


Cite this: *RSC Adv.*, 2021, 11, 17080

Adsorption process and mechanism of heavy metal ions by different components of cells, using yeast (*Pichia pastoris*) and Cu²⁺ as biosorption models

Xinggang Chen,^{ab} Zhuang Tian,^{ab} Haina Cheng,^{*ab} Gang Xu^{*c} and Hongbo Zhou^{*ab}

Microbial biomass has been recognized as an essential biosorbent to remove heavy metal ions, but the biosorption process and mechanism of different components of microbial cells have not been elucidated. In present study, *Pichia pastoris* X33 and Cu²⁺ was used as a biosorption model to reveal the biosorption process and mechanism of different components of microbial cells. For the biosorption of whole cells, the maximum removal efficiency was 41.1%, and the adsorption capacity was 6.2 mg g⁻¹. TEM-EDX analysis proved the existence of Cu²⁺ on the cell surface and cytoplasm. The maximum Cu²⁺ removal efficiency of the cell wall, cell membrane and cytoplasm were 21.2%, 20.7% and 18.5%, respectively. The optimum pH of Cu²⁺ biosorption of the *P. pastoris* cell, cell wall, cell membrane and cytoplasm was 6. Moreover, the maximum adsorption capacity of the cell, cell wall, cell membrane and cytoplasm was 16.13, 11.53, 10.97 and 8.87 mg g⁻¹, respectively. The maximum removal efficiencies of *P. pastoris* biomass treated with proteinase K and *P. pastoris* biomass treated with β-mannanase were 18.1% and 28.2%, respectively. The maximum removal efficiencies of mannan and glucan were 34% and 12%, respectively. The FTIR spectra showed that the amino group (N–H), hydroxyl (O–H), carbon oxygen bond (C–O), –CH, C–N and carbonyl group (C=O) of a ketone or aldehyde may interact with Cu²⁺. Thus, our work provides guidance for further understanding the effect of different cell components on biosorption.

Received 17th November 2020

Accepted 1st May 2021

DOI: 10.1039/d0ra09744f

rsc.li/rsc-advances

1. Introduction

With the development of industrialization, agricultural activities, and other human activity, various kinds of produced heavy metals have been discharged into water, water resources have been seriously polluted.^{1,2} Heavy metal contamination has been the focus for a long time due to its high toxicity and the difficulty of removal.^{3–7}

Many efforts have been made to find a way to effectively reduce and remove heavy metals before waste water is discharged into the environment, such as ion exchange,⁸ chemical precipitation,⁹ evaporation, flotation, membrane filtration,¹⁰ electrochemical,¹¹ coagulation–flocculation, and biosorption.¹² Although these technologies have proven effective, the sustainability and economic cost was hard to be underestimated.¹³ Biosorption, which mainly used the organisms as adsorbents, may be an alternative owing to strong adaptability, low cost, no secondary pollution, low energy consumption and

high efficiency. A large quantity of studies indicated that many organisms can efficiently adsorb heavy metal ions such as soybean meal waste, sugarcane bagasse, fungi and bacteria.¹⁴ However, most studies have only studied the biosorption properties of microbial intact cells, and limited reports have discussed the effects of microbial cell components on the biosorption process. Exploring the biosorption process and mechanism of different components of cells is very important for understanding biosorption.¹⁵

Element of Cu is a necessary element for biological growth. However, high concentrations of Cu are harmful to organisms.^{16–19} Excessive intake of Cu²⁺ leads to copper poisoning, diarrhea, epigastric pain, nausea and vomiting. Severe cases can lead to gastrointestinal mucosal ulcers, kidney damage, hemolysis, liver necrosis, shock, and even death.²⁰ Therefore, the World Health Organization (WHO) have recommended copper concentration in drinking water not to exceed 1.3 mg L⁻¹.²¹ However, copper is widely used in industrial productions such as electroplating, alloy manufacturing, refining processes and surface treatment industry, which inevitably resulted in the seriously pollution in water.^{15,22–26} Therefore, copper pollution is an urgent problem to be solved in heavy metal pollution, and the use of Cu²⁺ as the biosorption model of this study has certain representativeness.

^aSchool of Minerals Processing and Bioengineering, Central South University, Changsha, 410083, Hunan, China. E-mail: xgchencsuft@163.com; 1689256864@qq.com; zhouhb@csu.edu.cn; 443284336@qq.com

^bKey Laboratory of Biometallurgy of Ministry of Education, Changsha 410083, China

^cHunan Flag Bio-Tech Co., Ltd, Changsha, Hunan, 410083, China. E-mail: xugang55@hotmail.com



Yeast, a traditional model fungus, is commonly used in genetic engineering and fermentation engineering.²⁷ As a result, large quantities of waste yeast produced during these processes. Commonly, waste yeast biomass was used as organic fertilizer and feed.²⁸ Recently, consideration of yeast as an inexpensive biosorbents for the removal of metal ions becomes the focus.²⁹ The reason is there are many advantages of yeast, such as large-scale cultivation, low safety risks and easy to use.^{15,30} Many studies have studied the biosorption characteristics of yeast. Chen Can studied the morphological changes of *Saccharomyces cerevisiae* before and after biosorption of Ag^+ .³¹ M. Fadel explored the biosorption properties and optimum biosorption conditions of *Saccharomyces cerevisiae* for biosorption of manganese ions.³² Furthermore, Yunsong Zhang produced a bifunctional *Saccharomyces cerevisiae* as an adsorbent for Cd^{2+} or Pb^{2+} removal from aqueous solution.³³ Fatemeh Elahian demonstrated that the genetically modified *Pichia pastoris* is a cost-effective, high-throughput, robust, and facile system for metal biosorption.³⁴

All the above reports showed that yeast is an excellent and widely used adsorbent. However, the biosorption process is still unknown. *Pichia pastoris* was a typical model adsorbent, the exploration of the biosorption mechanism of different cell component can also provide a theoretical basis for understanding the biosorption mechanism of other adsorbents.

Here, *Pichia pastoris* X33 and Cu^{2+} were used as biosorption model. The biosorption process and mechanisms of *Pichia pastoris* were investigated. Firstly, through focusing on biosorption process of Cu^{2+} by *Pichia pastoris* biomass, the biosorption stage of Cu^{2+} by *Pichia pastoris* biomass was revealed. In addition, the relative biosorption ability of the main components of cell (cell wall, cell membrane and cytoplasm) and cell wall (glucan, protein, β -mannan) to Cu^{2+} was determined. Finally, the initial molecular mechanism of *Pichia pastoris* biosorption of Cu^{2+} was explored. Through this research we hope to provide a theoretical basis for the application of biological removal of heavy metals.

2. Materials and method

2.1 Preparation of solution

The copper sulphate pentahydrate was dissolved in deionized water to obtain a stock Cu^{2+} solution with a concentration of 400 mg L^{-1} . The test solution with different concentrations was prepared by appropriately diluting the stock solution.

2 g L^{-1} dicyclohexanone oxalyldihydrazone (BCO) solution: 1 g BCO was heated and dissolved in $400 \text{ mL } 50\% \text{ ethanol}$ solution (deionized water : ethanol = 1 : 1, V : V). After the solution was cooled, the volume was fixed to 500 mL .

$\text{pH } 9.0 \text{ NH}_4\text{Cl-NH}_3$ buffer: 35 g of NH_4Cl was dissolved in appropriate deionized water, and then 24 mL of 15 mol L^{-1} ammonia water was added. Finally, the volume of the solution was determined to 500 mL with deionized water.

500 g L^{-1} ammonium citrate solution: 250 g ammonium citrate was diluted to 500 mL with deionized water.

Proteinase K purchased from BioFroxx: enzyme activity $\geq 30 \text{ U mg}^{-1}$, $\text{pH } 6.2\text{--}6.8$.

β -Mannanase purchased from Hunan Lerkam Biology Corp., Ltd: enzyme activity $50\,000 \text{ U g}^{-1}$, $\text{pH } 5.0\text{--}6.0$.

Snailase purchased from Sigma-Aldrich (Shanghai) Trading Co., Ltd: $\text{pH } 5.2\text{--}7.2$. Hypertonic buffer: 0.8 mol L^{-1} mannitol prepared by phosphate buffer ($\text{pH } 6.0$).

Pre-treatment agent: 0.1 g EDTA-Na_2 and $0.1 \text{ mL } \beta$ -mercaptoethanol was dissolved in 100 mL deionized water. $2\% \text{ SDS}$ buffer: 2 g sodium dodecyl sulfate (SDS) was dissolved in 100 mL deionized water.

$\text{pH } 2.0\text{--}8.0 \text{ Na}_2\text{HPO}_4$ -citric acid buffer: different volumes of $0.2 \text{ mol L}^{-1} \text{ Na}_2\text{HPO}_4$ and 0.1 mol L^{-1} citric acid were dissolved to obtain different pH buffers. Different concentrations and pH of Cu^{2+} were prepared by dissolving copper sulphate pentahydrate in this buffer.

2.2 Preparation of biosorbents

Pichia pastoris X33 was acquired from Hunan Flag Bio-Tech Co., Ltd. It was preserved in Yeast Extract Peptone Dextrose (YPD) medium at 4°C . Cells were cultivated in liquid YPD medium using a rotating incubator at 250 rpm and 30°C for 12 h , the *P. pastoris* cells were harvested by centrifugation at 7000 rpm for 10 min . Each part of cells was obtained according to the following steps.

Cell wall: 0.5 g wet *P. pastoris* X33 biomass were suspended by 20 mL deionized water, frozen 1 h at -80°C and thawed at room temperature. The above procedure was repeated 3 times to slightly break the *P. pastoris* cells. Then ultrasonic cell-break was carried out using a JY92-IIN ultrasonic cell breaker (SCI-ENTZ, China) under ice bath conditions. The cells were broken repeatedly until intact cells were failed to be identified by microscopy. Then, let the broken cell suspension centrifuged $10\,000 \text{ rpm}$ for 15 min .^{35,36} The broken cell suspension was then divided into two parts, the precipitation was the cell wall and the membrane, the supernatant was the cytoplasm. In order to remove the cell membrane mixed in the cell wall, the precipitation was extracted by boiling water bath with $20 \text{ mL } 2\% \text{ SDS}$ buffer for 1 h and then washed with deionized water ($10\,000 \text{ rpm}$, 15 min) until the supernatant was clarified. The final precipitation was collected as the cell wall.^{37,38}

Cell membrane and cytoplasm: 0.5 g wet *P. pastoris* X33 biomass were firstly suspended 30 min at 30°C by 5 mL pre-treatment agent. Then, the 0.5 g wet *P. pastoris* X33 biomass which was treated by the pre-treatment agent were washed 2 times by hypertonic buffer (7000 rpm , 5 min). Next, $5 \text{ mL } 2\% \text{ snailase}$ solution was added to the 0.5 g wet *P. pastoris* X33 biomass, and the mixture which contained *P. pastoris* X33 biomass and snailase was then oscillated for 6 h in the rotary shaker. After these steps, the mixture was washed with hypertonic buffer for 2 times to obtain protoplasts. 4 mL deionized water was added to the protoplasts to break the protoplasts. After that, the break protoplasts were centrifuged at 10000 rpm for 15 min . Consequently, the precipitation was the cell membrane, the supernatant was the cytoplasm.

P. pastoris treated with enzymes: 0.5 g wet *P. pastoris* X33 biomass were suspended 30 min at 30°C by pre-treatment agent and then the 0.5 g wet *P. pastoris* X33 biomass which was treated by



the pre-treatment agent were washed 2 times by hypertonic buffer (7000 rpm, 5 min), after that, the 0.5 g wet *P. pastoris* X33 biomass were treated 6 hours with 2% β -mannase and 2% protease K, respectively. Finally, *P. pastoris* biomass that treated with enzymes was harvested by centrifugation at 7000 rpm for 10 min.

2.3 Biosorption studies

Unless otherwise noted, the biosorption experiments were carried out with 20 mL system (20 mL of 100 mg L⁻¹ Cu²⁺ and 0.5 g of wet biosorbents). Besides, the biosorption experiments were carried out on a mechanical shaker at 180 rpm, 25 °C. 1 mL samples were taken out at 5, 10, 15, 30, 60, 90, 120, 150 and 180 min. Samples were then centrifuged at 10 000 rpm for 5 min immediately to separate the biosorbents. After that, the supernatant was analyzed for residual Cu²⁺ concentration in the solution by spectrophotometry.

The procedure of spectrophotometric determination of Cu²⁺ concentration was as follows. Preparation of standard curve: transfer 0.0 mL, 0.2 mL, 0.4 mL, 0.8 mL, 1.2 mL, 1.6 mL, 2.0 mL and 2.4 mL of Cu²⁺ solution (10 mg L⁻¹) into 8 colorimetric tubes, add 200 μ L ammonium citrate solution, 500 μ L NH₄Cl-NH₃ buffer, 500 μ L BCO, and then add deionized water to 5 mL. Finally, the concentration of Cu²⁺ in the solution was analyzed using an Epoch UV-vis spectrophotometer (BioTek Inc, USA) at the wave length of 600 nm. The formula of standard curve was as follows:

$$y = 50.791x + 0.0559, R^2 = 0.9992 \quad (1)$$

where, y was the absorbance at 600 nm, x was the mass of Cu²⁺, mg. The Cu²⁺ concentration of the sample was also measured by the same method.

However, it is difficult to separate mannan, cytoplasm and glucan from Cu²⁺ solution by centrifugation. Therefore, the cytoplasm, glucan and mannan were injected into the dialysis bag respectively (Cu²⁺ could pass through the dialysis bag and the cytoplasm, glucan, mannan could not pass through the dialysis bag). Then put the dialysis bag into the 20 mL 100 mg mL⁻¹ (final concentration) Cu²⁺ solution. Finally, the biosorption experiment was carried. The experiment of different pH on the biosorption was also the same as above, only the 100 mg g⁻¹ Cu²⁺ was replaced with different pH 100 mg g⁻¹ Cu²⁺. After 180 min, the concentration of Cu²⁺ was determined.

The Cu²⁺ biosorption capacity (q) (2) and removal efficiency (R) (1) were calculated through the following equations, respectively:

$$R (\%) = \frac{C_0 - C}{C_0} \times 100\% \quad (2)$$

$$q (\text{mg g}^{-1}) = (C_0 - C) \times \frac{V}{m} \quad (3)$$

where, q (biosorption capacity) was the quantity of Cu²⁺ that be absorbed by the adsorbent per gram (mg g⁻¹), V was the total volume of Cu²⁺ solution (L), R (removal efficiency) was proportion of Cu²⁺ adsorbed by *P. pastoris* to total Cu²⁺. C and C_0 were the residual and initial concentrations of Cu²⁺, respectively. And m was the weight of biosorbents (g).

2.4 Isotherm studies

Biosorption equilibrium isotherms were performed in Erlenmeyer flasks with 0.5 g of dried biosorbents in 100 mL of a solution of Cu²⁺ at different concentrations (50 mg L⁻¹, 100 mg L⁻¹, 150 mg L⁻¹, 200 mg L⁻¹ and 250 mg L⁻¹) and agitated at 180 rpm for 3 h in a shaker at 25 °C. Langmuir eqn (3) and Freundlich eqn (4) models were utilized to describe and evaluate the experimental data.^{39,40}

Langmuir model:

$$\frac{C_e}{q_e} = \frac{1}{q_m} C_e + \frac{1}{q_m b} \quad (4)$$

Freundlich model:

$$\ln q_e = \ln K_F + \frac{1}{n} \ln C_e \quad (5)$$

where, q_e was the equilibrium adsorption capacity of Cu²⁺ (mg g⁻¹), q_m was the theoretical maximum adsorption capacity of Cu²⁺ (mg g⁻¹), C_e was the Cu²⁺ concentration (mg L⁻¹), C_0 was the initial Cu²⁺ concentration (mg L⁻¹), b was Langmuir constant (L mg⁻¹); K_F was the Freundlich constant which represents the adsorption capacity, and n was the Freundlich equation constant.

2.5 Analytic methods

Transmission electron micrograph (TEM) and energy dispersive X-ray spectroscopy (EDX) experiments (TEM-EDX) were performed on a Tecnai G2 20S-Twin transmission electron microscope (FEI Czech Republic s. r. o, Czech) and a GENES XM60S Energy Dispersive Spectrometer (EDAX Inc, USA). The functional groups of the biosorbents were determined using a spectrum two Fourier transform infrared spectroscopy (PerkinElmer, UK). FTIR spectroscopy was carried out in the infrared region ranging from 4000 cm⁻¹ to 450 cm⁻¹. The spectra for before and after biosorption of the prepared adsorbents in the previous step were compared.

3. Results and discussion

3.1 Biosorption of Cu²⁺ by *P. pastoris* biomass

The removal efficiency (R) of Cu²⁺ by *P. pastoris* increased with increasing contact time (Fig. 1). In the first place, the removal percent of Cu²⁺ by *P. pastoris* rose sharply in first 5 min. After that, it rose slowly within 15 min. The biosorption equilibrium was established at 15 min and the maximum removal efficiency was 41.1%. However, no dramatic increase of removal efficiency was observed with a further extension of contact time (>15 min). Further observation showed that the maximum biosorption capacity (q) was 1.10 mg g⁻¹ (6.2 mg g⁻¹ for dry biomass), which was similar to the reported results (Fig. 1).⁴¹

Based on above results, it is suggested that there were two steps for Cu²⁺ biosorption: the first fast step lasted for 15 min (short time) and the second slow stage (long time) lasted until the equilibrium was established.^{23,42} It indicated that the biosorption of Cu²⁺ on the cell surface played an important role in



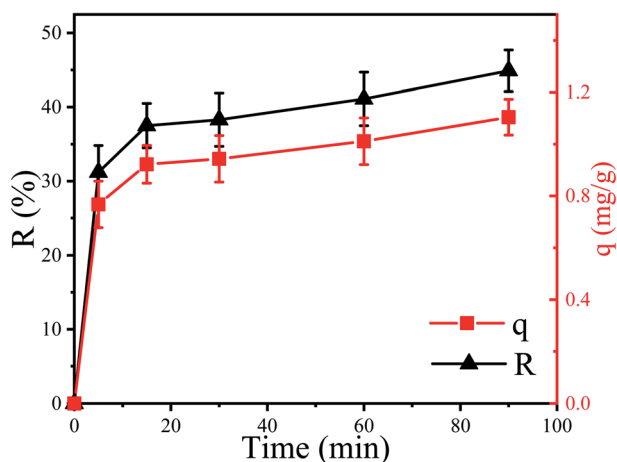


Fig. 1 Biosorption of Cu^{2+} by *P. pastoris* biomass, initial Cu^{2+} concentration: 100 mg L^{-1} , wet biomass dosage: 0.5 g , R – removal efficiency, q – biosorption capacity.

the initial biosorption process. This was because a variety of macromolecules constitute the cell wall of fungi, including mannans, proteins, glucans, lipids, *etc.*⁴³ These complex macromolecular structures provide potential binding sites for many different organic and inorganic molecules.⁴⁴ To sum up, it was speculated that Cu^{2+} was firstly adsorbed on the cell surface and then slowly entered the cell.

3.2 TEM-EDX analysis

To confirm that Cu^{2+} was firstly adsorbed on the cell surface and then slowly entered the cell, TEM-EDX experiments were performed. The TEM observation revealed that white irregular shaped precipitates (Cu^{2+}) were accumulated in the cytoplasm (Fig. 2A and D). Additionally, the locations and shapes of these precipitates were different (Fig. 2A and D, which was in line with prior results.⁴⁵ Therefore, we could preliminarily infer that Cu^{2+} entered the cell and distributed unevenly. Elemental mapping clearly revealed the distribution of Cu^{2+} in the cells (Fig. 2B). The purple dot corresponded to Cu^{2+} formed the shape of a cell (red circle) and was correlated with the brightness of Fig. 2A. In addition, the number of purple dots in the green circle increases significantly. The results indicated that Cu^{2+} was exactly entered into the *P. pastoris* cell.

TEM-EDX further confirmed that Cu^{2+} was adsorbed on the cell surface and inside. Six curves in Fig. 2C represented the abundance of C, O, Na, S, Cl, Cu elements from top to bottom. When the white line passed through the cell from the upper left to the lower right, the curves changed from left to right representing the change of element abundance. It could be seen that the white line entered the cell, the abundance of Cu^{2+} increased slightly. Then, the white line continued downward to the right, and entered the white precipitate. At this time, the abundance of Cu^{2+} was greatly improved. Moreover, when white lines passed through the white precipitates, the abundance of Cu^{2+} decreased dramatically. Finally, the white line passed through

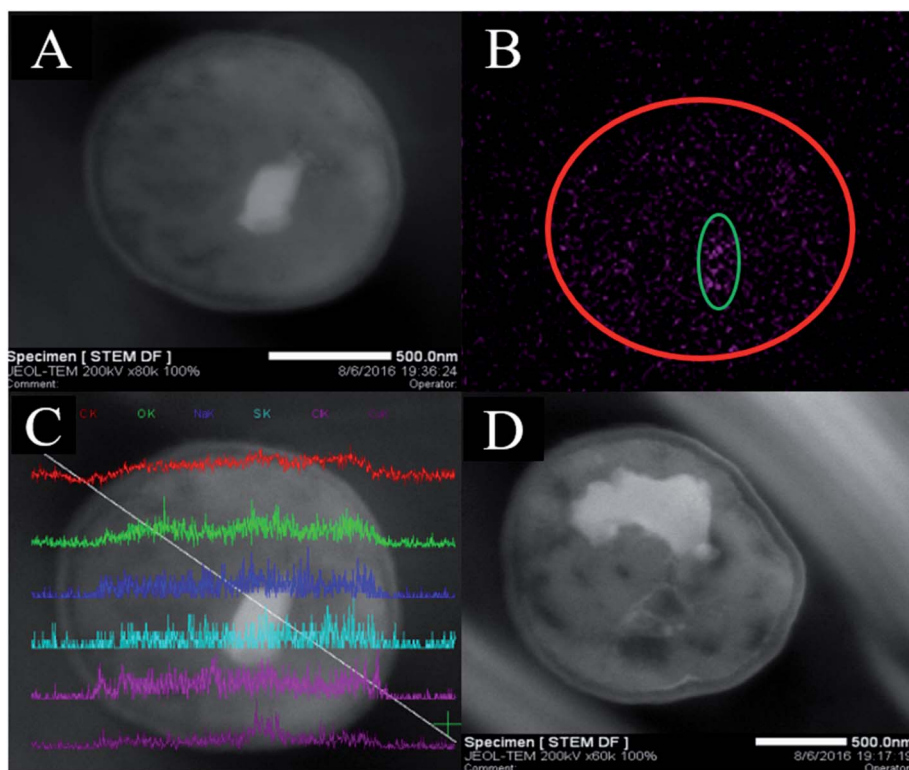


Fig. 2 TEM-EDX analyses of Cu^{2+} -loaded *P. pastoris* biomass after exposure to 100 mg L^{-1} Cu^{2+} for 1500 min, (A) TEM images of thin section of Cu^{2+} -loaded *P. pastoris* cell; (B) elemental mapping of Cu^{2+} -loaded *P. pastoris* cell, (C) line scan of Cu^{2+} -loaded *P. pastoris* cell; (D) TEM images of thin section of Cu^{2+} -loaded *P. pastoris* biomass.

the cell, the abundance of Cu^{2+} decreased. Above evidences indicated that Cu^{2+} existed in the interior and surface of *Pichia pastoris* cell. Previous studies based on scanning electron microscopy (SEM) and EDX analysis also indicated that heavy metal ions could be adsorbed on the surface of cells.⁴⁶

Therefore, biosorption and TEM-EDX experiments further explained that biosorption was carried out in two steps. Firstly, Cu^{2+} bound to functional groups on cell surface by ion-exchange or electrostatic interaction.^{47,48} Subsequently, Cu^{2+} passed through different layers of the cell wall (glucan, mannan and protein), then transported into the cell membrane, and eventually binds to the cytoplasm.

3.3 Biosorption of Cu^{2+} by *P. pastoris* cell components

The results of TEM-EDX showed that Cu^{2+} could be adsorbed on cell surface and cytoplasm. Nevertheless, the influence of the different components of the cell on the biosorption was not clear. Consequently, the biosorption experiments of different components of cells were carried out. It was observed that the maximum Cu^{2+} removal efficiency of cell wall, cell membrane and cytoplasm were 21.2%, 20.7% and 18.5% at 45 min, indicated that the maximum removal efficiency of cell wall was found to be slightly higher as compared to cell membrane and cytoplasm, respectively. Furthermore, it was found that the equilibrium time of the cell and cytoplasm was 30 min, while the equilibrium time of the cell wall and cell membrane was 15 min, which further proved that the cell wall and cell membrane on the cell surface have a significant impact on the initial stage of biosorption (Fig. 3). The reason was supposed that cell disruption increased the contact area with Cu^{2+} , cytoplasm, cell wall and cell membrane were not arranged neatly, so that the cytoplasm was exposed to the outside. E. Lopez Errasquin also believed that degraded cells, due to the destruction of cell membranes, provide more surface binding sites, greater available surface areas and expose intracellular components.⁴⁹ However, the maximum removal rate of Cu^{2+} by cell wall, cell membrane and cytoplasm was lower than that of cell. The reason for this phenomenon may be that the cell components

were prepared through a series of physical or chemical processes, such as ultrasound, Pre-treatment agent or SDS buffer, so that the cell components may lose some of the biosorption sites.

3.4 Effect of pH on biosorption of Cu^{2+} by *P. pastoris* cell components

The effect of pH on removal of Cu^{2+} by each part of cells was shown in Fig. 4. Under different pH, the removal efficiency of Cu^{2+} by cells was relatively higher than that of other cell components. However, the effects of pH on cells and cell components were roughly the same. At lower pH, the amount of biosorption to Cu^{2+} was small. Biosorption to Cu^{2+} increased with the increased of pH from 2.0 to 6.0. The highest removal efficiency was observed in the pH 6.0. These observations can be explained by the fact that at lower pH values, the surface charge of the biomass is positive, which is not favorable to cations biosorption. Meanwhile, hydrogen ions compete strongly with Cu^{2+} at the active sites, resulting in less biosorption. With increasing pH, electrostatic repulsions between cations and surface sites and the competing effect of hydrogen ions decrease. Consequently, the metal biosorption increases.⁵⁰ When pH was higher than 6, the OH^- competed with functional groups on the biosorbents to combine with Cu^{2+} . It precipitates, and it will no longer be able to bond to the functional groups present in or on the biosorbents, leading to a decrease in its biosorption capacity as observed in this study.⁵¹

3.5 Biosorption isotherms of *P. pastoris* cell components

The study of biosorption isotherm is of great significance in wastewater treatment as it provides valuable information for the pathway of biosorption reaction. In order to further explore the biosorption capacity of each cell component, the biosorption isotherm experiment was carried out. Fig. 5 showed the biosorption isotherms for intact cells, cell wall, membrane and

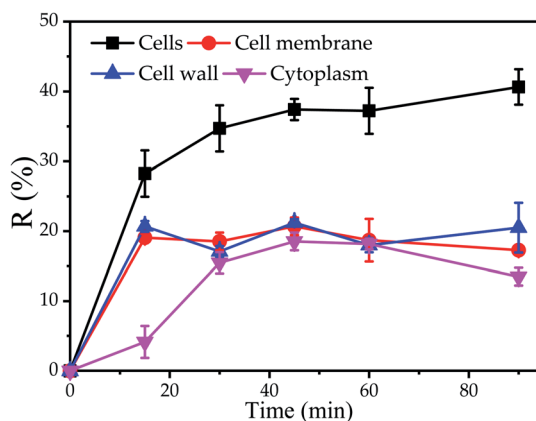


Fig. 3 Biosorption of Cu^{2+} by *P. pastoris* cell components, initial Cu^{2+} concentration: 100 mg L^{-1} , wet biomass dosage: 0.5 g, R – removal efficiency.

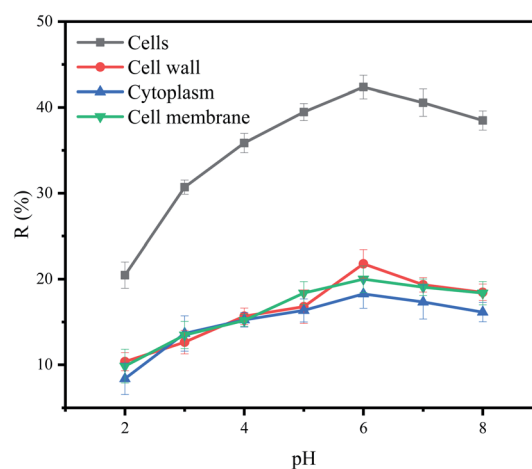


Fig. 4 Effect of pH on removal of Cu^{2+} by *P. pastoris* cell components, initial Cu^{2+} concentration: 100 mg L^{-1} , wet biomass dosage: 0.5 g, R – removal efficiency.



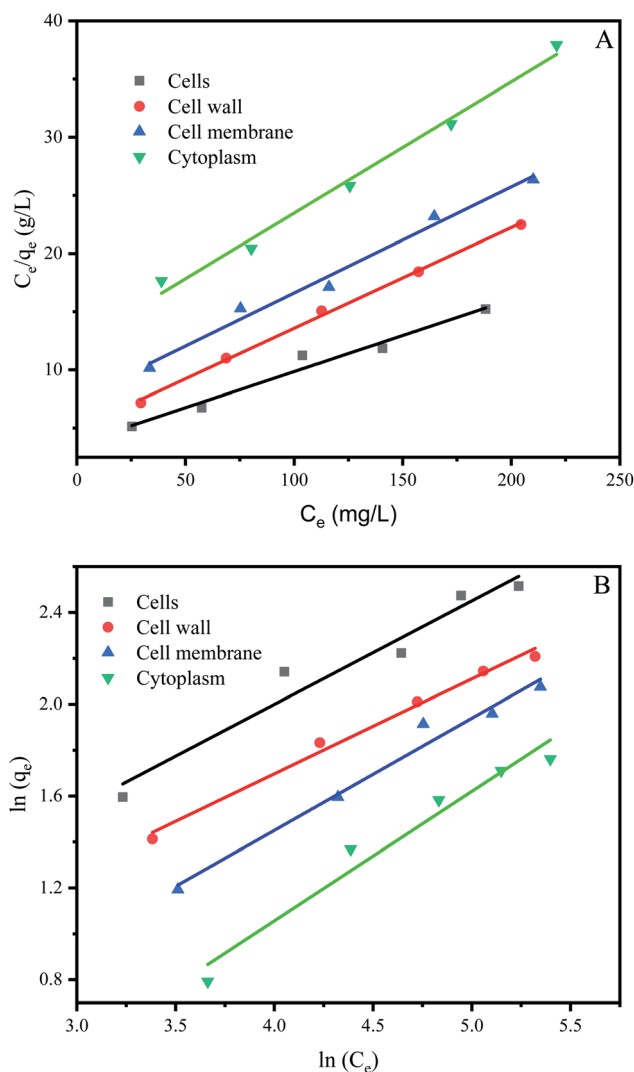


Fig. 5 Biosorption isotherms of *P. pastoris* cell components. (A) Langmuir biosorption isotherm of Cu^{2+} with *P. pastoris* cell components, (B) Freundlich biosorption isotherm of Cu^{2+} with *P. pastoris* cell components, initial Cu^{2+} concentration: 100 mg L^{-1} , wet biomass dosage: 0.5 g , R – removal efficiency.

cytoplasm. All the curve-fitting parameters were summarized in Table 1. A high correlation coefficient ($R^2 = 0.99$) indicated that Langmuir model could better fit the biosorption process of cells, cell membranes, cells and cytoplasm, while both models could fit the biosorption process of cell wall. Other studies on biosorption by yeast also observed that the biosorption equilibrium isotherm was set to the model described by Langmuir.^{51,52}

The affinity constant b obtained from the Langmuir model was 0.017 for cells, 0.018 for cell wall, 0.012 for cell membrane and 0.009 for cytoplasm, respectively (Table 1). Thus, cell wall has a greater affinity for Cu^{2+} than other cell components. This may be because the cell wall has more biosorption sites than the cell membrane and cytoplasm, in addition the cell wall has more contact area than the cell. The maximum biosorption capacity (q_m) of *Pichia pastoris* biomass was 16.13 mg g^{-1} . The biosorption capacity of various other yeasts was measured as 2.59 to 76.8 mg g^{-1} for Cu^{2+} (Table 2). Table 2 indicated that

Table 1 Isotherm model parameters for biosorption of Cu^{2+} on *P. pastoris* cell components using Langmuir and Freundlich isotherms

Parameters	Langmuir model			Freundlich model		
	$q_m (\text{mg g}^{-1})$	$b (\text{L mg}^{-1})$	R^2	$K_f (\text{L g}^{-1})$	n	R^2
Cells	16.13	0.017	0.97	1.22	2.22	0.95
Cell wall	11.53	0.018	0.99	1.04	2.42	0.99
Cell membrane	10.97	0.012	0.99	0.61	2.05	0.98
Cytoplasm	8.87	0.009	0.99	0.30	1.77	0.96

most yeast with strong biosorption capacity have been modified. Among the unmodified yeast, *Pichia pastoris* used in this study has higher biosorption capacity, indicated that *Pichia pastoris* is a promising biosorbent. These results agreed with previous reports that the difference species and moisture content of yeast or initial concentration of Cu^{2+} significant affected the removal efficiency.⁵³ However, the maximum adsorption capacity of cell, cell wall, cell membrane and cytoplasm were 16.13, 11.53, 10.97 and 8.87 mg g^{-1} , respectively. Different from the rank of affinity constant b , the cell had the highest biosorption capacity, followed by cell wall, cell membrane and cytoplasm. This result was consistent with the result in Section 3.3, which showed that although the contact area between cells and Cu^{2+} was smaller than that of other cell components, the biosorption capacity of cells was still larger than that of other cell components due to the long contact time. This result also proved that the biosorption capacity of cell wall was higher than that of cell membrane and cytoplasm, which indicated that cell wall played an important role in the biosorption process.

3.6 Effect of cell wall components of *Pichia pastoris* on biosorption of Cu^{2+}

Above results indicated that the cell wall played an important role in the initial stage of biosorption, which may be related to some material in the cell wall. The dried cell wall of yeast approximately contains 13% protein and 59.8% polysaccharide, which consist of about 31% mannan, 28.8% glucan and 8.1% of lipids.^{36,64} In order to investigate the effect of protein and mannan in the cell wall on biosorption, the *P. pastoris* biomass was treated 6 hours with β -mannanase (50 U) and protease K (30 U), respectively, and then used for biosorption experiments.

Results indicated that after treatment with protease K, the biosorption process had been greatly influenced. When the biosorption was finished, the maximum removal efficiencies of untreated *P. pastoris* biomass and treated with proteinase K were 43.2% and 18.1% respectively (Fig. 6A). The removal efficiency of *P. pastoris* biomass treated with proteinase K was only 41.9% of that in untreated *P. pastoris*. After treatment with β -mannanase, the removal efficiency was 28.2%, accounting for 65.3% of the removal efficiency of untreated *P. pastoris* biomass.

It proved that protein and mannan on the cell wall played an indispensable role in the biosorption of Cu^{2+} . Previous studies also demonstrated that Cu^{2+} could be adsorbed by protein which



Table 2 Biosorption parameter of Cu²⁺ on other yeasts biomass

Biosorbent	q_m (mg g ⁻¹)	pH	Equilibrium time	Ref.
Magnetically modified brewer's yeast	76.8	5–7	60 min	54
EDTAD-modified baker's yeast	65.0	6.0	20 min	55
baker's yeast treated with NaOH	27.79	5	20 min	56
Beer yeast	20.6	6	60 min	57
<i>S. cerevisiae</i> treated with NaOH	20	4.6	4 h	58
Heat pretreated baker's yeast	19.53	4.5	30 min	59
<i>Pichia stipitis</i>	16.94	4.5	10 d	60
<i>P. pastoris</i>	16.13	6.0	20 min	Present study
Baker's yeast treated with ethanol	15.64	5	20 min	56
<i>S. cerevisiae</i>	15.1	3.0	2 h	61
Baker's yeast	11.53	5	20 min	56
<i>S. cerevisiae</i> treated with ethanol	9.82	4.6	4 h	58
<i>S. cerevisiae</i> -Fe ₃ O ₄	8.30	5.5	10 min × 9	62
<i>S. cerevisiae</i>	4.73	5	24 h	53
<i>S. cerevisiae</i>	4.70	5.5	10 min × 9	62
Baker's yeast	4.5	6.0	20 min	55
<i>S. cerevisiae</i>	2.59	3	1440 min	63

directly influenced photosynthesis⁶⁵ and caused oxidative damage of cells.⁶⁶ Misbah Saleem explored the utilization of biomass from coconut copra meal (CCM) as a biosorbent to remove Ni²⁺ from aqueous solutions, and the main component of coconut powder is mannan, which also proved that mannan could adsorb heavy metal ions.⁶⁷ Xuegang Luo also proved that the modified mannan had excellent biosorption properties for Cu²⁺.⁶⁸

Not only that, the removal efficiency of Cu²⁺ by *Pichia pastoris* treated with protease K decreased more, which indicated that the protein had a greater impact on the entire biosorption process, even if the cell wall contained more mannans. The protein then appears to be the “glue” which affixes the two wall layers together, which results in the mannan-protein layer that may be almost completely removed by protease K, while the mannanase only removed some of the mannan and not the entire layer.⁶⁴

As mentioned above, the polysaccharide content in the cell wall was 59.8%, of which glucan was as high as 31%. To further determine the relative biosorption capacity of glucan and mannan in the cell wall for Cu²⁺. Biosorption of Cu²⁺ by glucan and mannan was carried out.

Results indicated that the removal efficiency of mannan and *P. pastoris* reached 34% and 43.3%, respectively (Fig. 6B). However, the removal efficiency of Cu²⁺ by mannan was better than that of *P. pastoris* in the early stage of biosorption which may due to that the mannan on the cell wall is a three-dimensional arrangement. Therefore, at the first stage of biosorption, the groups on the cell surface combined with Cu²⁺. After that, the groups inside the cell functioned with the penetration of Cu²⁺. In contrast, all groups of the broken cell wall were exposed outside, Cu²⁺ can be fully contacted with the cell walls. The fact further indicated that mannan on *P. pastoris* surface also played an indispensable role in biosorption of Cu²⁺.

However, it was noticed that the role of glucan in the biosorption process was mediocre. When the reaction was carried out to 10 min, the removal efficiency of glucan was 12% and desorption occurred subsequently, agreed with prior results that desorption occurs in some environments (Fig. 6B).⁶⁹

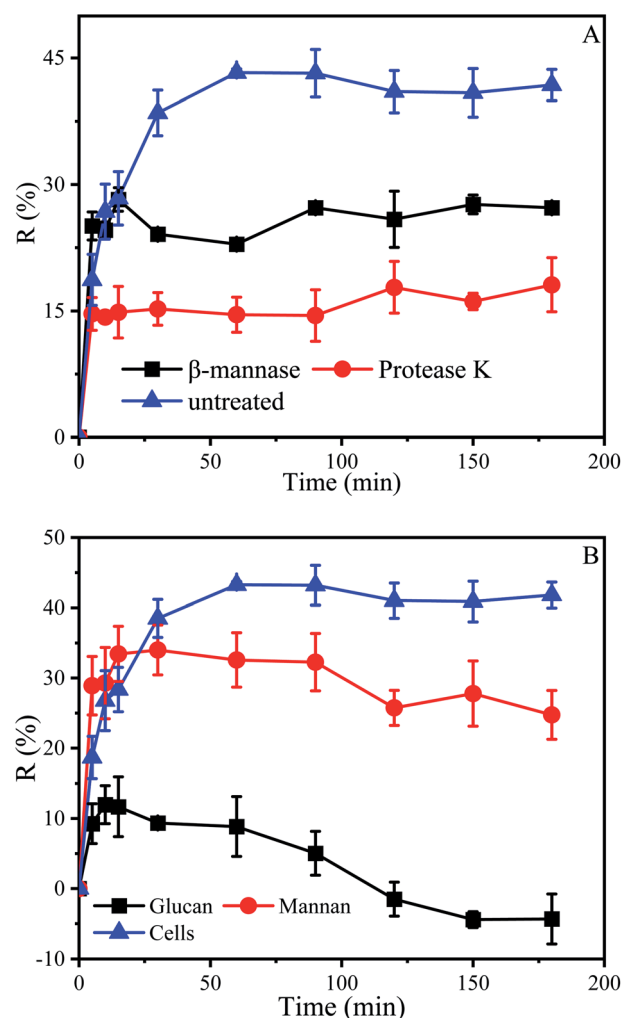


Fig. 6 Effect of cell wall components of *Pichia pastoris* on biosorption of Cu²⁺. (A) Biosorption of Cu²⁺ by *P. pastoris* treated with enzymes, (B) biosorption of Cu²⁺ by glucan and mannan, initial Cu²⁺ concentration: 100 mg L⁻¹, wet biomass dosage: 0.5 g, R – removal efficiency.



Removal efficiency is related to the properties of biosorbents, such as the amount of functional groups, molecular structure, swelling degree, surface area, and particle size.⁷⁰ For instance, the basic unit of chitosan is glucosamine, which is similar to glucan in structure. However, in contrast to glucan, many reports have shown that chitosan has highly efficient adsorption of heavy metal ions.⁷¹ In addition, chitosan could be obtained by *N*-deacetylation of chitin contained in yeast cell wall. This is also the reason why many studies tend to modify yeast to improve its adsorption capacity (Table 2). Mannose and glucose are isomers, which results in mannan having some functions that glucan does not have. Therefore, mannan and glucan had essential differences in structure, which was the reason for their different removal efficiency.⁷² These results also suggested that we could modify the mannan in the cell wall to improve the adsorption capacity of yeast, and the modification of yeast was inseparable from the exploration of adsorption sites.

3.7 FTIR analysis

The FTIR spectrum could effectively identify functional groups that may bind to Cu^{2+} .^{73–75} Each functional group has a specific absorption peak. When Cu^{2+} interacts with functional groups, the adsorption peaks of functional groups shift to higher or lower wave numbers.

Fig. 7A described the FTIR spectrum of *P. pastoris* biomass. The broad and intensely stretched peak at 3390 cm^{-1} due to the presence of hydroxyl (O–H) stretching in hydrogen bonds. After the biosorption process, the peak shifted to 3416 cm^{-1} , indicated that the hydroxyl from polysaccharides, fatty acids and protein were involved the biosorption of Cu^{2+} .^{76,77} The peaks at 2926 cm^{-1} and 1404 cm^{-1} represented the asymmetric stretching vibration of $-\text{CH}_2$ and bending vibration of $-\text{CH}_3$ and $-\text{CH}_2$. After the biosorption process, the peaks became more blunter, indicated that $-\text{CH}_2$ and CH_3 were the functional group to combine with Cu^{2+} . Slawomir Wierzbza also found that $-\text{CH}_2$ and CH_3 were the biosorption group in the yeast cell.⁷⁸ The peaks observed at 1650 cm^{-1} was attributed to amide I band (stretching vibration of $\text{C}=\text{O}$), 1540 cm^{-1} indicated the amide II band (stretching vibration of $\text{C}-\text{N}$ and bending vibration of $\text{N}-\text{H}$), 1248 cm^{-1} was attributed to amide III band (stretching vibration of $\text{C}-\text{N}$ and bending vibration of $\text{N}-\text{H}$).^{79,80} After biosorption, the wavenumber shifted to 1642 cm^{-1} , 1546 cm^{-1} , 1244 cm^{-1} respectively, further proved that protein played an important role in the biosorption of Cu^{2+} . The peak at 1074 cm^{-1} corresponded to stretching vibrations of $\text{C}-\text{O}$ which from polysaccharides, after biosorption, the wavenumber shifted to 1076 cm^{-1} .⁸¹

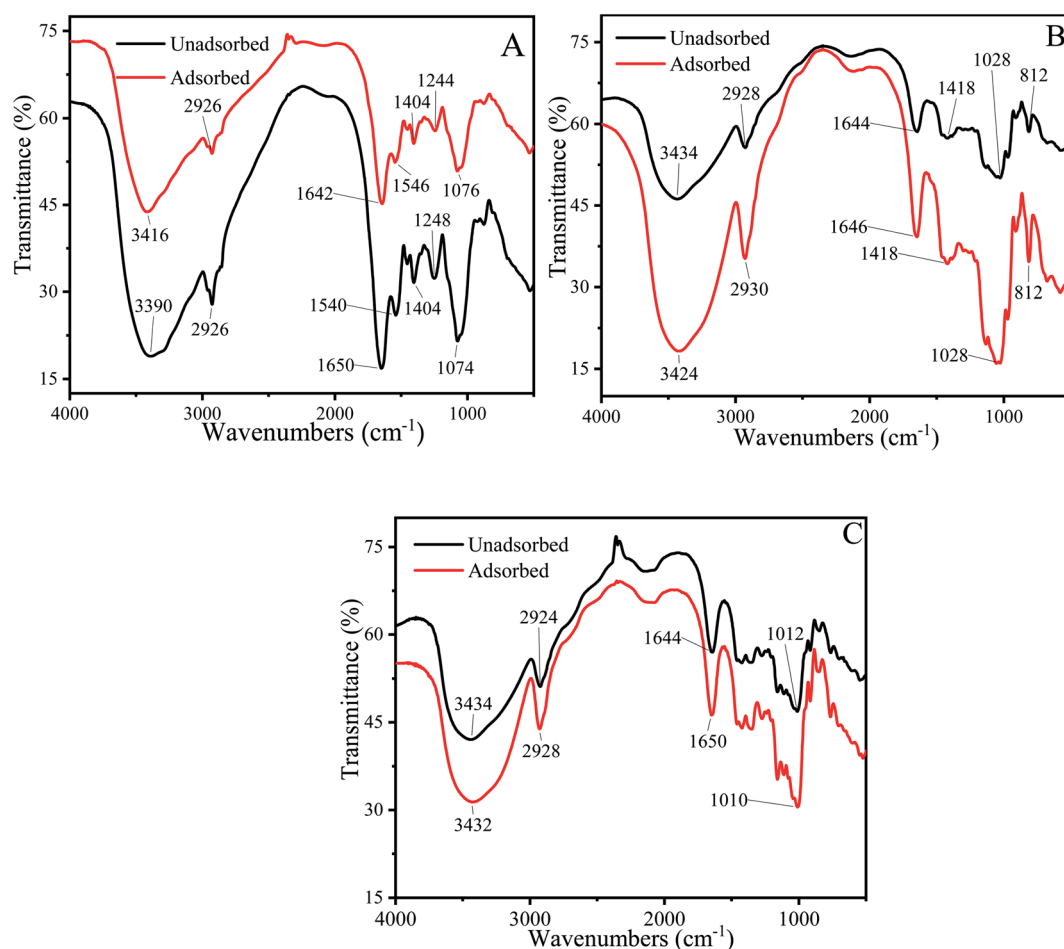


Fig. 7 FTIR spectrum of unadsorbed and Cu^{2+} -adsorbed biosorbents. (A) Cu^{2+} -Adsorbed and unadsorbed *P. pastoris* biomass. (B) Cu^{2+} -Adsorbed and unadsorbed mannan. (C) Cu^{2+} -Adsorbed and unadsorbed glucan.



Above results demonstrated that $-\text{OH}$, $\text{C}-\text{N}$, $\text{N}-\text{H}$, $\text{C}=\text{O}$, $\text{C}-\text{O}$ and $-\text{CH}$ from protein, polysaccharides and fatty acids were involved in Cu^{2+} biosorption process. It was consistent with previous biosorption experiment results. Haojie Huang also found that amide II, $-\text{CH}_2$, hydroxyl group, acetylamino group and amide I on the surface of *E. coli* cells participated in the biosorption of heavy metals.⁸²

To confirm the key groups of mannan and glucan to adsorb Cu^{2+} , FTIR spectrum of mannan and glucan before and after biosorption were determined. Fig. 7B and C showed the groups whose peak shape changed significantly after mannan and glucan adsorbed Cu^{2+} . Similar biosorption groups were present in glucan and mannan. However, the number of biosorption groups of mannan and glucan was significantly reduced.

Broad peak present in the 3434 cm^{-1} was attributed to the hydroxyl ($-\text{OH}$) stretching vibrations. After biosorption process, the wavenumber shifted to 3424 cm^{-1} and 3432 cm^{-1} respectively (Fig. 7B and C). It was suggested that the hydroxyl groups on mannan and glucan could also adsorb Cu^{2+} , which was consistent with the previous experimental results. The peaks at 2928 cm^{-1} and 2924 cm^{-1} shifted to 2930 cm^{-1} and 2928 cm^{-1} respectively (Fig. 7B and C). The peak at 1418 cm^{-1} was also the $-\text{CH}$ stretching vibration (Fig. 7B). After the biosorption process, the peak became acuter. It could be speculated that $-\text{CH}_2$ and $-\text{CH}_3$ in mannan and glucan were involved in biosorption of Cu^{2+} . Similar results were found in other adsorbents.^{77,83} The peak observed at 1644 cm^{-1} was attributed to the $\text{C}=\text{O}$ stretching vibration of mannan and glucan.⁸⁴ After biosorption process, the wavenumbers shifted to 1646 cm^{-1} and 1650 cm^{-1} respectively, demonstrated that $\text{C}=\text{O}$ in mannan and glucan could also bind to Cu^{2+} . The peak at 1028 cm^{-1} was attributed to stretching vibration of $\text{C}-\text{O}$ (Fig. 7B). After biosorption, the peak did not change, indicated that $-\text{C}-\text{O}$ was not the main group to combine with Cu^{2+} in mannan (Fig. 7B). However, the peak at 1012 cm^{-1} shifted to 1010 cm^{-1} demonstrated that the $\text{C}-\text{O}$ bond in glucan could bind to Cu^{2+} (Fig. 7C). The peak at 812 cm^{-1} was attributed to pyranose ring vibration band of mannose, which did not involve the biosorption of Cu^{2+} (Fig. 7B).⁸⁵

Above results showed that the polysaccharides on the cell wall also had a certain number of biosorption groups, which further proved the importance of polysaccharides for biosorption (Fig. 7B and C). However, when there were only polysaccharides and no protein in the cell wall, the number of biosorption sites also decreased, which also indicated the importance of protein in biosorption (Fig. 7).

The FTIR spectra indicated that $\text{C}=\text{O}$, $\text{N}-\text{H}$, $-\text{OH}$, $\text{C}-\text{O}$, $\text{C}-\text{N}$, and $-\text{CH}$ were responsible for biosorption of Cu^{2+} . Although the biosorption sites of mannan and glucan decreased, their biosorption sites were consistent with *P. pastoris* biomass. Many studies have reported that these similar groups can adsorb heavy metal ions.^{22,86–88}

4. Conclusions

Microorganism biomass could be used as an environmentally friendly and inexpensive adsorbent for heavy metal biosorption from wastewater. However, the effect of each

component of microbial cells on the adsorption of heavy metals is still unclear. In present study, *P. pastoris* and Cu^{2+} were used as models to explore the adsorption capacity of different microbial cell components for heavy metal ions. The biosorption experiment with $100\text{ mg L}^{-1}\text{ Cu}^{2+}$ showed that the biosorption equilibrium time of *P. pastoris* biomass was 15 min, the maximum removal efficiency was 41.1%, and the adsorption capacity was 6.2 mg g^{-1} . TEM-EDX showed that the associated Cu^{2+} heterogeneously present on the surface and inside of *P. pastoris* cells. These results indicated that the biosorption of Cu^{2+} on the cell surface played an important role in the initial biosorption process. The biosorption experiment with $100\text{ mg L}^{-1}\text{ Cu}^{2+}$ indicated that the maximum Cu^{2+} removal efficiency of cell wall, cell membrane and cytoplasm were 21.2%, 20.7% and 18.5%, respectively. The optimum pH of Cu^{2+} biosorption on *P. pastoris* cell, cell wall, cell membrane and cytoplasm was 6. Moreover, the maximum adsorption capacity of cell, cell wall, cell membrane and cytoplasm were 16.13, 11.53, 10.97 and 8.87 mg g^{-1} , respectively. Compared with other cell components, the adsorption capacity of cell wall was slightly higher, further indicated that cell wall played an important role in the initial biosorption process. In addition, the maximum removal efficiencies of *P. pastoris* biomass treated with proteinase K and *P. pastoris* biomass treated with β -mannanase were 18.1% and 28.2%, respectively, indicated that biosorption of Cu^{2+} was strongly influenced by the protein and mannan on the cell wall. The maximum removal efficiencies of mannan and glucan were 34% and 12%, respectively, indicated that the biosorption of mannan was stronger than that of glucan. Furthermore, the groups that could adsorb Cu^{2+} in *P. pastoris* biomass, mannan and glucan were hydroxyl ($\text{O}-\text{H}$), carbon oxygen bond ($\text{C}-\text{O}$), $-\text{CH}$, $\text{C}-\text{N}$ and carbonyl group ($\text{C}=\text{O}$). In addition, amino group ($\text{N}-\text{H}$), and $\text{C}-\text{N}$ in *P. pastoris* biomass could also adsorb Cu^{2+} . Finally, this study proved that in the process of heavy metal biosorption, the cell wall has the strongest adsorption capacity, and the protein and mannan in the cell wall have great influence on the biosorption due to its rich adsorption sites. The results will provide a theoretical basis for further understanding of the mechanism of biosorption of heavy metal ions. In addition, the results can provide guidance for the modification of yeast to improve its adsorption capacity.

Conflicts of interest

There are no conflicts to declare.

Acknowledgements

This work was supported by the Natural Science Foundation of Hunan Province, China (grant number 2015JJ5006); the National Natural Science Foundation of China (grant number 31870115) and the Hunan Flag Bio-Tech Co., Ltd, China.

References

- 1 S. J. Yu, X. X. Wang, W. Yao, *et al.*, Macroscopic, Spectroscopic, and Theoretical Investigation for the



- Interaction of Phenol and Naphthol on Reduced Graphene Oxide, *Environ. Sci. Technol.*, 2017, **51**(6), 3278–3286.
- 2 I. Ali, New Generation Adsorbents for Water Treatment, *Chem. Rev.*, 2012, **112**(10), 5073–5091.
 - 3 P. A. Xu, G. M. Zeng, D. L. Huang, *et al.*, Adsorption of Pb (II) by iron oxide nanoparticles immobilized *Phanerochaete chrysosporium*: Equilibrium, kinetic, thermodynamic and mechanisms analysis, *Chem. Eng. J.*, 2012, **203**, 423–431.
 - 4 T. Y. Jiang, J. Jiang, R. K. Xu, *et al.*, Adsorption of Pb (II) on variable charge soils amended with rice-straw derived biochar, *Chemosphere*, 2012, **89**(3), 249–256.
 - 5 Y. Cheng, C. P. Yang, H. J. He, *et al.*, Biosorption of Pb (II) Ions from Aqueous Solutions by Waste Biomass from Biotrickling Filters: Kinetics, Isotherms, and Thermodynamics, *J. Environ. Eng.*, 2016, **142**(9), 7.
 - 6 D. John Babu, P. King and Y. Prasanna Kumar, Optimization of Cu (II) biosorption onto sea urchin test using response surface methodology and artificial neural networks, *Int. J. Environ. Sci. Technol.*, 2019, **16**(4), 1885–1896.
 - 7 M. Banerjee, R. K. Basu and S. K. Das, Cu (II) removal using green adsorbents: kinetic modeling and plant scale-up design, *Environ. Sci. Pollut. Res.*, 2019, **26**(12), 11542–11557.
 - 8 A. Dąbrowski, Z. Hubicki, P. Podkościelny, *et al.*, Selective removal of the heavy metal ions from waters and industrial wastewaters by ion-exchange method, *Chemosphere*, 2004, **56**(2), 91–106.
 - 9 L. Chareerntanyarak, Heavy metals removal by chemical coagulation and precipitation, *Water Sci. Technol.*, 1999, **39**(10), 135–138.
 - 10 H. A. Qdais and H. Moussa, Removal of heavy metals from wastewater by membrane processes: a comparative study, *Desalination*, 2004, **164**(2), 105–110.
 - 11 C. Yuan and C. H. Weng, Electrokinetic enhancement removal of heavy metals from industrial wastewater sludge, *Chemosphere*, 2006, **65**(1), 88–96.
 - 12 S. M. Lee, C. Laldawngliana and D. Tiwari, Iron oxide nanoparticles-immobilized-sand material in the treatment of Cu(II), Cd(II) and Pb(II) contaminated waste waters, *Chem. Eng. J.*, 2012, **195**, 103–111.
 - 13 B. Volesky, Detoxification of metal-bearing effluents: biosorption for the next century, *Hydrometallurgy*, 2001, **59**(2), 203–216.
 - 14 I. S. Bădescu, D. Bulgariu, I. Ahmad, *et al.*, Valorisation possibilities of exhausted biosorbents loaded with metal ions – A review, *J. Environ. Manage.*, 2018, **224**, 288–297.
 - 15 J. Wang and C. Chen, Biosorption of heavy metals by *Saccharomyces cerevisiae*: a review, *Biotechnol. Adv.*, 2006, **24**(5), 427–451.
 - 16 R. Manohari and K. N. Yogalakshmi, Optimization of Copper (II) Removal by Response Surface Methodology Using Root Nodule Endophytic Bacteria Isolated from *Vigna unguiculata*, *Water, Air, Soil Pollut.*, 2016, **227**(8), 13.
 - 17 A. Verma, S. A. Shalu, *et al.*, Biosorption of Cu (II) using free and immobilized biomass of *Penicillium citrinum*, *Ecol. Eng.*, 2013, **61**, 486–490.
 - 18 M. R. Awual, New type mesoporous conjugate material for selective optical copper (II) ions monitoring & removal from polluted waters, *Chem. Eng. J.*, 2017, **307**, 85–94.
 - 19 M. R. Awual, M. M. Hasan, M. A. Khaleque, *et al.*, Treatment of copper (II) containing wastewater by a newly developed ligand based facial conjugate materials, *Chem. Eng. J.*, 2016, **288**, 368–376.
 - 20 A. A. Al-Homaidan, H. J. Al-Houri, A. A. Al-Hazzani, *et al.*, Biosorption of copper ions from aqueous solutions by *Spirulina platensis* biomass, *Arabian J. Chem.*, 2014, **7**(1), 57–62.
 - 21 J. Cotruvo, WHO Guidelines for Drinking Water Quality: First Addendum to the Fourth Edition, *J.-Am. Water Works Assoc.*, 2017, **109**, 44–51.
 - 22 X. J. Hu, H. D. Gu, T. T. Zang, *et al.*, Biosorption mechanism of Cu²⁺ by innovative immobilized spent substrate of fragrant mushroom biomass, *Ecol. Eng.*, 2014, **73**, 509–513.
 - 23 W. S. Wan Ngah and M. A. K. M. Hanafiah, Biosorption of copper ions from dilute aqueous solutions on base treated rubber (*Hevea brasiliensis*) leaves powder: kinetics, isotherm, and biosorption mechanisms, *J. Environ. Sci.*, 2008, **20**(10), 1168–1176.
 - 24 L. X. Huang, M. L. Li, G. C. Si, *et al.*, Assessment of microbial products in the biosorption process of Cu(II) onto aerobic granular sludge: Extracellular polymeric substances contribution and soluble microbial products release, *J. Colloid Interface Sci.*, 2018, **527**, 87–94.
 - 25 D. J. Babu, P. King and Y. P. Kumar, Optimization of Cu (II) biosorption onto sea urchin test using response surface methodology and artificial neural networks, *Int. J. Environ. Sci. Technol.*, 2019, **16**(4), 1885–1896.
 - 26 J. Wang and C. Chen, Biosorbents for heavy metals removal and their future, *Biotechnol. Adv.*, 2009, **27**(2), 195–226.
 - 27 J. Vina-Gonzalez, K. Elbl, X. Ponte, *et al.*, Functional expression of aryl-alcohol oxidase in *Saccharomyces cerevisiae* and *Pichia pastoris* by directed evolution, *Biotechnol. Bioeng.*, 2018, **115**(7), 1666–1674.
 - 28 R. Mata, S. Ratinho and D. Figueiro, Assessment of the Environmental Impact of Yeast Waste Application to Soil: An Integrated Approach, *Waste Biomass Valorization*, 2019, **10**(6), 1767–1777.
 - 29 E. Savastru, C. Zamfir and M. Diaconu, *et al.*, Biosorption of Cu (II) Ions from Aqueous Solution on *Saccharomyces Cerevisiae* Biomass: Isotherm and Kinetics Modelling, *Proceedings of the 2019 E-Health and Bioengineering Conference (EHB)*, Iași, Romania, F 21-23 Nov. 2019, 2019 [C]. EHB 2019 Grigore T, Popa University of Medicine and Pharmacy, Iași, Romania, 2019.
 - 30 E. V. Soares and H. Soares, Bioremediation of industrial effluents containing heavy metals using brewing cells of *Saccharomyces cerevisiae* as a green technology: a review, *Environ. Sci. Pollut. Res.*, 2012, **19**(4), 1066–1083.
 - 31 C. Chen, D. Wen and J. Wang, Cellular surface characteristics of *Saccharomyces cerevisiae* before and after Ag(I) biosorption, *Bioresour. Technol.*, 2014, **156**, 380–383.
 - 32 M. Fadel, N. M. Hassanein, M. M. Elshafei, *et al.*, Biosorption of manganese from groundwater by biomass of *Saccharomyces cerevisiae*, *HBRC J.*, 2017, **13**(1), 106–113.
 - 33 Y. S. Zhang, W. G. Liu, L. Zhang, *et al.*, Application of bifunctional *Saccharomyces cerevisiae* to remove lead(II)



- and cadmium(II) in aqueous solution, *Appl. Surf. Sci.*, 2011, **257**(23), 9809–9816.
- 34 F. Elahian, R. Heidari, V. R. Charghan, *et al.*, Genetically modified *Pichia pastoris*, a powerful resistant factory for gold and palladium bioleaching and nanostructure heavy metal biosynthesis, *Artif Cell Nanomed B*, 2020, **48**(1), 259–265.
 - 35 T. P. Lyons and J. S. Hough, Glycoproteins from yeast cell wall, *Biochem. J.*, 1970, **117**(2), 44P.
 - 36 P. A. Roelofsen, Yeast mannan, a cell wall constituent of baker's yeast, *Biochim. Biophys. Acta*, 1953, **10**(3), 477–478.
 - 37 P. Valachovic, A. Pechova and T. J. Mason, Towards the industrial production of medicinal tincture by ultrasound assisted extraction, *Ultrason. Sonochem.*, 2001, **8**(2), 111–117.
 - 38 Z. Hromadkova, A. Ebringerova and P. Valachovic, Comparison of classical and ultrasound-assisted extraction of polysaccharides from *Salvia officinalis* L, *Ultrason. Sonochem.*, 1999, **5**(4), 163–168.
 - 39 I. Langmuir, The constitution and fundamental properties of solids and liquids, *J. Am. Chem. Soc.*, 1916, **38**(11), 2221–2295.
 - 40 H. Freundlich, Over the adsorption in solution, *J. Phys. Chem.*, 1906, **57**, 385–471.
 - 41 A. Dutta, L. P. Zhou, C. O. Castillo-Araiza, *et al.*, Cadmium (II), Lead (II), and Copper (II) Biosorption on Baker's Yeast (*Saccharomyces cerevisiae*), *J. Environ. Eng.*, 2016, **142**(9), 7.
 - 42 R. Razmovski and M. Šćiban, Biosorption of Cr(VI) and Cu(II) by waste tea fungal biomass, *Ecol. Eng.*, 2008, **34**(2), 179–186.
 - 43 M. Fomina and G. M. Gadd, Biosorption: current perspectives on concept, definition and application, *Bioresour. Technol.*, 2014, **160**, 3–14.
 - 44 A. P. Stafussa, G. M. Maciel, A. G. da Silva Anthero, *et al.*, Biosorption of anthocyanins from grape pomace extracts by waste yeast: kinetic and isotherm studies, *J. Food Eng.*, 2016, **169**, 53–60.
 - 45 I. Letnik, R. Avrahami, R. Port, *et al.*, Biosorption of copper from aqueous environments by *Micrococcus luteus* in cell suspension and when encapsulated, *Int. Biodeterior. Biodegrad.*, 2017, **116**, 64–72.
 - 46 T. S. Wang, X. Y. Zheng, X. Y. Wang, *et al.*, Different biosorption mechanisms of Uranium(VI) by live and heat-killed *Saccharomyces cerevisiae* under environmentally relevant conditions, *J. Environ. Radioact.*, 2017, **167**, 92–99.
 - 47 L. Z. Huang, G. M. Zeng, D. L. Huang, *et al.*, Biosorption of cadmium (II) from aqueous solution onto *Hydrilla verticillata*, *Environ. Earth Sci.*, 2010, **60**(8), 1683–1691.
 - 48 A. Mishra, B. D. Tripathi and A. K. Rai, Biosorption of Cr(VI) and Ni(II) onto *Hydrilla verticillata* dried biomass, *Ecol. Eng.*, 2014, **73**, 713–723.
 - 49 E. Lopez Errasquin and C. Vazquez, Tolerance and uptake of heavy metals by *Trichoderma atroviride* isolated from sludge, *Chemosphere*, 2003, **50**(1), 137–143.
 - 50 Z. Aksu, Equilibrium and kinetic modelling of cadmium(II) biosorption by *C. vulgaris* in a batch system: effect of temperature, *Sep. Purif. Technol.*, 2001, **21**(3), 285–294.
 - 51 Q. Peng, Y. Liu, G. Zeng, *et al.*, Biosorption of copper(II) by immobilizing *Saccharomyces cerevisiae* on the surface of chitosan-coated magnetic nanoparticles from aqueous solution, *J. Hazard. Mater.*, 2010, **177**(1), 676–682.
 - 52 S. Amirnia, M. B. Ray and A. Margaritis, Heavy metals removal from aqueous solutions using *Saccharomyces cerevisiae* in a novel continuous bioreactor-biosorption system, *Chem. Eng. J.*, 2015, **264**, 863–872.
 - 53 J. M. do Nascimento, J. D. de Oliveira, A. C. L. Rizzo, *et al.*, Biosorption Cu (II) by the yeast *Saccharomyces cerevisiae*, *Biotechnol. Rep.*, 2019, **21**, e00315.
 - 54 L. Uzun, N. Sağlam, M. Safarikova, *et al.*, Copper Biosorption on Magnetically Modified Yeast Cells Under Magnetic Field, *Sep. Sci. Technol.*, 2011, **46**(6), 1045–1051.
 - 55 J. Yu, M. Tong, X. Sun, *et al.*, Enhanced and selective adsorption of Pb²⁺ and Cu²⁺ by EDTAD-modified biomass of baker's yeast, *Bioresour. Technol.*, 2008, **99**(7), 2588–2593.
 - 56 Y. Zhang, W. Liu, M. Xu, *et al.*, Study of the mechanisms of Cu²⁺ biosorption by ethanol/caustic-pretreated baker's yeast biomass, *J. Hazard. Mater.*, 2010, **178**(1–3), 1085–1093.
 - 57 A. Stanila, T. Mihaiescu, C. Socaciu, *et al.*, Removal of Copper and Lead Ions from Aqueous Solution Using Brewer Yeast as Biosorbent, *Rev. Chim.*, 2016, **67**(7), 1276–1280.
 - 58 S. Tonk, B. Nagy, A. Toeroek, *et al.*, Cd(II), Zn(II) and Cu(II) Bioadsorption on Chemically Treated Waste Brewery Yeast Biomass: The Role of Functional Groups, *Acta Chim. Slov.*, 2015, **62**(3), 736–746.
 - 59 A. M. Stanescu, L. Stoica, C. Constantin, *et al.*, Physicochemical Characterization and Use of Heat Pretreated Commercial Instant Dry Baker's Yeast as a Potential Biosorbent for Cu(II) Removal, *Clean: Soil, Air, Water*, 2014, **42**(11), 1632–1641.
 - 60 P. Yilmazer and N. Saracoglu, Bioaccumulation and biosorption of copper(II) and chromium(III) from aqueous solutions by *Pichia stipitis* yeast, *J. Chem. Technol. Biotechnol.*, 2009, **84**(4), 604–610.
 - 61 I. Zinicovscaia, N. Yushin, D. Grozdov, *et al.*, Metal Removal From Complex Copper Containing Effluents by Waste Biomass of *saccharomyces cerevisiae*, *Ecol. Chem. Eng. S*, 2020, **27**(3), 415–435.
 - 62 J. C. Jos, K. B. Debs, G. Labuto, *et al.*, Synthesis, characterization, and application of yeast-based magnetic bionanocomposite for the removal of Cu(II) from water, *Chem. Eng. Commun.*, 2019, **206**(11), 1581–1591.
 - 63 C. Cojocar, M. Diaconu, I. Cretescu, *et al.*, Biosorption of copper(II) ions from aqua solutions using dried yeast biomass, *Colloids Surf., A*, 2009, **335**(1–3), 181–188.
 - 64 D. Brady, A. D. Stoll, L. Starke, *et al.*, Chemical and enzymatic extraction of heavy metal binding polymers from isolated cell walls of *Saccharomyces cerevisiae*, *Biotechnol. Bioeng.*, 1994, **44**(3), 297–302.
 - 65 X. Li, W. L. Yang, H. J. He, *et al.*, Responses of microalgae *Coelastrella* sp to stress of cupric ions in treatment of anaerobically digested swine wastewater, *Bioresour. Technol.*, 2018, **251**, 274–279.
 - 66 O. D. Iseri, D. A. Korpe, E. Yurtcu, *et al.*, Copper-induced oxidative damage, antioxidant response and genotoxicity in *Lycopersicon esculentum* Mill. and *Cucumis sativus* L, *Plant Cell Rep.*, 2011, **30**(9), 1713–1721.



- 67 M. Saleem, N. Wongsrisujarit and S. Boonyarattanakalin, Removal of nickel (II) ion by adsorption on coconut copra meal biosorbent, *Desalin. Water Treat.*, 2015, **57**(12), 5623–5635.
- 68 X. G. Luo, F. Liu, Z. F. Deng, *et al.*, Removal of copper(II) from aqueous solution in fixed-bed column by carboxylic acid functionalized deacetylated konjac glucomannan, *Carbohydr. Polym.*, 2011, **86**(2), 753–759.
- 69 X. C. Chen, Y. P. Wang, Q. Lin, *et al.*, Biosorption of copper(II) and zinc(II) from aqueous solution by *Pseudomonas putida* CZ1, *Colloids Surf., B*, 2005, **46**(2), 101–107.
- 70 F. Liu, X. G. Luo, X. Y. Lin, *et al.*, Removal of copper and lead from aqueous solution by carboxylic acid functionalized deacetylated konjac glucomannan, *J. Hazard. Mater.*, 2009, **171**(1–3), 802–808.
- 71 G. Z. Kyzas and D. N. Bikiaris, Recent Modifications of Chitosan for Adsorption Applications: A Critical and Systematic Review, *Mar. Drugs*, 2015, **13**(1), 312–337.
- 72 H. Saito, Y. Yoshioka, N. Uehara, *et al.*, Relationship between conformation and biological response for (1→3)-beta-D-glucans in the activation of coagulation factor G from limulus amebocyte lysate and host-mediated antitumor activity. Demonstration of single-helix conformation as a stimulant, *Carbohydr. Res.*, 1991, **217**, 181–190.
- 73 M. Kostic, M. Radovic, J. Mitrovic, *et al.*, Using xanthated *Lagenaria vulgaris* shell biosorbent for removal of Pb (II) ions from wastewater, *J. Iran. Chem. Soc.*, 2014, **11**(2), 565–578.
- 74 M. R. Lasheen, N. S. Ammar and H. S. Ibrahim, Adsorption/desorption of Cd(II), Cu(II) and Pb(II) using chemically modified orange peel: Equilibrium and kinetic studies, *Solid State Sci.*, 2012, **14**(2), 202–210.
- 75 M. Bansal, U. Garg, D. Singh, *et al.*, Removal of Cr(VI) from aqueous solutions using pre-consumer processing agricultural waste: A case study of rice husk, *J. Hazard. Mater.*, 2009, **162**(1), 312–320.
- 76 A. Celekli and H. Bozkurt, Bio-sorption of cadmium and nickel ions using *Spirulina platensis*: Kinetic and equilibrium studies, *Desalination*, 2011, **275**(1–3), 141–147.
- 77 M. R. Pouya and S. Behnam, Adsorption behavior of copper ions on alga *Jania adhaerens* through SEM and FTIR analyses, *Sep. Sci. Technol.*, 2017, **52**(13), 2062–2068.
- 78 S. Wierzba, Biosorption of nickel (II) and zinc (II) from aqueous solutions by the biomass of yeast *Yarrowia lipolytica*, *Pol. J. Chem. Technol.*, 2017, **19**(1), 1–10.
- 79 C. Chen, J. Hu and J. L. Wang, Uranium biosorption by immobilized active yeast cells entrapped in calcium-alginate-PVA-GO-crosslinked gel beads, *Radiochim. Acta*, 2020, **108**(4), 273–286.
- 80 Q. W. Dai, F. Q. Dong and W. Zhang, Biosorption of Lead Ions on Dried Waste Beer Yeast and the Analysis by FTIR, *Spectrosc. Spectral Anal.*, 2009, **29**(7), 1788–1792.
- 81 H. L. Peng, D. Li, J. Ye, *et al.*, Biosorption behavior of the *Ochrobactrum* MT180101 on ionic copper and chelate copper, *J. Environ. Manage.*, 2019, **235**, 224–230.
- 82 H. Huang, Q. Jia, W. Jing, *et al.*, Screening strains for microbial biosorption technology of cadmium, *Chemosphere*, 2020, **251**, 126428.
- 83 J. Yang, F. Q. Dong, Q. W. Dai, *et al.*, Biosorption of Radionuclide Uranium by *Deinococcus Radiodurans*, *Spectrosc. Spectral Anal.*, 2015, **35**(4), 1010–1014.
- 84 A. K. Lakra, L. Domdi, Y. M. Tilwani, *et al.*, Physicochemical and functional characterization of mannan exopolysaccharide from *Weissella confusa* MD1 with bioactivities, *Int. J. Biol. Macromol.*, 2020, **143**, 797–805.
- 85 A. Acemi, O. Cobanoglu and S. Turker-Kaya, FTIR-based comparative analysis of glucomannan contents in some tuberous orchids, and effects of pre-processing on glucomannan measurement, *J. Sci. Food Agric.*, 2019, **99**(7), 3681–3686.
- 86 S. Nigam, K. Gopal and P. S. Vankar, Biosorption of arsenic in drinking water by submerged plant: *Hydrilla verticillata*, *Environ. Sci. Pollut. Res.*, 2013, **20**(6), 4000–4008.
- 87 Y. Khambhaty, K. Mody, S. Basha, *et al.*, Biosorption of Cr(VI) onto marine *Aspergillus niger*: experimental studies and pseudo-second order kinetics, *World J. Microbiol. Biotechnol.*, 2009, **25**(8), 1413–1421.
- 88 H. F. Li, Y. B. Lin, W. M. Guan, *et al.*, Biosorption of Zn(II) by live and dead cells of *Streptomyces ciscaucasicus* Strain CCNWHX 72-14, *J. Hazard. Mater.*, 2010, **179**(1–3), 151–159.

