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Copper-bottomed: electrochemically active bacteria exploit conductive sulphide networks for enhanced electrogenicity†

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In this study, we demonstrate that anodic electroactive bacteria like *Geobacter sulfurreducens* generate copper(I) and copper(II) sulphides when grown on copper electrodes. The insoluble copper sulphides form a conductive network within the biofilms, strongly enhancing the biofilm electrogenicity – i.e., the ability of the biofilm to produce electric currents. Compared to biofilms grown on graphite, the average relative current density of copper-based biofilms was 237%, with a maximum geometric current density of $1.59 \pm 0.23 \text{ mA cm}^{-2}$. An additional electrochemical CuS deposition prior to biofilm cultivation further increased the bioelectrocatalytic current generation to 2 mA cm^{-2} . The chemical deposition of CuS onto graphite allowed cultivating biofilms with current densities 134% higher than at unmodified graphite. This approach – the chemical CuS deposition onto inexpensive electrode materials – thus represents a promising pathway for the development of scalable, high-performance electrode materials for microbial electrochemical technologies.

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Broader context

An increase in the performance of biofilm electrodes is decisive for a future success of microbial fuel cells and related microbial electrochemical technologies. Although the development of 3D-electrode designs has already substantially increased the current production of biofilm electrodes, it is essential to understand and to overcome intrinsic limitations in electroactive biofilms to enhance their bioelectrochemical turnover rates. Here we demonstrate that, when grown on metallic copper, *Geobacter* dominated electroactive biofilms can form a matrix of highly conductive copper sulphides and exploit it for substantially enhanced current generation. The findings underline that the ohmic resistance represents a major bottleneck in electroactive biofilms that can be overcome by simple chemical electrode modifications.

1. Introduction

Biocompatible and biocidal – in technical or biotechnological applications these antipodal material attributes are decisive for either allowing microbial life on a given surface or preventing it. Some metals like copper and its alloys have gained a distinct, general perception as efficient antimicrobial materials, a perception that goes back to antiquity.¹ In naval history, the cladding of wooden ship hulls with copper sheets was the first efficient measure against biofouling in the marine environment,² creating the basis of the idiomatic phrase “to be copper-bottomed”. Today, the metal’s antimicrobial properties, which are based on the

antimicrobial action of traces of copper ions on the metal surface,³ are being exploited in a wide range of applications in healthcare, food processing, biotechnology and many other areas.^{4,5}

The general perception of copper’s antimicrobial properties has prevented its consideration for applications that require – instead of forbid – its colonization by bacteria, for example as electrode material for microbial fuel cells (MFCs) and related microbial electrochemical technologies (METs). In these systems, for their great biocompatibility, carbonaceous materials like graphite have been considered as the materials of choice – although their low specific conductivity may cause substantial electrical (ohmic) energy losses, thus reducing their applicability.

In a recent publication we reported the surprising finding that electrogenic (anodic electrochemically active) bacteria not only colonize copper electrodes readily, but also deliver current densities higher than biofilms grown on graphite or gold surfaces.⁶ Whereas the biocompatibility of copper for the growth of electrogenic bacteria could be explained by the low

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redox potential in the anaerobic environment, which prevents the liberation of free, antimicrobial copper ions, the reasons for the enhanced electrogeneity remained unsolved. By performing a comparative electrochemical and spectroscopic analysis of electroactive biofilms grown on copper and graphite electrodes, we are now able to explain the mechanisms of the enhanced electrogeneity (*i.e.*, the ability of electrogenic bacteria to deliver electrons to an electrode) of copper-based, *Geobacter* dominated mixed culture biofilms. The study underlines major physical limitations in electroactive biofilms and derives strategies to develop low-cost materials for an improved microbial current generation.

2. Methods

2.1 Electrochemical setup and general experimental conditions

All chemicals used in this study were purchased from Roth or Sigma-Aldrich and were of analytical grade. The bioelectrochemical measurements were performed under strictly anaerobic conditions at a temperature of 35 °C. The measurements were carried out in half-cell setups under potentiostatic control (MPG 2, BioLogic, France) using round-bottom flasks (250 mL) as electrochemical cells. Stainless steel mesh (Ludwig Ohlendorf KG, Germany) and Ag/AgCl electrodes (sat. KCl, 0.197 V *vs.* SHE; Sensortechnik Meinsberg GmbH, Germany) served as counter and reference electrodes, respectively. All electrode potentials in this article refer to the above Ag/AgCl electrode.

For a comparative evaluation of the biofilm growth and electrochemical biofilm performance at graphite and copper electrodes under identical environmental and electrochemical conditions, each electrochemical cell contained at least one graphite and one copper working electrode, addressed individually.

Copper electrodes were prepared from copper foil (0.2 mm thickness; >99.9% purity, Chempur, Germany), which was cut to a geometric surface area of 2.25 cm². A copper wire (ϕ 1 mm) was soldered to the back side as electrical connection. For the preparation of graphite electrodes, graphite rods (ϕ 1.5 cm, polycrystalline graphite, CP Graphite GmbH) were cut into slices with a thickness of 5 mm. A hole (ϕ 1 mm) was drilled 5 mm into the graphite disk from the lateral area. It was filled with conductive silver paint (Busch silver paint, Busch 27 GmbH & Co. KG, Germany), and a copper wire (ϕ 1 mm) was pushed into the hole. For all prepared electrodes, the back sides (and in case of the graphite electrodes also the cylindrical surface) of the electrodes were isolated using two component epoxy glue (R&G Faserverbundwerkstoffe GmbH, Germany). All connecting wires were isolated using heat shrinking tubes. Prior to the electrochemical measurements the electrodes were rinsed with isopropanol and ethanol to remove organic residues and were sonicated (Emmi 12HC, EMAG AG, Germany) in deionised water for 30 minutes.

For electrochemical CuS deposition on the surfaces of the copper electrodes, the respective electrodes were immersed into 250 mL of a 0.1 M NaNO₃ solution and a potential of −0.2 V was applied until stationary currents were achieved. 1 g of sodium sulphide hydrate (purity 60%) was added to the solution and the potential was held overnight (approximately 16 h) to finish the

reaction.⁷ To precipitate CuS onto graphite, the polycrystalline graphite electrodes were placed into a 0.1 M CuSO₄ solution and vacuum was applied in order to allow the copper sulphate to penetrate into the porous graphite structure. Afterwards, the electrodes were dipped into diluted Na₂S solution for 20 minutes for CuS precipitation. The electrodes were subsequently stored in an evacuated desiccator to allow the electrodes to dry. To confirm the presence of CuS, Raman mappings of the electrodes were recorded prior to their use (see Fig. SI-1A and B in the ESI†).

2.2 Biofilm cultivation

The electrochemically active biofilms in this study were cultivated in a standard growth medium using 10 mM of sodium acetate as the sole carbon source.⁸ The growth medium contained phosphate buffer at pH 7 (Na₂HPO₄ (4.33 g L^{−1}), NaH₂PO₄·H₂O (2.69 g L^{−1}), NH₄Cl (0.31 g L^{−1}), KCl (0.13 g L^{−1}), trace metal (12.5 mL L^{−1}) and vitamin (12.5 mL L^{−1}) solutions. The trace metal solution for the standard experiments (presence of sulphate) contained the following metal salts: MgSO₄·7H₂O (3.0 g L^{−1}), MnSO₄·H₂O (0.45 g L^{−1}), FeSO₄·7H₂O (0.1 g L^{−1}), ZnSO₄ (0.1 g L^{−1}), CuSO₄ (6.4 mg L^{−1}), AlK(SO₄)₂ (10 mg L^{−1}). For preparing a sulphate-free growth medium, all metal sulphates were replaced by the respective chloride or nitrates. To achieve anaerobic conditions, the solution was purged with nitrogen for at least 20 min prior to use.

Acetate fed, *Geobacter* dominated electrochemically active biofilms,⁹ enriched at graphite rod electrodes using primary wastewater of the wastewater treatment plant Steinhof, Braunschweig (Germany), served as the microbial inoculum for all mixed culture experiments. The biofilms were scratched off from their electrodes into a falcon tube filled with 3 mL buffer solution using a sterile spatula and were dispersed with a vortex mixer (Vortex Genie 2, Scientific Industries, USA) until no clump fractions were left. Afterwards, the suspension was used as inoculum to cultivate biofilms on the studied electrodes.

For biofilm cultivation, a constant potential of −0.2 V was applied. The cultivation experiments were conducted under semi-batch conditions for an average of 6 batch cycles, corresponding to a duration of about 50 days. The currents were recorded for each electrode, with the maximum electrode current of the individual batch cycles serving as a key parameter, reflecting the catalytic biofilm performance without substrate limitation. These maximum currents can be considered as values that would be achieved in continuous mode experiments with a short hydraulic retention time of the substrate solution.

All geometric current densities in this manuscript refer to the geometric surface area of the individual biofilm electrodes. The calculation of relative current densities is based on the simultaneous use of at least two working electrodes in every cultivation experiment, where one graphite electrode always served as the benchmark electrode. The relative current density was calculated as the ratio of the geometric current density of the target biofilm electrode (j_{target}) and the geometric current density of the graphite based biofilm electrode (j_{graphite}):

$$j_{\text{relative}} = \frac{j_{\text{target}}}{j_{\text{graphite}}} \quad (1)$$



Based on this calculation, the relative current density of biofilm electrodes based on unmodified graphite is 1.

All biofilm cultivation experiments were performed at least in triplicates.

2.3 Chemical and spectroscopic analysis

Elementary biofilm analysis was performed by inductively coupled plasma optical emission spectrometry (ICP-OES, Varian, Germany). For this purpose, mature biofilms were carefully removed from the respective electrode surface and were rinsed into quartz beakers with pure water. After drying overnight at 100 °C and weighing, they were digested in 40 mL of 65% nitric acid (Fischer Scientific UK, Loughborough, UK), evaporated to near dryness, followed by reuptake in 20 mL 65% nitric acid, re-evaporation to near-dryness, and the uptake in 15 mL of a 10% nitric acid overnight. The ICP-OES analysis was carried out using the acidic sample solution.

For speciation analysis of electrode surfaces and biofilm components we used Raman spectroscopy (InVia REFLEX, Renishaw®, UK) in combination with a laser wavelength of 532 nm (Nd:Yag laser). Microscopic surface images and 3D surface mapping were measured using an adapted Leica-Microscope DM 2700 with a depth resolution of 20 µm at 50 fold magnification.

3. Results and discussion

3.1 Biofilms grown on copper show a superior bioelectrochemical performance

Acetate fed, mixed culture anodic biofilms, usually dominated by *Geobacter sulfurreducens* or *anodireducens*, can be considered as the model electrogenic biofilms for a majority of MFC/MET studies.^{10–12} Fig. 1A shows the bioelectrocatalytic current generation of such biofilms, grown under semi-batch conditions on copper and on graphite, over a cultivation period of 50 days.

The depicted current density pattern is typical for the applied experimental conditions: Thus, the use of preselected bacteria^{13,14} leads to a short incubation time, with stationary conditions reached within ten days after inoculation. A subsequent decrease in current density is observed, which is likely due to an increasing mass transfer limitation by the steadily growing biofilm.^{15,16} The biofilms grown on graphite electrodes (black symbols) reach a maximum geometric current density of $0.87 \pm 0.14 \text{ mA cm}^{-2}$, which is in good agreement with literature values for flat, 2D electrodes.^{17,18} Previous studies suggested that a further increase of the bioelectrochemical substrate turnover of these biofilms is restricted by increasing proton and substrate gradients within the biofilms.¹⁹ Yet, as shown in Fig. 1A (red symbols) for biofilm grown on copper, geometric current density values of $1.59 \pm 0.23 \text{ mA cm}^{-2}$ can be reached, representing an increase in current density by 82% as compared to graphite electrodes.

To compare the performance of graphite and copper-based biofilm electrodes at different stages of biofilm growth and maturation, we normalized the geometric current densities presented in Fig. 1A against the values of graphite biofilm electrodes.

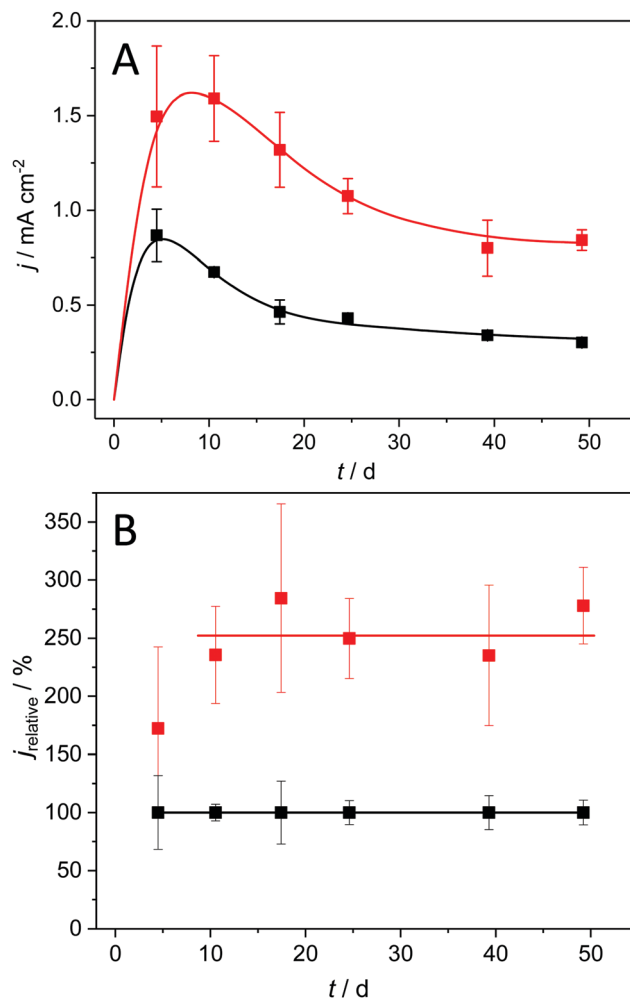


Fig. 1 Development of the anodic bioelectrocatalytic currents at mixed culture biofilms grown in semi-batch experiments using graphite (black symbols) and copper (red symbols) as electrode materials. (A) Geometric current densities; (B) relative current densities.

The resulting relative currents, which are depicted in Fig. 1B, clearly show that the copper electrodes retain their superior bioelectrocatalytic performance over the entire experimental duration, with an average relative current of 256%.

3.2 Biofilms grown on copper contain copper sulphides

In order to understand the reasons for performance difference between graphite and copper-based biofilms, we analysed the respective biofilms using Raman spectroscopy. As shown in Fig. 2A, at Raman shifts above 500 cm^{-1} the spectroscopic features of copper and graphite based biofilms are virtually identical. Based on the specific composition of the cellular components, bacterial species can be identified by means of their distinct Raman spectra.^{20,21} Since previous studies in our lab showed a dominance of *Geobacter* species under the chosen environmental conditions,²² we compared the Raman spectra depicted in Fig. 2A with the Raman spectrum of a pure culture of *Geobacter sulfurreducens* (Fig. 2B, black spectrum). All spectra show an almost 100% agreement – including their fine structures.



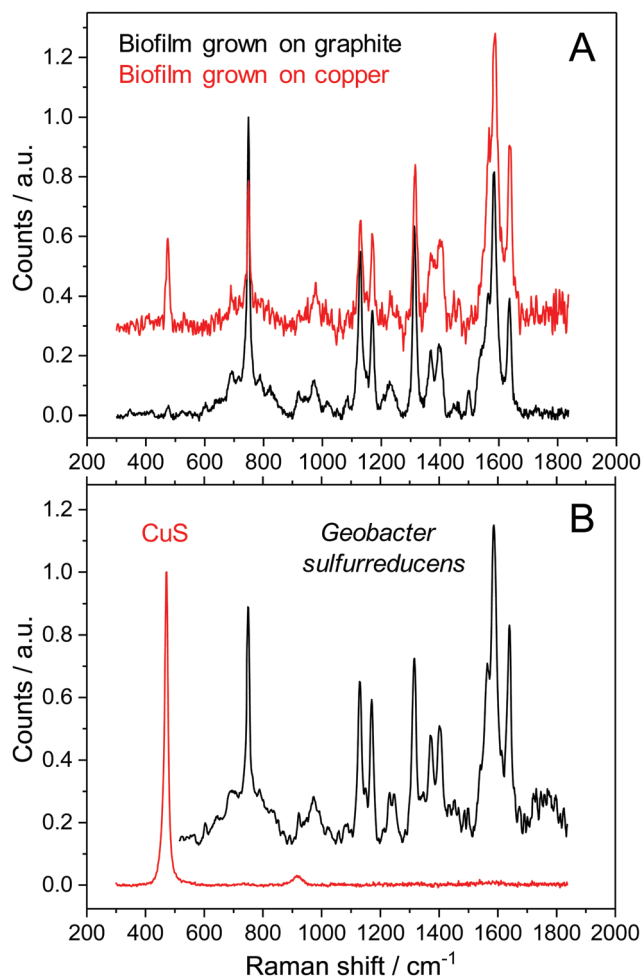


Fig. 2 (A) Raman spectra of a graphite-based biofilm electrode (black) and a copper-based biofilm electrode (red); (B) Raman spectra of a comparable *G. sulfurreducens* pure culture biofilm electrode (black) and of chemically precipitated CuS (red).

This accordance indicates a strong dominance of *G. sulfurreducens* or of a closely related *Geobacter* species in the studied biofilms.

There is, however, a clear difference at Raman shifts below 500 cm^{-1} : copper-grown biofilms show an additional signal at a Raman shift of 473 cm^{-1} . A comparative analysis of different sparingly soluble copper salts that may have been formed by precipitation of copper ions by components in the microbial growth medium (see Fig. SI-02, ESI† for further information) reveals a strong consistency with copper(II) sulphide, CuS (Fig. 2B, red spectrum). A Raman mapping of CuS thereby revealed the distribution of this species throughout the entire biofilm (Fig. 3).

In order to confirm this finding, we performed comparative ICP-OES based elementary analyses of biofilms grown on copper and on graphite electrodes. The results, presented in Table 1, reveal the occurrence of significant amounts of copper and sulphur in the copper-grown biofilms, constituting about 56% of the total biofilm mass. The relative amount of both species leads to a stoichiometry of the copper sulphide species of Cu_2S – corresponding to copper(I) sulphide.

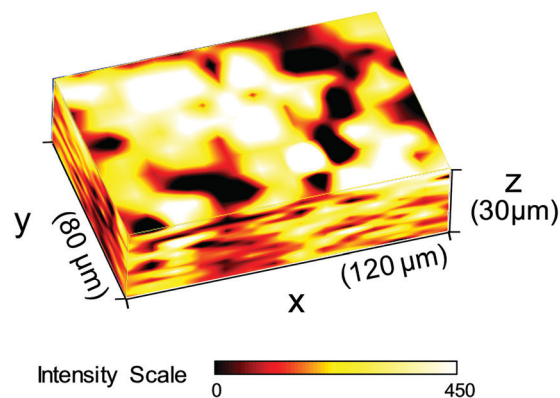


Fig. 3 Raman mapping of a copper-based biofilm electrode, performed at a Raman shift of 473 cm^{-1} .

Table 1 ICP-OES based elementary analysis of biofilms grown on graphite and copper (all data in $\mu\text{g cm}^{-2}$)

| | Biofilms@graphite | Biofilms@copper |
|------------------|-------------------|-----------------------------------|
| Total dry weight | 5994 ± 597 | 13196 ± 2108 |
| Ca | 3.1 ± 1.9 | 1.4 ± 0.02 |
| Cu | 5.3 ± 4.0 | 5904 ± 1325 |
| Fe | 6.3 ± 0.05 | 19.0 ± 1.6 |
| K | 16.2 ± 0.73 | 17.0 ± 1.5 |
| Mg | 12.6 ± 1.5 | 13.3 ± 1.4 |
| Mn | 1.6 ± 0.05 | 2.3 ± 0.05 |
| Na | 113.9 ± 6.9 | 80.0 ± 2.0 |
| P | 131.2 ± 8.1 | 109.8 ± 2.5 |
| S | 29.7 ± 1.8 | 1491 ± 327 |
| Zn | 1.1 ± 0.2 | 6.7 ± 1.9 |

The above findings provoke a series of questions: Where do the sulphide ions originate, when they were not contained in the original growth medium? In order to form copper sulphides, metallic copper needs to become oxidized. What is the thermodynamic basis for the oxidation of this metal? And why do we see two different copper sulphide species – CuS and Cu_2S ?

Answers to these questions are provided in Fig. 4: The presence of sulphide ions in the biofilms can be explained by the microbial reduction of sulphate ions, contained in the growth medium (see composition in the experimental part), by sulphate reducing bacteria such as *Geobacter sulfurreducens*.²³ In a bioelectrochemical system, this reaction is a generally unwanted side reaction that reduces the number of electrons transferrable from an electrogenic bacterium to the anode *via* extracellular electron transfer, EET.

Using the solubility equilibria for copper(I) sulphide ($K_{\text{SP,Cu}_2\text{S}} = a_{\text{Cu}^+}^2 a_{\text{S}^{2-}}$) and copper(II) sulphide ($K_{\text{SP,CuS}} = a_{\text{Cu}^{2+}} a_{\text{S}^{2-}}$), the Nernst equations for the redox transitions of copper in the presence of sulphide ions can be derived (eqn (2)–(4)):

$$E_{\text{Cu/Cu}^+} = E_{\text{Cu/Cu}^+}^{\theta} + \frac{RT}{F} \ln \left(\frac{K_{\text{SP,Cu}_2\text{S}}^{1/2}}{a_{\text{S}^{2-}}^{1/2}} \right) \quad (2)$$

$$E_{\text{Cu/Cu}^{2+}} = E_{\text{Cu/Cu}^{2+}}^{\theta} + \frac{RT}{2F} \ln \left(\frac{K_{\text{SP,CuS}}}{a_{\text{S}^{2-}}} \right) \quad (3)$$



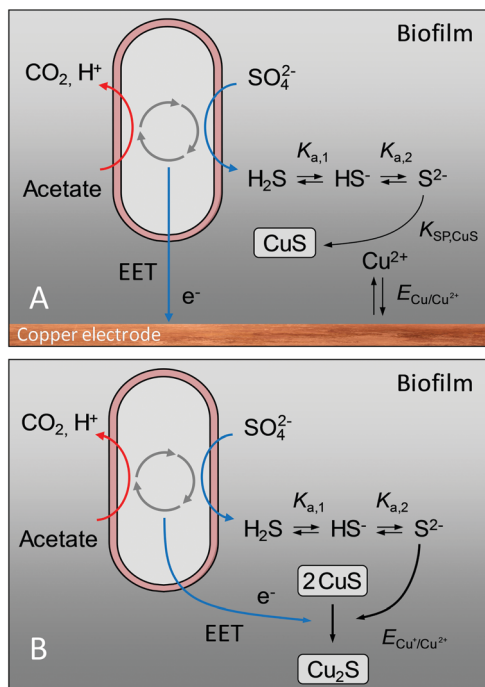


Fig. 4 Schematic illustration of the proposed processes that lead to (A) the formation of copper(II) sulphide, CuS, at the biofilm–copper interface and (B) the conversion of CuS to Cu₂S in the bulk of the electrochemically active biofilm.

$$E_{\text{Cu}^+/\text{Cu}^{2+}} = E_{\text{Cu}^+/\text{Cu}^{2+}}^{\theta} + \frac{RT}{F} \ln \left(\frac{K_{\text{SP,CuS}}}{K_{\text{SP,Cu}_2\text{S}}^{1/2} a_{\text{S}^{2-}}^{1/2}} \right) \quad (4)$$

The growth medium used in our experiments contained about $2 \times 10^{-4} \text{ mol L}^{-1}$ of sulphate ions (see experimental part). Assuming that in the presence of 10 mM electron donor (acetate) the entire sulphate pool in the biofilm may become reduced, a total sulphide level of $2 \times 10^{-4} \text{ mol L}^{-1}$ can be estimated. It is important to point out that the total sulphide concentration is not equal to the concentration of deprotonated sulphide ions (S^{2-}). Here, the protonation equilibria of the $\text{H}_2\text{S}/\text{HS}^-/\text{S}^{2-}$ system with the acid dissociation constants $\text{p}K_{\text{a}1} = 6.99$ and $\text{p}K_{\text{a}2} = 12.35$ have to be taken into account. For the determination of $a_{\text{S}^{2-}}$, we used a graphical tool – the Bjerrum plot (see Fig. SI-03, ESI†). Assuming an average biofilm pH of 6.5,²⁴ the activity of deprotonated sulphide ions was determined as $a_{\text{S}^{2-}} = 1.8 \times 10^{-11} \text{ mol L}^{-1}$. Using this value in combination with the solubility products of Cu₂S and CuS

(see Table 2) and the above Nernst equations (eqn (2)–(4)), the redox potentials of copper in the presence of sulphide can be calculated. Table 2 clearly shows that, based on the extremely low solubility of the copper sulphides, the presence of sulphide ions leads to a strong shift of the oxidation potential of all copper redox transitions towards negative values, transforming copper into a reactive metal species.

Based on the calculated redox potentials provided in the right column of Table 2, an applied electrode potential of -0.2 V (vs. Ag/AgCl) during turnover conditions favours the formation of CuS at the biofilm–electrode interface. During cell proliferation, which in *Geobacter* biofilms takes place close to the electrode–biofilm interface,²⁶ older cell layers including incorporated insoluble compounds like CuS, are moved outwards. This process explains the occurrence of copper sulphide in the entire biofilm. With increasing distance from the electrode surface, the redox potential in electroactive biofilms shifts to negative values. Thus, as shown by Babauta *et al.*²⁴ for bioelectrochemical turnover conditions, the redox potential inside *Geobacter* biofilms dropped from $+0.45 \text{ V}$ – the applied electrode potential – to -0.5 V within $10 \mu\text{m}$ distance from the electrode surface, further decreasing with increasing distance to the electrode surface. At such potentials, copper(I) sulphide, Cu₂S, becomes the thermodynamically favoured species. The bulk stoichiometry of Cu₂S, as found by ICP-OES analysis, can thus be explained by a microbially induced reduction of Cu(II) to Cu(I), as illustrated in Fig. 4B. The reduction may not be thorough, so that CuS remains spectroscopically detectable throughout the biofilm (Fig. 3).

3.3 Copper sulphides promote electron transfer

The ICP-OES based elementary analysis, depicted in Table 1, yields an interesting detail: when subtracting the masses of copper and sulphur from the total surface related dry mass, copper and graphite-based biofilms possess a virtually identical dry mass of 5801 and 5994 $\mu\text{g cm}^{-2}$, respectively. We can thus assume that the total biomass is unaffected by the choice of the electrode material.

Physical insights into the enhanced bioelectrochemical activity of the copper-based biofilms are provided by the steady state voltammograms, depicted in Fig. 5. The fundamental shape of the turnover voltammograms of copper and graphite-based biofilms (Fig. 5A) is highly similar, with identical inflection points, as illustrated by the maxima of the first derivatives of the voltammetric curves (Fig. 5B). Thus, the fundamental EET mechanisms remain unaffected by the nature of the underlying electrode material.

Table 2 Standard potentials of Cu(0)/(I), (0)/(II) and (I)/(II); the solubility products of Cu(I) and (II) sulphide and the redox potentials of copper in the presence of sulphide

| | Standard potentials/V (vs. SHE) | Solubility products ²⁵ | Redox potentials in the presence of sulphide/V vs. SHE (vs. Ag/AgCl) ^a |
|--------------|---|--|---|
| Cu/Cu(I) | $E_{\text{Cu}^0/\text{Cu}^+}^{\theta} = 0.521$ | $K_{\text{SP,Cu}_2\text{S}} = 2.5 \times 10^{-48}$ | $E_{\text{Cu}/\text{Cu}^+} = -0.605 (-0.802)$ |
| Cu/Cu(II) | $E_{\text{Cu}^0/\text{Cu}^{2+}}^{\theta} = 0.342$ | $K_{\text{SP,CuS}} = 6 \times 10^{-37}$ | $E_{\text{Cu}/\text{Cu}^{2+}} = -0.437 (-0.634)$ |
| Cu(I)/Cu(II) | $E_{\text{Cu}^+/\text{Cu}^{2+}}^{\theta} = 0.153$ | | $E_{\text{Cu}^+/\text{Cu}^{2+}} = -0.278 (-0.475)$ |

^a Calculated for a total concentration of sulphide species of $2 \times 10^{-4} \text{ mol L}^{-1}$, pH 6.5, $a_{\text{S}^{2-}} = 1.8 \times 10^{-11} \text{ mol L}^{-1}$ and a temperature of $35 \text{ }^{\circ}\text{C}$.



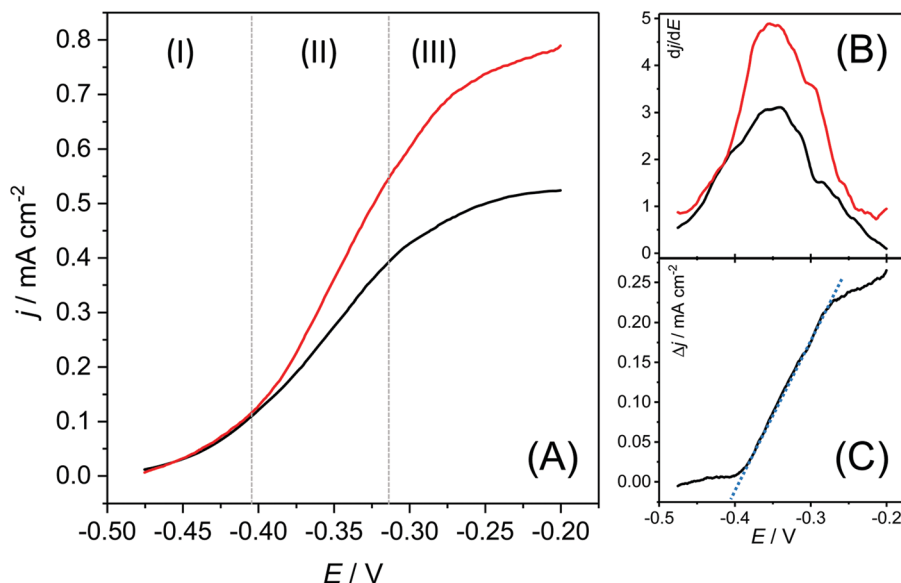


Fig. 5 (A) Biofilm turnover steady state voltammograms of a copper-based biofilm electrode (red curve) and a graphite-based electrode (black curve) calculated from the respective cyclic voltammograms via averaging of forward and backward currents (see also Fig. SI-04, ESI†). The scan rate was 1 mV s^{-1} . (B) First derivatives of both voltammograms. (C) Difference of the current densities of the voltammetric curves, calculated as $\Delta j = j_{\text{copper electrode}} - j_{\text{graphite electrode}}$.

The voltammograms show three distinct sections: The onset of the voltammetric curve (Section (1)) is dominated by the kinetics of the interfacial electron transfer (Butler–Volmer–Kinetics).²⁷ Here both biofilm electrodes show a very similar behaviour. The similarity can be expected, since the voltammetry of *Geobacter* biofilms grown on graphite is already electrochemically highly reversible.²⁸ A major difference appears in Section (2): here, the copper-based electrode shows a considerably steeper slope than the graphite electrode. In order to analyse the difference in both curves, we subtracted both steady state voltammograms from each other (Fig. 5C). The slope in the resulting residual voltammogram shows an almost perfect linearity indicating that the difference in both voltammograms lies predominantly in the conductivity of the biofilms. From the slope in the potential window between -380 and -280 mV it can be estimated that the ohmic resistance of the copper-grown biofilm is about $535 \Omega \text{ cm}^2$ lower than that of the graphite-based biofilm. The increased conductivity of the copper-based biofilm can be attributed to the incorporated copper sulphides. Thus, with a maximum specific conductivity of $\sigma_{\text{Cu}_2\text{S}} = 10^2 \text{ S cm}^{-1}$,²⁹ copper(I) sulphide is a semiconductor, whereas, with a conductivity of up to $\sigma_{\text{CuS}} = 10^5 \text{ S cm}^{-1}$,²⁹ copper(II) sulphide possesses metallic conductivity.

In Section (3) in the voltammetric curve, the biofilm reaches the stationary current density. In alignment with the data from the constant potential measurement (see Fig. 1), the copper-based biofilm clearly outperforms the graphite-based biofilm electrodes. This finding can be explained by means of the metabolic stratification within the biofilms. Thus, Chadwick *et al.* recently showed that the metabolic activity of *G. sulfurreducens* biofilms is restricted to the cells within a $10 \mu\text{m}$ biofilm layer most adjacent to the electrode surface.²⁶ A steep drop of the redox potential to values of about -0.5 V beyond this biofilm layer²⁴ leads to redox conditions unfavourable for bioelectrochemical

substrate oxidation. Hence, all cells more distant to the electrode surface remain metabolically inactive. The reduced ohmic resistance of copper-based biofilms may now lead to a less steep potential drop in the biofilm, allowing the active cell layer thickness to grow. The resulting increase in the number of active cells leads to a proportional increase in bioelectrochemical current generation.

In order to evaluate to which extent the copper oxidation itself may contribute to the increased bioelectrochemical current densities, we calculated the copper oxidation current based on the amount of copper accumulated in the biofilm in the course of the biofilm cultivation (see Table 1). Assuming that copper oxidation takes place in the turnover phases of the semi-batch cultivation, a current density of about $8 \mu\text{A cm}^{-2}$ was calculated. This current density is insignificant compared to the overall increase in current density of the copper-based biofilms.

One may argue that an increased microscopic surface area of the electrode material due to the microbially induced copper corrosion or due to the copper sulphide deposition can be made responsible for the enhanced biofilm electrogenicity. However, as we have recently demonstrated, electrode structures that are smaller in dimension than the thickness of electroactive biofilm do not lead to a significant or lasting current enhancement.^{15,30}

To confirm the role of copper sulphides in the promotion of the bioelectrochemical turnover, we performed two different experiments. In experiment 1, we removed all sulphates from the growth medium by replacing the respective salts by nitrate and chloride components and used the growth medium for cultivation. Under these conditions, sulphide formation, as depicted in Fig. 4, would not be possible. As shown in Fig. 6, red curve, the relative current of the resulting copper-based biofilm electrode was now exactly as high as that of the corresponding graphite electrodes. This proves that the presence of copper sulphides is responsible and essential for the current increase.



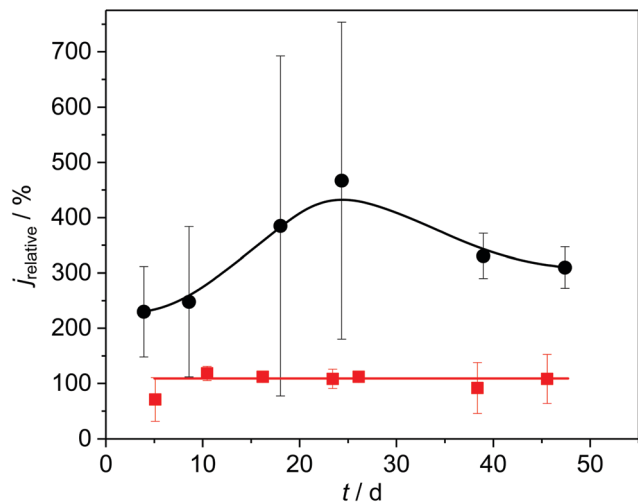


Fig. 6 Relative bioelectrocatalytic currents of copper-based biofilms electrodes. Red symbols: biofilms grown in sulphur-free medium; Black symbols: biofilms grown on copper sulphide modified copper electrodes.

In experiment 2 we deposited copper(II) sulphides onto copper electrodes prior to biofilm cultivation. As the black curve in Fig. 6 illustrate, this measure had an additional positive effect on the current generation. Thus, the relative current increased to a value of 460% (corresponding to a geometric current density of up to 2 mA cm^{-2}) at day 25 after inoculation (see Fig. SI-05, ESI†).

Is the use of metallic copper thus essential for a current density enhancement? To answer this question, we deposited copper sulphide onto graphite electrodes and used these modified electrodes for biofilm cultivation. The development of the resulting bioelectrocatalytic current generation is depicted in Fig. 7.

Fig. 7A shows that biofilm electrodes based on CuS-modified graphite reach current densities of 234% relative to unmodified graphite. Hereby, as shown in Fig. 7B, the current enhancement is not an immediate effect but increases over the first 10 days of operation, whereas the biofilm grown on unmodified graphite already loses activity during that period. This supports the hypothesis that the penetration of the conductive copper sulphide(s) into the biofilm during the biofilm growth increases the thickness of the metabolically active cell layer, an effect that requires the biofilm to reach a certain total thickness. These results clearly highlight that a copper sulphide deposition on graphite can substitute the use of metallic copper as electrode base material, thus creating a high-performance but low-cost electrode material.

The large standard deviation in Fig. 6 and 7 (as well as Fig. SI-05, ESI†) can be attributed to variations in the quality of the copper sulphide deposition. The inhomogeneity becomes visible in the Raman mappings (Fig. SI-01A and B, ESI†). Here we see a considerable potential for a further improvement *via* more sophisticated deposition procedures. This clearly needs to be addressed in a follow-up study.

When using copper sulphide as an electrode modifier, a possible oxidative decomposition (*via* oxidation of the sulphide ion) has to be taken into account. The redox potential for this process depends on environmental variables such as the

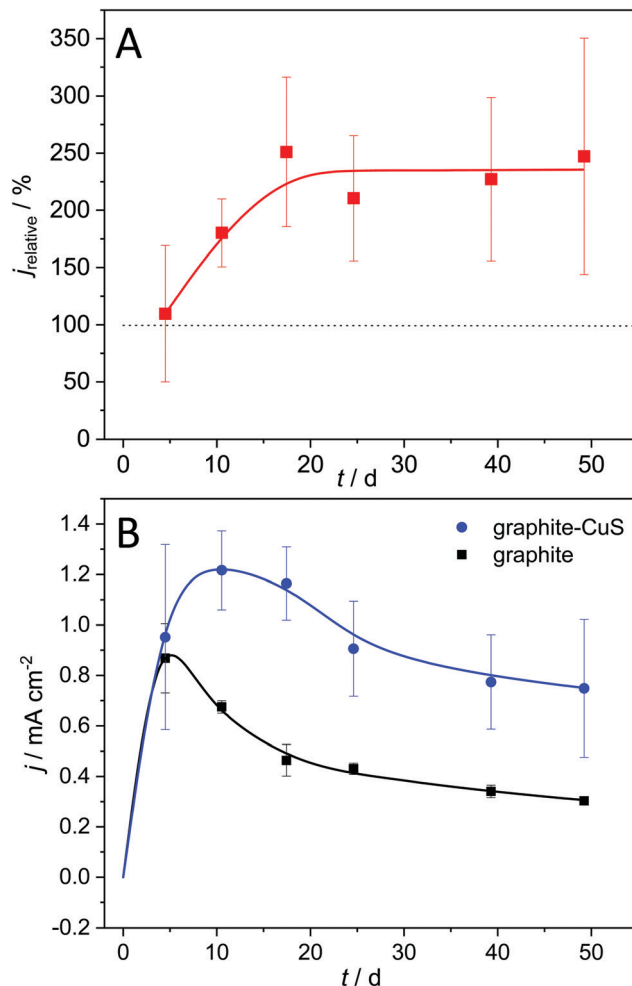


Fig. 7 (A) Development of the relative anodic current densities of mixed culture biofilms grown at CuS modified graphite electrodes; (B) underlying geometric current densities of the biofilms grown at CuS modified graphite (blue symbols) versus unmodified graphite (black symbols).

surrounding pH and can be derived from a respective Pourbaix diagram see, *e.g.*, ref. 31. We have determined the threshold for the oxidation process using cyclic voltammetry. Thus, as depicted in Fig. SI-06 (ESI†) the onset potential is about 0.1 V (vs. Ag/AgCl). For a microbial fuel cell, this onset potential would have to be seen in relation to the operating potential of an oxygen cathode. In Fig. SI-06 (ESI†), the oxygen reduction reaction at a platinum electrode is shown to have an onset potential of 0.1 V, which coincides with the copper sulphide oxidation (as well as with the oxidation of metallic copper oxidation, as shown in ref. 6). In an operating fuel cell, when avoiding extended short circuit conditions, the anode potential usually remains well below the cathode potential. It can thus be assumed that copper sulphide can be safely used well below its oxidation potential.

4. Conclusions

In this study, we have shown that using copper electrodes systematically and substantially enhances the current generation



of *G. sulfurreducens* biofilm electrodes. The current enhancement is based on the bioelectrochemical formation of copper(I) and copper(II) sulphides, which are incorporated into the biofilms – serving as conductive network, supporting extracellular electron transfer. Thereby, the extremely low solubility of the copper sulphides prevents antimicrobial effects of the copper compounds. As compared to graphite-based biofilm electrodes, the current density increased to an average of 237%. An additional electrochemical copper sulphide deposition prior to the biofilm cultivation further increased the current densities to a maximum relative current of 460% after 25 days of operation.

The use of metallic copper can be completely avoided when CuS-modified graphite is used for biofilm cultivation. With this strategy, relative currents of about 234% are achieved, rendering this composite electrode material a promising material for bioelectrochemical applications.

Conflicts of interest

There are no conflicts to declare.

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References

- G. Borkow and J. Gabbay, *Curr. Chem. Biol.*, 2009, **3**, 272–278.
- J. R. Harris, *Econ. Hist. Rev.*, 1966, **19**, 550–568.
- G. Grass, C. Rensing and M. Solioz, *Appl. Environ. Microbiol.*, 2011, **77**, 1541–1547.
- K. Page, M. Wilson and I. P. Parkin, *J. Mater. Chem.*, 2009, **19**, 3818–3831.
- G. Borkow, *Curr. Chem. Biol.*, 2012, **6**, 93–103.
- A. Baudler, I. Schmidt, M. Langner, A. Greiner and U. Schröder, *Energy Environ. Sci.*, 2015, **8**, 2048–2055.
- L. Fotouhi and M. Rezaei, *Microchim. Acta*, 2009, **167**, 247–251.
- W. E. Balch, G. E. Fox, L. J. Magrum, C. R. Woese and R. S. Wolfe, *Microbiol. Rev.*, 1979, **43**, 260–296.
- A. Baudler, S. Riedl and U. Schröder, *Front. Energy Res.*, 2014, **2**, 1–6.
- F. Harnisch, C. Koch, S. A. Patil, T. Huebschmann, S. Mueller and U. Schroeder, *Energy Environ. Sci.*, 2011, **4**, 1265–1267.
- C. Koch and F. Harnisch, *ChemElectroChem*, 2016, **3**, 1282–1295.
- X. Zhu, M. D. Yates, M. C. Hatzell, H. Ananda Rao, P. E. Saikaly and B. E. Logan, *Environ. Sci. Technol.*, 2014, **48**, 1352–1358.
- Y. Liu, F. Harnisch, K. Fricke, R. Sietmann and U. Schröder, *Biosens. Bioelectron.*, 2008, **24**, 1012–1017.
- S. Riedl, R. K. Brown, S. Klöckner, K. J. Huber, B. Bunk, J. Overmann and U. Schröder, *ChemElectroChem*, 2017, **4**, 3081–3090.
- C. Moß, A. Behrens and U. Schröder, *ChemSusChem*, 2020, **13**, 582–589.
- D. Sun, S. Cheng, A. Wang, F. Li, B. E. Logan and K. Cen, *Environ. Sci. Technol.*, 2015, **49**, 5227–5235.
- C. I. Torres, R. Krajmalnik-Brown, P. Parameswaran, A. K. Marcus, G. Wanger, Y. A. Gorby and B. E. Rittmann, *Environ. Sci. Technol.*, 2009, **43**, 9519–9524.
- D. Pocaznoi, A. Calmet, L. Etcheverry, B. Erable and A. Bergel, *Energy Environ. Sci.*, 2012, **5**, 9645–9652.
- C. I. Torres, A. K. Marcus and B. E. Rittmann, *Biotechnol. Bioeng.*, 2008, **100**, 872–881.
- R. M. Jarvis and R. Goodacre, *Anal. Chem.*, 2004, **76**, 40–47.
- L. Ashton, K. Lau, C. L. Winder and R. Goodacre, *Future Microbiol.*, 2011, **6**, 991–997.
- I. Schmidt, A. Pieper, H. Wichmann, B. Bunk, K. Huber, J. Overmann, P. J. Walla and U. Schröder, *ChemElectroChem*, 2017, **4**, 2515–2519.
- Z. Qian, H. Tianwei, H. R. Mackey, M. C. M. van Loosdrecht and C. Guanghao, *Water Res.*, 2019, **150**, 162–181.
- J. T. Babauta, H. D. Nguyen, T. D. Harrington, R. Renslow and H. Beyenal, *Biotechnol. Bioeng.*, 2012, **109**, 2651–2662.
- K. Rauscher, J. Voigt, I. Wilke, K. T. Wilke and R. Friebe, *Chemische Tabellen und Rechentafeln für die analytische Praxis*, Verlag Europa-Lehrmittel, Haan-Gruiten, 2000.
- G. L. Chadwick, F. J. Otero, J. A. Gralnick, D. R. Bond and V. J. Orphan, *Proc. Natl. Acad. Sci. U. S. A.*, 2019, **116**, 20716–20724.
- A. J. Bard and L. R. Faulkner, *Electrochemical Methods – Fundamentals and Applications*, John Wiley & Sons, Inc., 2001.
- K. Fricke, F. Harnisch and U. Schröder, *Energy Environ. Sci.*, 2008, **1**, 144–147.
- D. R. Lide, *CRC handbook of chemistry and physics*, CRC Press, Boca Raton, FL, 84th edn, 2003.
- C. Moß, S. A. Patil and U. Schröder, *Front. Energy Res.*, 2019, **7**, 1–10.
- I. Puigdomenech and C. Taxén, *Thermodynamic Data for Copper, Implications for the Corrosion of Copper Under Repository Conditions*, 2000.

