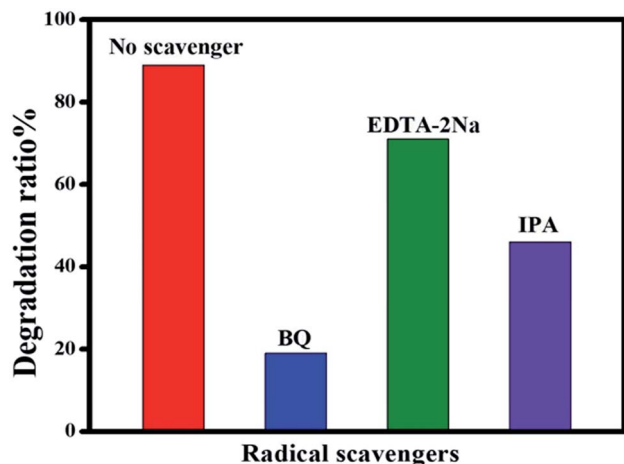


Fig. 12 Effect of scavengers on the photocatalytic degradation of CIP over Ag@p-g-C₃N₄/Bi₂MoO₆ nanocomposite.

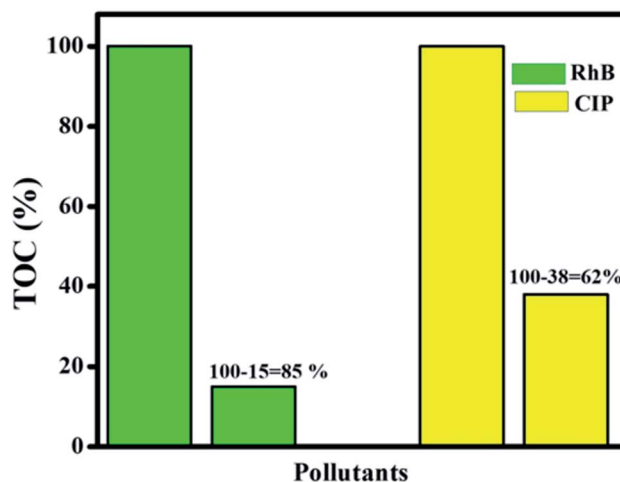


Fig. 14 Mineralization efficiency for RhB and CIP of the final Ag@p-g-C₃N₄/Bi₂MoO₆ catalyst.

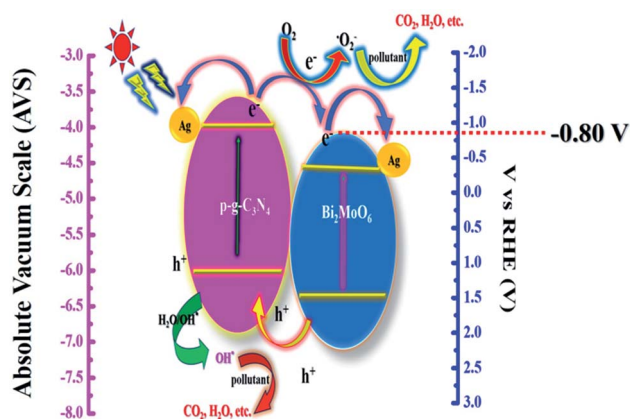


Fig. 13 Schematic representation of the probable photocatalytic mechanism for Ag@p-g-C₃N₄/Bi₂MoO₆.

efficiency development.^{43,44} Importantly, in addition to the obvious increase in surface area, an increase in pore volume (inset of Fig. 9) was found, rising from 0.15 cm³ g⁻¹ for as-prepared p-g-C₃N₄ to 0.46 cm³ g⁻¹ for Ag@p-g-C₃N₄/Bi₂MoO₆, facilitating charge separation in the entire composite structure. Additionally, the addition of effective Ag and Bi₂MoO₆ components increased the pore diameter from 10.6 nm for as-prepared p-g-C₃N₄ to 15.2 nm for Ag@p-g-C₃N₄/Bi₂MoO₆.

3.9 Photoelectrochemical analysis

Photocurrent measurements and electrochemical impedance spectroscopy (EIS) are valuable methods for characterizing the separation and migration potential of photoexcited charges. As shown in Fig. 10a, the transient photocurrent measurements of Ag@p-g-C₃N₄/Bi₂MoO₆ indicate a high photocurrent density compared with pure samples, implying a more effective separation of the photoexcited charges (electron-hole pairs) and limitation of their recombination. Several repeat on-off cycles display a similar photocurrent reaction. Significantly, it shows

that the samples have high photostability. EIS electrochemical measurements of the as-synthesized samples were also used to analyze the charge-transfer resistance between the hetero-structures. Fig. 10b depicts the smaller semicircle radius of prepared nanocomposites than the pure p-g-C₃N₄ and Bi₂MoO₆. This indicates the higher efficacy in separation and charge transfer of photogenerated electron-hole pairs.

3.10 Photocatalytic performance

The photocatalytic activity of Ag@p-g-C₃N₄/Bi₂MoO₆ was investigated in the degradation of CIP and RhB under visible light irradiation. The results are shown in Fig. 11a and b. The results of the degradation of CIP and RhB using Ag@p-g-C₃N₄/Bi₂MoO₆ show excellent photocatalytic performance, with values of 85% and 99.7%, which are much higher than when using pure p-g-C₃N₄, Bi₂MoO₆ and p-g-C₃N₄/Bi₂MoO₆ combination. When pure p-g-C₃N₄ is used under visible light, it shows poor photocatalytic

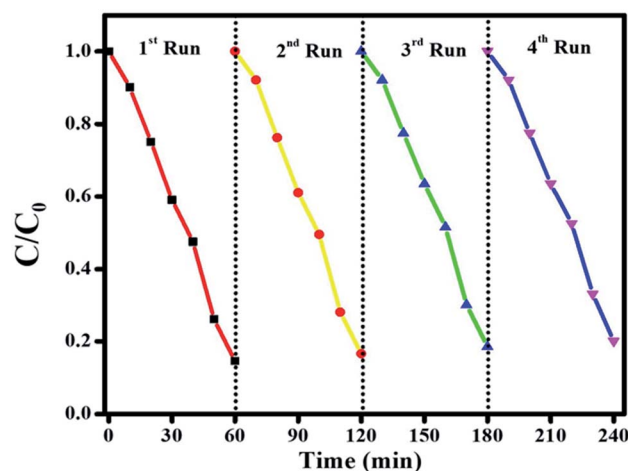


Fig. 15 The stability test for CIP degradation activity for the final Ag@p-g-C₃N₄/Bi₂MoO₆ catalyst.



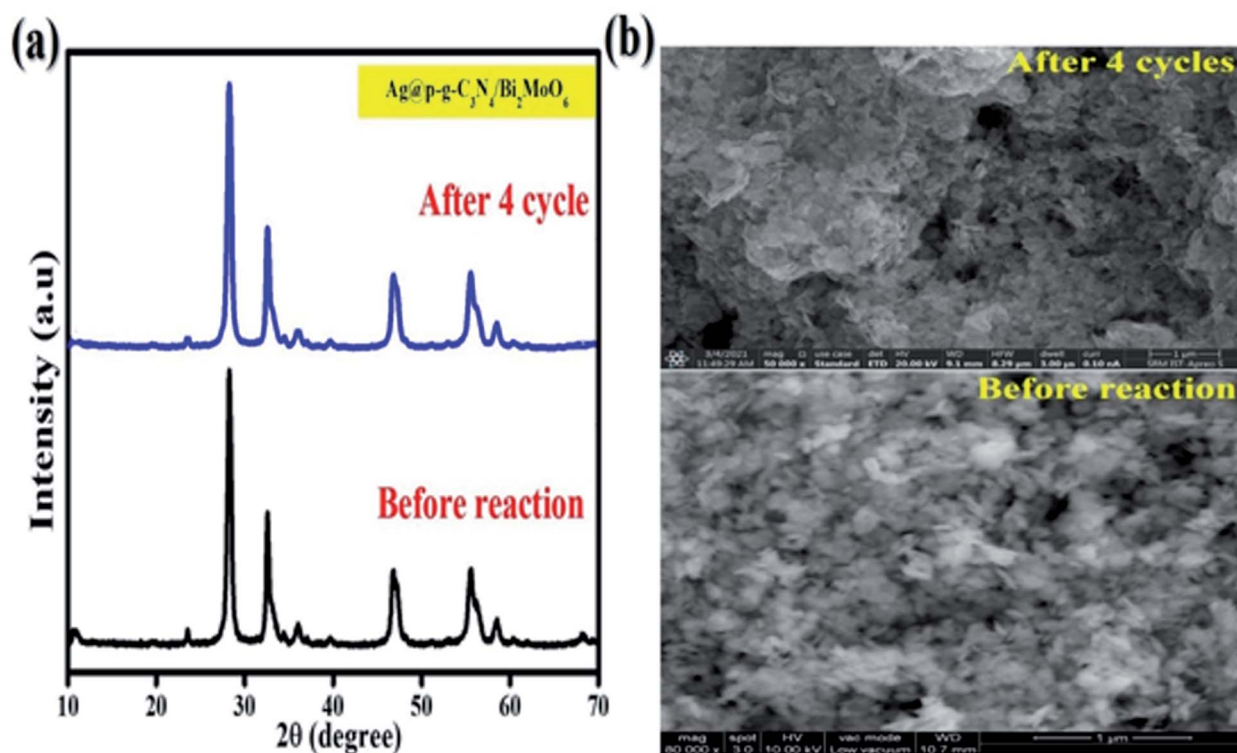


Fig. 16 (a) XRD patterns and (b) SEM images of fresh and used Ag@p-g-C₃N₄/Bi₂MoO₆ catalyst.

activity and this may be due to the inhibitory effect on the surface area of graphene sheets, while g-C₃N₄ loaded on the metallic surface of Bi₂MoO₆ would allow effective separation of the electron-hole pairs. In addition, Ag nanoparticles enwrapped on p-g-C₃N₄/Bi₂MoO₆ would further increase the photocatalytic degradation effect due to the surface plasmon resonance interaction between Ag and p-g-C₃N₄. All synthesized catalyst activities are shown in Fig. 11c and d. To understand the decomposition rate, the kinetic behavior of CIP and RhB degradation is represented as a plot of C/C_0 versus time (min) in accordance with the pseudo first-order reaction kinetics based on the following equation:

$$-\ln(C/C_0) = Kt$$

where C is the concentration of pollutants at time t , C_0 is the initial concentration of CIP and RhB, and K is the reaction rate constant as shown in Fig. 11e and f.

3.11 Radical trapping experiments

To identify the reactive species involved in the photodegradation of CIP, a series of radical trapping experiments were carried out using various radical scavengers. IPA was chosen as a quencher of hydroxyl radicals ($\cdot\text{OH}$), EDTA-2Na was selected as a quencher of holes (h^+) and BQ was chosen as a quencher of the superoxide radical anion ($\cdot\text{O}_2^-$). These quenchers were added to the photocatalyst during the degradation process. As can be seen in Fig. 12, when BQ was added into the reaction mixture the degradation rate achieved was only 19%. This clearly shows that BQ has a strong effect on CIP degradation, which demonstrates that

($\cdot\text{O}_2^-$) plays an essential role in the degradation process. In contrast, the addition of IPA and EDTA-2Na has not much effect on CIP degradation, clearly indicating that (h^+) and ($\cdot\text{OH}$) do not have much impact on the photocatalytic system.

3.12 Plausible degradation mechanism

The suggested photocatalytic mechanism of the synthetic heterojunctions is depicted in Fig. 13. Upon visible light irradiation, both semiconductors p-g-C₃N₄ and Bi₂MoO₆ are excited to produce photogenerated electron-hole pairs. The diagram demonstrates clearly that $\cdot\text{OH}$ plays a less active role in the dye degradation process due to the band gap level of photogenerated h^+ having an energy of 2.48 V in the valence band level showing a lower energy gap than $\cdot\text{OH}$ with an average energy of 2.70 V, suggesting that it is difficult for $\cdot\text{OH}$ to interact from H₂O during photodegradation. It is worth noting that h^+ has a strong oxidation ability which directly decomposes the CIP molecule. More particularly, e^- at the conduction band is much smaller (-0.80 V) than that of $\cdot\text{O}_2^-$ which is 1.23 V, resulting in a simple reduction of dissolved O₂ to $\cdot\text{O}_2^-$ which degrades various organic matter.⁴⁵ More particularly, the decorated Ag nanoparticles used in this work play an important role in serving as a solid-state electron mediator. The e^- developed on the conduction band of p-g-C₃N₄ heterojunctions is easily transferred to Ag particles and Bi₂MoO₆ microplates, effectively promoting electron transfer and suppressing electron-hole recombination efficiency.³³ Furthermore, the formed h^+ could migrate from Bi₂MoO₆ to g-C₃N₄ in the heterojunctions, reducing recombination efficiency significantly. As a result, the photocatalytic dye degradation efficiency of Ag@p-g-



$\text{C}_3\text{N}_4/\text{Bi}_2\text{MoO}_6$ is higher than those of pure p-g- C_3N_4 , Bi_2MoO_6 , and p-g- $\text{C}_3\text{N}_4/\text{Bi}_2\text{MoO}_6$ composites.

3.13 TOC analysis

The TOC technique was employed to assay the mineralization of CIP and RhB using the optimum photocatalyst. Fig. 14 shows that mineralization efficiency values of 62% and 85% were achieved for CIP and RhB, respectively, while the photocatalytic efficiency values were 85% and 99.7% for CIP and RhB under 60 min of irradiation. This difference between mineralization and photocatalytic efficiency is due to the intermediates present during the photocatalytic process that can be mineralized in a longer reaction time.

3.14 Recyclability and stability

To check the stability of the photocatalyst, the recyclability is a vital parameter to determine for the final nanocomposite. Fig. 15 shows the results of stability tests for CIP degradation activity for the final $\text{Ag}@p\text{-g-C}_3\text{N}_4/\text{Bi}_2\text{MoO}_6$ catalyst. After 4 consecutive runs, the degradation efficiency of the catalyst was found to be 92%, which indicates the high stability of the prepared photocatalyst. For the sake of comparison, the stability of the photocatalyst was evaluated by powder XRD and SEM after four successive repeat runs and the results are displayed in Fig. 16a and b. XRD patterns and SEM images of fresh and used samples reveal that the catalyst's crystalline nature and morphological features are not destroyed after several cycles.

4. Conclusion

To conclude, we developed a simple and environmentally friendly approach of using a ternary photocatalyst, $\text{Ag}@p\text{-g-C}_3\text{N}_4/\text{Bi}_2\text{MoO}_6$, for environmental contaminant degradation. The unique characterization techniques revealed that the prepared $\text{Ag}@p\text{-g-C}_3\text{N}_4/\text{Bi}_2\text{MoO}_6$ nanocomposite possessed high purity and crystallinity. It is also to be noted that the interfacial charge separation and efficiency are fully controlled due to the formation of interfacial bonds between metallic Ag and p-g- $\text{C}_3\text{N}_4/\text{Bi}_2\text{MoO}_6$. Moreover, the $\text{Ag}@p\text{-g-C}_3\text{N}_4/\text{Bi}_2\text{MoO}_6$ catalyst shows enough durability for the degradation of CIP. Ag nanoparticles were decorated on p-g- $\text{C}_3\text{N}_4/\text{Bi}_2\text{MoO}_6$ which leads to a higher surface area of the photocatalyst for the photodegradation of CIP and RhB under visible light irradiation for a duration of 1 h. After four cycles, the photocatalyst maintains its effectiveness and exhibits excellent structural stability. The pollutant undergoes a pseudo first-order reaction. The mechanistic pathway suggests that ($\cdot\text{O}_2^-$) plays a significant role in the photodegradation of the various pollutants. Thus, the present study demonstrates that capped $\text{Ag}@p\text{-g-C}_3\text{N}_4/\text{Bi}_2\text{MoO}_6$ photocatalysts have a high potential for use in environmental protection.

Conflicts of interest

There are no conflicts to declare.

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