


 Cite this: *RSC Adv.*, 2021, **11**, 13867

Fabrication of a novel magnetic metal–organic framework functionalized with 2-aminothiophenol for preconcentration of trace silver amounts in water and wastewater†

 Seyyed Hossein Mousavi,^a Mahboobeh Manoochehri  ^{*a} and Faramarz Afshar Taromi^b

Herein, a novel magnetic metal–organic framework functionalized (MMOF) with 2-aminothiophenol (2-ATP) was fabricated and employed for separation/preconcentration of trace silver amounts. At first magnetite nanoparticles (Fe_3O_4 NPs) were synthesized and then coated with SiO_2 . Thereafter, the $\text{Fe}_3\text{O}_4@\text{SiO}_2$ nanoparticles were modified with 2-ATP. Finally, the functionalized MMOF was prepared by the fabrication of MIL-101(Cr) in the presence of $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ NPs. MIL-101(Cr)/ $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ nanocomposite was characterized with FT-IR, SEM, elemental analysis, XRD and VSM and then utilized in the separation/determination of silver ions in various real samples. The effects of diverse experimental variables such as pH, uptake time, adsorbent amount, desorption time, eluent concentration and volume were studied comprehensively employing experimental design methodology. After optimization, LOD and linearity were 0.05 ng mL^{-1} and $0.2\text{--}200 \text{ ng mL}^{-1}$, respectively. Repeatability of the new method was determined based on RSD value for 5, 50, 150 ng mL^{-1} ($n = 5$) concentrations which was 9.3%, 6.8% and 4.5%, respectively. Ultimately, the outlined method was utilized in the separation/determination of silver ions in various water and wastewater samples satisfactorily.

 Received 18th January 2021
 Accepted 2nd April 2021

DOI: 10.1039/d1ra00420d

rsc.li/rsc-advances

1. Introduction

Silver is a commercially important metal which is utilized in the fabrication of high strength/corrosion resistant alloys as well as in jewelry. Moreover, silver alloys and compounds are widely utilized in pharmaceutical/dental preparations and in implanted prostheses owing to their well-known antibacterial nature.^{1,2} Besides, silver is extensively applied in imaging and photographic applications.^{1,2} The wide utilization of silver compounds and silver containing procedures in industry has led to the increase of silver in environmental samples.^{2,3} Silver can enter into the environmental samples *via* industrial discharges owing to the fact that it often occurs as an impurity in Cu, Sb, As and Zn ores. Besides, exposure to low levels of silver compounds is common because of the use of its soluble compounds for disinfection of drinking water.^{2,4,5} Accordingly, the accurate determination of silver ions in wastewater and environmental water samples is important and challenging in

analytical⁶ and developing new separation/extraction method is vital.^{7,8} Flame atomic absorption spectrometry (FAAS),^{2,3,7,9,10} flame furnace atomic absorption spectrometry¹¹ and inductively coupled plasma-atomic emission spectrometry^{12,13} are popular techniques for quantification of silver. These analytical techniques are not suitable for quantification of silver in environmental samples due to the low sensitivity as well as the matrix complexity. Accordingly, performing a separation/preconcentration step before the instrumental analysis is mandatory.¹⁰

Various extraction/preconcentration methods such as solid-phase extraction (SPE),¹⁴ liquid–liquid extraction,¹⁵ solvent-assisted dispersive-SPE,¹⁶ hollow fiber liquid phase micro-extraction,¹⁷ single drop microextraction,¹⁸ and cloud point extraction¹⁹ have been reported for silver determination. Among these sample preparation methods, SPE is extensively employed approach for the extraction of metals ions. The SPE benefits from the advantages such as low consumption of solvent, high sensitivity, simplicity and rapidity.² Different types of adsorbent materials have been reported for the extraction of silver such as chelating resin,² modified silica gel,⁴ cellulose nitrate membrane,⁶ modified magnetic nanoparticles^{9,13} modified magnetic graphene oxide,¹⁰ carbon nanotubes,²⁰ Amberlite XAD-16 resin²¹ and modified magnetic halloysite nanotubes.²²

^aDepartment of Chemistry, Central Tehran Branch, Islamic Azad University, Tehran, 1467686831, Iran. E-mail: dr.manoochehri@yahoo.com; Fax: +98 2188385798; Tel: +98 9127242698

^bDepartment of Polymer Engineering and Color Technology, Amirkabir University of Technology, 424 Hafez Avenue, P. O. Box: 15875-4413, Tehran, Iran

† Electronic supplementary information (ESI) available. See DOI: 10.1039/d1ra00420d



MOFs, a class of hybrid porous crystalline materials, are fabricated by the metal ion centers and organic linkers. The excellent structural diversity of MOFs as well as their inimitable features, led to their wide application in the analytical chemistry.^{23,24} MOFs have effectively been applied as adsorbents in diverse extraction approaches.^{25–27} Various types of MOFs have been reported for the extraction of heavy metal ions such MOF-199,^{28,29} MIL-101(Fe),^{30,31} MIL-96(Al)³² and aluminium-fumarate MOF.³³ Among divers type of MOFs, MIL-101(Cr) is one of the widely utilized adsorbent owing to its high chemical stability in the wide range of pH, large specific high surface area and high metal content.³⁴

Herein, a novel MMOF functionalized with 2-aminothiophenol (2-ATP) was fabricated and employed for the separation/preconcentration of trace silver amounts. At first magnetite nanoparticles (Fe_3O_4 NPs) was synthesized and then coated with SiO_2 . Thereafter, the $\text{Fe}_3\text{O}_4@\text{SiO}_2$ nanoparticles were modified with 2-ATP. Finally, the functionalized MMOF was fabricated by the fabrication of MIL-101(Cr) in the presence of $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ NPs. MIL-101(Cr)/ $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ nanocomposite was characterized with various methods and then utilized to separation/quantification of silver ions in various real matrixes. Modification of MIL-101 with $\text{Fe}_3\text{O}_4@2\text{-aminothiophenol}$ NPs not only improves the selectivity of this sorbent toward silver ions but also facilitates the separation and extraction process.

2. Experimental

2.1 Reagents and solutions

All reagents ($\text{Cr}(\text{NO}_3)_3 \cdot 9\text{H}_2\text{O}$, $(\text{NH}_4)_2\text{Fe}(\text{SO}_4)_2 \cdot 6\text{H}_2\text{O}$, FeCl_3 , hydrochloric acid, nitric acid, sodium hydroxide, (3-chloropropyl)-triethoxysilane (3-CPTS), 1,4-benzene dicarboxylic acid (H₂BDC), methanol, tetraethyl orthosilicate (TEOS), *N,N*-dimethylformamide (DMF), ammonium hydroxide (28% w/v), toluene, triethylamine (TEA), 2-aminothiophenol (2-ATP), and ethanol) were of analytical grade and purchased from Merck (Darmstadt, Germany) or Fluka (Seelze, Germany). Stock solution (1000 mg L⁻¹) of Ag(i) was supplied from Merck. All working solutions were prepared in deionized water. Stock solutions of other species were prepared in deionized water containing 2% (v/v) nitric acid.

River water and well water samples (Tehran, Iran) were collected in clean polyethylene bottles. Radiological wastewater sample was collected in clean polyethylene bottles from a hospital (Karaj, Iran). Photographic and electroplating wastewater samples were collected in clean polyethylene bottles (local workshops, Karaj, Iran). All the samples were filtered *via* cellulose membrane (Millipore, 0.45 μm) and then were stored in cleaned polyethylene bottles.

2.2 Instrumentation

An AA-680 flame atomic absorption spectrometer (FAAS) from Shimadzu (Kyoto, Japan) consists of an air-acetylene flame; a silver hollow cathode lamp (radiation source) with the wavelength of 328.1 nm was used for silver quantification. The instrument was set according to the manufacturer's guides. The

morphology of nanocomposite was explored *via* a MIRA3 TESCAN field emission scanning electron microscope (FESEM, Czech Republic). FT-IR spectrum of the sorbent was recorded using KBr pellet method and by employing a Bomem MB-Series spectrophotometer (USA). Magnetic features of the fabricated materials were studied using a vibrating sample magnetometer (VSM) model AGFM/VSM117 3886 (Kashan, Iran). The pH was measured utilizing a Metrohm pH-meter model 827 consist of a glass calomel electrode. X-ray diffraction (XRD) pattern of MIL-101/ $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ nanocomposite was recorded using a Philips-PW 12C diffractometer (Amsterdam, The Netherlands) consist of a copper K α radiation source. CHNS content of MIL-101/ $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ nanocomposite was explored using a Thermo Finnigan Flash EA112 elemental analyzer (Okehampton, UK).

2.3 Sorbent fabrication

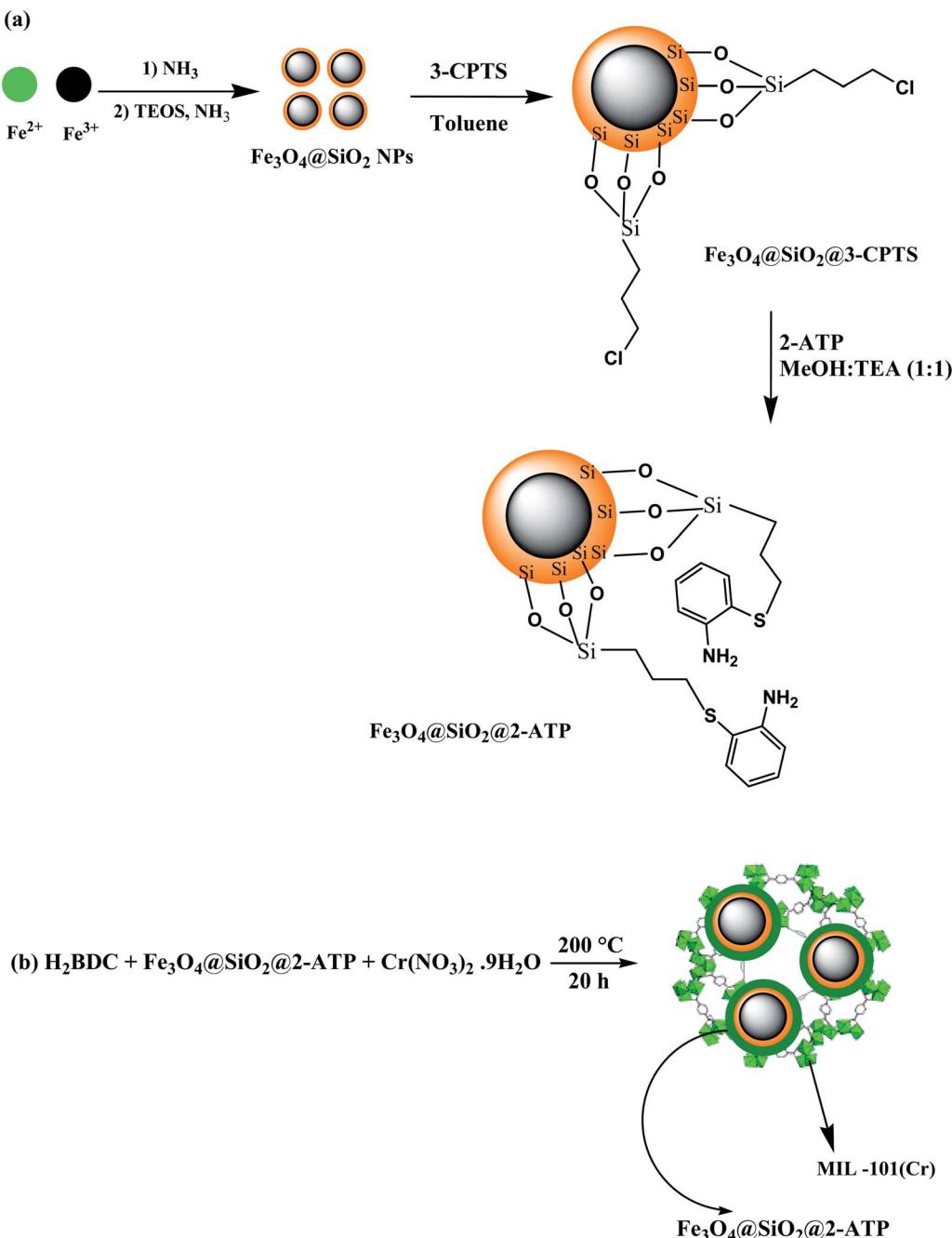
Fabrication of $\text{Fe}_3\text{O}_4@\text{SiO}_2$ NPs was conducted according to our previously reported works.^{8,34} To functionalize the $\text{Fe}_3\text{O}_4@\text{SiO}_2$ NPs, 2.0 g fabricated $\text{Fe}_3\text{O}_4@\text{SiO}_2$ NPs were added to 150 mL dried toluene, and stirred for 30 min. Then, 3.0 mL 3-CPTS was added to the mixture and it was refluxed for 15 h at 130 °C. Thereafter, the product was magnetically separated and then washed with ethanol and acetone. Finally, the product was dried at room temperature. In the next step, 1.0 g $\text{Fe}_3\text{O}_4@\text{SiO}_2@3\text{-CPTS}$ was suspended in 100 mL of triethylamine and methanol mixture (1 : 1 v/v), then 1.5 g 2-ATP was added and the mixture was refluxed for 20 h. Ultimately, $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ NPs were washed with ethanol, and dried at ambient temperature. A scheme for the synthesis process is exhibited in Fig. 1a.

To synthesize MIL-101/ $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ nanocomposite, 2.0 mmol H₂BDC was added to 20 mL of DI water (solution 1). After that, 0.5 g $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ NPs was added to 30 mL aquatic solution containing 2.0 mmol $\text{Cr}(\text{NO}_3)_3 \cdot 9\text{H}_2\text{O}$ (solution 2). Afterward, two mixtures (1 and 2) were transferred into an autoclave and it was heated at 200 °C for 20 h.^{8,34} Finally, MIL-101(Cr)/ $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ NPs was isolated magnetically and washed with deionized water (4 × 25 mL) and ethanol (5 × 25 mL), respectively, and then dried at ambient temperature. MIL-101(Cr)/ $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ was characterized by FT-IR spectroscopy, VSM, CHNS, FESEM, and XRD techniques. A scheme of the fabrication process is illustrated in Fig. 1b.

2.4 Separation/preconcentration procedure

At first, the pH of sample was fixed at 6.2 by using and 0.1 mol L⁻¹ HCl and NH₄OH solutions. Then, MIL-101(Cr)/ $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ adsorbent (21 mg) was added to the sample and the suspension was stirred for 13 min. After that, the solution was magnetically separated from the adsorbent and the uptake percentage determined after FAAS quantification. To desorb Ag(i) ions, the adsorbent was suspended in 1.7 mL 0.56 mol L⁻¹ nitric acid solution and the mixture was stirred for 16 min. Afterward, the adsorbent was isolated from the extraction medium and the supernatant was introduced to FAAS into determine its silver content.





TEOS: tetraethyl orthosilicate
 2-ATP: 2-aminothiophenol
 3-CPTS: 3-chloropropyl triethoxysilane
 H₂BDC: benzene-1,4-dicarboxylic acid

Fig. 1 A scheme for the synthesis of (a) $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ nanoparticles, and (b) MIL-101(Cr)/ $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ nanocomposite.

3. Results and discussion

3.1 Characterization studies

The FT-IR spectra of MIL-101(Cr) and MIL-101(Cr)/ $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ NPs composite were recorded to specify the functional moieties of these materials. The characteristics peaks of MIL-101(Cr) are observable at 1372 and 1569 cm^{-1} ($\text{O}-\text{C}-\text{O}$,

dicarboxylate groups), 1624 cm^{-1} ($\text{C}=\text{O}$), and 3389 cm^{-1} ($\text{O}-\text{H}$ of adsorbed water) (Fig. 2a). In the spectrum of the MIL-101(Cr)/ $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ composite, the characteristics bands at 588 cm^{-1} ($\text{Fe}-\text{O}$), 1046 cm^{-1} ($\text{Si}-\text{O}-\text{Si}$), 2929 cm^{-1} ($\text{C}-\text{H}$ aliphatic), 1500 – 1650 cm^{-1} ($\text{C}=\text{N}$, $\text{C}=\text{C}$), 3394 cm^{-1} ($\text{O}-\text{H}$), and 3429 cm^{-1} ($\text{N}-\text{H}$) proved the modification of MIL-101(Cr) with 2-ATP (Fig. 2b).



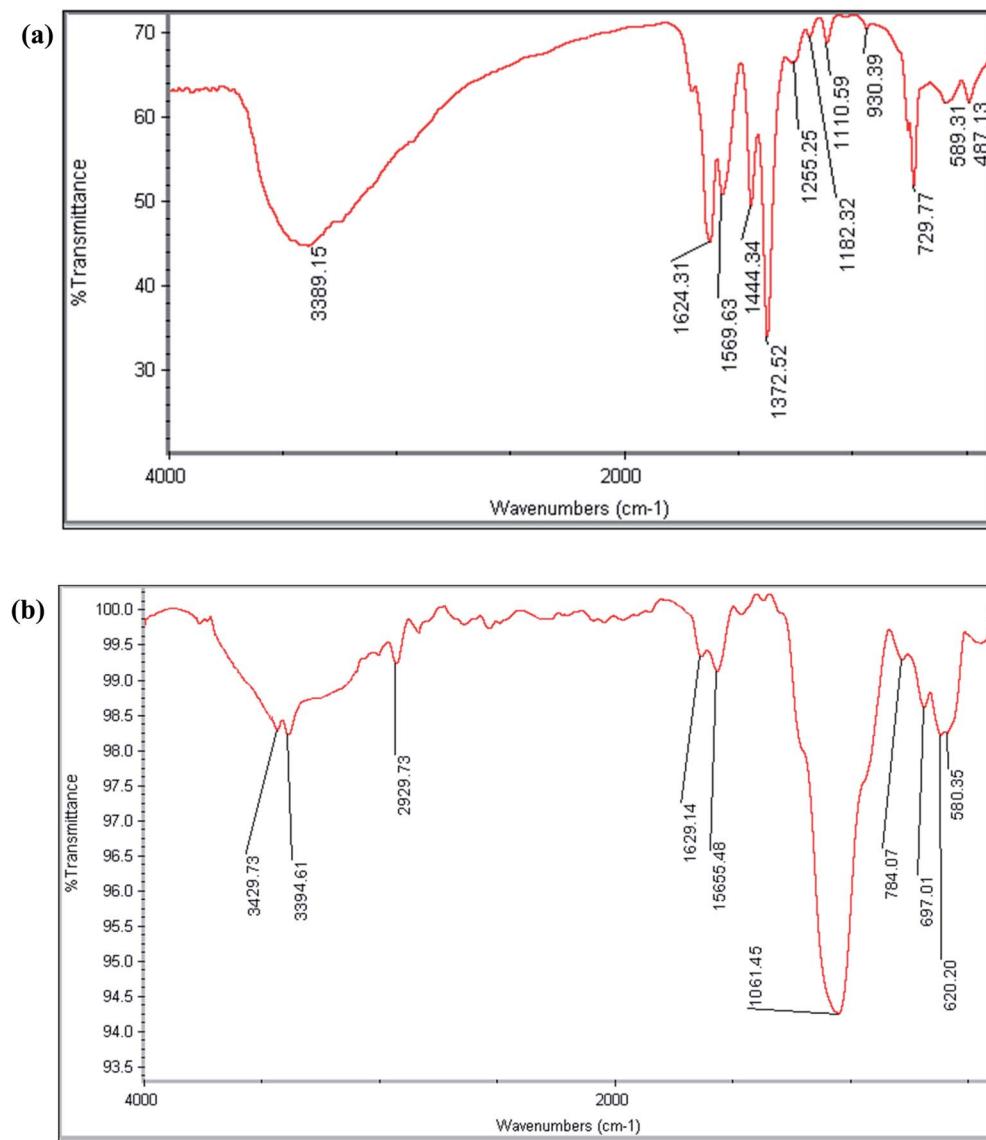


Fig. 2 FT-IR spectra of (a) MIL-101(Cr) and (b) MIL-101(Cr)/ $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ nanocomposite.

Elemental analysis was accomplished to study the CHNS composition of the MIL-101(Cr)/ $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ composite. The results showed 29.5% C, 2.1% N, 2.7% S and 1.5% H for MIL-101(Cr)/ $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ NPs composition. The advent of nitrogen and sulfur elements in the composition of MIL-101(Cr)/ $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ approved the functionalization of the nanocomposite with 2-APT because only this reagent contain N and S elements among the utilized materials.

The surface and dimension of MIL-101(Cr)/ $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ was explored by FESEM method (Fig. 3a). The SEM micrograph of MIL-101(Cr)/ $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ nanocomposite exhibited a rough surface which can be attributed to the presence of $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ NPs. $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ NPs are spherical in shape as it is observable in the SEM micrograph the average size of these particles is 25 nm that is in good agreement with the previous reported MMOFs.^{28,34}

To explore magnetic characteristics of MNPs, $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ and MIL-101(Cr)/ $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ nanocomposites, VSM plots of these materials were recorded as are show in Fig. 3b. Saturation magnetization value for MNPs, $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ and MIL-101(Cr)/ $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ composite was 67, 41 and 26 emu g⁻¹, respectively. The obtained plots and values exhibit that these materials are superparamagnetic and can be magnetically isolated by a common external magnetic field.²³

The structure of MIL-101(Cr)/ $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ composite was characterized by XRD method and the obtained pattern is exhibited in Fig. 4. Five characteristic peaks at $2\theta = 30.8^\circ$, 36° , 58° , 64° and 75.1° are attributed to of $\text{Fe}_3\text{O}_4@\text{SiO}_2@2\text{-ATP}$ NPs and approve the presence of these particles in the composite structure. Besides, four characteristic peaks which are corresponded to MIL-101(Cr) are observable at $2\theta = 2^\circ$, 17.6° , 20° and 28.3° prove the presence of MOF in the composite.³⁴ These



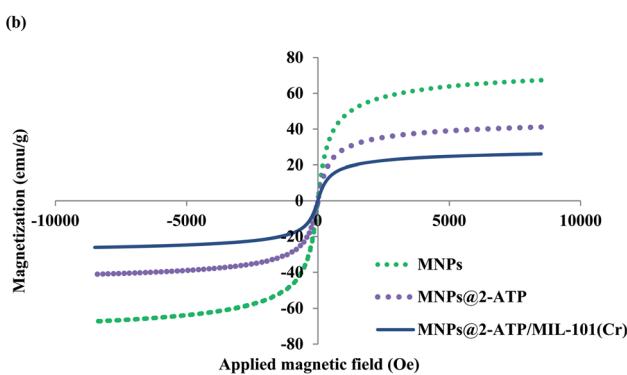
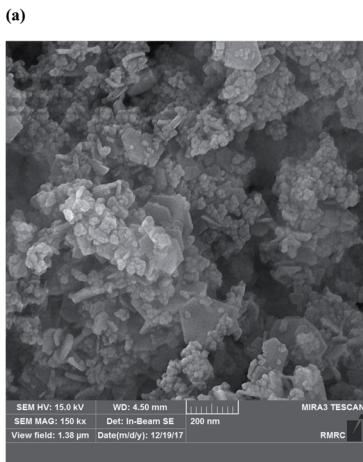


Fig. 3 (a) FESEM of MIL-101(Cr)/Fe₃O₄@SiO₂@2-ATP nanocomposite and (b) VSM curves Fe₃O₄ NPs, Fe₃O₄@SiO₂@2-ATP NPs and MIL-101(Cr)/Fe₃O₄@SiO₂@2-ATP composite.

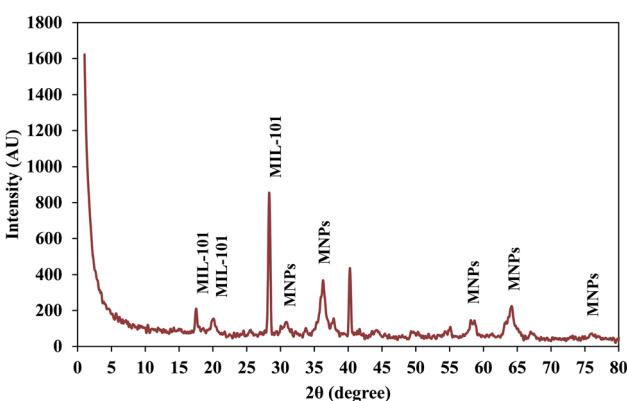


Fig. 4 XRD pattern of MIL-101(Cr)/Fe₃O₄@SiO₂@2-ATP nanocomposite.

results confirm that the fabricated material is a hybrid of MIL-101(Cr) and Fe₃O₄@SiO₂@2-ATP NPs. Moreover, the average crystallite size of Fe₃O₄@SiO₂@2-ATP NPs in the MIL-101(Cr)/Fe₃O₄@SiO₂@2-ATP composite was determined from the XRD pattern by using Scherrer equation and this value was about 20 nm, which is in good agreement with the value of FESEM method.

3.2 Optimization of the affecting factors

3.2.1 Uptake step. In this step, three experimental variables (pH of solution, MIL-101(Cr)/Fe₃O₄@SiO₂@2-ATP dosage, and uptake time) were considered as the main affecting factors. In this regards, experimental design method was utilized to investigate the effect of each variable and their possible interactions. Accordingly, Box-Behnken design (BBD) was employed. According to the number of experiments test equation ($N = 2K(K - 1) + C_0$) 17 tests were designed by setting K (number of factors) at 3 and C_0 (number of center points number) at 5.³⁵ The experimental variables along with their levels are depicted in Table 1. The results exhibit a model *F*-value equal to 193.27 with the *p*-value lower than 0.0001 (at 95% confidence). The *p*-value is lower than 0.05 which approve that the suggested model is significant and cannot be occur as a result of noise. Moreover, the lack of fit is not significant relative to the pure error because its *p*-value is 0.0638.^{24,35} Based on the ANOVA results (Table S1, ESI†), all variables exhibit significant effect on the extraction of silver ions. Among these variables, the pH of solution is the most significant parameter while sorbent dosage and uptake time are the second and third significant factors. The uptake of silver ions increases by increasing the pH of solution. At the lower pH values the heteroatoms of the MIL-101(Cr)/Fe₃O₄@SiO₂@2-ATP NPs are protonated and hence a repulsion force between silver ions and active sites of the sorbent is occur. By increasing the pH, heteroatoms of the MIL-101(Cr)/Fe₃O₄@SiO₂@2-ATP NPs are deprotonated and suitable condition for coordination interaction between Ag(i) ions and active sites of the sorbent will be provided.³⁴ Behind the pH value of 6.2 the extractability of silver ions decreased that can be corresponded to the Ag(i) hydrolysis and competition of hydroxide ion to silver ion. Moreover, Ag(i) is more stable at acidic pH² (Fig. 5). The highest uptake percentage was observed by utilizing a pH value of 6.2, an uptake time of sorption time 13 min and a MIL-101(Cr)/Fe₃O₄@SiO₂@2-ATP NPs composite dosage of 21 mg.

3.2.2 Elution step. In this step, selection of proper elution solution is very important. Thereby, effect of three different acidic solutions (1 mol L⁻¹) of HCl, nitric acid, and sulfuric acid was examined to desorb the silver ions from the MIL-101(Cr)/Fe₃O₄@SiO₂@2-ATP NPs. Other variables including pH of sample, uptake time, MIL-101(Cr)/Fe₃O₄@SiO₂@2-ATP dosage, eluent volume and elution time were set at 6.2, 13 min, 21 mg, 2.0 mL and 15 min, respectively. The highest recovery percentage for silver ions was observed results when nitric acid was employed so it was selected as the best desorbing solution.

After selection of desorbing solution, the effect of elution time, HNO₃ concentration, and eluent volume were investigated and opted *via* BBD. Accordingly, 17 tests were designed and performed (Table 1). The ANOVA results (Table S2, ESI†) exhibited that the model is significant due to the *p*-value lower than (0.0001). The lack of fit *p*-value is 0.0528 that implies there is no pure error. The eluent volume exhibited the most significant positive effect on the recovery and the two other parameters showed a significant positive effect too. The *R*-squared, and adjusted *R*-squared were 0.9877 and 0.9719, respectively that are suitable value for a constructed model. Besides, the predicted *R*-



Table 1 Experimental variables and levels of the Box–Behnken design (BBD)

			Level		
			Lower	Central	Upper
Sorption step	A: pH of sample		3.0	5.0	7.0
	B: uptake time (min)		8.0	12.0	16.0
	C: sorbent dosage (mg)		10.0	17.5	25.0
Elution step	A: elution time (min)		10.0	15.0	20.0
	B: HNO_3 concentration (mol L^{-1})		0.25	0.5	0.75
	C: HNO_3 volume (mL)		1.0	1.5	2.0

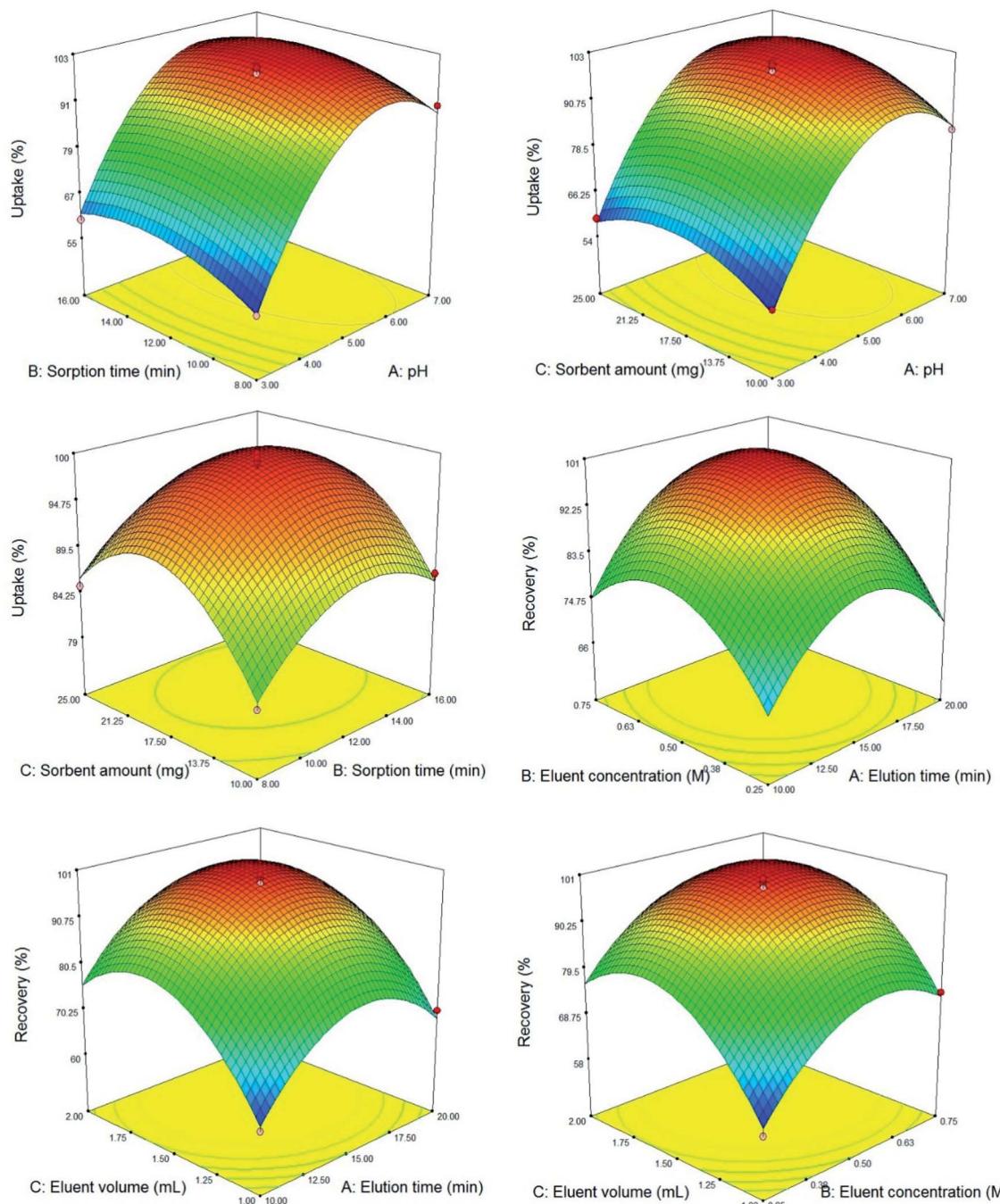


Fig. 5 3D response surface plots of optimization assay.



Table 2 Effect of potentially interfering species on the recovery of silver ions

Species	Tolerance ratio	Recovery (%) \pm SD ^a
Na ⁺	15 000	101.2 \pm 4.3
K ⁺	15 000	99.4 \pm 5.1
Ca ²⁺	5000	98.0 \pm 3.7
Mg ²⁺	5000	96.9 \pm 3.5
Cr ³⁺	1500	95.2 \pm 5.0
Fe ³⁺	1500	98.5 \pm 6.2
Mn ²⁺	1500	97.8 \pm 4.5
Al ³⁺	1500	99.2 \pm 6.0
Au ³⁺	1000	95.3 \pm 3.9
Zn ²⁺	1000	96.4 \pm 4.6
Ni ²⁺	1000	95.7 \pm 5.6
Cd ²⁺	1000	97.7 \pm 5.1
Pb ²⁺	1000	98.5 \pm 4.8
Hg ²⁺	750	95.0 \pm 4.0
Pd ²⁺	400	95.6 \pm 3.6

^a Standard deviation. Conditions: pH, 6.2; uptake time, 13 min; adsorbent dosage, 21 mg; elution time, 16 min; 1.7 mL 0.56 mol L⁻¹ nitric acid as the eluent.

squared of 0.8342 is in good agreement with the adjusted *R*-squared. Adequate precision as the S/N measurement tool was 21.485. A ratio greater than 4 is desirable and exhibits the adequate signal. As a result the suggested model can be employed to navigate the design space. To explore the concurrent effect of the variables; response surface plots (Fig. 5) were utilized. As depicted in 3D plots, all the studied variables are optimized properly with a curvature surface. Finally, the best silver ions recovery was observed by employing 1.7 mL 0.56 mol L⁻¹ nitric acid solution as the eluent and an elution time of 16 min.

3.3 Effect of possible interferences and durability study

To determine the possible interfering species, effect of various species on the extraction recovery of silver ions was studied. The tolerance limit of the various species is tabulated in Table 2. The species concentration causing $\pm 5\%$ deviation in the recovery of silver ions was assumed as the tolerance limit. As summarized

in Table 2, the presence of other constituent has no significant effect on the extraction recovery of Ag(I) ions. It can be concluded that the MIL-101(Cr)/Fe₃O₄@SiO₂@2-ATP has good selectivity towards silver ions under the opted extraction conditions.

The durability of MIL-101(Cr)/Fe₃O₄@SiO₂@2-ATP sorbent was studied by investigating the decrease in the Ag(I) ions recovery by performing several uptake–elution cycles under the opted condition. The results revealed that the MIL-101(Cr)/Fe₃O₄@SiO₂@2-ATP can be employed up to 6 times without notable decrease in the silver ions recovery.

3.4 Adsorption capacity and breakthrough volume

The adsorption capacity of MIL-101(Cr)/Fe₃O₄@SiO₂@2-ATP composite toward silver ions was explored by employing a standard solution containing appropriate concentration of Ag(I) ions. The maximum adsorption capacity (MAC) was obtained as the adsorbed amount of silver ions (mg) per MIL-101(Cr)/Fe₃O₄@SiO₂@2-ATP weight (g). Accordingly, the MAC for adsorption of silver ions on MIL-101(Cr)/Fe₃O₄@SiO₂@2-ATP adsorbent is 103 mg g⁻¹.

Effect of breakthrough was explored by varying the solution volume in the range of 50–600 which contains 0.01 mg silver ions. Thereafter, the extraction process was accomplished to determine the breakthrough. The results exhibited that up to 500 mL the recovery of silver ions is quantitative and the dilution effect has no significant. Thereby, an enrichment factor of 294 fold is attained regarding the breakthrough (600 mL) and eluent (1.7 mL) volumes.

3.5 Analytical characteristics and comparison

The analytical characteristics of new method including linear range, LOD and precision were evaluated under the opted condition. The instrument response was linear in the range of 0.2–200 ng mL⁻¹ with the *r*² value of 0.9963. LOD was obtained by $3S_b/m$ equation that was 0.05 ng mL⁻¹. In this equation, *S_b* is the standard deviation of 6 replicate signal of blank and *m* represents the slope of calibration plot. Precision of the new method was explored as RSD% value at three concentration of 5, 50, 150 ng mL⁻¹ (*n* = 5) and which was 9.3%, 6.8% and 4.5%, respectively.

Table 3 Analytical characteristics of the new method compared to the former reported SPE methods for separation/determination of silver

Analytical instrument	Adsorbent	LOD (ng mL ⁻¹)	Linear range (ng mL ⁻¹)	EF ^a	MAC ^b	RSD (%)	Ref.
ICP-OES ^c	Polythiophene-coated Fe ₃ O ₄ NPs	0.2–2.0	0.75–100	114	—	4.2	13
ICP-OES	Modified magnetic NPs	0.12	—	194	10	5.31	36
FI ^d -FAAS	Silica gel based chelating sorbent	1.3	—	15	24	3.0	37
FAAS	mGO@SiO ₂ @PPy-PTh ^e nanocomposite	0.1	0.5–500	125	49	2.7	10
FAAS	Fe ₃ O ₄ NPs@murexide	0.15	0.5–400	225	48	5.0	9
FAAS	MIL-101(Cr)/Fe ₃ O ₄ @SiO ₂ @2-ATP nanocomposite	0.05	0.2–200	294	103	<9.4	Current research

^a Enrichment factor. ^b Maximum adsorption capacity in mg g⁻¹. ^c Inductively coupled plasma optical emission spectrometry. ^d Flow injection.

^e Polypyrrole–polythiophene.



Table 4 Determination of silver ions in diverse real matrixes (mean \pm SD^a)

Sample	Real value	Added value	Found value	Recovery (%)
Radiological wastewater	190 \pm 10	200	381 \pm 28	95.5
Photographical wastewater	210 \pm 13	200	390 \pm 38	90.0
Electroplating wastewater	158 \pm 8	200	345 \pm 23	93.5
River water	10.2 \pm 0.6	10.0	20.5 \pm 1.7	103
Well water	21.8 \pm 1.5	20.0	39.6 \pm 2.5	89.0

^a Standard deviation ($n = 3$). All concentrations are based on ng mL⁻¹.

In order to explore the accuracy of the current method, ore polymetallic gold Zidarovo-PMZrZ (206 BG 326) was analyzed by the recommended extraction procedure. The experimental value (16.9 ± 0.3 mg kg⁻¹) is in good accordance with the certified value of silver (17.2 mg kg⁻¹). To confirm the accuracy of new method, *t*-test was accomplished. The *t*_{experimental} for the three replicates is 1.7 that is lower than *t*_{critical} ($t_{0.05, 2} = 4.30$). The results affirm that there is no significant difference between the experimental value and the real value. Accordingly, the new method can be employed as a reliable procedure for the extraction of silver ions from complex matrices.

The analytical characteristics of the new method including LOD, linearity, precision, MAC and enrichment factor was compared to the former reported methods for separation/determination of silver (Table 3). As summarized in Table 3, LOD and precision of the new method are better than or comparable those of the reported methods. A high MAC of 103 mg g⁻¹ and an excellent enrichment factor as high as 294 was achieved for this method which is higher than the previously methods. Herein, FAAS was employed as the detection system which is a well-known, relatively inexpensive, widely used, and simple to operate instrument.

3.6 Applicability to real matrixes

To explore the analytical applicability of the optimized procedure to real matrixes various real samples such as river water, well water and electroplating, radiological and photographic wastewaters were analyzed. The results of this assay (recoveries and RSD values) are summarized in Table 4. The RSD and recovery values were located in the range of 89–103% and 5.0–9.7%, respectively. These results affirm that the new method is the reliable and accurate for the determination of silver ions in real matrixes and MIL-101(Cr)/Fe₃O₄@SiO₂@2-ATP NPs sorbent has the high performance for preconcentration/separation of Ag(I).

4. Conclusion

In this research, a MMOF functionalized with 2-aminothiophenol (2-ATP) was fabricated and employed for separation/preconcentration of trace silver amounts. At first magnetite nanoparticles (Fe₃O₄ NPs) was synthesized and then coated with SiO₂. Thereafter, the Fe₃O₄@SiO₂ NPs were modified with 2-

ATP. Finally, the functionalized MOF fabricated by the synthesis of MIL-101(Cr) in the presence of Fe₃O₄@SiO₂@2-ATP NPs. MIL-101(Cr)/Fe₃O₄@SiO₂@2-ATP nanocomposite was characterized with various methods and then utilized to separation/quantification of silver ions in various real samples. Modification of MIL-101 with Fe₃O₄@2-APT NPs not only improves the selectivity of this sorbent toward silver ions but also facilitates the separation and extraction process. A high adsorption capacity (103 mg g⁻¹), a low LOD (0.05 ng mL⁻¹) and an excellent enrichment factor as high as 294 was achieved. The method was validated by analyzing as a certified reference material and performing *t*-test. Also, the precision of method was located in the range of 4.5–9.3%. As a result, the analytical features of the new extraction method are acceptable. Ultimately, the outlined method was employed for extraction/quantification of silver ions in various water and wastewater samples satisfactorily.

Conflicts of interest

There are no conflicts to declare.

References

- 1 T. Rohani and M. A. Taher, *Talanta*, 2010, **80**, 1827–1831.
- 2 T. Çetin, S. Tokaloğlu, A. Ülgen, S. Şahan, İ. Özentürk and C. Soykan, *Talanta*, 2013, **105**, 340–346.
- 3 C. K. Christou and A. N. Anthemidis, *Talanta*, 2009, **78**, 144–149.
- 4 T. Madrakian, A. Afkhami, M. A. Zolfigol and M. Solgi, *J. Hazard. Mater.*, 2006, **128**, 67–72.
- 5 T. W. Purcell and J. J. Peters, *Environ. Toxicol. Chem.*, 1998, **17**, 539–546.
- 6 M. Soylak and R. S. Cay, *J. Hazard. Mater.*, 2007, **146**, 142–147.
- 7 A. A. Asgharinezhad, N. Jalilian, H. Ebrahimzadeh and Z. Panjali, *RSC Adv.*, 2015, **5**, 45510–45519.
- 8 S. Rezabek and M. Manoochehri, *RSC Adv.*, 2020, **10**, 36897–36905.
- 9 S. Karami, H. Ebrahimzadeh and A. A. Asgharinezhad, *Anal. Methods*, 2017, **9**, 2873–2882.
- 10 N. Jalilian, H. Ebrahimzadeh, A. A. Asgharinezhad and K. Molaei, *Microchim. Acta*, 2017, **184**, 2191–2200.
- 11 F. Gerondi and M. A. Z. Arruda, *Talanta*, 2012, **97**, 395–399.



12 Y. S. Chung and R. M. Barnes, *J. Anal. At. Spectrom.*, 1988, **3**, 1079–1082.

13 E. Tahmasebi and Y. Yamini, *Microchim. Acta*, 2014, **181**, 543–551.

14 H. Ebrahimzadeh, N. Shekari, N. Tavassoli, M. M. Amini, M. Adineh and O. Sadeghi, *Microchim. Acta*, 2010, **170**, 171–178.

15 O. Ortet and A. P. Paiva, *Sep. Sci. Technol.*, 2010, **45**, 1130–1138.

16 J. A. López-López, J. A. Jönsson, M. García-Vargas and C. Moreno, *Anal. Methods*, 2014, **6**, 1462–1467.

17 M. Chamsaz, M. H. Arbab-Zavar and J. Akhondzadeh, *Anal. Sci.*, 2008, **24**, 799–801.

18 G. Absalan, M. Akhond, L. Sheikhian and D. M. Goltz, *Anal. Methods*, 2011, **3**, 2354–3259.

19 J. B. Chao, J. F. Liu, S. J. Yu, Y. D. Feng, Z. Q. Tan, R. Liu and Y. G. Yin, *Anal. Methods*, 2011, **83**, 6875–6882.

20 Q. Ding, P. Liang, F. Song and A. Xiang, *Sep. Sci. Technol.*, 2006, **41**, 2723–2732.

21 A. Tunçeli and A. R. Türker, *Talanta*, 2000, **51**, 889–894.

22 M. Amjadi, A. Samadi, J. L. Manzoori and N. Arsalani, *Anal. Methods*, 2015, **7**, 5847–5853.

23 N. Jalilian, H. Ebrahimzadeh and A. A. Asgharinezhad, *J. Chromatogr. A*, 2019, **1608**, 460426.

24 N. Jalilian, H. Ebrahimzadeh and A. A. Asgharinezhad, *Microchim. Acta*, 2019, **186**, 597.

25 S. Rezabek and M. Manoochehri, *Microchim. Acta*, 2018, **185**, 365.

26 Z. Li, M. Qi, C. Tu, W. Wang, J. Chen and A. J. Wang, *Food Anal. Methods*, 2017, **10**, 4094–4103.

27 T. Liang, S. Wang, L. Chen and N. Niu, *Food Anal. Methods*, 2019, **12**, 217–228.

28 A. Tadjarodi and A. Abbaszadeh, *Microchim. Acta*, 2016, **183**, 1391–1399.

29 M. Taghizadeh, A. A. Asgharinezhad, M. Pooladi, M. Barzin, A. Abbaszadeh and A. Tadjarodi, *Microchim. Acta*, 2013, **180**, 1073–1084.

30 M. Babazadeh, R. Hosseinzadeh-Khanmiri, J. Abolhasani, E. Ghorbani-Kalhor and A. Hassanpour, *RSC Adv.*, 2015, **5**, 19884–19892.

31 E. Ghorbani-Kalhor, R. Hosseinzadeh-Khanmiri, J. Abolhasani, M. Babazadeh and A. Hassanpour, *J. Sep. Sci.*, 2015, **38**, 1179–1186.

32 A. Mehdinia, D. Jahedi Vaighan and A. Jabbari, *ACS Sustainable Chem. Eng.*, 2018, **6**, 3176–3186.

33 R. Kashanaki, H. Ebrahimzadeh and M. Moradi, *New J. Chem.*, 2018, **42**, 5806–5813.

34 M. Mehraban, M. Manoochehri and F. A. Taromi, *New J. Chem.*, 2018, **42**, 17636–17643.

35 A. A. Asgharinezhad and H. Ebrahimzadeh, *J. Chromatogr. A*, 2015, **1412**, 1–11.

36 M. H. Mashhadizadeh and Z. Karami, *J. Hazard. Mater.*, 2011, **190**, 1023–1029.

37 P. Liu, Q. Pu and Z. Su, *Analyst*, 2000, **125**, 147–150.

