


 Cite this: *RSC Adv.*, 2020, 10, 1971

Exposure risk assessment of nine metal elements in Chongqing hotpot seasoning

 Lianji Zheng,^{ab} Qi Zhang,^a Zhi Li,^b Rui Sun^a and Geng Zhong^{ID}*^a

Atomic absorption spectrometry (AAS) and atomic fluorescence spectrometry (AFS) were used to analyze the contents of nine metal elements (Pb, As, Hg, Cd, Cr, Fe, Mn, Cu, Zn) in 100 groups of Chongqing hotpot seasoning (CHS). Meanwhile, Crystal Ball software based on Monte Carlo simulation technology was used to assess the exposure risk of the nine metal elements in CHS for people of different ages in Chongqing. In general, the average Hazard Index (HI) of the nine metal elements is $0.306 < 1$, indicating no non-carcinogenic risks from these nine elements for inhabitants of Chongqing under the current consumption level of CHS. Children (ages 7–13) and adult women have higher chronic daily intake (CDI) than adult males. The carcinogenic risk of Pb, As and Cd are within the acceptable risk level (10^{-6} to 10^{-4}). The sensitivity analysis suggests that the contents of the nine metal elements and daily intake (PIR) in CHS were positively correlated with the risk index, while the body weight was negatively correlated with the risk index. This study provides a scientific basis for guiding the safe consumption of Chongqing hotpot, and provides a theoretical basis for the development of safety-compliant CHS quality standards.

Received 30th November 2019

Accepted 1st January 2020

DOI: 10.1039/c9ra10028h

rsc.li/rsc-advances

1. Introduction

The pursuit of gourmet food is very important in Chinese cuisine culture. As a branch of Chinese cuisine culture, the Hotpot culture is becoming more and more popular in China, especially the spicy hotpot originating in Chongqing.¹ The production of Hotpot seasoning as the main source of hotpot flavor, with the increasing popularity of hotpot, has been industrialized. Chongqing hotpot seasoning (CHS) is cooked by boiling animal oil or vegetable oil with spicy foods (such as chilli, zanthoxylum, ginger, garlic) and other spices by fire. When eating, the hotpot seasoning is put into a metal pot and boiled with water to make the spicy, hot spicy and delicious taste spill over and dissolve in the soup, then cook the food materials.² Finally, cooked food is dipped in sauce and eaten. This whole process named as “Tang huoguo” in China.³

At present, CHS has been sold at home and abroad, and will be produced in the world in the future. Therefore, it is necessary to establish quality standards for safety-compliant CHS. The content analysis of metal elements in CHS, which may come from raw materials, productive process and use phase, is an important part of constructing standards and ensuring food safety. In China, the safe contents of metals such as Pb and As in CHS are determined according to the national food safety standards,⁴ while other metals such as Hg, Cd, Cr, Mn, Cu, Zn and Fe have no relevant limit standards, and there is no relevant

research report on the content and exposure risk of metal elements in the CHS.

Metal elements, especially heavy metals, has the characteristics of toxicity, bioaccumulation and non-degradability.⁵ It has been confirmed that the main reason for the contamination of food by metal is man-made activities.⁶ Heavy metals accumulate in the soil, leading to the decrease of crop yield and quality.⁷ Eating contaminated crops will have a negative impact on humans, animals and ecosystems.⁸

The metal elements in CHS are mainly from the raw materials. At present, there are many reports about metal enrichment in food at home and abroad, but different food in different countries and regions are polluted by metals differently.⁹ Jafarian-Dehkordi A.¹⁰ detected vegetables (cucumbers, tomatoes, cabbage, lettuce, potatoes, onions, carrots, leeks, dill, spinach, cilantro, parsley) grown in the suburbs of Isfahan for Cd, Cr and Pb levels all above the WHO/FAO limit. Arora M.¹¹ showed that vegetables irrigated with wastewater from different sources had large accumulation of Zn, Mn and Fe. Luo C.¹² found that local soil and vegetables that dispose of e-waste carelessly were heavily polluted by metals (Cd, Cu, Pb, Zn). In China, the pepper cultivated in Yanqi basin of Xinjiang is heavily polluted with Pb and Zn, while slightly polluted with Cd and Cr.¹³ Chilies grown around the smelter in Guizhou are rich in Cd.¹⁴ There are different levels of Zn and As pollution in Lanzhou vegetable base.¹⁵ The pollution of Cu and Cr in Zanthoxylum base on the suburb of Weinan is serious,¹⁶ and the Pb pollution of chile in Shaanxi exceeded 1 mg kg^{-1} .¹⁷ In general, it is of practical significance to carry out metal content detection in CHS.

^aCollege of Food Science, Southwest University, Chongqing 400715, PR China. E-mail: 26819079@qq.com

^bFood Industry Research Institute of Chongqing, Chongqing 400042, PR China



Exposure assessment is to calculate the data of pollutant exposure level survey from the perspective of human health, so as to directly evaluate the harm of pollutants to residents' health and establish an objective risk assessment model.¹⁸ There are two main risk assessment methods for chemical hazards in the world. One is a risk assessment method based on quantitative model,¹⁹ and the other is a risk assessment method based on uncertainty model (such as Monte Carlo method). Compared with the two methods, the method based on the uncertainty model is more intuitively and more in line with the uncertainty nature of risk, so it has more advantages.²⁰ Fakhri, Yadolah used Monte Carlo method to evaluate the exposure of Pb and Cd in onion and soil from Iran.²¹ Qu C.²² based on the Monte Carlo method, took Qixia mining area as an example to evaluate the health risks of soil contaminated by heavy metals.

In this paper, 100 groups of CHS obtained from the local supermarkets and hotpot restaurants were taken as research objects. Nine kinds of metal element (Pb, As, Hg, Cd, Cr, Fe, Mn, Cu, Zn) were detected. Based on Monte Carlo simulation technology, the Crystal ball risk assessment software was used to evaluate chronic daily intake (CDI), carcinogenic risk, non-carcinogenic risk and sensitivity analysis. The exposure risk of nine metal elements in CHS for people of different ages in Chongqing were evaluated. The risk assessment results of this study provides scientific basis for guiding the safe consumption of Chongqing hotpot, and provides a theoretical basis for the development of safety-compliant CHS quality standards.

2. Experimental

2.1. Materials

One hundred different groups of samples including pre-packaged and fresh spicy hotpot seasoning were purchased from the local market (12 groups) and hotpot restaurants (88 groups). Pb, As, Hg, Cd, Cr, Fe, Mn, Cu and Zn standard solutions were purchased from the National Standard Substances Center. Nitric acid (guarantee reagent), perchloric acid (analytical reagent), and 30% hydrogen peroxide (analytical reagent) were purchased from Chongqing Chuandong Chemical Industry (Group) Co., Ltd.

2.2. Methods

2.2.1. Determination of metal elements. The main paragraph text follows directly on here. Add 1 g CHS, 8 mL HNO₃ and 0.5 mL HClO₄ into the high-pressure digestion tank, and after standing overnight, the mixture was heated for 2.5 h at 120–130 °C.²³ After the solution was clarified, the volume was fixed to 25 mL and was to be measured. Cu, Fe, Zn, Mn were determined according to the methods of flame atomic absorption spectrometry (FAAS) and Pb, Cr, Cd were determined according to the methods of graphite furnace atomic absorption spectrometry (GF-AAS). The above seven elements were determined by ICE3000 atomic absorption spectrometry (ThermoFisher, USA). During GF-AAS analysis, argon was used as inert gas, pyrolytic-coated graphite tubes with a platform and 5 μL matrix modifier were used. The matrix modifier was a mixture of 0.5% (w/v)

ammonium dihydrogen phosphate (H₂PO₄NH₄) and 1% (v/v) HNO₃. Most of the matrix was removed before atomization with less interference. As and Hg were determined by AFS-970 two-channel atomic fluorescence spectrometer (Haiguang Instrument Comp., Beijing, China). During AFS analysis, 5% (v/v) HCl was used as carrier liquid, and a mixture of 2% (w/v) potassium borohydride (NaBH₄) and 0.5% (w/v) sodium hydroxide (NaOH) was used as reducing agent.

In order to ensure the accuracy of data, each sample was repeated three times, and blank and standard substance (GBW-25) (from China National Standards Research Center) were compared. For every 20 samples measured, two standard samples were measured to verify the accuracy of the instrument. The standard curve shows a linear relationship within the concentration range, the regression coefficient (R^2) > 0.999, and the relative standard deviation <10%, indicating that the elemental analysis method is accurate and reliable. The average recovery of metal elements is 90–110%, which verifies the effectiveness of the method. All metal concentrations are measured on a natural weight basis (mg kg⁻¹). The detection limits (LOD) of Pb, As, Hg, Cd, Cr, Fe, Mn, Cu and Zn were 0.02, 0.001, 0.003, 0.001, 0.01, 0.75, 0.2, 0.2 and 1 mg kg⁻¹, respectively.

2.2.2. Consumption data and population weight data sources. Body weight, age, and consumption rate are important parameters for evaluating dietary exposure assessment.²⁴ This study conducted a face-to-face survey of 2000 hotpot consumers in Chongqing, including 1000 males and 1000 females. According to age, the respondents were divided into children group (7–13 years old), adolescents group (14–17 years old), adults group (18–49 years old), and elderly group (50–78 years old). The survey data were processed by SPSS virson19 software. The consumption level of CHS of different groups was shown in Table 1.

2.2.3. Risk assessment

2.2.3.1. Exposure assessment. Exposure assessment is one of the most important components of the risk assessment process.²⁵ In this study, nine metal elements in CHS were respectively used as the single source of chemical substances exposure. The chronic daily intake (CDI) was calculated according to the formula (1) using the Monet Carlo method.^{26,27} The daily CHS intake per capita (PIR) is based on the CHS consumption level survey data of different populations.

$$CDI = \frac{C_f \times PIR \times ABS \times EF \times ED}{bw \times AT} \quad (1)$$

Table 1 CHS consumption level and body weight of different groups

Age	Consumption level (g d ⁻¹)		Body weight (kg)	
	male	female	male	female
7–13	33.4–95.8	30.6–91.6	24.2–46.4	22.2–44.2
14–17	37.7–107.9	29.7–105.3	46.3–68.1	34.7–58.3
18–49	37–134.8	33.3–111.5	52.5–84.3	43–72.2
50–75	40.7–123.1	33.3–107.5	55.7–82.5	46.5–68.9



Table 2 Daily reference dose (RfD) of metal elements in CHS

Metal elements	Reference	RfD (mg per kg per bw per d)
Pb	JECFA, PWVI = 0.025 mg per kg per bw per d	0.0035
As	JECFA, BMDL0.5 = 3 µg per kg per bw per d	0.003
Hg	JECFA, PWVI = 4 µg per kg per bw	0.00057
Cd	JECFA, PTMI = 0.025 mg per kg per bw per month	0.00083
Cr	CNS, UL = 500 µg d ⁻¹	0.0083
Fe	CNS, UL = 40 mg d ⁻¹	0.667
Mn	CNS, UL = 11 mg d ⁻¹	0.183
Cu	CNS, UL = 8 mg d ⁻¹	0.133
Zn	CNS, UL = 40 mg d ⁻¹	0.667

In the formula: CDI is the chronic daily intake, mg per kg per bw per d; C_f is the exposure concentration of 9 metal elements in 100 groups of CHS, mg kg⁻¹; PIR is the daily CHS intake per capita, kg d⁻¹; EF is the exposure frequency, day per year; ED is the exposure duration (determined by different age groups), year; BW is the body weight, kg; AT (AT = ED year * 365 days per year) is the pulling time, d; ABS is the intestinal and gastric absorption coefficient (default: 1).

2.2.3.2. *Non-carcinogenic risk.* Non-carcinogenic risk of element metals in CHS was evaluated by calculating hazard quotient (HQ).²⁸ In the formula (2): HQ is the food exposure safety index of metal elements in CHS, indicating the impact of food safety; RfD is the daily reference dose of metal elements in CHS. When the calculated HQ is equal to or less than 1, it is impossible to produce non-carcinogenic effects; If the HQ is greater than 1, it may have a non-carcinogenic effect on human health.^{25,26}

$$HQ = \frac{CDI}{RfD} \quad (2)$$

Reference dose (RfD) refers to a dose that an individual can be continuously exposed to this level for a long time without being harmed. Joint FAO/WHO Expert Committee on Food Additives (JECFA) recommended tolerable daily intake (TDI) is 0.0035 mg per kg per bw per d for Pb, 0.00083 mg per kg per bw per d for Cd, 0.003 mg per kg per bw per d for As, 0.00057 mg per kg per bw per d for Hg, 0.0083 mg per kg per bw per d for Cr.²⁹ According to the regulations of Chinese Nutrition Society (CNS), the tolerable upper intake levels (UL) of Fe, Zn, Cu and Mn in adults with 60 kg are 40 mg per d, 40 mg per d, 8 mg d⁻¹ and 11 mg d⁻¹ respectively.³⁰ Therefore, the RfD used in the HQ calculation of this study is shown in Table 2.

At the current stage of quantitative risk assessment of multiple chemicals, exposure to two or more chemicals may result in additive and/or interactions effect,^{31,32} so risk-increasing assumptions must be applied. In this study, the risk-increasing assumptions was determined using the hazard index (HI) and calculated according to formula (3).^{33–36}

$$HI = HQ_1 + HQ_2 \dots + HQ_n \quad (3)$$

2.2.3.3. *Carcinogenic risk.* Carcinogenic risk refers to an individual's increased likelihood of developing cancer over a lifetime due to exposure to potentially carcinogenic substances. Formula (4) shows how to calculate the carcinogenic risk

$$\text{Cancer risk} = CDI \times SF \quad (4)$$

In the formula (4): CDI is the same as formula (1), mg per kg per bw per d; SF is the slope factor of carcinogenic elements, mg kg⁻¹ d⁻¹. Environmental protection agency (EPA) recommend that the SF of As, Cd, Pb are 1.5 mg kg⁻¹ d⁻¹, 6.3 mg kg⁻¹ d⁻¹, 0.009 mg kg⁻¹ d⁻¹, respectively.³⁷

2.2.3.4. *Sensitivity analysis.* Sensitivity analysis is an important part of risk assessment and the basis of risk management.³⁸ The influence of each variable parameter in the evaluation model on the simulation results is analyzed, so as to take control measures for the main influencing factors to reduce the risk.³⁹ Crystal Ball risk assessment software was used to analyze the sensitivity of consumers aged 7–78 years to the consumption of nine metal elements in CHS. The main factors considered in this experiment include the content of metal elements in CHS, daily intake of CHS, body weight, the exposure duration and pulling time.

2.3. Data analysis

Firstly, Monte Carlo technology Crystal ball software was used to determine the optimal fitting distribution of the

Table 3 Content and detection rate of nine metal elements in 100 groups of CHS^a

Metal elements	Content of metal element/mg kg ⁻¹	LOD/mg kg ⁻¹	Detection rate/%
Pb	0.01–1.130	0.02	92
As	0.003–0.045	0.001	100
Hg	0.0015–0.002	0.003	0
Cd	0.0005–0.160	0.001	45
Cr	0.005–3.080	0.01	99
Zn	0.500–35.664	1	92
Mn	0.1–11.603	0.2	98
Cu	0.100–5.819	0.2	99
Fe	0.375–74.358	0.75	92

^a Note: undetected results were expressed as 1/2 LOD.



Table 4 Probabilistic estimation of CDI (mg per kg per bw per d) to nine metal elements

	Ages		Mean		50%	75%	95%	Min	Max	
Pb	7–13	Male	4.71×10^{-4}	3.94×10^{-4}	2.41×10^{-4}	5.32×10^{-4}	1.55×10^{-2}	2.97×10^{-6}	2.99×10^{-2}	
		Female	3.17×10^{-4}		1.56×10^{-4}	3.57×10^{-4}	6.87×10^{-3}	1.59×10^{-6}	1.30×10^{-2}	
	14–17	Male	2.11×10^{-4}	2.28×10^{-4}	1.07×10^{-4}	2.37×10^{-4}	7.81×10^{-3}	1.84×10^{-6}	1.51×10^{-2}	
		Female	2.44×10^{-4}		1.24×10^{-4}	2.85×10^{-4}	5.73×10^{-3}	1.54×10^{-6}	1.09×10^{-2}	
	18–49	Male	2.29×10^{-4}	2.29×10^{-4}	1.06×10^{-4}	2.54×10^{-4}	5.51×10^{-3}	6.60×10^{-7}	1.05×10^{-2}	
		Female	2.30×10^{-4}		1.06×10^{-4}	2.55×10^{-4}	7.70×10^{-3}	9.86×10^{-7}	1.17×10^{-2}	
	50–78	Male	2.00×10^{-4}	2.02×10^{-4}	1.02×10^{-4}	2.27×10^{-4}	4.24×10^{-3}	9.24×10^{-7}	8.03×10^{-3}	
		Female	2.05×10^{-4}		1.05×10^{-4}	2.34×10^{-4}	3.63×10^{-3}	1.36×10^{-6}	6.80×10^{-3}	
		All		2.63×10^{-4}	1.31×10^{-4}	2.98×10^{-4}	6.92×10^{-3}	1.48×10^{-6}	1.32×10^{-2}	
	As	7–13	Male	2.56×10^{-5}	2.58×10^{-5}	1.92×10^{-5}	3.41×10^{-5}	1.47×10^{-4}	1.46×10^{-6}	2.40×10^{-4}
			Female	2.60×10^{-5}		1.94×10^{-5}	3.42×10^{-5}	1.62×10^{-4}	1.33×10^{-6}	2.70×10^{-4}
		14–17	Male	1.75×10^{-5}	1.87×10^{-5}	1.37×10^{-5}	2.33×10^{-5}	7.70×10^{-5}	1.56×10^{-6}	1.19×10^{-4}
Female			1.99×10^{-5}		1.50×10^{-5}	2.65×10^{-5}	9.60×10^{-5}	1.40×10^{-6}	1.51×10^{-4}	
18–49		Male	1.83×10^{-5}	1.85×10^{-5}	1.27×10^{-5}	2.39×10^{-5}	1.18×10^{-4}	7.23×10^{-7}	1.96×10^{-4}	
		Female	1.87×10^{-5}		1.32×10^{-5}	2.43×10^{-5}	1.19×10^{-4}	7.04×10^{-7}	1.99×10^{-4}	
50–78		Male	1.66×10^{-5}	1.69×10^{-5}	1.26×10^{-5}	2.22×10^{-5}	7.47×10^{-5}	1.02×10^{-6}	1.15×10^{-4}	
		Female	1.72×10^{-5}		1.31×10^{-5}	2.27×10^{-5}	7.73×10^{-5}	1.39×10^{-6}	1.19×10^{-4}	
		All		2.00×10^{-5}	1.49×10^{-5}	2.64×10^{-5}	1.09×10^{-4}	1.20×10^{-6}	1.76×10^{-4}	
Hg		7–13	Male	2.92×10^{-6}	2.92×10^{-6}	2.46×10^{-6}	4.15×10^{-6}	1.14×10^{-5}	1.67×10^{-10}	1.69×10^{-5}
			Female	2.92×10^{-6}		2.46×10^{-6}	4.11×10^{-6}	1.17×10^{-5}	4.86×10^{-11}	1.74×10^{-5}
		14–17	Male	1.93×10^{-6}	2.08×10^{-6}	1.72×10^{-6}	2.81×10^{-6}	5.68×10^{-6}	7.09×10^{-10}	7.54×10^{-6}
	Female		2.23×10^{-6}		1.92×10^{-6}	3.23×10^{-6}	6.66×10^{-6}	5.09×10^{-10}	8.77×10^{-6}	
	18–49	Male	2.10×10^{-6}	2.10×10^{-6}	1.62×10^{-6}	2.94×10^{-6}	9.47×10^{-6}	3.05×10^{-10}	1.44×10^{-5}	
		Female	2.10×10^{-6}		1.65×10^{-6}	2.93×10^{-6}	9.33×10^{-6}	5.64×10^{-10}	1.42×10^{-5}	
	50–78	Male	1.87×10^{-6}	1.90×10^{-6}	1.62×10^{-6}	2.69×10^{-6}	6.00×10^{-6}	2.68×10^{-12}	8.26×10^{-6}	
		Female	1.94×10^{-6}		1.68×10^{-6}	2.79×10^{-6}	6.02×10^{-6}	4.53×10^{-10}	8.11×10^{-6}	
		All		2.25×10^{-6}	1.89×10^{-6}	3.21×10^{-6}	8.29×10^{-6}	3.45×10^{-10}	1.19×10^{-5}	
	Cd	7–13	Male	1.24×10^{-5}	1.24×10^{-5}	3.20×10^{-6}	1.06×10^{-5}	7.31×10^{-4}	5.85×10^{-9}	1.44×10^{-3}
			Female	1.25×10^{-5}		3.15×10^{-6}	1.02×10^{-5}	9.78×10^{-4}	8.38×10^{-9}	1.93×10^{-3}
		14–17	Male	8.17×10^{-6}	8.74×10^{-6}	2.17×10^{-6}	6.86×10^{-6}	6.04×10^{-4}	3.93×10^{-9}	1.19×10^{-3}
Female			9.31×10^{-6}		2.42×10^{-6}	7.68×10^{-6}	6.67×10^{-4}	5.39×10^{-9}	1.31×10^{-3}	
18–49		Male	8.45×10^{-6}	8.63×10^{-6}	2.09×10^{-6}	6.78×10^{-6}	4.62×10^{-4}	8.71×10^{-10}	9.06×10^{-4}	
		Female	8.80×10^{-6}		2.17×10^{-6}	7.24×10^{-6}	4.08×10^{-4}	1.60×10^{-9}	7.98×10^{-4}	
50–78		Male	8.55×10^{-6}	8.42×10^{-6}	2.05×10^{-6}	6.60×10^{-6}	1.74×10^{-3}	8.14×10^{-9}	3.47×10^{-3}	
		Female	8.29×10^{-6}		2.10×10^{-6}	6.79×10^{-6}	4.13×10^{-4}	3.50×10^{-9}	8.08×10^{-4}	
		All		9.56×10^{-6}	2.42×10^{-6}	7.84×10^{-6}	7.51×10^{-4}	4.71×10^{-9}	1.48×10^{-3}	
Cr		7–13	Male	1.49×10^{-3}	1.51×10^{-3}	1.13×10^{-3}	1.94×10^{-3}	9.60×10^{-3}	-1.82×10^{-4}	1.62×10^{-2}
			Female	1.53×10^{-3}		1.14×10^{-3}	2.01×10^{-3}	1.31×10^{-2}	-8.66×10^{-5}	2.30×10^{-2}
		14–17	Male	1.01×10^{-3}	1.08×10^{-3}	7.97×10^{-4}	1.34×10^{-3}	5.57×10^{-3}	-8.74×10^{-5}	9.11×10^{-3}
	Female		1.16×10^{-3}		8.78×10^{-4}	1.51×10^{-3}	1.11×10^{-2}	-1.22×10^{-4}	1.98×10^{-2}	
	18–49	Male	1.09×10^{-3}	1.09×10^{-3}	7.75×10^{-4}	1.41×10^{-3}	9.08×10^{-3}	-1.52×10^{-4}	1.59×10^{-2}	
		Female	1.09×10^{-3}		7.72×10^{-4}	1.42×10^{-3}	1.30×10^{-2}	-1.64×10^{-4}	2.38×10^{-2}	
	50–78	Male	9.58×10^{-4}	9.80×10^{-4}	7.49×10^{-4}	1.25×10^{-3}	4.88×10^{-3}	-6.68×10^{-5}	7.84×10^{-3}	
		Female	1.00×10^{-3}		7.77×10^{-4}	1.30×10^{-3}	5.94×10^{-3}	-1.04×10^{-4}	9.88×10^{-3}	
		All		1.17×10^{-3}	8.76×10^{-4}	1.52×10^{-3}	9.03×10^{-3}	-1.21×10^{-4}	1.57×10^{-2}	
	Zn	7–13	Male	1.81×10^{-2}	1.82×10^{-2}	1.43×10^{-2}	2.47×10^{-2}	1.05×10^{-1}	-1.82×10^{-2}	1.73×10^{-1}
			Female	1.83×10^{-2}		1.44×10^{-2}	2.46×10^{-2}	1.24×10^{-1}	-1.17×10^{-2}	2.10×10^{-1}
		14–17	Male	1.22×10^{-2}	1.30×10^{-2}	1.00×10^{-2}	1.66×10^{-2}	6.05×10^{-2}	-8.16×10^{-3}	9.66×10^{-2}
Female			1.38×10^{-2}		1.13×10^{-2}	1.89×10^{-2}	6.32×10^{-2}	-9.34×10^{-3}	9.83×10^{-2}	
18–49		Male	1.29×10^{-2}	1.30×10^{-2}	9.41×10^{-3}	1.74×10^{-2}	1.01×10^{-1}	-1.11×10^{-2}	1.74×10^{-1}	
		Female	1.30×10^{-2}		9.65×10^{-3}	1.74×10^{-2}	1.10×10^{-1}	-1.02×10^{-2}	1.92×10^{-1}	
50–78		Male	1.16×10^{-2}	1.18×10^{-2}	9.42×10^{-3}	1.58×10^{-2}	5.44×10^{-2}	-1.01×10^{-2}	8.51×10^{-2}	
		Female	1.19×10^{-2}		9.57×10^{-3}	1.62×10^{-2}	6.88×10^{-2}	-9.68×10^{-3}	1.14×10^{-1}	
		All		1.40×10^{-2}	1.10×10^{-2}	1.89×10^{-2}	8.58×10^{-2}	-1.11×10^{-2}	1.43×10^{-1}	
Mn		7–13	Male	3.34×10^{-3}	3.36×10^{-3}	2.35×10^{-3}	4.49×10^{-3}	2.44×10^{-2}	1.81×10^{-5}	4.15×10^{-2}
			Female	3.38×10^{-3}		2.35×10^{-3}	4.47×10^{-3}	2.39×10^{-2}	1.48×10^{-5}	4.02×10^{-2}
		14–17	Male	2.24×10^{-3}	2.42×10^{-3}	1.66×10^{-3}	3.06×10^{-3}	1.53×10^{-2}	1.31×10^{-5}	2.58×10^{-2}
	Female		2.60×10^{-3}		1.88×10^{-3}	3.51×10^{-3}	1.54×10^{-2}	1.40×10^{-5}	2.52×10^{-2}	
	18–49	Male	2.20×10^{-3}	2.30×10^{-3}	1.60×10^{-3}	2.96×10^{-3}	1.52×10^{-2}	1.29×10^{-5}	2.56×10^{-2}	
		Female	2.40×10^{-3}		1.58×10^{-3}	3.17×10^{-3}	2.40×10^{-2}	1.00×10^{-5}	4.26×10^{-2}	
	50–78	Male	2.15×10^{-3}	2.21×10^{-3}	1.55×10^{-3}	2.90×10^{-3}	1.68×10^{-2}	1.47×10^{-5}	2.90×10^{-2}	
		Female	2.27×10^{-3}		1.63×10^{-3}	3.09×10^{-3}	1.50×10^{-2}	7.62×10^{-6}	2.50×10^{-2}	
		All		2.57×10^{-3}	1.82×10^{-3}	3.46×10^{-3}	1.88×10^{-2}	1.32×10^{-5}	3.19×10^{-2}	



Table 4 (Contd.)

	Ages		Mean		50%	75%	95%	Min	Max
Cu	7–13	Male	2.22×10^{-3}	2.26×10^{-3}	1.65×10^{-3}	2.80×10^{-3}	3.43×10^{-2}	1.58×10^{-5}	6.40×10^{-2}
		Female	2.30×10^{-3}		1.73×10^{-3}	2.94×10^{-3}	1.99×10^{-2}	5.30×10^{-5}	3.53×10^{-2}
	14–17	Male	1.45×10^{-3}	1.58×10^{-3}	1.12×10^{-3}	1.85×10^{-3}	8.72×10^{-3}	4.84×10^{-5}	1.46×10^{-2}
		Female	1.71×10^{-3}		1.31×10^{-3}	2.20×10^{-3}	1.70×10^{-2}	9.41×10^{-6}	3.05×10^{-2}
	18–49	Male	1.55×10^{-3}	1.54×10^{-3}	1.10×10^{-3}	1.99×10^{-3}	1.69×10^{-2}	4.20×10^{-5}	3.06×10^{-2}
		Female	1.54×10^{-3}		1.11×10^{-3}	1.99×10^{-3}	1.03×10^{-2}	3.60×10^{-5}	1.74×10^{-2}
	50–78	Male	1.40×10^{-3}	1.42×10^{-3}	1.07×10^{-3}	1.82×10^{-3}	9.95×10^{-3}	3.98×10^{-5}	1.71×10^{-2}
		Female	1.44×10^{-3}		1.09×10^{-3}	1.85×10^{-3}	1.02×10^{-2}	6.36×10^{-6}	1.76×10^{-2}
		All		1.70×10^{-3}	1.27×10^{-3}	2.18×10^{-3}	1.59×10^{-2}	3.14×10^{-5}	2.84×10^{-2}
	Fe	7–13	Male	1.73×10^{-2}	1.75×10^{-2}	9.04×10^{-3}	2.02×10^{-2}	2.51×10^{-1}	1.50×10^{-4}
Female			1.77×10^{-2}		9.04×10^{-3}	2.02×10^{-2}	1.07×10^0	1.46×10^{-4}	2.11×10^0
14–17		Male	1.18×10^{-2}	1.24×10^{-2}	6.34×10^{-3}	1.39×10^{-2}	3.34×10^{-1}	8.81×10^{-5}	6.41×10^{-1}
		Female	1.30×10^{-2}		7.19×10^{-3}	1.55×10^{-2}	2.82×10^{-1}	1.20×10^{-4}	5.34×10^{-1}
18–49		Male	1.23×10^{-2}	1.23×10^{-2}	6.03×10^{-3}	1.41×10^{-2}	2.61×10^{-1}	7.26×10^{-5}	4.94×10^{-1}
		Female	1.23×10^{-2}		6.02×10^{-3}	1.37×10^{-2}	2.90×10^{-1}	7.24×10^{-5}	5.53×10^{-1}
50–78		Male	1.11×10^{-2}	1.13×10^{-2}	5.92×10^{-3}	1.30×10^{-2}	2.25×10^{-1}	9.01×10^{-5}	4.26×10^{-1}
		Female	1.15×10^{-2}		6.21×10^{-3}	1.36×10^{-2}	4.60×10^{-1}	1.32×10^{-4}	8.95×10^{-1}
		All		1.34×10^{-2}	6.97×10^{-3}	1.55×10^{-2}	3.97×10^{-1}	1.09×10^{-4}	7.64×10^{-1}

determination results of 9 metal elements in 100 groups of CHS; secondly, the simulation results of the mean and different exposure sites (50%, 75%, 95%) were obtained by 10 000 iterations; finally, the above simulation results were compared with RfD and SF.

3. Results and discussion

3.1. Metal element contents

Content and detection rate of 9 metal elements in 100 groups of CHS were shown in Table 3. The results showed that the detection rate of Hg was 0%, the detection rate of Cd was 45%, and the detection rate of other seven element metals was more than 90%. The contents of Pb, As, Hg, Cd, Cr, Zn, Mn, Cu and Fe were as follows: 0.01–1.130 mg kg⁻¹, 0.003–0.045 mg kg⁻¹, 0.0015–0.002 mg kg⁻¹, 0.0005–0.160 mg kg⁻¹, 0.005–3.080 mg kg⁻¹, 0.500–35.664 mg kg⁻¹, 0.1–11.603 mg kg⁻¹, 0.100–5.819 mg kg⁻¹, 0.375–74.358 mg kg⁻¹. Among them, the maximum value of Pb in one group exceeded the maximum value (1 mg kg⁻¹) stipulated in the food pollutant limit of China's national food safety standard GB2762-2017.⁴ GB2762-2017 stipulated that the maximum limit of As was 0.5 mg kg⁻¹, and the determination value of As content in 100 groups of CHS were all lower than the maximum limit. However, the maximum limit values of Cd, Cr, Fe, Mn, Cu and Zn in CHS have not been found at home and abroad, so it is necessary to conduct risk assessment.

3.2. Risk assessment

The fitting distribution of the above metal elements content determination results was performed to determine the probability distribution. The results showed that Pb, Cd, Cu, Cr and Fe were lognormal distribution, As was beta distribution, Zn was max extreme distribution and Mn was gamma distribution.

Determination of Hg were all below the detection limit, so treated with uniform distribution.

According to different age groups, formula (1) is used to calculate the CDI of 9 metals in CHS. The results of 10 000 iterations using Monte Carlo simulation technology were shown in Table 4. The average CDI of Pb, As, Hg, Cd, Cr, Zn, Mn, Cu and Fe were as follows: 2.63×10^{-4} mg per kg per bw per d, 2.00×10^{-5} mg per kg per bw per d, 2.25×10^{-6} mg per kg per bw per d, 9.56×10^{-6} mg per kg per bw per d, 1.17×10^{-3} mg per kg per bw per d, 1.40×10^{-2} mg per kg per bw per d, 2.57×10^{-3} mg per kg per bw per d, 1.70×10^{-3} mg per kg per bw per d, 1.34×10^{-2} mg per kg per bw per d. All the average CDI were lower than the RfD of each metal elements. The order of CDI was Zn > Fe > Mn > Cu > Cr > Pb > As > Cd > Hg. Zn had the highest average CDI value and Hg had the lowest average CDI value.

The CDI values of Pb, Hg and Cr ingested by different age groups through CHS were in the order of 7–13 > 18–49 > 14–17 > 50–78. The CDI values of As, Cd, Zn, Mn, Cu and Fe were in the order of 7–13 > 14–17 > 18–49 > 50–78. Compared with different genders in the same age group, the CDI of females was generally higher than males.

3.3. Non-carcinogenic risk

The non-carcinogenic risk in this study was assessed with HQ values at 50%, 75%, and 95% exposure sites. Non-carcinogenic risks of element metals from ingestion of CHS were shown in Table 5 and Fig. 1. Fig. 1 shown that mean HQ value of nine metal elements: Cr > Pb > Zn > Fe > Mn > Cu > Cd > As > Hg. Table 5 shown that 50% and 75% exposure sites HQ value were all less than 1. At the exposure site, 95% HQ values of As, Hg, Zn, Mn, Cu and Fe were all less than 1, Pb was all more than 1, Cd was more than 1 at 7–13 years old (female) and 50–78 years old (male), Cr was more than 1 at 7–13 years old (male and female),





Table 5 Probabilistic estimation of HQ and calculated HI for investigated elements

Metal elements	Exposure site	7-13		14-17		18-49		50-78		Total HQ	Contribution to HI (%)	HI
		Male	Female	Male	Female	Male	Female	Male	Female			
Pb	50%	6.90×10^{-2}	4.46×10^{-2}	3.04×10^{-2}	3.54×10^{-2}	3.03×10^{-2}	3.03×10^{-2}	3.03×10^{-2}	2.91×10^{-2}	3.74×10^{-2}	25%	3.06×10^{-1}
	75%	1.26×10^{-1}	8.30×10^{-2}	5.58×10^{-2}	6.60×10^{-2}	5.86×10^{-2}	5.86×10^{-2}	5.86×10^{-2}	5.35×10^{-2}	6.95×10^{-2}		
	95%	4.43×10^0	1.96×10^0	2.23×10^0	1.64×10^0	1.57×10^0	1.74×10^0	1.74×10^0	1.21×10^0	1.98×10^0		
As	50%	6.40×10^{-3}	6.46×10^{-3}	4.57×10^{-3}	5.00×10^{-3}	4.24×10^{-3}	4.39×10^{-3}	4.39×10^{-3}	4.20×10^{-3}	4.95×10^{-3}	2%	
	75%	9.98×10^{-3}	9.98×10^{-3}	6.85×10^{-3}	7.75×10^{-3}	6.86×10^{-3}	7.04×10^{-3}	7.04×10^{-3}	6.52×10^{-3}	7.71×10^{-3}		
	95%	4.89×10^{-2}	5.39×10^{-2}	2.57×10^{-2}	3.20×10^{-2}	3.93×10^{-2}	3.98×10^{-2}	3.98×10^{-2}	2.49×10^{-2}	3.63×10^{-2}		
Hg	50%	4.32×10^{-3}	4.32×10^{-3}	3.01×10^{-3}	3.36×10^{-3}	2.84×10^{-3}	2.90×10^{-3}	2.90×10^{-3}	2.85×10^{-3}	3.32×10^{-3}	1%	
	75%	6.50×10^{-3}	6.42×10^{-3}	4.45×10^{-3}	5.08×10^{-3}	4.55×10^{-3}	4.49×10^{-3}	4.49×10^{-3}	4.27×10^{-3}	5.02×10^{-3}		
	95%	2.00×10^{-2}	2.05×10^{-2}	9.97×10^{-3}	1.17×10^{-2}	1.66×10^{-2}	1.64×10^{-2}	1.64×10^{-2}	1.05×10^{-2}	1.45×10^{-2}		
Cd	50%	3.86×10^{-3}	3.80×10^{-3}	2.61×10^{-3}	2.92×10^{-3}	2.52×10^{-3}	2.61×10^{-3}	2.61×10^{-3}	2.47×10^{-3}	2.91×10^{-3}	4%	
	75%	9.51×10^{-3}	9.12×10^{-3}	6.12×10^{-3}	6.90×10^{-3}	6.07×10^{-3}	6.43×10^{-3}	6.43×10^{-3}	5.96×10^{-3}	7.02×10^{-3}		
	95%	8.81×10^{-1}	1.18×10^0	7.27×10^{-1}	8.03×10^{-1}	5.57×10^{-1}	4.92×10^{-1}	4.92×10^{-1}	2.10×10^0	9.04×10^{-1}		
Cr	50%	1.36×10^{-1}	1.37×10^{-1}	9.60×10^{-2}	1.06×10^{-1}	9.33×10^{-2}	9.30×10^{-2}	9.30×10^{-2}	9.03×10^{-2}	1.06×10^{-1}	46%	
	75%	2.05×10^{-1}	2.12×10^{-1}	1.43×10^{-1}	1.62×10^{-1}	1.47×10^{-1}	1.49×10^{-1}	1.49×10^{-1}	1.34×10^{-1}	1.61×10^{-1}		
	95%	1.16×10^0	1.57×10^0	6.71×10^{-1}	1.33×10^0	1.09×10^0	1.57×10^0	1.57×10^0	5.88×10^{-1}	1.09×10^0		
Zn	50%	2.15×10^{-2}	2.16×10^{-2}	1.51×10^{-2}	1.69×10^{-2}	1.41×10^{-2}	1.45×10^{-2}	1.45×10^{-2}	1.41×10^{-2}	1.65×10^{-2}	7%	
	75%	3.28×10^{-2}	3.26×10^{-2}	2.21×10^{-2}	2.53×10^{-2}	2.25×10^{-2}	2.25×10^{-2}	2.25×10^{-2}	2.10×10^{-2}	2.50×10^{-2}		
	95%	1.58×10^{-1}	1.86×10^{-1}	9.07×10^{-2}	9.47×10^{-2}	1.51×10^{-1}	1.64×10^{-1}	1.64×10^{-1}	8.16×10^{-2}	1.29×10^{-1}		
Mn	50%	1.28×10^{-2}	1.28×10^{-2}	9.05×10^{-3}	1.03×10^{-2}	8.74×10^{-3}	8.63×10^{-3}	8.63×10^{-3}	8.47×10^{-3}	9.96×10^{-3}	5%	
	75%	2.12×10^{-2}	2.10×10^{-2}	1.46×10^{-2}	1.66×10^{-2}	1.40×10^{-2}	1.47×10^{-2}	1.47×10^{-2}	1.37×10^{-2}	1.63×10^{-2}		
	95%	1.33×10^{-1}	1.31×10^{-1}	8.36×10^{-2}	8.43×10^{-2}	8.30×10^{-2}	8.30×10^{-2}	8.30×10^{-2}	8.20×10^{-2}	1.02×10^{-1}		
Cu	50%	1.24×10^{-2}	1.30×10^{-2}	8.42×10^{-3}	9.88×10^{-3}	8.30×10^{-3}	8.32×10^{-3}	8.32×10^{-3}	8.07×10^{-3}	9.57×10^{-3}	4%	
	75%	1.84×10^{-2}	1.95×10^{-2}	1.23×10^{-2}	1.47×10^{-2}	1.31×10^{-2}	1.30×10^{-2}	1.30×10^{-2}	1.21×10^{-2}	1.44×10^{-2}		
	95%	2.58×10^{-1}	1.50×10^{-1}	6.55×10^{-2}	1.27×10^{-1}	1.27×10^{-1}	7.72×10^{-2}	7.72×10^{-2}	7.48×10^{-2}	7.68×10^{-2}		
Fe	50%	1.36×10^{-2}	1.36×10^{-2}	9.50×10^{-3}	1.08×10^{-2}	9.04×10^{-3}	9.03×10^{-3}	9.03×10^{-3}	8.88×10^{-3}	1.05×10^{-2}	7%	
	75%	2.50×10^{-2}	2.46×10^{-2}	1.71×10^{-2}	1.91×10^{-2}	1.71×10^{-2}	1.68×10^{-2}	1.68×10^{-2}	1.60×10^{-2}	1.91×10^{-2}		
	95%	3.76×10^{-1}	1.61×10^0	5.01×10^{-1}	4.22×10^{-1}	3.91×10^{-1}	4.35×10^{-1}	4.35×10^{-1}	3.38×10^{-1}	6.90×10^{-1}		

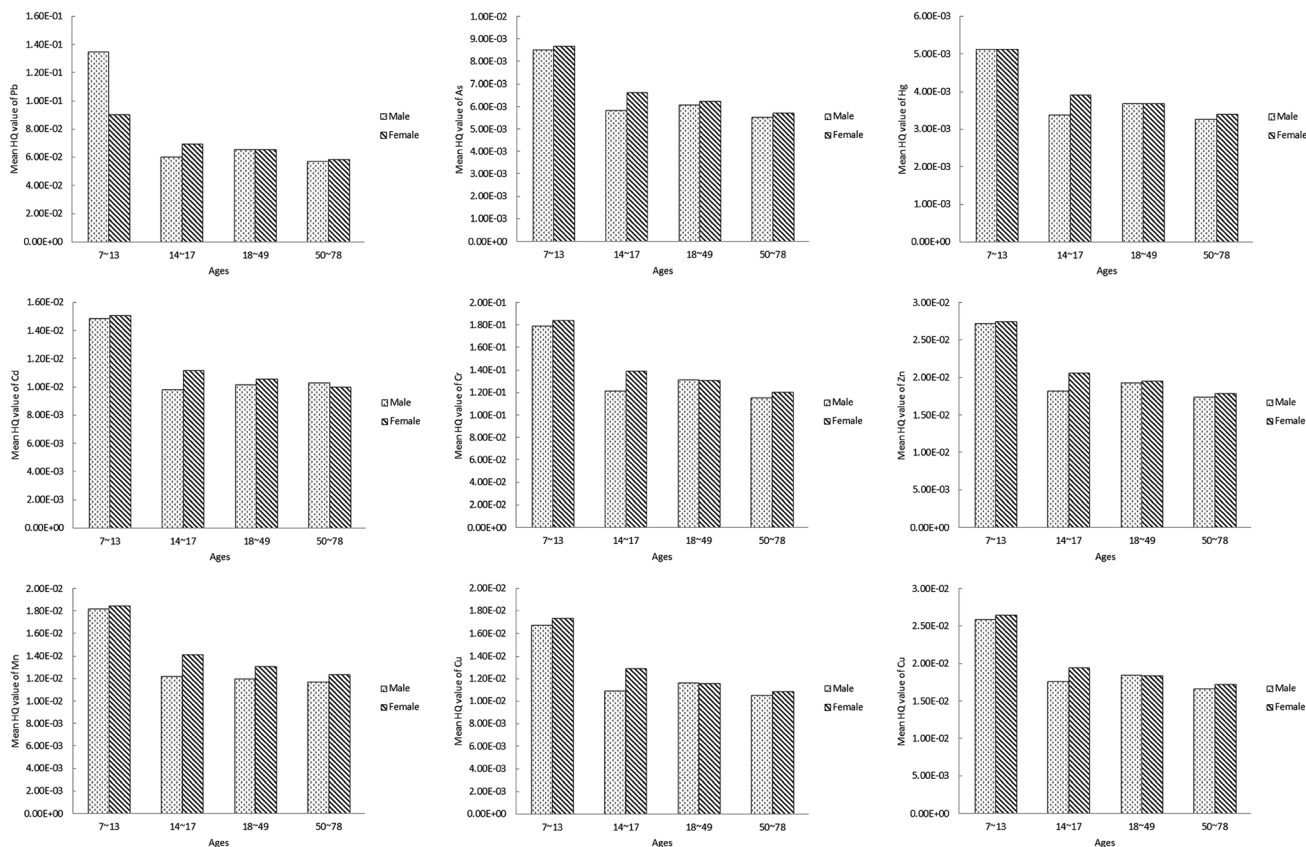


Fig. 1 Mean HQ value of nine metal elements.

14–17 years old (female) and 18–49 years old (male and female). The results showed that humans will not have harmful effects on the 75% probability of ingesting these element metals through CHS, and 95% probability will only have harmful effects on some people. The average HI value of the nine element metals was $0.306 < 1$, indicating that the overall risk of the nine element metals was safe. The contribution rate of Cr to HI value was the largest 46%, followed by Pb (25%), Zn and Fe (7%), and other element metals (5% or less).

3.4. Carcinogenic risk

Due to the lack of carcinogenic slope factors of Hg, Cr, Cu, Fe, Mn and Zn, this experiment only estimated the carcinogenic risk of three metal elements (As, Cd and Pb) in CHS. The mean carcinogenic risk values (standard deviation) As, Cd and Pb of children, adolescents, adults and the elderly were respectively as following: 3.87×10^{-5} (1.92×10^{-4}), 7.84×10^{-5} (2.60×10^{-4}), 3.55×10^{-6} (7.43×10^{-6}), 2.80×10^{-5} (1.21×10^{-4}), 5.51×10^{-5} (1.75×10^{-4}), 2.05×10^{-6} (3.30×10^{-6}), 2.78×10^{-5} (1.55×10^{-4}), 5.43×10^{-5} (1.69×10^{-4}), 2.06×10^{-6} (3.64×10^{-6}), 2.53×10^{-5} (1.17×10^{-4}), 5.31×10^{-5} (2.76×10^{-4}), 1.82×10^{-6} (3.10×10^{-6}). All cancer risks ranged from 1×10^{-6} to 1×10^{-4} . In general, risks in the range of 1×10^{-6} to 1×10^{-4} were considered tolerable.⁴⁰ Therefore, on the whole, the carcinogenic risk levels of As, Cd and Pb in CHS were acceptable.

3.5. Sensitivity analysis

Sensitivity analysis was performed on AT, ED, Cf, PIR and BW with Crystal ball software to determine the most important variables affecting the amount of health risk. Fig. 2 showed the results of a sensitivity analysis used to assess the non-carcinogenic risk of nine metal elements in CHS. The analysis results showed that Cf, PIR and BW were the main variables affecting the risk index except AT and ED (this is a pair of complementary parameters), among which the metal element content in CHS was the most sensitive (see blue bar in Fig. 1). The contribution rates of Pb, As, Hg, Cd, Cr, Mn, Cu, Zn and Fe to risk assessment were 55.5%, 31.3%, 39.6%, 74.1%, 33.2%, 47.2%, 31.6%, 39.7% and 54.4%, respectively. It can be seen that the content of each of the 9 metal elements in CHS is the most important factor that determines the risk index of ingesting metal elements. PIR was positively correlated with risk index, and its influence degree ranged from 4.1% to 12.0%, while body weight was negatively correlated with risk index. In general, it is particularly important to control the contamination of hotpot base materials and processing process to reduce the content of 9 element metals in CHS to ensure the safety of the exposure of dietary metal elements in hotpot.

3.6. Uncertainty analysis

Monte Carlo simulation was used to analyze the uncertainty of the estimated parameters. Through sensitivity analysis, it is



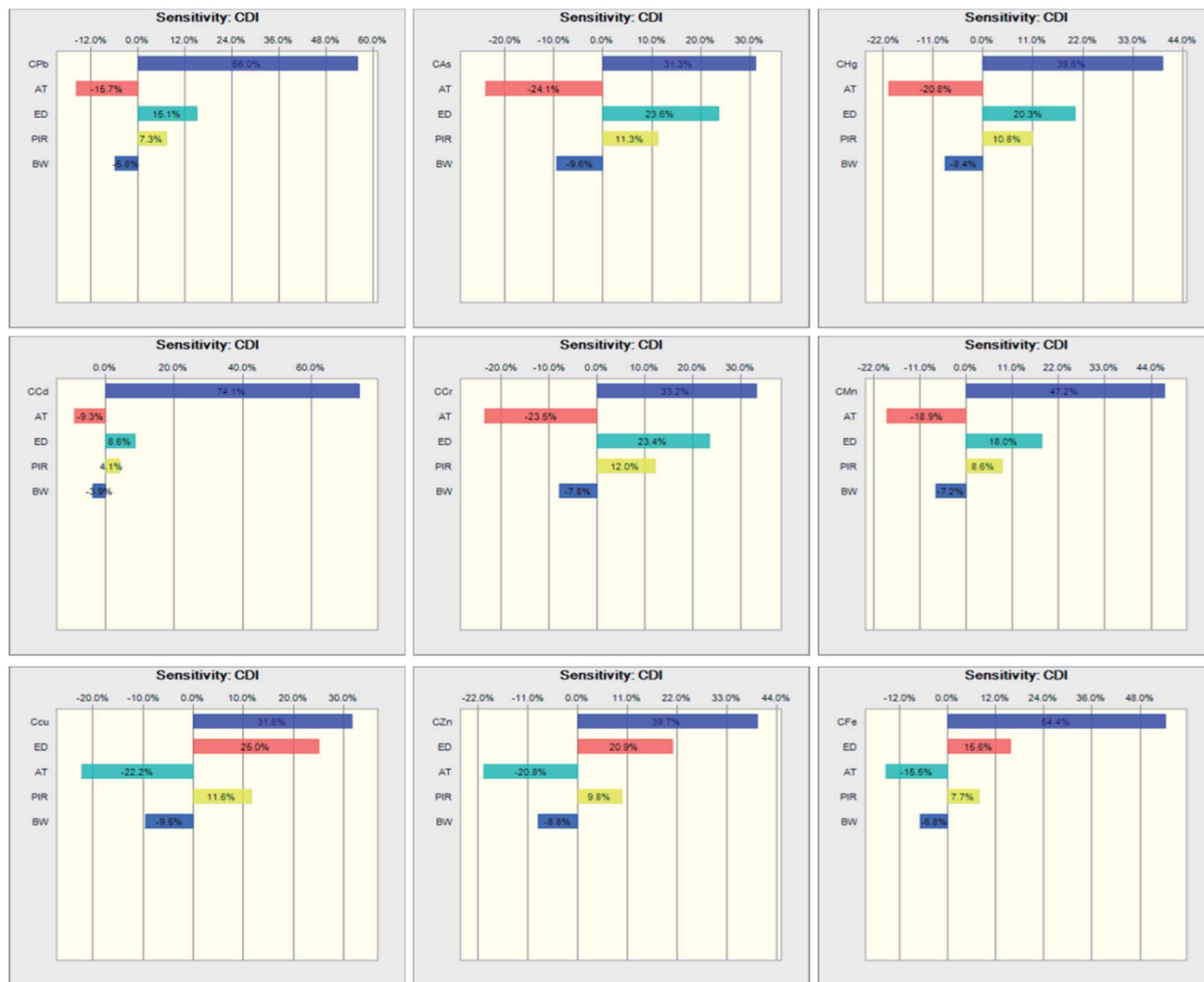


Fig. 2 Sensitivity analysis of metal element risk factors in CHS.

determined that the content of metal elements in the CHS, the daily dietary intake per capita and the per capita body weight of local residents are the main factors affecting CDI. The content of metal elements in the CHS depends on the brand, sampling area, selection and accuracy of measurement method. Due to the influence of sampling conditions and cost, this study only collected 100 groups of CHS in Chongqing for the measurement of metal element content. The limitation of sample size may lead to the uncertainty of metal element exposure risk assessment. All of the above results were based on the assumption that the diners will absorb all the metal elements contained in the CHS. However, when actually eating hotpot, the CHS was used after dissolving in water in a certain proportion, and the CHS was not directly eaten. Therefore, this hypothesis maximized the risk of experimental results. In addition, the selection of exposed population was only considered from the age difference and gender difference, and special groups such as pregnant women and sensitive groups with physical health defects were not considered, so the evaluation results were uncertain and limited to some extent. In future studies, it is

necessary to further expand the sampling scope, sampling scale, collection times, consumption data and other influencing factors, so as to make the conclusion of risk assessment of CHS more comprehensive and convincing.

4. Conclusions

Food safety risk assessment results are used in many ways, the most important of which is to serve as the scientific basis for the formulation and revision of food safety standards and the implementation of food safety supervision and management. In this study, Crystal ball software of Monte Carlo simulation technology was used to evaluate the risk of nine metal elements exposed to CHS pose no chronic toxicity risk to 90% of Chongqing residents. Pb, Cr and Cd bring less than 10% hazard risk to some people. It is worth noting that these probabilistic risks are the result of risk maximization. The variable that contributes the most to the sensitivity of risk index is the metal element content of CHS, which indicates that controlling the



contamination of CHS is especially important to ensure the safety of exposed food. The result of this risk assessment provides the theoretical basis for the formulation and revision of the CHS limit standard.

Conflicts of interest

There are no conflicts to declare.

Acknowledgements

This work was supported by Special Performance Incentive Guidance Program for Chongqing Scientific Research Institutions in 2018 (100125) and Performance Incentive Program for Chongqing Institutions (4411800216).

References

- M. Wu, P. Guo, S. W. Tsui, H. Chen and Z. Zhao, An ethnobotanical survey of medicinal spices used in Chinese hotpot, *Food Res. Int.*, 2012, **48**, 226–232.
- M. Wang, Y. Wang, B. Zhao, N. Zhao, Z. Xiong and L. Zhao, Multi-class determination of steroid hormones and antibiotics in fatty hotpot ingredients by pressurized liquid extraction and liquid chromatography–tandem mass spectrometry, *J. Pharm. Biomed. Anal.*, 2019, **171**, 193–203.
- H. Dong, X. Zeng and W. Bai, Solid phase extraction with high polarity Carb/PSA as composite fillers prior to UPLC-MS/MS to determine six bisphenols and alkylphenols in trace level hotpot seasoning, *Food Chem.*, 2018, **258**, 206–213.
- GB 2762-2017 Food safety national standard of food pollutant limit, Standards Press of China, 2017.
- X. Liu, G. Liu, L. Tao, J. Hou, A. Wang and R. Li, Heavy metal speciation and pollution of agricultural soils along Jishui River in non-ferrous metal mine area in Jiangxi Province, China, *J. Geochem. Explor.*, 2013, **132**, 156–163.
- X. Zhang, T. Zhong, L. Liu and X. Ouyang, Impact of Soil Heavy Metal Pollution on Food Safety in China, *PLoS One*, 2015, **10**, e135182.
- W. Kong, Q. Li, J. Liu, X. Li, L. Zhao, Y. Su, Q. Yue and B. Gao, Adsorption behavior and mechanism of heavy metal ions by chicken feather protein based semi-interpenetrating polymer networks super adsorbent resin, *RSC Adv.*, 2016, **6**, 83234–83243.
- Y. Guo, Y. Fu, S. Yan, Y. Yang, Y. Guo and Z. Bao, Heavy metal distribution between parent soil and pepper in an unpolluted area, Hainan Island, China, *Environ. Earth Sci.*, 2012, **66**, 1083–1089.
- A. Singh and S. M. Prasad, Reduction of heavy metal load in food chain: technology assessment, *Rev. Environ. Sci. Bio/Technol.*, 2011, **10**, 199–214.
- A. Jafarian-Dehkordi and M. Alehashem, Heavy metal contamination of vegetables in Isfahan, Iran, *Res. Pharm. Sci.*, 2013, **8**, 51–58.
- M. Arora, B. Kiran, S. Rani, A. Rania, B. Kaur and N. Mittal, Heavy metal accumulation in vegetables irrigated with water from different sources, *Food Chem.*, 2008, **111**, 811–815.
- C. Luo, C. Liu, X. Liu, X. Liu, F. Li, G. Zhang and X. Li, Heavy metal contamination in soils and vegetables near an e-waste processing site, south China, *J. Hazard. Mater.*, 2011, **186**, 481–490.
- M. Eziz, A. Mamut and A. Mohmmad, Soil heavy metal pollution and ecological risk warning assessment of pepper field in Yanqi Basin, Xinjiang, *Acta Ecol. Sin.*, 2018, **38**, 1075–1086.
- S. Xiong, D. Xing, Y. Wang, A. Zhang, G. Peng, F. Liao and M. Hu, Distribution and Risk Assessment of Heavy Metals in a Pepper-Growing Area—A Case Study in Baiyi, Guizhou, *J. Southwest Univ., Nat. Sci. Ed.*, 2018, **40**, 35–41.
- L. Tao, J. Ren and G. Qiao, Analyses on Heavy Metal of Soil in Vegetable Bases of Lanzhou, *Chin. J. Soil Sci.*, 2010, **6**, 1479–1483.
- X. Su, Z. Liu, M. Hu and C. Han, Assessment of Heavy Metal Pollution in the Suburbs of Weinan City in Shaanxi Province, *Hubei Agric. Sci.*, 2016, **55**, 3293–3297.
- C. Xu, S. Guo, Y. Zheng, Q. Dong, J. Liu and N. Liu, Determination and Correlation Analysis of Eight Heavy Metals in Chinese Prickly Ash and Soil of Shaanxi, *Food Ferment. Technol.*, 2014, **50**, 92–95.
- M. Smith, Food Safety in Europe (FOSIE): risk assessment of chemicals in food and diet: overall introduction, *Food Chem. Toxicol.*, 2002, **40**, 141–144.
- T. Rycroft, K. Hamilton, C. N. Haas and L. Linkov, A quantitative risk assessment method for synthetic biology products in the environment, *Sci. Total Environ.*, 2019, **696**, 133940.
- K. K. Sharma and A. Seal, Modeling uncertain data using Monte Carlo integration method for clustering, *Expert Syst. Appl.*, 2019, **137**, 100–116.
- Y. Fakhri, A. Mousavi Khaneghah, G. O. Conti, M. Ferrante, A. Khezri, A. Darvishi, M. Ahmadi, V. Hasanzadeh, A. Rahimizadeh, H. Keramati, B. Moradi and N. Amanidaz, Probabilistic risk assessment (Monte Carlo simulation method) of Pb and Cd in the onion bulb (*Allium cepa*) and soil of Iran, *Environ. Sci. Pollut. Res.*, 2018, **25**, 30894–30906.
- C. Qu, K. Sun, S. Wang, L. Huang and J. Bi, Monte Carlo Simulation-Based Health Risk Assessment of Heavy Metal Soil Pollution: A Case Study in the Qixia Mining Area, China, *Hum. Ecol. Risk Assess. Int. J.*, 2012, **18**, 733–750.
- H. Cao, L. Qiao, H. Zhang and J. Chen, Exposure and risk assessment for aluminium and heavy metals in Puerh tea, *Sci. Total Environ.*, 2010, **408**, 2777–2784.
- R. A. Fallahzadeh, M. T. Ghaneian, M. Miri and M. M. Dashti, Spatial analysis and health risk assessment of heavy metals concentration in drinking water resources, *Environ. Sci. Pollut. Res.*, 2017, **24**, 24790–24802.
- M. Indrani, J. R. A. Clark, P. J. Dobson and O. Richard, Expert perspectives on potential environmental risks from nanomedicines and adequacy of the current guideline on environmental risk assessment, *Environ. Sci.: Nano*, 2018, **5**, 1873–1889.
- B. Djahed, M. Taghavi, M. Farzadkia, S. Norzaee and M. Mirid, Stochastic exposure and health risk assessment



- of rice contamination to the heavy metals in the market of Iranshahr, Iran, *Food Chem. Toxicol.*, 2018, **115**, 405–412.
- 27 Y. Mengliang, *Studies on Residue and Dietary Exposure Risk Assessment of Pesticides Applied in Apple*, Chinese academy of agricultural sciences, 2016.
- 28 A. Gruszecka-Kosowska, Assessment of the Kraków inhabitants' health risk caused by the exposure to inhalation of outdoor air contaminants, *Stoch. Environ. Res. Risk Assess.*, 2018, **32**, 485–499.
- 29 W. Xu, *Study on Vegetable Heavy Metals Risk Assessment in Guangdong Province*, Huazhong agricultural university, 2012.
- 30 C. S. O. Nutrition, *Dietary nutrient intake for Chinese residents*, China Light Industry Press, 2013.
- 31 USEPA, Guidelines for the health risk assessment of chemical mixtures, *Fed. Regist.*, 1986, **51**, 34014–34025.
- 32 USEPA, *Technical Support Document on Health Risk Assessment of Chemical Mixtures*, EPA/600/8-90/064, U.S. Environmental Protection Agency, Office of Research and Development, Washington DC, 1990.
- 33 USEPA, *Supplementary Guidance for Conducting Health Risk Assessment of Chemical Mixtures*, EPA/630R-00/002, U.S. Environmental Protection Agency, Risk Assessment Forum, Washington DC, 2000.
- 34 K. A. Heys, R. F. Shore, M. Glória Pereira, K. C. Jones and F. L. Martin, Risk assessment of environmental mixture effects, *RSC Adv.*, 2016, **6**, 47844.
- 35 USEPA, Guidelines for the health risk assessment of chemical mixtures, *Fed. Regist.*, 1986, **51**, 34014–34025.
- 36 S. C. Sofuoğlu and P. Kavcar, An exposure and risk assessment for fluoride and trace metals in black tea, *J. Hazard. Mater.*, 2008, **158**(2), 392–400.
- 37 USEPA, *Guidelines for Carcinogen Risk Assessment*, NCEA-F-0644, Review Draft, U.S. Environmental Protection Agency, Risk Assessment Forum, Washington DC, 1999.
- 38 Y. Gong, L. M. Nunes, B. K. Greenfield, Q. Zhen, Q. Yang, L. Huang, W. Bu and H. Zhong, Bioaccessibility-corrected risk assessment of urban dietary methylmercury exposure via fish and rice consumption in China, *Sci. Total Environ.*, 2018, **630**, 222–230.
- 39 X. Xiang, *Detection and exposure assessment of 3-MCPD ester in edible oils and optimization of refining process of camellia oil*, Zhejiang university, 2018.
- 40 X. Hu, Y. Zhang, Z. Ding, T. Wang, H. Lian, Y. Sun and J. Wu, Bioaccessibility and health risk of arsenic and heavy metals (Cd, Co, Cr, Cu, Ni, Pb, Zn and Mn) in TSP and PM_{2.5} in Nanjing, China, *Atmos. Environ.*, 2012, **57**, 146–152.

