


Cite this: *RSC Adv.*, 2019, 9, 40536

Environmental risk assessment in livestock manure derived biochars

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Livestock-manure-derived biochar is one of major products obtained from the pyrolysis of livestock manure. This study quantitatively assesses the pollution level and ecological risks associated with heavy metals in livestock manure and the biochar produced by its pyrolysis. The geo-accumulation index (GAI) values of heavy metals in livestock manure were significantly decreased ($P < 0.05$) and indicated to be at the grade of uncontaminated expected for Zn in pig-manure-derived biochar (PMB, 0.94, 800 °C) *via* pyrolysis. Therefore, Zn should be paid more attention in PMB. The risk factors (E_r^i) result shows that heavy metals in biochars were significantly decreased ($P < 0.05$) with increasing pyrolysis temperature. Potential ecological risk index values revealed that the integrated risks from the heavy metals were significantly decreased ($P < 0.05$) after pyrolysis. Similarly, the risk assessment code values indicated that the risks from the heavy metals in livestock-manure-derived biochars were significantly decreased ($P < 0.05$) after pyrolysis. In summary, pyrolysis represents an effective treatment method for livestock manure and can provide an effective method to reduce the risks of environmental pollution.

Received 9th October 2019
Accepted 25th November 2019

DOI: 10.1039/c9ra08186k

rsc.li/rsc-advances

1. Introduction

The vigorous development of the livestock industry in China has brought enormous economic benefits but also causes severe environmental pollution. The annual production of livestock (including poultry) manure in China has reached 3.264 billion tons, which is 1.6 times the total emission of industrial solid wastes.¹ Discharges of manure lead to environmental problems, such as odors rooting in the emission of toxic gases and pollution of waterways because of leaching and runoff of nutrients and heavy metals.² Meanwhile, feedstock additives were used to prevent disease of animals and promote the growth of animals, resulting in high content of heavy metals in manures.³ Long term application of animal manures sustaining high concentrations of heavy metals can lead to elevating concentrations of soil heavy metals,⁴ which subsequently harms crop and humans through the food chain.⁵ Therefore, the appropriate management and treatment of livestock manure are issues of increasing concern.

There are numerous treatment methods that can mitigate most of the toxic gas emissions and water pollution due to leaching of nutrients and heavy metals from livestock manure,

such as composting, calcination,^{6,7} gasification,⁸ and pyrolysis.⁹ However, some of these techniques suffer from limitations and require further improvement, such as high temperature combustion and gasification, where the high contents of moisture and alkali metals in manure cause ash agglomeration.¹⁰ In contrast, pyrolysis operating at lower temperature compared with combustion and gasification, which reduces the risk of ash agglomeration and toxic emissions such as NO_x, SO₂, and particulate matter.¹¹ And manure can be converted *via* pyrolysis into available energy and value-added biochar.⁹ Heavy metals in animal manures can be reapportion into the bio-oil, gas and solid (biochar) phases during pyrolysis, and this reapportion of heavy metals is dependent on experimental conditions.¹² Moreover, the chemical speciation of heavy metals has been shown to be altered from active to more stable states during pyrolysis.¹³

Biochars prepared by livestock manure is the solid carbonaceous residue, which comes from the pyrolysis of livestock manure. The conversion of livestock manure to biochar has a variety of potential economic effectiveness, including agricultural waste reduction, fuel production, and carbon sink.¹⁴ Biochar can weaken the leaching of nutrients and heavy metals from soils,^{15,16} improve soil quality, and increase the pH of acidic soils.¹⁷ Furthermore, it can effectively adsorb pollutants from wastewater.¹⁸ However, the heavy metals already present in manure are not destroyed during pyrolysis and remain problematic.¹⁹ In addition, in terms of pollution adsorption, composite materials are exhibited specific selectivity, non-toxic, robust, and suitable material for high adsorption than

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biochar.^{20–22} Therefore, there is still a need to further improve the pyrolysis process of biochar.

Typical evaluation methods, such as bioavailability and ecotoxicity, geo-accumulation index (GAI), potential ecological risk index (PERI), and risk assessment code (RAC) were used to quantitative evaluation of heavy metals, which have been applied to the analysis of soil, groundwater,²³ natural rivers, PM_{2.5} and PM₁₀, mining areas, brick kiln bottom ashes, atmospheric deposition, and road-deposited sediments.²⁴ These techniques have also been utilized to analyze residues from municipal sewage sludge and its liquefaction, compost, and pyrolysis²⁵ in small-scale lab research, and have been rarely used to livestock-manure-derived biochar.²⁶

To obtain more widespread grasp of the pollution level and underlying ecological risks of heavy metals in biochars, it is essential to perform a systematic and quantificational evaluation of heavy metals toxicity associated with this material. Therefore, the main objectives of the present study were (1) to assess the biological availability and ecotoxicity of heavy metals in livestock-manure-derived biochar, (2) to assess the degree of pollution by heavy metals using the GAI, and (3) to assess the underlying ecological risk using the RAC and PERI.

2. Materials and methods

2.1 Sampling collection and preprocessing

This study investigated two species of manure: pig manure (PM) and sheep manure (SM). The PM was obtained from a large pig farm in Changsha, Hunan. Most of pigs, alongside a small number of sows and piglets, are bred on this farm. The SM was collected from a peasant farm enterprise in Hunan. Sheep are representative reared in a decentralized manner, resulting in a low heavy metal concentration in the SM. Two manure samples were dried to constant weight in a dryer at 60 °C. Then, the samples after dried were mill and through a 100-mesh sieve. The obtained samples were stored in a drying apparatus before analysis.

2.2 Pyrolysis procedure

Pyrolysis was implemented at a level of laboratory-scale cracking furnace (SK-G08123K, China). Pyrolysis was carried out according to previous research.⁹ The pyrolysis chamber was made up of a quartz tube (internal diameter = 72 mm, length = 1000 mm) and an electrically heating furnace with a programmable temperature controller. Temperature profiles were acquired using a thermocouple placed in the center of the reactor. The pyrolysis was conducted in the temperature range from 200 °C to 800 °C and pyrolysis samples were assigned codes according to their pyrolysis temperatures.

The resulting biochar was ground and through a 100-mesh sieve. The produced samples were stored in a drying apparatus before analysis.

2.3 Sequential extraction

The sequential extraction method was used to investigate the chemical speciation distribution of Cd, Cr, Pb, Cu, Zn, Ni, and

Mn in PM, SM, PMB, and SMB. A improved European Community Bureau of Reference (BCR) extraction procedure was employed as reported previously.²⁷ The fractions obtained were as follows: fraction one (F1), acid soluble, fraction associated with carbonate and cation exchange sites; fraction two (F2), reducible, fraction associated with Fe–Mn oxides; fraction three (F3), oxidizable, fraction associated with organic matter and sulfides; fraction four (F4), residual residue. The total heavy metal concentration in PM, SM, livestock-manure-derived biochar, and F4 was determined using the same digestion method (hydrochloric acid, nitric acid, and perchloric acid). Reagent blanks and sample analytical duplicates were applied regularly to ensure the precision of the analysis. Heavy metals from BCR extraction were detected by high-performance Inductively Coupled Plasma Optic Emission Spectrometer (ICP-OES, Optima 8300, USA).

2.4 Analytical procedures

2.4.1 Bioavailability and ecotoxicity. The mobility, bioavailability, and ecotoxicity of heavy metals in PM, SM, PMB, and SMB were determined by their total concentration and speciation.²⁸ According to the BCR sequential extraction procedure used, the biological availability and ecotoxicity of the heavy metals can be represent as follows: F1 + F2, concentrations of heavy metals with direct biological availability and ecotoxicity (CDBio); F3 and F4, concentrations of heavy metals with no toxicity (CNbio).²⁹

2.4.2 Geo-accumulation index. The degree of contamination of biochar can be evaluated using the GAI, as described by Liu and co-workers.³⁰ This pollution index is generally used by researchers for environmental analysis, soil,³¹ and road dust,³² and can be defined by the following formula:

$$\text{GAI} = \log_2 \frac{C_n}{1.5B_n} \quad (1)$$

where C_n is the measured content of a particular heavy metal in PM, SM, PMB, or SMB, and B_n is the background value for the heavy metal. The B_n values of Cd, Cr, Zn, Cu, Pb, Ni, and Mn used in this study were 0.08, 64, 90, 25, 26, 30, and 380 mg kg^{−1}, respectively.³³ The constant factor of 1.5 is included to better analyze the natural variance in the content of a particular substance in the environment that is not markedly influenced by artificial elements. The GAI values were used to classify the PM, SM, PMB, and SMB samples into seven grades based on the degree of contamination: uncontaminated (UC) rank refers to samples with GAI < 0; uncontaminated to moderately contaminated (UMC) rank refers to samples with 0 < GAI < 1; moderately contaminated (MC) rank refers to samples with 1 < GAI < 2; moderately to heavily contaminated (MHC) rank refers to samples with 2 < GAI < 3; heavily contaminated (HC) rank refers to samples with 3 < GAI < 4; heavily to extremely contaminated (HEC) rank refers to samples with 4 < GAI < 5, and extremely contaminated (EC) rank refers to samples with GAI > 5.³⁴

2.4.3 Potential ecological risk index. The potential ecological risk from the heavy metals in the biochar samples was assessed by calculating the PERI as described by Xu and co-



workers.³⁵ The PERI accounts for the toxicity, sensitivity, and content of heavy metals. The PERI is applicable to an enormous all kind of research domains, *e.g.*, ecology, marine pollution, environmental science, environmental chemistry, and biological toxicology, and is a valuable parameter for comprehensively evaluating the ecological risks of heavy metals.^{36,37} The PERI can be calculated using the following formulas:

$$C_f^i = \frac{C_D^i}{C_R^i} \quad (2)$$

$$E_r^i = T_r^i \times C_f^i \quad (3)$$

$$\text{PERI} = \sum_{i=1}^n E_r^i \quad (4)$$

where C_f^i is the pollution coefficient for a given heavy metal, C_D^i is the content of the heavy metal in a PM, SM, PMB, or SMB sample, C_R^i is the background content of the heavy metal and defined as B_n , T_r^i is the toxicity response factor of the heavy metal, and the PERI is the underlying ecological risk caused by the entire pollution. According to Hakanson, the toxicity factors (T_r^i) for Cd, Cr, Cu, Pb, Zn, Ni, and Mn are 30, 2, 5, 5, 1, 5, and 1, respectively.³⁴ The soil background concentrations of Cd, Cr, Cu, Pb, Zn, Ni, and Mn are 0.08, 64, 25, 26, 90, 30, and 380 mg kg⁻¹, separately.³³ To exactly represent the degrees of contamination and underlying ecological risks from heavy metals in the PM, SM, PMB, and SMB samples, the total of the acid-soluble/exchangeable fraction (F1) and the reducible fraction (F2) for individual heavy metals, rather than the total content, was used as the estimated value to calculate the GAI and PERI. The risk factors (E_r^i) were classified according to Ren *et al.*,³⁸ where $E_r^i < 5$ represents a low risk (LR), $5 < E_r^i < 10$ represents a moderate risk (MR), $10 < E_r^i < 20$ represents a considerable risk (CR), $20 < E_r^i < 40$ represents a high risk (HR), and $E_r^i > 40$ represents a very high risk (VHR). For the purposes of this article, a PERI of insufficient 30 stand for a low risk (LR), a PERI between 30 and 60 stand for a moderate risk (MR), a PERI between 60 and 120 stand for a considerable risk (CR), and a PERI of greater than 120 stand for a very high risk (VHR).³⁹

2.4.4 Risk assessment code. The RAC³² has been widely used to evaluate the heavy metal contamination of sediments and soils.^{40,41} In this research, the RAC was applied to assess the availability of heavy metals in the PM, SM, PMB, and SMB samples using a proportion based on the percentage of heavy metals consist in the exchangeable and carbonate fraction (F1). A proportion of heavy metals in F1 of less than 1% represents no risk (NR), a value in the scope of 1% to 10% represents low risk (LR), a value in the scope of 10% to 30% represents medium risk (MR), a value in the scope of 30% to 50% represents high risk (HR), and a value of higher than 50% represents very high risk (VHR).⁴²

3. Results and discussion

3.1 Bioavailability and ecotoxicity assessment

The bioavailabilities and ecotoxicities of the heavy metals and properties in the PM, SM, PMB, and SMB samples are

presented in Tables 1 and 2. For PMB, the CDbio of Cr tended to decline with increasing pyrolysis temperature. However, the CDbio of Cd, Pb, Zn, Cu, Ni, and Mn increased with increasing temperature from 200 °C to 350 °C and then declined from 500 °C to 800 °C. For SMB, the CDbio of Cd, Cr, Pb, Ni, Zn, and Cu tended to decline with increasing pyrolysis temperature. The CDbio of Mn increased with increasing temperature from 200 °C to 350 °C and then declined from 500 °C to 800 °C.

In this study, the CDbio of Cr, Pb, and Ni tended to decrease with increasing pyrolysis temperature. Some previous studies have indicated that heavy metals are fractionated and volatilized during pyrolysis.⁴³ The CDbio of heavy metals in biochar tends to increase with increasing pyrolysis temperature. This phenomenon may be attributable to the more stable organic matter generated with an increasing pyrolysis temperature and the stronger adsorption of the livestock-manure-derived biochars.^{44,45} Because pigs and sheep are fed differently, the CDbio values of the heavy metals in PM and PMB are different from those in SM and SMB. This finding is compliance with the results of a previous research.⁴⁶ The CDbio of Cd in biochar tended to decrease with increasing pyrolysis temperature from 500 °C to 800 °C. This result may be due to fractional distillation and devolatilization of Cd. Previous studies have reported that Cd evaporates with increasing pyrolysis temperature.⁴⁷ Cadmium chloride, produced by the combination of Cd and chloride, is the major Cd species in the vapor phase during pyrolysis.^{48,49} In this study, the bioavailability and ecotoxicity of Pb in livestock-manure-derived biochars was reduced after pyrolysis. This result may be due to Pb make up more chlorides than other elements.^{50–52} The CDbio of Cr tended to decrease with increasing pyrolysis temperature. However, the CNbio of Cr appeared to increase due to agglomeration. This is likely attributable to the low chlorine content during the pyrolysis process and the high concentration of Cr in the concentrated material.^{53–55} Similar results were observed for Cu and Ni as for Pb. The bioavailability and ecotoxicity of Cu reduced in PMB. The bioavailability and ecotoxicity of Ni reduced in SMB. This is likely attributable to sulfur in the feedstock may combine with Cu and Ni to produce metal sulfides and sulfates.⁵⁶ Cu and Ni Sulfides have a lower volatility,⁵⁴ and are likely to cause Cu and Ni to agglomerate in the condensed phase. The manure is subjected to pyrolysis, the surface properties and material affinity of the material affect its adsorption capacity and adsorption process.^{57–59} Therefore, the bioavailability and ecotoxicity of heavy metal in biochars was affected. In summary, the results demonstrate that pyrolysis is an effectually method for reducing the toxicity of heavy metals.

3.2 Contamination degree based on GAI

The GAI was used to assess the degree of metal pollution in PM, SM, PMB, and SMB. The GAI value of the heavy metals in PM, PMB, SM, and SMB are presented in Fig. 1 and 2. The GAI results for Ni, Pb, and Cr in PM and PMB were below 0, corresponding to UC. Similarly, the GAI values for Cr, Zn, Cu, Pb, and Ni in SM and SMB were below 0 (UC).



Table 1 Concentrations of heavy metals in direct bioavailable and non-bioavailable fractions in PM, SM and manure-derived biochars (mg kg⁻¹)

	Cd	Cr	Zn	Cu	Pb	Ni	Mn
CDbio							
PM	0.45 ± 0.01	3.94 ± 0.51	476.16 ± 12.01	227.44 ± 10.01	7.92 ± 1.21	25.12 ± 0.25	342.48 ± 11.1
PM 200	0.46 ± 0.02	3.69 ± 0.52	544.32 ± 11.01	292.48 ± 10.06	7.64 ± 0.12	25.60 ± 0.52	453.76 ± 10.5
PM 350	0.67 ± 0.01	2.91 ± 0.41	402.48 ± 9.21	64.96 ± 0.56	8.24 ± 0.24	25.04 ± 0.48	343.84 ± 9.02
PM 500	0.37 ± 0.03	2.28 ± 0.38	364.48 ± 5.06	17.36 ± 0.09	8.08 ± 0.26	24.08 ± 0.32	356.88 ± 8.21
PM 650	0.21 ± 0.05	2.06 ± 0.29	350.20 ± 8.12	17.48 ± 0.08	7.20 ± 0.08	24.42 ± 0.26	311.28 ± 6.21
PM 800	0.11 ± 0.01	1.39 ± 0.25	259.28 ± 10.05	12.96 ± 0.05	5.48 ± 0.06	22.76 ± 0.29	257.84 ± 5.31
SM	0.54 ± 0.06	2.94 ± 0.26	75.28 ± 0.85	1.60 ± 0.02	9.52 ± 0.01	24.72 ± 0.18	558.40 ± 10.1
SM 200	0.73 ± 0.02	2.46 ± 0.19	80.16 ± 0.54	0.40 ± 0.01	9.30 ± 0.02	24.48 ± 0.19	552.40 ± 12.3
SM 350	0.39 ± 0.01	2.52 ± 0.16	60.88 ± 0.48	— ^a	6.96 ± 0.05	23.28 ± 0.15	581.88 ± 11.2
SM 500	0.06 ± 0.12	2.05 ± 0.23	46.88 ± 0.39	— ^a	6.00 ± 0.01	23.00 ± 0.23	632.64 ± 9.2
SM 650	0.05 ± 0.04	1.64 ± 0.12	34.82 ± 0.26	— ^a	6.02 ± 0.01	22.80 ± 0.16	496.56 ± 7.12
SM 800	— ^a	0.97 ± 0.13	24.02 ± 0.54	— ^a	2.66 ± 0.02	20.02 ± 0.17	302.48 ± 6.12
CNbio							
PM	0.11 ± 0.01	9.53 ± 0.26	196.88 ± 1.2	350.00 ± 1.3	7.60 ± 0.05	25.24 ± 0.06	134.96 ± 0.12
PM 200	0.15 ± 0.01	11.25 ± 0.32	207.20 ± 1.1	347.46 ± 1.2	9.36 ± 0.05	30.48 ± 0.07	78.48 ± 0.11
PM 350	0.31 ± 0.02	20.77 ± 0.31	784.64 ± 1.0	944.48 ± 0.9	18.62 ± 0.07	63.52 ± 0.05	497.84 ± 1.0
PM 500	0.75 ± 0.02	25.32 ± 0.28	976.56 ± 0.8	1160.66 ± 1.0	22.88 ± 0.08	77.28 ± 0.05	603.96 ± 0.9
PM 650	0.89 ± 0.03	26.30 ± 0.19	1079.76 ± 1.2	1190.64 ± 0.8	24.72 ± 0.06	80.10 ± 0.05	671.68 ± 0.8
PM 800	0.96 ± 0.01	28.13 ± 0.19	1179.60 ± 2.1	1250.80 ± 1.2	27.24 ± 0.06	85.48 ± 0.09	770.04 ± 0.86
SM	0.71 ± 0.03	10.38 ± 0.21	25.68 ± 0.09	15.84 ± 0.08	14.72 ± 0.09	26.56 ± 0.04	69.60 ± 0.75
SM 200	0.65 ± 0.02	12.38 ± 0.18	32.32 ± 0.21	19.46 ± 0.07	16.72 ± 0.06	32.88 ± 0.05	143.44 ± 0.26
SM 350	1.65 ± 0.11	20.14 ± 0.23	110.16 ± 1.0	29.56 ± 0.09	32.88 ± 0.05	62.88 ± 0.04	476.00 ± 1.1
SM 500	2.15 ± 0.02	23.93 ± 0.17	144.20 ± 2.1	33.88 ± 0.06	39.48 ± 0.06	75.80 ± 0.07	583.52 ± 0.96
SM 650	2.38 ± 0.11	27.47 ± 0.15	178.46 ± 1.3	37.60 ± 0.05	44.50 ± 0.04	87.56 ± 0.06	839.32 ± 0.86
SM 800	2.48 ± 0.15	29.36 ± 0.16	198.44 ± 0.2	59.44 ± 0.06	50.47 ± 0.06	95.14 ± 0.08	1096.08 ± 1.1
BV	0.08	64	90	25	26	30	380

^a Below the detection limits: CDbio = F1 + F2, CNbio = F3 + F4.**Table 2** Properties of PM, SM and manure-derived biochars^a

Material	T (°C)	Char yield (%)
PM	Material	100
	200	87.58
	350	54.68
	500	46.80
	650	45.10
	800	42.95
SM	Material	100
	200	86.83
	350	56.11
	500	48.66
	650	43.27
	800	41.07

^a Char yield = (biochar mass/raw stuff mass) × 100.

For PM, the GAI results for the heavy metals were in the order of Cu (2.60) > Cd (1.87) > Zn (1.81), which correspond to grades of MC for Cd and Zn and MHC for Cu. For PMB, the GAI values for Cu, Cd, and Zn increased after pyrolysis at 200 °C compared with those in PM. The GAI values of Cu and Zn decreased with increasing pyrolysis temperature from 200 °C to 800 °C. However, the GAI of Cd initially increased with increasing pyrolysis temperature from 1.91 at 200 °C to 2.45 at 350 °C and thereafter started to decrease. The GAI values of Cu in PMB

samples obtained at pyrolysis temperatures from 500 °C to 800 °C were all less than zero, corresponding to UC. Meanwhile, the GAI values of Cd in PMB were 1.59 and 0.78 at 500 °C and 650 °C, respectively. These results correspond to grades of MC for Cd and UMC for Cd in PMB. The GAI values of Zn in PMB samples obtained at pyrolysis temperatures of 500 °C, 650 °C, and 800 °C were 1.43 (MC), 1.37 (MC), and 0.94 (UMC), respectively. These results indicate that the levels of Zn in PMB should be monitored.

For SMB, the GAI of Cd in SMB increased after pyrolysis at 200 °C compared with that in SM. The GAI of Cd in SMB decreased from 1.68 at 350 °C to −1.31 at 650 °C. Upon further increasing the pyrolysis temperature to 800 °C, the GAI of Cd in SMB increased. In general, the GAI values of Cd in SMB samples obtained at pyrolysis temperatures from 500 °C to 800 °C were all under 0 (UC).

The reason that the GAI value of Cd in biochars was increased at 200 °C and thereafter decreased with increasing pyrolysis temperature was because of the decrease in volume of biochar during the pyrolysis at 200 °C result in the content of Cd increased. Subsequently, the Cd volatilization with increasing pyrolysis temperature.⁶⁰ Overall, the degree of contamination by heavy metals was mitigated by pyrolysis treatment. The GAI results for Cd, Cu, and Zn in PM were categorized as MC, MHC, and MC, separately. In contrast, the GAI results for Cd, Cu, and Zn in PMB after pyrolysis at 800 °C were classified as UC, UC,



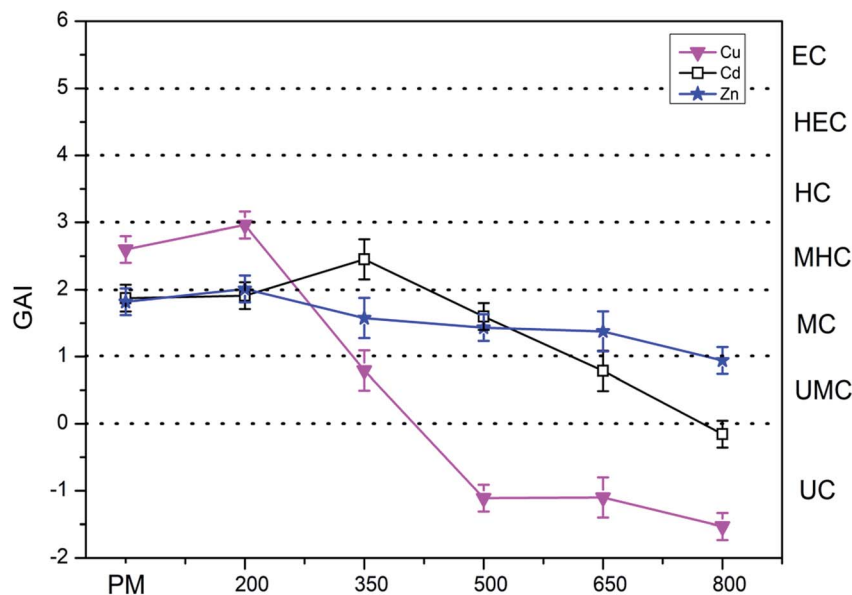


Fig. 1 Geo-accumulation index (GAI) of heavy metals in PM and biochars.

and UMC. Furthermore, the degree of contamination by Cd in SMB changed from UMC to UC upon increasing the pyrolysis temperature. Therefore, pyrolysis can effectively decrease the contamination grade at 800 °C.

3.3 Potential ecological risk assessment based on PERI

The ecological risk assessment consequence of the heavy metals in the samples are shown in Fig. 3 and 4. The result shown that the risk indices (E_r^i) of the heavy metals in PM were ranked in the following order: Cd > Cu > Zn > Ni > Pb > Mn > Cr. The E_r^i values of Cd and Cu in PM were 164.85 and 45.48, respectively, corresponding to VHR, which should be of major concern. The E_r^i of Zn in PM was 5.29, indicating MR to the local

environment. The E_r^i values of Ni, Pb, Mn, and Cr in PM were all below 5, suggesting LR to the local environment. However, the E_r^i values of Ni, Pb, Mn, and Cr in PMB were all below 5, indicating LR to the local environment. The E_r^i of Zn in PMB at 200 °C was 6.04, suggesting MR to the local environment. However, at higher pyrolysis temperatures, the E_r^i values of Zn in PMB were all below 5, indicating LR. The E_r^i values of Cu in PMB at 200 °C and 350 °C were 58.49 and 12.99, respectively, corresponding to VHR and CR to the local environment. The E_r^i values of Cd in PMB were all above 40, indicating VHR. Therefore, the treatment of Cd requires urgent attention. For SMB, the E_r^i values of all of the other heavy metals except Cd were below 5, suggesting LR to the environment. The E_r^i values of Cd

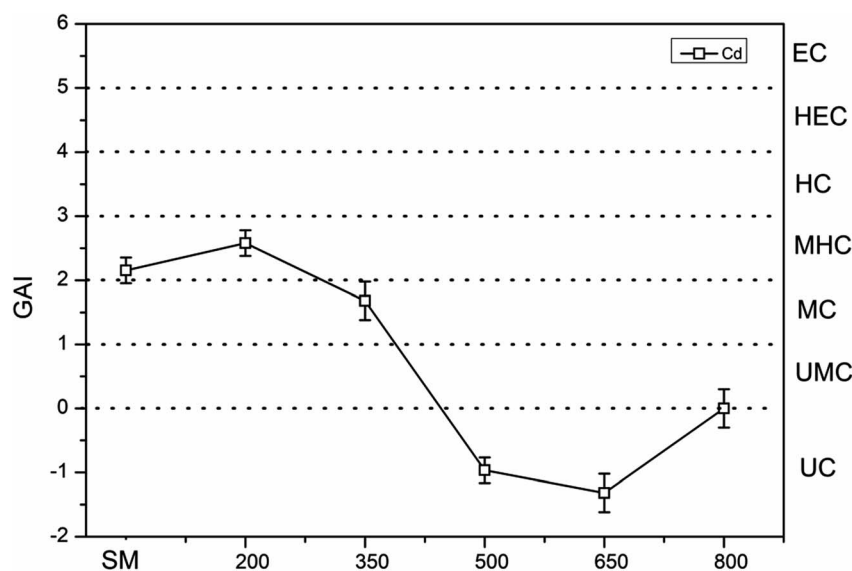


Fig. 2 Geo-accumulation index (GAI) of heavy metals in SM and biochars.



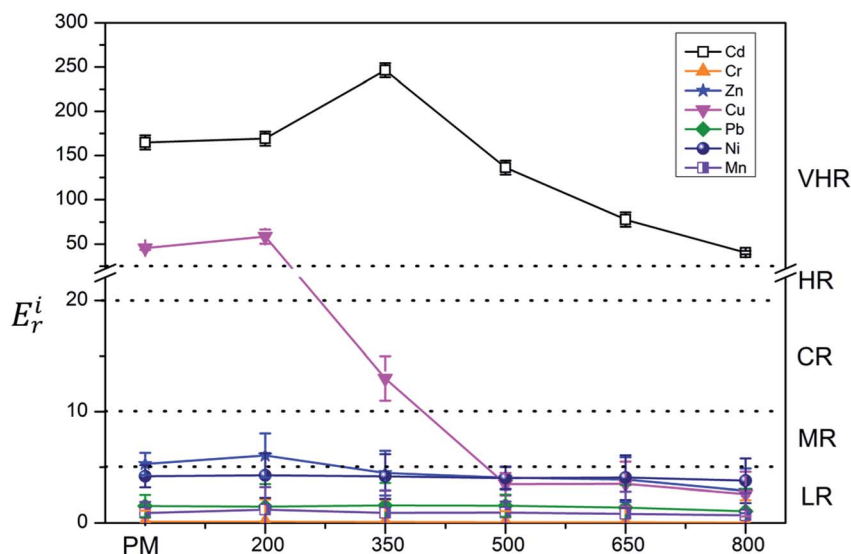


Fig. 3 Potential ecological risk factor of individual heavy metal (E_r^i) in PM and biochars.

at 200 °C, 350 °C, 500 °C, and 650 °C were 269.18 (VHR), 144.23 (VHR), 23.05 (HR), and 18.04 (CR), respectively. The reasons that the E_r^i values of Cd came under very high risk was because of relatively higher concentration of heavy metals in biochars with increasing pyrolysis temperature. Cd began to transfer to the off-gas and could not be prevented, while Cr, Ni, Cu, Zn, and Pb were retained quantitatively in the char during pyrolysis at temperatures up to 750 °C.⁶¹ In conclusion, the E_r^i values of the heavy metals in the biochars declined with increasing pyrolysis temperature. Pyrolysis temperature above 350 °C be regarded as the optimum condition. Therefore, pyrolysis temperature at 800 °C can provide an available method for reducing the risks of heavy metals.

To quantitate the overall underlying ecological risk from the heavy metals in PM, PMB, SM, and SMB, the PERI was

calculated as the total of all seven risk factors. Descriptive statistics of PERI are also summarized in Fig. 5 and 6. The PERI values of the heavy metals in PM and SM were 222.37 and 208.99, respectively, which correspond to VHR. After pyrolysis, the PERI values of PMB remained VHR at 200 °C, 350 °C, and 500 °C, although they decreased with increasing pyrolysis temperature from 650 °C to 800 °C. At pyrolysis temperatures of 650 °C and 800 °C, the PERI values were classified as CR and MR, respectively. Similarly, after pyrolysis treatment, the PERI values of SMB remained VHR at 200 °C and 350 °C and declined with increasing pyrolysis temperature from 500 °C to 800 °C. At pyrolysis temperatures of 500 °C, 650 °C and 800 °C, the PERI values were classified as LR. These results demonstrate that the PERI of biochar also decreases with increasing pyrolysis temperature. Pyrolysis temperature above 650 °C be regarded as

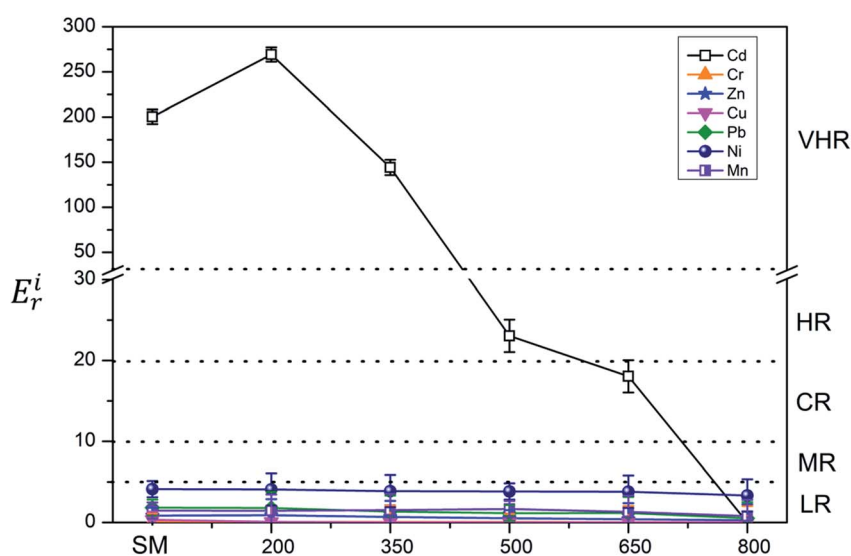


Fig. 4 Potential ecological risk factor of individual heavy metal (E_r^i) in SM and biochars.

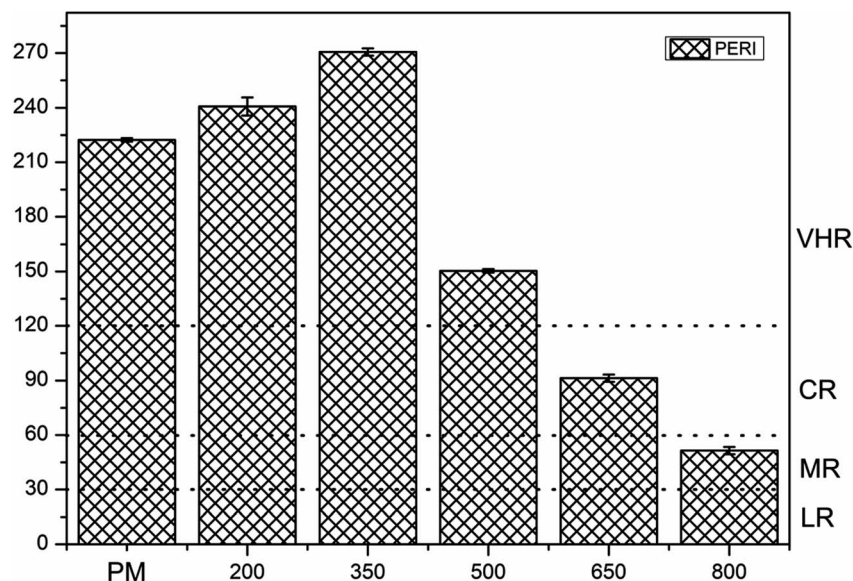


Fig. 5 Potential ecological risk index (PERI) of heavy metals in PM and biochars.

the optimum condition. PERI values can be used to describe the susceptibility of the local ecosystem to toxic heavy metals and indicate the ecological risk resulting from the overall pollution.⁶² For PM, SM, PMB, and SMB, the element Cd was found to be responsible for most of the total risk. It can be concluded that the high ecological risk of PM, SM, PMB, and SMB can be primarily attributed to Cd.

3.4 Environmental risk assessment based on RAC

Fig. 7 and 8 present the environmental risk evaluation results according to the RAC. For PM, the RAC of Mn was classified as VHR, whereas the RAC values of Cr, Pb, Zn, Cu, and Ni indicated MR and the RAC of Cd indicated HR. The RAC values of PMB

also decreased with increasing pyrolysis temperature. After pyrolysis, the classification of the RAC of Mn in PMB decreased to HR. The RAC values of Cd, Cr, Zn, Ni, and Pb in PMB were 9.90%, 1.62%, 1.05%, 9.74%, and 5.67%, indicating LR to the local environment. The RAC of Cu in PMB was 0.97%, suggesting NR to the local environment.

For SM, the RAC of Mn indicated VHR, whereas the RAC values of Zn, Cd, Ni, and Pb indicated MR and the RAC values of Cu and Cr indicated LR. Similarly, the RAC values of SMB decreased with increasing pyrolysis temperature. After pyrolysis, the RAC of Mn in SMB was 15.39%, suggesting MR to the environment. Meanwhile, the RAC values of Ni, Zn, Pb, and Cr in SMB were 9.32%, 3.47%, 2.33%, and 2.02%, respectively,

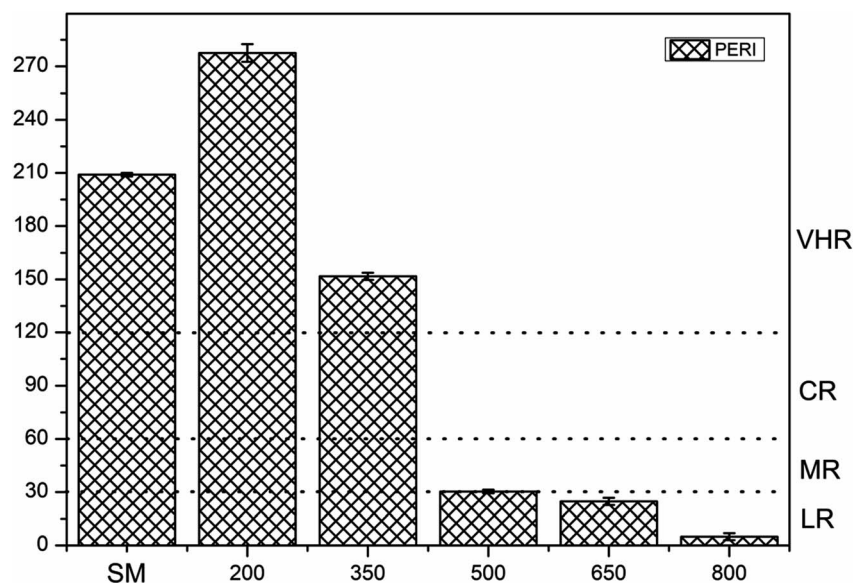


Fig. 6 Potential ecological risk index (PERI) of heavy metals in SM and biochars.



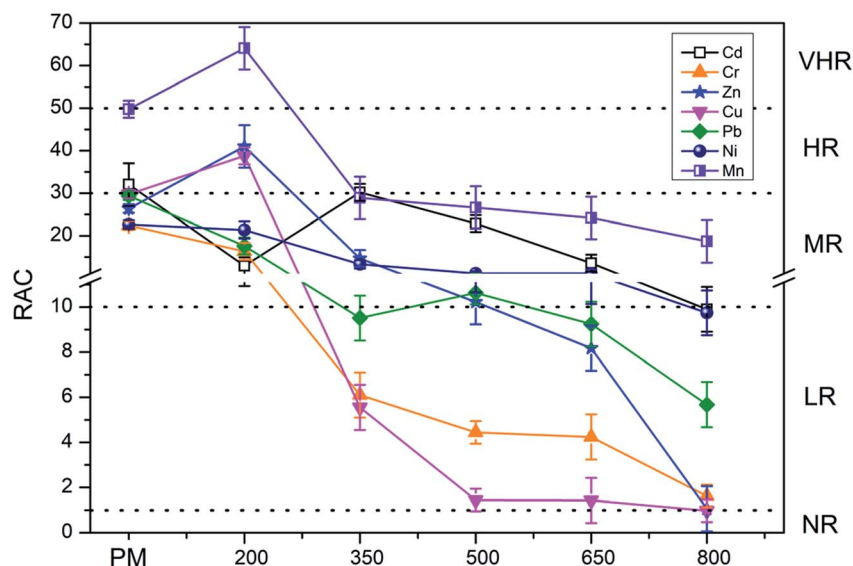


Fig. 7 Risk assessment code (RAC) of heavy metals in PM and biochars.

indicating LR to the environment, and the RAC values of Cu and Cd were 0, suggesting NR to the local environment. Heavy metals were categorized as the F2 fraction when bounded with Fe and Mn oxides and as the F3 fraction when bounded with organic matters and sulfides. When the content of Fe and Mn oxides, organic matter, and sulfides vary in biochars with increasing pyrolysis temperature, the heavy metal F2 and F3 extraction fractions can convert to other fractions, potentially leading to changes in the F1 fraction (the most unstable fraction).⁴¹ Overall, the RAC values indicate that the risk from heavy metals decreases with increasing pyrolysis temperature. Pyrolysis temperature above 500 °C be regarded as the optimum condition.

3.5 Comparison of assessment results based on GAI, PERI, and RAC

Comparison of the assessment results according to GAI, PERI, and RAC values revealed several inconsistencies between the three classifications with respect to evaluating heavy metal pollution in PM, SM, and livestock-manure-derived biochar. According to the GAI evaluation, the Cu in PM was classified as MHC. However, the E_r^i of the Cu in PM was classified as VHR. Furthermore, according to the GAI evaluation, the Cd in SM was classified as MHC, whereas the RAC indicated a HR due to the high percentage of Cd in the acid-soluble/exchangeable fraction (F1). In addition, the Cd in PMB was classified as VHR based on E_r^i and as MR based on the RAC. Similarly, the Cd in SMB was classified as LR based on E_r^i and as NR based on the RAC.

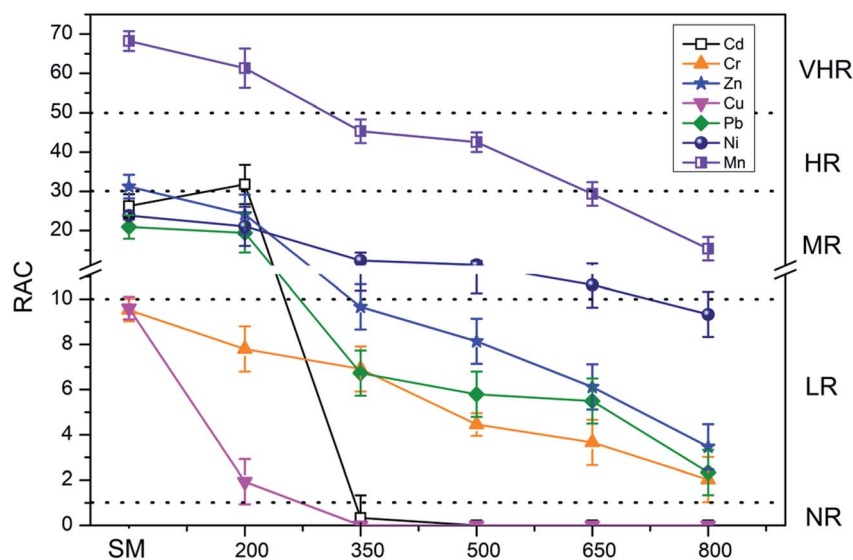


Fig. 8 Risk assessment code (RAC) of heavy metals in SM and biochars.

These inconsistencies can be explained as follows. The GAI considers only the accumulation levels of individual heavy metals without regard to their toxicity response factor. The PERI can be used to describe both the ecological risks caused by a single contamination and the integrated risk or pollution from multiple contaminations.³⁹ The RAC considers only the percentage of each pollutant in the F1 fraction, ignoring the percentage in the reducible fraction (F2). In contrast with the RAC, both the GAI and PERI consider the sum of metals bound to F2. Thus, it is essential to consider the results of all three assessment methods when attempting to evaluate the environmental impact, as this allows more integrated and accurate evaluation results to be obtained.

4. Conclusion

According to the results of bioavailability, ecotoxicity, and GAI, the risks of heavy metals in livestock-manure-derived biochar were decreased after pyrolysis. The PERI results demonstrated that increasing the pyrolysis temperature, results in the content of heavy metal in biochars decreased, reducing its the risk of contamination. The toxicity of the heavy metals also decreased with increasing pyrolysis temperature. However, according to the potential ecological risk analysis, the Cd in PMB still presents a VHR to the local environment. Therefore, the removal of Cd from PMB should be addressed. In summary, pyrolysis can be used to mitigate the toxicity and environmental risk of heavy metals in livestock-manure-derived biochar when pyrolysis temperature above 350 °C. Furthermore, pyrolysis temperature at 800 °C can provide an effective method for treating agricultural waste.

Conflicts of interest

There are no conflicts to declare.

Acknowledgements

This research was financially supported by the National Key Research and Development Programs of China (Grant No. 2017YFD0301505 and 2018YFC0213300), the Natural Science Foundation of Hunan Province, China (Grant No. 2018JJ3222 and 2017NK2383), the Special Fund for Local Scientific and Technological Development under the Leadership of the Central Government (Grant No. 2017CT5010), the Science Foundation of Hunan Education Department, China (Grant No. CX2017B386, 18B097, CX2018B405 and S201910537053) and the Double First-class Construction Project of Hunan Agricultural University (Grant No. kxk201801007, YB2018005, 18QN19, SYL2019025, SYL2019026).

References

- 1 Z. Tian, B. U. Meidong and G. Wei, *Chinese J. Ecol.*, 2012, **31**, 1241–1249.
- 2 N. Zhu, *Bioresour. Technol.*, 2006, **97**, 1870–1875.

- 3 W. Ouyang, X. Xu, Z. Hao and X. Gao, *J. Hydrol.*, 2017, **546**, 71–80.
- 4 M. Hamidpour, E. Khadivi and M. Afyuni, *Environ. Earth Sci.*, 2016, **75**, 1037.
- 5 X. Zeng, D. Zou, A. Wang, Y. Zhou, Y. Liu, Z. Li, F. Liu, H. Wang, Q. Zeng and Z. Xiao, *J. Environ. Manage.*, 2020, **255**, 109885.
- 6 Z. Xiao, X. Yuan, H. Li, L. Jiang, L. Leng, X. Chen, G. Zeng, F. Li and L. Cao, *Sci. Total Environ.*, 2015, **536**, 774–783.
- 7 Z. Xiao, X. Yuan, L. Jiang, X. Chen, H. Li, G. Zeng, L. Leng, H. Wang and H. Huang, *Energy*, 2015, **91**, 441–450.
- 8 C. Liu, Q. Tong, Y. Li, N. Wang, B. Liu and X. Zhang, *RSC Adv.*, 2019, **9**, 4488–4498.
- 9 X. Zeng, Z. Xiao, G. Zhang, A. Wang, Z. Li, Y. Liu, H. Wang, Q. Zeng, Y. Liang and D. Zou, *J. Anal. Appl. Pyrolysis*, 2018, **132**, 82–93.
- 10 Y. Niu, H. Tan and S. Hui, *Prog. Energy Combust. Sci.*, 2016, **52**, 1–61.
- 11 W.-J. Liu, W.-W. Li, H. Jiang and H.-Q. Yu, *Chem. Rev.*, 2017, **117**, 6367–6398.
- 12 Q. Lin, X. Xu, L. Wang, Q. Chen, J. Fang, X. Shen, L. Lou and G. Tian, *Front. Environ. Sci. Eng.*, 2017, **11**, 5.
- 13 J. Meng, L. Wang, L. Zhong, X. Liu, P. C. Brookes, J. Xu and H. Chen, *Chemosphere*, 2017, **180**, 93–99.
- 14 K. B. Cantrell, T. Ducey, K. S. Ro and P. G. Hunt, *Bioresour. Technol.*, 2008, **99**, 7941–7953.
- 15 Y. Wang, B. Zhong, M. Shafi, J. Ma, J. Guo, J. Wu, Z. Ye, D. Liu and H. Jin, *Chemosphere*, 2019, **219**, 510–516.
- 16 L. He, H. Zhong, G. Liu, Z. Dai, P. C. Brookes and J. Xu, *Environ. Pollut.*, 2019, **252**, 846–855.
- 17 R. SHI, J. LI, N. NI and R. XU, *J. Integr. Agric.*, 2019, **18**, 1508–1517.
- 18 H. Nacke, A. C. Gonçalves, M. A. Campagnolo, G. F. Coelho, D. Schwantes, M. G. Dos Santos, D. L. Briesch and J. Zimmermann, *Open Chem.*, 2016, **14**, 103–117.
- 19 L. Li, D. Zou, Z. Xiao, X. Zeng, L. Zhang, L. Jiang, A. Wang, D. Ge, G. Zhang and F. Liu, *J. Cleaner Prod.*, 2019, **210**, 1324–1342.
- 20 M. R. Awual, N. H. Alharthi, M. M. Hasan, M. R. Karim, A. Islam, H. Znad, M. A. Hossain, M. E. Halim, M. M. Rahman and M. A. Khaleque, *Chem. Eng. J.*, 2017, **324**, 130–139.
- 21 M. R. Awual, *J. Environ. Chem. Eng.*, 2019, **7**, 103124.
- 22 M. R. Awual, *Mater. Sci. Eng., C*, 2019, **101**, 686–695.
- 23 Z. Xiao, X. Yuan, L. Leng, L. Jiang, X. Chen, W. Zhibin, P. Xin, Z. Jiachao and G. Zeng, *Environ. Sci. Pollut. Res.*, 2016, **23**, 3934–3942.
- 24 F. Li, J. Zhang, W. Liu, J. Liu, J. Huang and G. Zeng, *Sci. Total Environ.*, 2018, **644**, 649–660.
- 25 F. Li, J. Zhang, C. Liu, M. Xiao and Z. Wu, *Process Saf. Environ. Prot.*, 2018, **116**, 169–179.
- 26 A. Makara, Z. Kowalski and A. Saeid, *Open Chem.*, 2017, **15**, 19–27.
- 27 Z. Xiao, X. Yuan, H. Li, L. Jiang, L. Leng, X. Chen, G. Zeng, F. Li and L. Cao, *Sci. Total Environ.*, 2015, **536**, 774–783.
- 28 T. G. Kazi, M. K. Jamali, A. Siddiqui, G. H. Kazi, M. B. Arain and H. I. Afridi, *Chemosphere*, 2006, **63**, 411–420.



- 29 M. Chen, X. ming Li, Q. Yang, G. ming Zeng, Y. Zhang, D. xiang Liao, J. jin Liu, J. mei Hu and L. Guo, *J. Hazard. Mater.*, 2008, **160**, 324–329.
- 30 X. Liu, Y. Wang, C. Gui, P. Li, J. Zhang, H. Zhong and Y. Wei, *RSC Adv.*, 2016, **6**, 101960–101967.
- 31 F. Li, Y. Cai and J. Zhang, *Sustainability*, 2018, **10**, 91.
- 32 Z. Sun and J. Chen, *Int. J. Environ. Res. Public Health*, 2018, **15**, 85.
- 33 J. Zhang, *Hunan Geol.*, 1994, **12**, 185–191.
- 34 L. Hakanson, *Water Res.*, 1980, **14**, 975–1001.
- 35 X. Xu, X. Hu, Z. Ding and Y. Chen, *Chemosphere*, 2017, **189**, 76–85.
- 36 L. Wu, G. Liu, C. Zhou, R. Liu, S. Xi, C. Da and F. Liu, *Environ. Sci. Pollut. Res.*, 2018, **25**, 588–600.
- 37 L. Mao, H. Ye, F. Li, H. Tao, L. Shi, L. Chen, W. Shi, G. Yan and H. Chen, *Mar. Pollut. Bull.*, 2017, **125**, 389–398.
- 38 X. Z. Ren Fumin, Z. Yusong and N. Muchen, *J. Beijing Jiaotong Univ.*, 2007, **31**, 102–105.
- 39 H. Huang, X. Yuan, G. Zeng, H. Zhu, H. Li, Z. Liu, H. Jiang, L. Leng and W. Bi, *Bioresour. Technol.*, 2011, **102**, 10346–10351.
- 40 Y. Lin, J. Ma, Z. Zhang, Y. Zhu, H. Hou, L. Zhao, Z. Sun, W. Xue and H. Shi, *Sci. Total Environ.*, 2018, **610–611**, 944–950.
- 41 G. Liang, B. Zhang, M. Lin, S. Wu, H. Hou, J. Zhang, G. Qian, X. Huang and J. Zhou, *Sci. Total Environ.*, 2017, **607–608**, 1339–1347.
- 42 S. K. Sundaray, B. B. Nayak, S. Lin and D. Bhatta, *J. Hazard. Mater.*, 2011, **186**, 1837–1846.
- 43 Z. Gu, M. Wu, K. Li and P. Ning, *Arabian J. Chem.*, 2017, **10**, S2196–S2204.
- 44 J. Paz-Ferreiro, P. Plasencia, G. Gascó and A. Méndez, *Land Degrad. Dev.*, 2017, **28**, 355–360.
- 45 F. Wang, Q. Bu, X. Xia and M. Shen, *Environ. Pollut.*, 2011, **159**, 1905–1913.
- 46 A. Zielińska and P. Oleszczuk, *Biomass Bioenergy*, 2015, **75**, 235–244.
- 47 K. Lin, N. Lai, J. Zeng and H. Chiang, *Environ. Sci. Pollut. Res.*, 2018, **25**, 6487–6496.
- 48 M. Benito Abascal, M. Bläsing, Y. Ninomiya and M. Müller, *Ind. Eng. Chem. Res.*, 2016, **55**, 6911–6921.
- 49 J. Liu, Z. Zhuo, S. Sun, W. Xie, S. Lu, J. Sun, J. Kuo and W. Yujie, *Water Environ. Res.*, 2016, **88**, 2245–2256.
- 50 H. Belevi and M. Langmeier, *Environ. Sci. Technol.*, 2000, **34**, 2507–2512.
- 51 S. Abanades, G. Flamant and D. Gauthier, *J. Hazard. Mater.*, 2001, **88**, 75–94.
- 52 H. Okyere, R. B. Voegborlo and S. E. Agorku, *Food Chem.*, 2015, **179**, 331–335.
- 53 J.-G. Jang, W.-H. Kim, M.-R. Kim, H.-S. Chun and J.-K. Lee, *Korean J. Chem. Eng.*, 2001, **18**, 506–511.
- 54 I. S. Mkilaha, H. Yao and I. Naruse, *J. Mater. Cycles Waste Manage.*, 2002, **4**, 143–149.
- 55 D. Pudasainee, Y. C. Seo, J. H. Kim and H. N. Jang, *J. Mater. Cycles Waste Manage.*, 2013, **15**, 202–209.
- 56 J. Luan, R. Li, Z. Zhang, Y. Li and Y. Zhao, *Waste Manage. Res.*, 2013, **31**, 1012–1018.
- 57 M. R. Awual, *Chem. Eng. J.*, 2016, **303**, 539–546.
- 58 M. R. Awual, *Chem. Eng. J.*, 2017, **307**, 85–94.
- 59 M. R. Awual, *Chem. Eng. J.*, 2017, **307**, 456–465.
- 60 J. Du, L. Zhang, T. Liu, R. Xiao, R. Li, D. Guo, L. Qiu, X. Yang and Z. Zhang, *Bioresour. Technol.*, 2019, **274**, 73–82.
- 61 X. Yuan, L. Leng, H. Huang, X. Chen, H. Wang, Z. Xiao, Y. Zhai, H. Chen and G. Zeng, *Chemosphere*, 2015, **120**, 645–652.
- 62 G. Shi, Z. Chen, C. Bi, Y. Li, J. Teng, L. Wang and S. Xu, *Environ. Pollut.*, 2010, **158**, 694–703.

