


 Cite this: *RSC Adv.*, 2020, 10, 4427

# Immobilized $\text{Ag}_3\text{PO}_4/\text{GO}$ on 3D nickel foam and its photocatalytic degradation of norfloxacin antibiotic under visible light

 Bang Ji,<sup>ab</sup> Wenfeng Zhao,<sup>c</sup> Jieli Duan,<sup>\*a</sup> Lanhui Fu,<sup>a</sup> Lizhe Ma<sup>a</sup> and Zhou Yang <sup>\*ad</sup>

In this study, a series of  $\text{Ag}_3\text{PO}_4/\text{graphene oxide (GO)}$  films were dip-coated on a metal nickel foam. The immobilized catalysts were characterized by X-ray diffraction, scanning electron microscopy, X-ray photoelectron spectroscopy, ultraviolet-visible spectroscopy, Raman spectroscopy, high-resolution transmission electron microscopy and photoluminescence spectroscopy. The results show that  $\text{Ag}_3\text{PO}_4/\text{GO}$  was successfully supported on a nickel foam. The photocatalytic activity of the film catalyst under visible light was investigated by the degradation of norfloxacin, an antibiotic. Photocatalytic stability of this catalyst was also investigated. An optimized film exhibited superior activity and stability, the degradation rate of norfloxacin was about 83.68% in 100 min and the reaction rate constant  $k$  was 1.9 times that of pristine  $\text{Ag}_3\text{PO}_4$ . Further investigation found that photo-generated holes ( $h^+$ ) and superoxide anion radicals ( $\cdot\text{O}_2^-$ ) are the main active species in the photodegradation process. The result indicates that the addition of GO inhibits the recombination of photogenerated electron-hole pairs, and thus has improved the photocatalytic activity and cyclic stability under visible light. The photocatalytic mechanism of the film catalyst was proposed. The prepared  $\text{Ag}_3\text{PO}_4/\text{GO}$  film catalyst is a promising candidate for treatment of wastewater containing antibiotics.

 Received 23rd October 2019  
 Accepted 4th December 2019

DOI: 10.1039/c9ra08678a

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## 1. Introduction

Norfloxacin is a quinolone antibiotic widely used in clinical medicine and aquaculture. However, it cannot be completely metabolized and absorbed by humans and animals. To make it worse, more than 80% of norfloxacin is excreted in the original form of the drug. With increasing human activity and migration, antibiotics are widespread and are contaminating soil and water. In the natural environment, antibiotics are generally difficult to degrade. Norfloxacin antibiotic has significant mutagenicity, teratogenicity, and embryotoxicity, which could not only cause serious harm to the ecological balance, but also pose a serious threat to human health. Therefore, how to effectively reduce antibiotic residues in the natural environment has attracted wide attention.

Conventional methods to remove antibiotic residues from the environment include membrane filtration,<sup>1</sup> physical adsorption,<sup>2</sup> ultraviolet photodegradation,<sup>3</sup> biological contact

oxidation,<sup>4</sup> wet oxidation,<sup>5</sup> electrochemical oxidation<sup>6</sup> and so on. Photocatalytic oxidation has been proved to be a promising method for decontamination and purification of wastewater.<sup>7–10</sup> Some photocatalysts have been applied to photocatalytic degradation of antibiotics residues in water.<sup>11–15</sup> In 2010, Ye's research group found that  $\text{Ag}_3\text{PO}_4$  has highly photocatalytic performance under the visible light for photodegradation of organic pollutants.<sup>16</sup> However,  $\text{Ag}_3\text{PO}_4$  is prone to photocorrosion, which limits its practical application. Photocorrosion in  $\text{Ag}_3\text{PO}_4$  is caused by the accumulation of photogenerated electrons ( $e^-$ ) in the conduction band that reduce  $\text{Ag}^+$  to elementary  $\text{Ag}^0$ . A lot of effort has been made to overcome this problem with  $\text{Ag}_3\text{PO}_4$  through compositing with carbon nanotubes,<sup>17</sup> activated carbon,<sup>18</sup> silver halide,<sup>19</sup> and graphene oxide.<sup>20</sup> The results show that these composite materials can prevent photocorrosion of  $\text{Ag}_3\text{PO}_4$ , and improve the photocatalytic stability. As a photo-generated electron transmitter, GO has the advantage of large surface area and high carrier mobility. It could expand the light absorption range, and also improve the stabilities by inhibiting the recombination of electron-hole pairs. Although  $\text{Ag}_3\text{PO}_4/\text{GO}$  composite has been reported to be able to improve the photocatalytic performance, the work was limited to the particle form. Because of powders are difficult to recycle, it hampers the practical application of  $\text{Ag}_3\text{PO}_4/\text{GO}$  catalyst for photodegradation of organic pollutants. A solution to the challenge of recycled use is through the immobilization of the catalyst in a film form.

<sup>a</sup>College of Engineering, South China Agricultural University, Guangzhou 510000, China. E-mail: duanjieli@scau.edu.cn; yangzhou@scau.edu.cn

<sup>b</sup>School of Materials Science and Engineering, Nanyang Technological University, Singapore 639798, Singapore

<sup>c</sup>College of Electronic Engineering, South China Agricultural University, Guangzhou 510000, China

<sup>d</sup>Guangdong Provincial Key Laboratory of Conservation and Precision Utilization of Characteristic Agricultural Resources in Mountainous Areas, Jiaying University, Meizhou 514021, China



To prepare a coating of photocatalyst, the selection of the substrate is important, which would directly affect the photocatalytic performance of the catalyst. 3D metal nickel foam has attracted considerable attention due to its high porosity, large surface area, excellent mechanical properties, corrosion resistance and permeability.<sup>21</sup> As a substrate of photocatalyst, it has potential application value in wastewater treatment. A series of catalysts, for example, graphene/graphite/TiO<sub>2</sub>,<sup>22</sup> graphene/nanosheets,<sup>23</sup> BiFeO<sub>3</sub>,<sup>24</sup> TiO<sub>2</sub>/SiO<sub>2</sub> (ref. 25) have been prepared using nickel foam as the catalyst support. However, nickel foam supported Ag<sub>3</sub>PO<sub>4</sub>/GO film catalyst and its photocatalytic degradation of antibiotics have rarely been reported. Nickel mesh has stable properties, corrosion resistance and unique advantages in wastewater treatment.

In this study, a series of Ag<sub>3</sub>PO<sub>4</sub>/GO film catalysts with different GO loading were prepared by dip-coating onto a nickel foam. The prepared film catalysts were characterized by X-ray diffraction (XRD), ultraviolet-visible spectroscopy (UV-vis), Raman spectroscopy (Raman), scanning electron microscopy (SEM), and X-ray photoelectron spectroscopy (XPS), high-resolution transmission electron microscopy (HR-TEM) and photoluminescence spectroscopy (PL). The photocatalytic activity and stability of the film catalyst were evaluated by the visible light degradation of norfloxacin, a common antibiotic. The kinetics of norfloxacin degradation over Ag<sub>3</sub>PO<sub>4</sub>/GO (denoted as AG-*X*, where *X* stands for the weight percent of GO in the composite). For comparison, Ag<sub>3</sub>PO<sub>4</sub> without GO (denoted as AP) was also examined. Moreover, the photocatalytic mechanism of Ag<sub>3</sub>PO<sub>4</sub>/GO film was discussed in this work.

## 2. Experimental

### 2.1 Material preparation

Nickel foam with a porosity of 95% was supplied by Jiangsu Taili Foam Metal Co., Ltd (China). The foam was cut into a rectangular shape with a size of 50 mm × 60 mm × 3 mm. Before the catalyst coating, the nickel foam was pretreated with concentrated nitric acid (HNO<sub>3</sub>) for 1 min, and then washing with absolute ethanol (C<sub>2</sub>H<sub>5</sub>OH) and deionized water to remove residual acid. The nickel foam was dried in a vacuum oven for 30 min.

The film catalyst was prepared by dip-coating method. The specific process is as follows: graphite oxide (GO) was prepared from graphite powder by a modified Hummers' method. A certain amount of GO (0.5%, 1%, 2%, and 4 wt%) was added to 150 mL of silver nitrate (AgNO<sub>3</sub>) solution (0.9 M) under vigorous stirring to obtain solution A. Then, 150 mL of disodium hydrogen phosphate (Na<sub>2</sub>HPO<sub>4</sub>) solution (0.3 M) was added dropwise into the solution A under stirring for 30 min, the obtained solution is denoted as solution B. After that, 8 g of inorganic binder sodium silicate (Na<sub>2</sub>SiO<sub>3</sub>) was immediately added into the solution B. The obtained mixture was intensively stirred for 150 min, which denoted as solution C. Finally, the pretreated nickel foam was soaked into the solution C and then withdrew at a rate of 9 mm s<sup>-1</sup>. After the dip-coating, the nickel foam was placed in a vacuum oven at 80 °C for 9 h to obtain the Ag<sub>3</sub>PO<sub>4</sub>/GO film catalysts supported on the metal nickel foam.

For comparison, Ag<sub>3</sub>PO<sub>4</sub> (denoted as AP) without addition of GO was prepared following the same procedure. All the reagents in the experiment were purchased from National Medicine Group Chemical Reagent Co., Ltd. They are of analytical grade and used without any pre-treatment.

The loading of Ag<sub>3</sub>PO<sub>4</sub>/GO photocatalysts on nickel foam was calculated by using the following equation:<sup>26</sup>

$$L = (W_t - W_{t_0})/W_{t_0} \times 100\% \quad (1)$$

where *L* is the loading amount of Ag<sub>3</sub>PO<sub>4</sub>/GO photocatalysts. *W<sub>t</sub>* and *W<sub>t<sub>0</sub></sub>* respectively represent the weights of nickel foam supported Ag<sub>3</sub>PO<sub>4</sub>/GO and nickel foam. The loading amounts of the samples AG-0.5, AG-1, AG-2, AG-4, and AP were 14.56 wt%, 15.89 wt%, 16.78 wt%, 17.23 wt%, 12.38 wt%, respectively.

### 2.2 Analytical method

The crystal structure of the photocatalysts was analyzed using an X-ray diffractometer (XRD, D/Max-2500 Rigaku, Japan). The surface morphology was analyzed by a scanning electron microscope (SEM, TDCLS-4800, Japan) and High-resolution transmission electron microscopy (HR-TEM, JEM-2100 F, Japan). The composition and chemical status of the photocatalysts were analyzed with X-ray photoelectron spectroscopy (XPS, Thermo Scientific ESCALAB 250 Xi). The optical properties of the photocatalysts were analyzed in a UV-vis spectrophotometer (UV-vis, Shimadzu, UV-2600, Japan). The molecular structure of the photocatalysts was analyzed by Raman spectroscopy using a 532 nm laser (Raman, Voyage BWS435-785SY, America). Photoluminescence spectra were recorded on a fluorescence spectrophotometer (PL, F-7000, Hitachi, Japan).

### 2.3 Photocatalytic activity

In order to test the photocatalytic performance of the prepared samples, the nickel foam was placed in a photocatalytic reactor first (schematic in Fig. 1). After that, 120 mL of 15 mg L<sup>-1</sup>

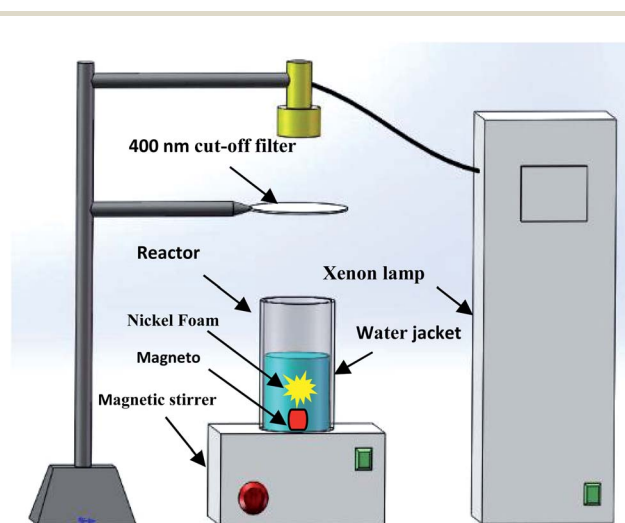


Fig. 1 Schematic of a device for photocatalytic performance evaluation.



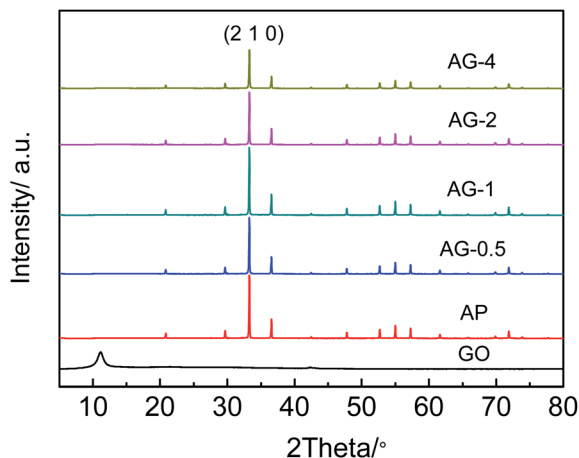


Fig. 2 XRD spectra of the GO, AP, and AG-X ( $X = 0.5, 1, 2$  and  $4$ ).

norfloxacin solution was added to the photocatalytic reactor. A 250 W xenon lamp with a cut-off filter of 400 nm was used as the visible light source. The intensity of the light source is  $100 \text{ mW cm}^{-2}$ . Before the experiment, the photocatalytic reactor was wrapped with a tin foil for 30 min to achieve adsorption

equilibrium of the norfloxacin solution in the dark. During the photo reaction, the temperature of the norfloxacin solution was maintained at a constant using a fan that introduces circulating cooling water to the water jacket. In the experiment, a 5 mL aliquot was sampled in every 10 min, and the concentration of norfloxacin was determined by measuring the characteristic peak at 272 nm using the UV-vis spectrophotometer. The photocatalytic degradation rate,  $K$ , is calculated by:<sup>27</sup>

$$K\% = (1 - C/C_0) \times 100\% \quad (2)$$

where  $C_0$  and  $C$  are the concentrations of norfloxacin solution at the start and time  $t$ , respectively.

#### 2.4 Photocatalytic reactive species

In order to study the photocatalytically active species in the experiment, a radical capture experiment in the photo-degradation process of norfloxacin was carried out. The detailed experimental processes are the same for the photocatalytic activity experiment. During the experiment, isopropanol (IPA) was chosen as hydroxyl radical ( $\cdot\text{OH}$ ) scavenger, sodium oxalate (AO) was chosen as hole ( $\text{h}^+$ ) scavenger, benzoquinone (BQ) was chosen as radicals ( $\cdot\text{O}_2^-$ ) scavenger, the photocatalytic degradation rate of

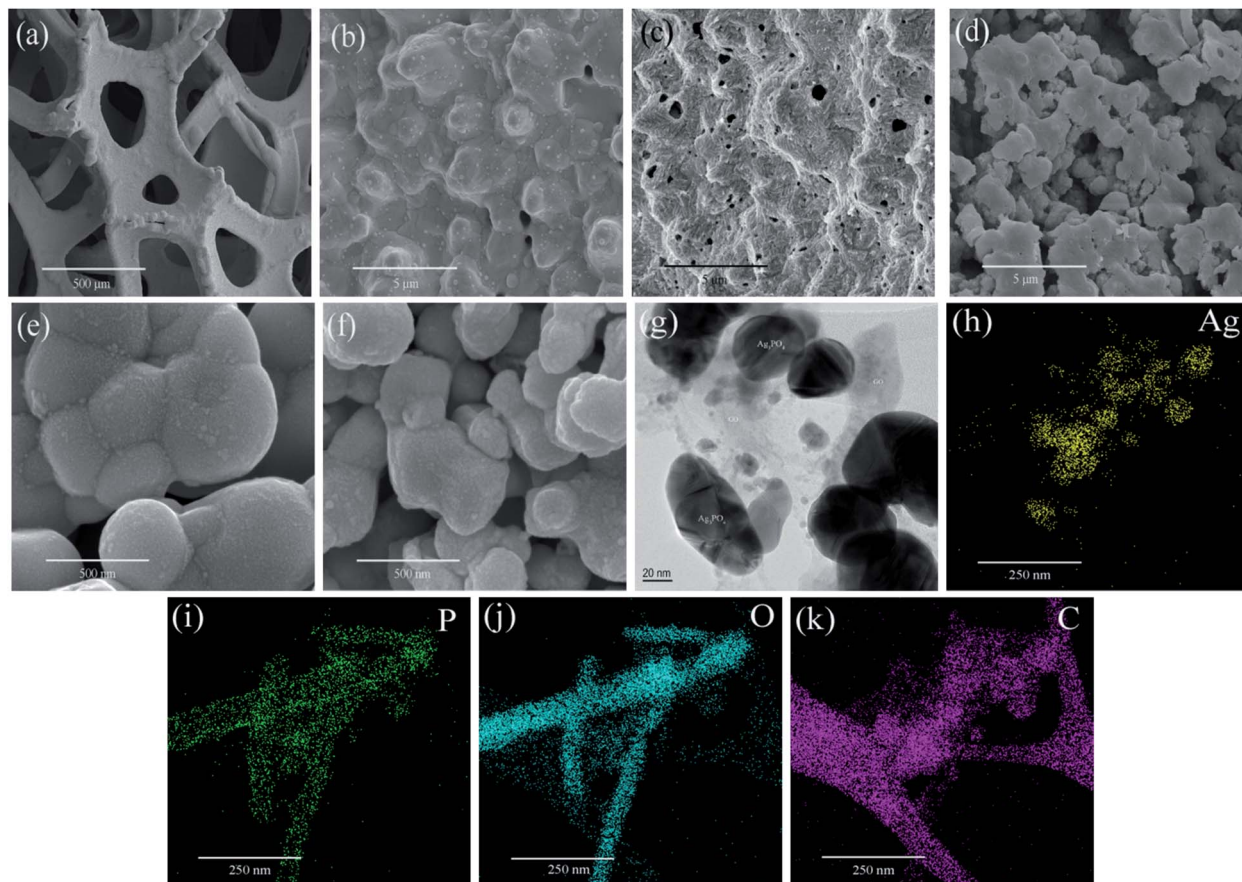


Fig. 3 SEM images: (a)  $200\times$  magnification of untreated nickel foam, (b)  $20\,000\times$  magnification of untreated nickel foam, (c)  $20\,000\times$  magnification of pre-treated nickel foam, (d)  $20\,000\times$  magnification of AG-1, (e)  $20\,000\times$  magnification of AP, (f)  $20\,000\times$  magnification of AG-1. (g) TEM of AG-1, (h–k) EDS mapping of AG-1.





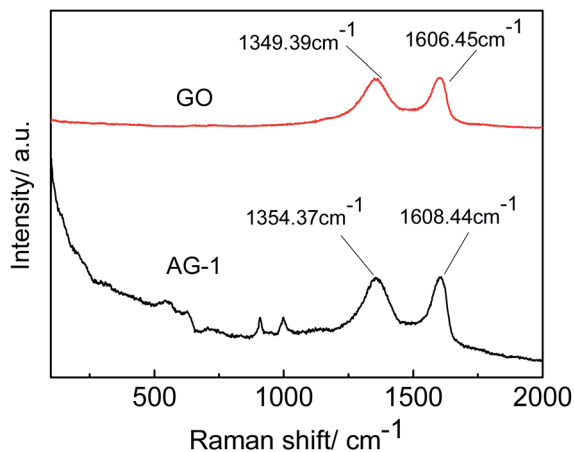


Fig. 4 Raman spectra of GO and AG-1.

each set of experiments was analyzed, and the active agents in the photocatalytic degradation process were then analyzed.

### 2.5 Photocatalytic stability

The prepared sample of AP and AG-2 (*i.e.*,  $\text{Ag}_3\text{PO}_4$  with 2% GO) were used to evaluate the stability of the film catalyst. After each photocatalytic experiment, the film catalysts were rinsed with deionized water and then dried in a vacuum for the next cycle of the same test. The photo degradation experiment was repeated for 4 cycles.

## 3. Results and discussion

### 3.1 Crystal structure

Fig. 2 shows the XRD spectra of the GO,  $\text{Ag}_3\text{PO}_4$ ,  $\text{Ag}_3\text{PO}_4/\text{GO}$  with addition different percentage of GO. The peaks corresponding to  $\text{Ag}_3\text{PO}_4$  (JCPDS-06-0505) were observed in AP, AG-0.5, AG-1, AG-2 and AG-4. The diffraction angles at  $20.88^\circ$ ,  $29.69^\circ$ ,  $33.29^\circ$ ,  $36.58^\circ$ ,  $42.48^\circ$ ,  $55.02^\circ$ ,  $57.28^\circ$  and  $71.89^\circ$  are indexed to the (1 1 0), (2 0 0), (2 1 0), (2 1 1), (2 2 0), (3 2 0), (3 2 1) and (4 2 1) planes.<sup>28</sup> Pure GO sample shows a characteristic diffraction peak at  $10.95^\circ$ , confirming that the material prepared is indeed GO. However, the characteristic diffraction peak of GO was not observed in the AG-X samples, due to the low concentration of GO in the composites.<sup>29</sup> The introduction of GO does not change the crystal structure of  $\text{Ag}_3\text{PO}_4$ . As the amount of added GO increased, the intensity of the characteristic diffraction peak at the (2 1 0) crystal plane of  $\text{Ag}_3\text{PO}_4$  decreased, consistent with an early report.<sup>30</sup>

### 3.2 Surface morphology

Fig. 3(a) shows the SEM image of untreated nickel foam with a three-dimensional skeleton structure. It has good porosity and permeability, which could make the catalyst full of the reaction chamber and is conducive to the improvement of catalytic performance. Fig. 3(b) is an enlarged view of the untreated nickel foam showing that the surface of the foam is very smooth. Fig. 3(c) is an SEM image of a pretreated nickel foam with a magnification of 20 000 times. From the picture we could

know that after the acidification and oxidation treatment, the organic impurities on the surface of the nickel foam are removed. The surface is corroded, showing a porous structure clearly, which is favorable for the catalyst load. Fig. 3(d) is an SEM image of 20 000 times the magnification of the AG-1 sample. It can be seen from the diagram that the  $\text{Ag}_3\text{PO}_4/\text{GO}$  catalyst is uniformly loaded on the surface of the nickel foam, and the dispersion is very well. Fig. 3(e) and (f) are 200 000 times the magnification of sample AP and AG-1, respectively. It can be seen from the diagram that when GO was introduced into the synthetic system, the diameter of  $\text{Ag}_3\text{PO}_4$  decreased. It was mainly due to the self-assembly of positively-charged  $\text{Ag}^+$  on negatively charged GO sheets driven by electrostatic action, which prevents the growth of  $\text{Ag}_3\text{PO}_4$  particles.<sup>31,32</sup> Fig. 3(g) is TEM images of the AG-1 sample, suggested that  $\text{Ag}_3\text{PO}_4$  particles were distributed on the GO sheets. This structure may provide sufficient contacting surface area between  $\text{Ag}_3\text{PO}_4$  particles and GO sheets, so it's convenient for the transmission of carriers. Further analysis of the EDS elemental mappings (Fig. 3(h–k)) showed that Ag (silver), phosphorus (P), oxygen (O) and carbon (C) homogeneously distributed in the sample AG-1, which further confirms that the  $\text{Ag}_3\text{PO}_4$  particles distributed on the GO sheets.

### 3.3 Molecular structure analysis

In order to characterize the molecular structure of the prepared materials, Raman analysis was performed on the GO and AG-1 samples. As shown in Fig. 4, vibrational bands at  $542.37\text{ cm}^{-1}$ ,  $627.58\text{ cm}^{-1}$ ,  $908.02\text{ cm}^{-1}$ , and  $999.68\text{ cm}^{-1}$  are attributed to  $\text{Ag}_3\text{PO}_4$ , while the ones at  $1354.37\text{ cm}^{-1}$  and  $1608.44\text{ cm}^{-1}$  correspond to GO in the AG-1 sample. The results confirm that the composite materials contain both  $\text{Ag}_3\text{PO}_4$  and GO, however the GO content was too low to be detected by XRD as explained earlier. In the spectra of the GO, the bands at  $1349.39\text{ cm}^{-1}$  and  $1606.45\text{ cm}^{-1}$  correspond to the D peak and G peak of GO, and the intensity ratio of D to G peaks ( $I_D/I_G$ ) is 0.97. The D peak represents the lattice defect of the carbon atoms, and the G peak is related to the in-plane stretching vibration of

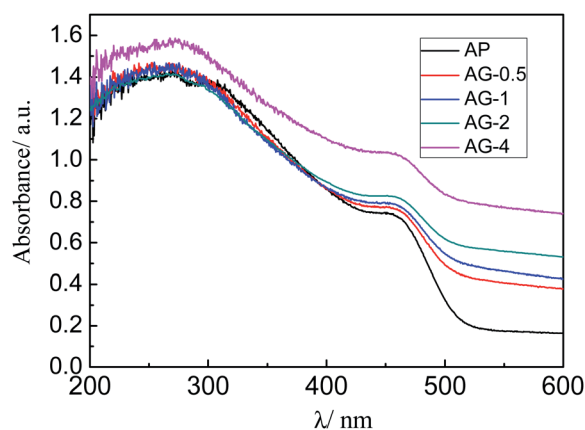


Fig. 5 UV-vis diffuse reflectance spectroscopy of AP, AG-0.5, AG-1, AG-2 and AG-4 samples.



the carbon atom in the  $sp^2$  hybrid. The  $I_D/I_G$  ratio is generally used to evaluate the degree of defects in the GO. After GO addition to the  $Ag_3PO_4$ , the  $I_D/I_G$  value is increased to 0.99, indicating that the degree of defects increases, which is consistent with the previous study.<sup>33,34</sup> The result further demonstrates the successful composition of  $Ag_3PO_4$  and GO.

### 3.4 Optical absorption

The UV-vis diffuse reflectance spectra of the prepared samples (AP, AG-0.5, AG-1, A-2, and AG-4) are shown in Fig. 5. Pure  $Ag_3PO_4$  (AP) shows an absorption edge at about 520 nm. After GO was added, the absorbance in the visible region has increased with the amount of GO added, which is potentially beneficial for improving the photocatalytic performance.<sup>35</sup>

However, the absorption edges of the (AG-X) samples are not shifted as compared with the AP sample.

### 3.5 Composition and chemical bonding state analyses

To understand the composition and chemical state of the materials, samples of GO, AP, and AG-1 were analyzed by XPS. Fig. 6(a) is the full spectrum of the samples. C and O elements were found in GO, while C, Ag, O, and P elements were detected in AP and AG-1 samples. Fig. 6(b) is the high-resolution Ag 3d spectra of AG-1 sample. The electron binding energy of Ag 3d appears at 367.9 and 374.1 eV, which are attributed to  $Ag\ 3d_{5/2}$  and  $Ag\ 3d_{3/2}$ , respectively. It confirms that Ag exists as  $Ag^+$  in AG-1 sample. Fig. 6(c) is a high-resolution P 2p spectrum of AG-1 sample. The peak at 132.7 eV indicates that P exists as  $P^{5+}$ . Fig. 6(d) shows the high-resolution O 1s spectra of samples GO,

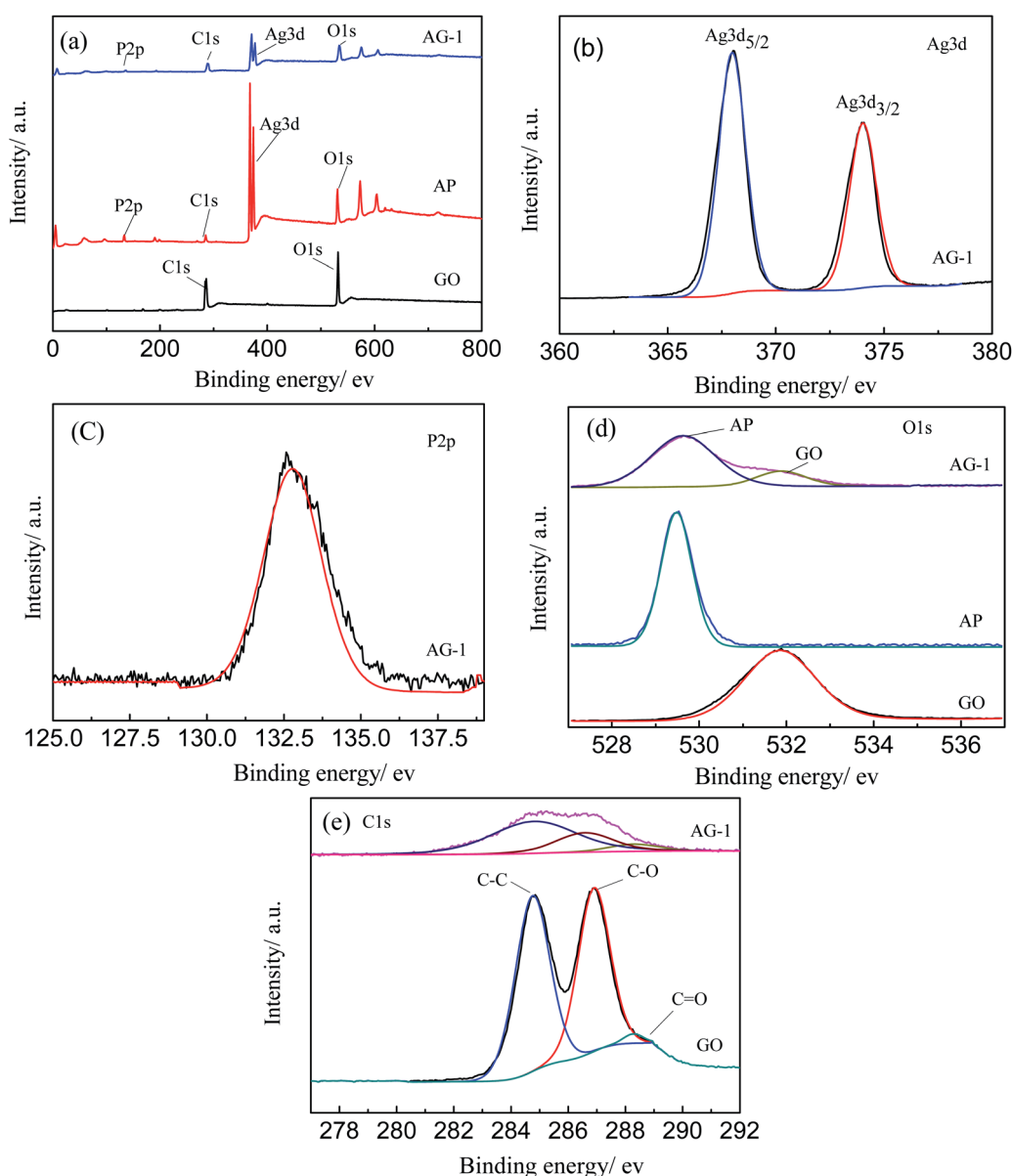


Fig. 6 XPS spectra of (a) AG-1, AP, and GO; (b) Ag 3d of AG-1; (c) P 2p of AG-1; (d) O 1s of GO, AP and AG-1; (e) C 1s of AG-1 and GO.



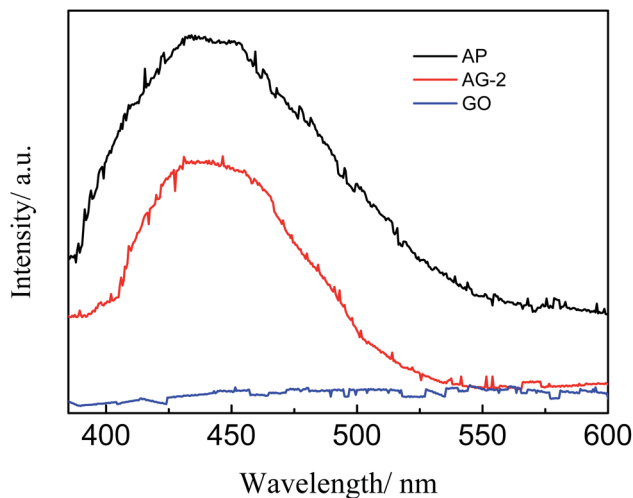


Fig. 7 Photoluminescence spectra of AP, AG-2 and GO samples.

AP and AG-1. The two peaks at 529.5 and 531.8 eV in AG-1 sample correspond to the binding energy of O 1s in  $\text{Ag}_3\text{PO}_4$ , and O 1s in GO, respectively. Fig. 6(e) is a high-resolution C 1s spectral peak of AG-1 and GO. Both samples show peaks at 284.8, 286.9, and 288.3 eV, which can be ascribed to binding energies of C-C, C-O and C=O bonds in GO, respectively. All these results indicate presence of  $\text{Ag}_3\text{PO}_4$  and/or GO in the analyzed samples.<sup>36–38</sup>

### 3.6 Photoluminescence

It is well known that a lower photoluminescence spectra intensity can be indicative of a lower photo charge carrier recombination rate, implying a faster separation and transfer of photo generated electron-hole pairs.<sup>39</sup> Fig. 7 shows the PL intensity of the AP, AG-2 and GO. The PL curve has a similar shape, but is much lower than that of AP in the intensity. This result suggests that GO can inhibit the recombination of photo-generated charge carriers under the visible light. Therefore, addition GO is expected to improve the photocatalytic activity of  $\text{Ag}_3\text{PO}_4$ , which will be discussed further later.

### 3.7 Photocatalytic degradation of norfloxacin

Fig. 8(a) shows the UV-vis spectral changes of photocatalytic degradation of norfloxacin with AG-1 composites under the visible light. The absorption peak of norfloxacin located at 272 nm decreased gradually with increasing time. The intensity change, which implies the photocatalytic degradation rate, was fastest around 60 min. Fig. 8(b) displays the photocatalytic degradation curves of norfloxacin over AP, AG-0.5, AG-1, A-2 and AG-4 composites under visible light irradiation. The adsorption rates of AP, AG-0.5, AG-1, AG-2 and AG-4 samples were 9.41%, 14.37%, 19.01%, 30.24%, and 38.73% respectively after reaching the adsorption equilibrium. With the increase of GO content, the adsorption rate increased, mainly because GO has a large specific surface area, and there are many functional groups on the surface. The high adsorption performance of the norfloxacin molecule on the  $\text{Ag}_3\text{PO}_4/\text{GO}$  favors the

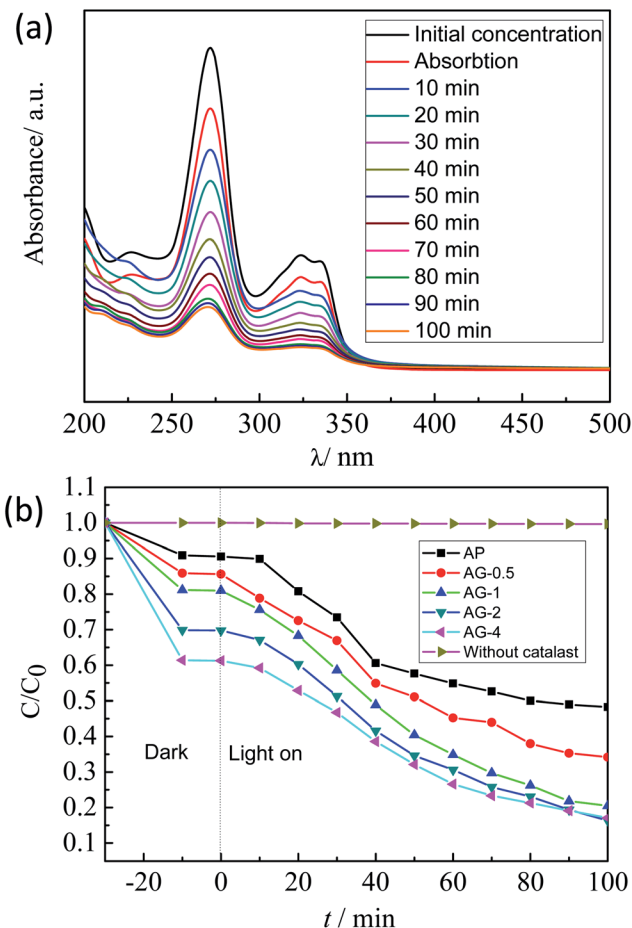


Fig. 8 (a) UV-vis spectral changes of photocatalytic degradation of norfloxacin with AG-1 composites at different irradiation times. (b) Photocatalytic degradation curves of norfloxacin over AP, AG-0.5, AG-1, A-2 and AG-4 composites under visible light irradiation.

photocatalytic degradation of the norfloxacin. In addition, the comparative experiments showed that norfloxacin would not be degraded under visible light without the  $\text{Ag}_3\text{PO}_4$  photocatalyst.

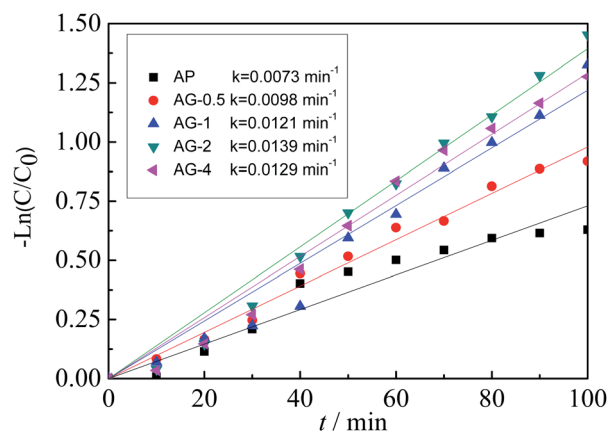


Fig. 9 Kinetics of norfloxacin degradation over AG-X, AP under visible light irradiation.



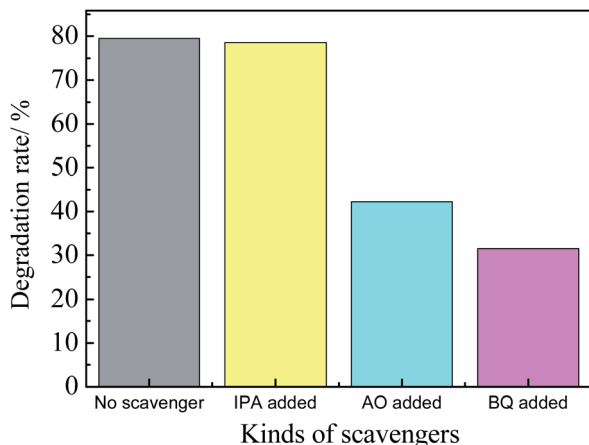


Fig. 10 Photocatalytic degradation rate of different kinds of scavengers over AG-1 composites.

The photocatalytic performance of the  $\text{Ag}_3\text{PO}_4$  is significantly improved by the addition of GO. There are two benefits with incorporating GO in the photocatalyst. First, the relatively large

surface area of the GO can provide more active sites to adsorb norfloxacin to be closer to the photocatalytically active sites. Second, photo-generated electrons of  $\text{Ag}_3\text{PO}_4$  can be transferred to GO, resulting in more effective charge separation.

According to the Langmuir–Hinshelwood model, photocatalytic degradation of norfloxacin was studied using the pseudo first-order kinetic model:<sup>40</sup>

$$-\ln(C/C_0) = kt \quad (3)$$

where  $t$  is the light irradiation time, and  $k$  the reaction rate constant. The photocatalytic performance of each sample can be compared by the  $k$  values, and the higher the  $k$ , the better the photocatalytic performance.

Fig. 9 shows the kinetic of norfloxacin photocatalytic degradation over  $\text{Ag}_3\text{PO}_4$  and  $\text{Ag}_3\text{PO}_4/\text{GO}$  under visible light irradiation. With increasing GO content, the  $k$  value increases initially and then decreases. The  $k$  value of the sample AG-2 is the highest at  $0.0139 \text{ min}^{-1}$ , about 1.9 times that of AP. This result reveals that adding appropriate amount of GO could enhance the photocatalytic activity of norfloxacin under visible light irradiation.

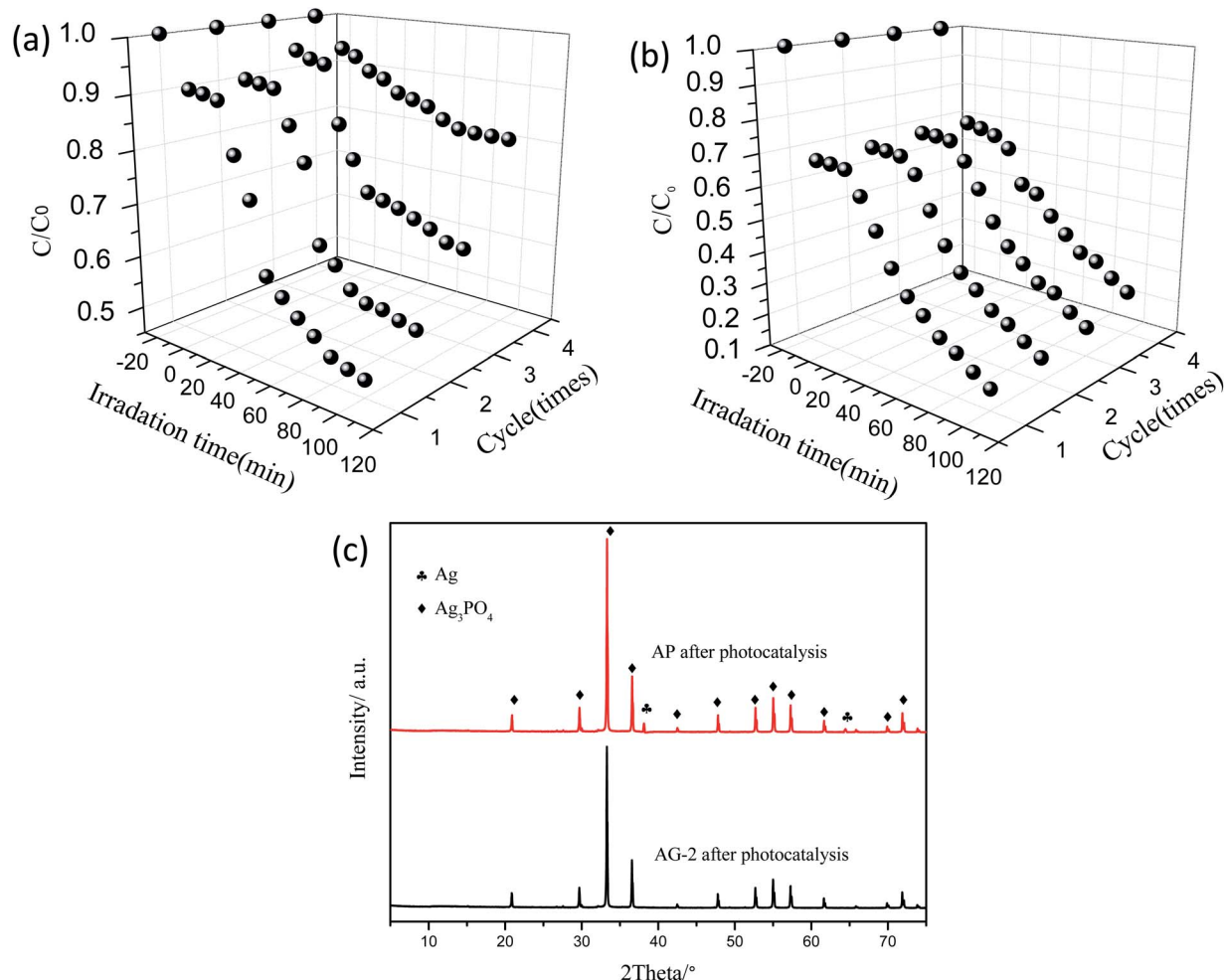


Fig. 11 Repeated photodegradation experiment by (a) AP and (b) AG-2, (c) XRD patterns of the AP and AG-2 samples after 4 cycles.





In order to study the photocatalytic reaction mechanism of the composite film, a radical capture experiment was carried out. Fig. 10 shows the photocatalytic degradation rate of different kinds of scavengers over AG-1 composites. It shows that when AO and BQ were added to the norfloxacin solution, the photocatalytic degradation rate decreased significantly. When IPA was added, the photocatalytic degradation rate hardly changed. The results imply that  $h^+$  and  $\cdot O_2^-$  are the main active species in the photocatalytic reaction of  $Ag_3PO_4/GO$ , and  $\cdot OH$  is not responsible for the photocatalytic degradation of norfloxacin.

In this study, GO was added mainly to increase photocatalytic stability and reduce the cost of  $Ag_3PO_4$ . Fig. 11(a) and (b) are the results of the cyclic experiment of the sample AP, AG-2 photocatalytic degradation of  $15 \text{ mg L}^{-1}$  of norfloxacin solution. It is clear that after four cycles, the photocatalytic degradation efficiency of sample AG-2 did not change much, but the AP sample has completely lost its activity. Furthermore, the crystalline structures of the sample AP and AG-2 after four recycle times were investigated to understand the structure stability (Fig. 11(c)). It was clear that a relatively weak XRD diffraction peak at  $38.11^\circ$  and  $64.44^\circ$  for the AP sample after four recycles can be indexed to the (1 1 1) and (2 2 0) lattice plane of metallic  $Ag^0$  (JCPDS-65-2871), respectively. There was no distinct metallic  $Ag^0$  peak in the XRD patterns of the AG-2 sample after four cycles, confirming higher stability of the AG-2 sample. The main reason is that in the photocatalytic process the photo-generated electrons are enriched on the surface of  $Ag_3PO_4$ , which will reduce  $Ag^+$  to  $Ag^0$ . With the loading of GO, the electrons can be transferred away much quickly, as evidenced by the PL measurement discussed earlier. All these results have demonstrated that the immobilized  $Ag_3PO_4/GO$  composite is an efficient and stable visible light photocatalytic catalyst.

Based on the discussions above, a possible mechanism for the photodegradation of norfloxacin over  $Ag_3PO_4/GO$  composite film catalyst has been proposed, as illustrated by Fig. 12 and eqn (4)–(8). Under visible light irradiation, electrons ( $e^-$ ) in the valence band (VB) of  $Ag_3PO_4$  are excited to conduction band (CB), which produces holes ( $h^+$ ) in the valence band (eqn (4)). With the addition of GO, electrons ( $e^-$ ) in the conduction band of  $Ag_3PO_4$  are transferred to GO, which inhibits the recombination of electrons ( $e^-$ ) and holes ( $h^+$ ) (eqn (5)). This will also inhibit the photo-corrosion of  $Ag_3PO_4$  as discussed before. With holes ( $h^+$ ) formed in the valence band (VB) of  $Ag_3PO_4$ , hydroxyl radicals ( $\cdot OH$ ) are formed through the reaction of holes ( $h^+$ ) with water (eqn (6)). The electrons ( $e^-$ ) accumulated on the GO surface and  $Ag_3PO_4$  conduction band reacts with the adsorbed oxygen to form superoxide anion radical ( $\cdot O_2^-$ ) (eqn (7)). The  $\cdot OH$  and  $\cdot O_2^-$  can be used as strong oxidizers to oxidize norfloxacin to  $CO_2$  and  $H_2O$  (eqn (8)). GO has strong adsorption capacity of norfloxacin molecules and electrons ( $e^-$ ) capture capability, it accelerates the capture of norfloxacin molecules onto the catalyst surface. It also inhibits the recombination of photogenerated electron hole pairs, thus providing more active free radicals to participate in the photocatalytic degradation of norfloxacin. Therefore, the AG-X composite film exhibited higher photocatalytic activity and cyclic stability than AP.

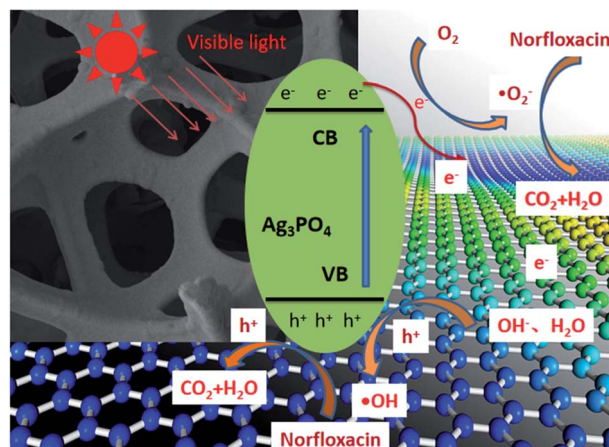
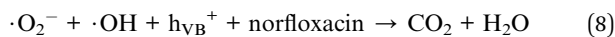
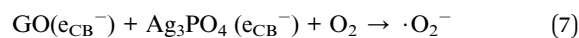
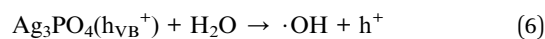
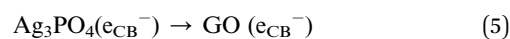
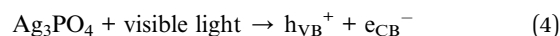


Fig. 12 Schematic diagram of photocatalytic enhancement mechanism by  $Ag_3PO_4/GO$  composite film catalyst.



## 4. Conclusion

In summary, a series of  $Ag_3PO_4/GO$  composite film catalysts were prepared by dip coating onto to the surface of nickel foam. The optical properties, chemical composition, surface morphology, molecular structure, and crystalline structure of the films were characterized and analyzed. The results showed that  $Ag_3PO_4/GO$  was successfully loaded onto the surface of the nickel foam. The photocatalytic activity and stability of the film catalyst were investigated by the degradation of norfloxacin antibiotic under visible light. The results showed that the GO has a great influence on the activities. With the increase of GO, the adsorption and photocatalytic efficiency increased. The AG-2 sample exhibited an optimum photocatalytic activity and stability under visible light. Furthermore, photocatalytic radical capture experiments proved that holes ( $h^+$ ) and superoxide anion radicals ( $\cdot O_2^-$ ) are the main active groups in the photocatalytic degradation process. Accordingly, the photocatalytic reaction mechanism was proposed. The immobilized  $Ag_3PO_4/GO$  composite film is a promising candidate in treatment of antibiotics in wastewater.

## Conflicts of interest

The authors declare that they have no conflict of interest.





## Acknowledgements

This paper is financially supported by China Agriculture Research System (CARS-31), Science and Technology Planning Project of Guangdong Province, China (2018A050506076).

## References

- Q. Jia, Z. Geng, Y. Liu, W. Wang, C. Han, G. Yang, H. Li and L. Qu, *J. Mater. Sci.*, 2018, **53**, 14989–14997.
- D. Yin, Z. Xu, J. Shi, L. Shen and Z. He, *J. Water Reuse Desalin.*, 2018, **8**, 350–359.
- S. Donsingha and K. Assatarakul, *Food Control*, 2018, **92**, 162–168.
- Q. Sui, C. Jiang, J. Zhang, D. Yu, M. Chen, Y. Wang and Y. Wei, *Environ. Int.*, 2018, **118**, 274–281.
- R. S. Ribeiro, Z. Frontistis, D. Mantzavinos, D. Venieri, M. Antonopoulou, I. Konstantinou, A. M. T. Silva, J. L. Faria and H. T. Gomes, *Appl. Catal., B*, 2016, **199**, 170–186.
- L. Zhu, B. Santiago-Schübel, H. Xiao, H. Hollert and S. Kueppers, *Water Res.*, 2016, **102**, 52–62.
- Y. Hong, C. Li, G. Zhang, Y. Meng, B. Yin, Y. Zhao and W. Shi, *Chem. Eng. J.*, 2016, **299**, 74–84.
- R. Wang, J. Tang, X. Zhang, D. Wang, X. Wang, S. Xue, Z. Zhang and D. D. Dionysiou, *J. Hazard. Mater.*, 2019, **375**, 161–173.
- Y. Zeng, D. Chen, T. Chen, M. Cai, Q. Zhang, Z. Xie, R. Li, Z. Xiao, G. Liu and W. Lv, *Chemosphere*, 2019, **227**, 198–206.
- N. Tang, Y. L. Li, F. Chen and Z. Han, *RSC Adv.*, 2018, **8**, 42233–42245.
- Z. Huang, X. Dai, Z. Huang, T. Wang, L. Cui, J. Ye and P. Wu, *Chemosphere*, 2019, **221**, 824–833.
- J. H. O. S. Pereira, A. C. Reis, D. Queirós, O. C. Nunes, M. T. Borges, V. J. P. Vilar and R. A. R. Boaventura, *Sci. Total Environ.*, 2013, **463–464**, 274–283.
- B. Wu, Y. Li, K. Su, L. Tan, X. Liu, Z. Cui, X. Yang, Y. Liang, Z. Li, S. Zhu, K. W. K. Yeung and S. Wu, *J. Hazard. Mater.*, 2019, **377**, 227–236.
- Q. Yan, C. Li, C. Lin, Y. Zhao and M. Zhang, *J. Mater. Sci.: Mater. Electron.*, 2018, **29**, 2517–2524.
- W. Li, Q. Chen, X. Lei and S. Gong, *RSC Adv.*, 2019, **9**, 51–519.
- Z. Yi, J. Ye, N. Kikugawa, T. Kako, S. Ouyang, H. Stuart-Williams, H. Yang, J. Cao, W. Luo, Z. Li, Y. Liu and R. L. Withers, *Nat. Mater.*, 2010, **9**, 559–564.
- D. Zhao, A. Li, M. Wu and M. Du, *React. Kinet., Mech. Catal.*, 2018, **124**, 347–362.
- H. Wang, Z. Ye, C. Liu, J. Li, M. Zhou, Q. Guan, P. Lv, P. Huo and Y. Yan, *Appl. Surf. Sci.*, 2015, **353**, 391–399.
- L. Cai, T. Xu, J. Shen and W. Xiang, *Mater. Sci. Semicond. Process.*, 2015, **37**, 19–28.
- G. Chen, M. Sun, Q. Wei, Y. Zhang, B. Zhu and B. Du, *J. Hazard. Mater.*, 2013, **244–245**, 86–93.
- J. Lu, T. Xiong, W. Zhou, L. Yang, Z. Tang and S. Chen, *ACS Appl. Mater. Interfaces*, 2016, **8**, 5065–5069.
- J. Jia, D. Li, X. Cheng, J. Wan and X. Yu, *Appl. Catal., A*, 2016, **525**, 128–136.
- P. Miao, J. He, Z. Sang, F. Zhang, J. Guo, D. Su, X. Yan, X. Li and H. Ji, *J. Alloys Compd.*, 2018, **732**, 613–623.
- S. Li, G. Zhang, H. Zheng, Y. Zheng and P. Wang, *J. Power Sources*, 2018, **386**, 21–27.
- H. Hu, W. Xiao, J. Yuan, J. Shi and W. Shangguan, *Int. J. Photoenergy*, 2008, **2008**, 1–8.
- S. Li, G. Zhang, H. Zheng, Y. Zheng and P. Wang, *J. Power Sources*, 2018, **386**, 21–27.
- M. Chahkandi and M. Zargazi, *J. Hazard. Mater.*, 2019, **380**, 120879.
- Z. Diao, S. Pu, W. Qian, S. Liang, L. Kong, D. Xia, Z. Lei, J. Du, H. Liu and J. Yang, *Chemosphere*, 2019, **221**, 511–518.
- Q. Xiang, D. Lang, T. Shen and F. Liu, *Appl. Catal., B*, 2015, **162**, 196–203.
- Q. Yan, X. Xie, C. Lin, Y. Zhao, S. Wang and Y. Liu, *J. Mater. Sci.: Mater. Electron.*, 2017, **28**, 16696–16703.
- W. Chen, X. Niu and J. Wang, *J. Photochem. Photobiol., A*, 2018, **356**, 304–311.
- X. Yang, J. Qin, Y. Jiang, K. Chen, X. Yan, D. Zhang, R. Li and H. Tang, *Appl. Catal., B*, 2015, **166–167**, 231–240.
- C. Cui, S. Li, Y. Qiu, H. Hu, X. Li, C. Li, J. Gao and W. Tang, *Appl. Catal., B*, 2017, **200**, 666–672.
- H. Wang, L. Zou, Y. Shan and X. Wang, *Mater. Res. Bull.*, 2018, **97**, 189–194.
- P. Wang, T. Chen, B. Yu, P. Tao and Y. Bai, *J. Taiwan Inst. Chem. Eng.*, 2016, **62**, 267–274.
- L. Zhou, O. G. Alvarez, C. S. Mazon, L. Chen, H. Deng and M. Sui, *Catal. Sci. Technol.*, 2016, **6**, 5972–5981.
- X. Yang, H. Cai, M. Bao, J. Yu, J. Lu and Y. Li, *Chem. Eng. J.*, 2018, **334**, 355–376.
- Z. Liu, H. Feng, S. Xue, P. Xie, L. Li, X. Hou, J. Gong, X. Wei, J. Huang and D. Wu, *Appl. Surf. Sci.*, 2018, **458**, 880–892.
- X. Song, Y. Li, Z. Wei, S. Ye and D. D. Dionysiou, *Chem. Eng. J.*, 2017, **314**, 443–452.
- X. Chen, Y. Dai, X. Wang, J. Guo, T. Liu and F. Li, *J. Hazard. Mater.*, 2015, **292**, 9–18.

