



Cite this: RSC Adv., 2019, 9, 32462

Adsorption mechanism of polyethyleneimine modified magnetic core–shell $\text{Fe}_3\text{O}_4@\text{SiO}_2$ nanoparticles for anionic dye removal

Binyan Huang,^{ab} Yunguo Liu,^{cd} Bin Li,^e Hui Wang^f and Guangming Zeng^{cd}

In this paper, polyethyleneimine modified magnetic core–shell $\text{Fe}_3\text{O}_4@\text{SiO}_2$ nanoparticles ($\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$) were innovatively synthesized and investigated using various techniques such as TEM, TGA, FT-IR, XRD, VSM and XPS. The adsorption performance based on the removal of the anionic dyes Methyl orange and Congo red from aqueous solution was studied systematically. The results showed that the adsorption rate of anionic dyes MO and CR increased rapidly then decreased gradually as the pH increased, the adsorption capacity of $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ for MO was better than that for CR, and the maximum adsorption capacity for MO and CR was 231.0 mg g^{-1} at pH 4 and 134.6 mg g^{-1} at pH 6, respectively. The equilibrium adsorption capacities for MO and CR increased rapidly in the initial 40 min, and reached equilibrium in approximately 180 min, while the adsorption capacity for MB was relative low even negligible, demonstrating the strong adsorptive affinity of $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ toward anionic compounds. Both of the adsorption processes followed the pseudo-second-order kinetics model and the Freundlich isotherm model. The mechanism of adsorption was mainly related to electrostatic attraction and the number of active sites occupied by anionic dyes. This study provides valuable guidance and is an effective method for the removal of anionic dyes from aquatic environments.

Received 13th August 2019
Accepted 6th October 2019DOI: 10.1039/c9ra06299h
rsc.li/rsc-advances

Introduction

With the rapid growth of modern industry, organic dyes have found increasing application in wide-ranging industrial fields, including textiles, plastics, paper, cosmetics and leather manufacturing, as well as in laboratory research.^{1,2} However, the application of organic dyes also causes serious water pollution and potential toxicity, mutagenicity, even carcinogenicity to humans and the environment.^{3,4} In particular, anionic dyes are a major cause for concern because of their high solubility, carcinogenic and mutagenic activity and their high consumption in the textile industry.⁵ For example, Methyl orange (MO) is a typical anionic dye for both industrial products and laboratory research. It is also difficult to degrade due to its complex structure.^{6,7} Congo red (CR), containing two azo functional groups, is an industrial synthetic colorant, which has been

known to be metabolised to benzidine and to cause allergic reaction. Meanwhile, due to the poor biodegradability, this utilization also causes serious environmental pollution.^{8,9} Various physical-chemical processes have already been established for the separation of toxic dyes from the aqueous environment.^{10,11} However, these processes are not always economical and effective because of the complex structure of the dyes and their low biodegradability and recalcitrance.¹² Among various possible methods to remove dyes from industrial effluents, the adsorption is considered as the most promising alternative than others in terms of the initial cost, ease of operation and the insensitive to toxic substances.^{13,14}

Polyethyleneimine (PEI), is known to contain a large amount of nitrogen atoms of the amino groups on its macromolecular chains and it is protonated at $\text{pH} < 10$.¹⁵ Therefore, it is an ideal adsorbent of anionic contaminants by electrostatic interaction.^{16,17} Nevertheless, the PEI is preferably soluble in water, so a solid support is required for the attachment of PEI to realize the adsorption performance.¹⁸ Despite the surface of SiO_2 has abundance of hydroxyl groups, which are able to link the special functional groups of PEI, it is hard to be separated from the aqueous solution. To overcome this drawback, introducing magnetic performance to the adsorbent can be a research hot-spot, and magnetic nanoparticles, especially the Fe_3O_4 , have attracted growing concern because of the outstanding characteristics such as low toxicity and easy separation.¹⁹ Recently, the magnetic microspheres with core–shell structured have been

^aCollege of Life Sciences and Environment, Hengyang Normal University, Hengyang 421008, P. R. China

^bInstitute of Bio-technology for Livestock, Hengyang Normal University, Hengyang 421008, P. R. China

^cCollege of Environmental Science and Engineering, Hunan University, Changsha 410082, P. R. China. E-mail: liuyunguo@163.com

^dKey Laboratory of Environmental Biology and Pollution Control (Hunan University), Ministry of Education, Changsha 410082, P. R. China

^eChangsha Environment Protection Agency, Changsha 410001, P. R. China

^fCollege of Environmental Science and Engineering, Central South University of Forestry and Technology, Changsha 410004, P. R. China



widely investigated and applied in catalysis and adsorption.^{20,21} The iron oxide as the core, which could offer high efficiency of separation for convenient recycling, and the silica as the shell, which could offer the active sites of modification, and protect the iron oxide from certain bad reaction conditions. At the same time, it is also avoid the problem that a large amount of adsorption sites are occupied due to the introduction of magnetic particles.²²

In this study, a $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ nanocomposite was innovatively synthesized by using Fe_3O_4 as the core and SiO_2 with abundance of surface hydroxyl groups as the shell to form typical core–shell magnetic nanoparticles, on which polymeric PEI was coated on the surface of the magnetic nanoparticles through a cross-linking reaction. To our best knowledge, the analysis of adsorption mechanism by the magnetic PEI to remove anionic dyes from aqueous solutions is still rare, especially for the difference in adsorption of anionic dyes. Therefore, the objective of this paper were to: (1) prepare and characterize polyethylenimine modified magnetic core–shell $\text{Fe}_3\text{O}_4@\text{SiO}_2$ nanoparticles ($\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$) and apply it as an adsorbent for removal of anionic dyes from aqueous solution; (2) investigate the relevant parameters, including pH, adsorption kinetics and adsorption isotherms to analyze the adsorption performance of $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ for MO and CR in aqueous solutions; and (3) discuss the $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ adsorption mechanisms for MO and CR, especially the difference in adsorption between them, furthermore, the regeneration and reutilization of $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ were also evaluated.

Materials and methods

Materials

A branched polyethylenimine (PEI) with a molecular weight of 25 000 was obtained from Sigma-Aldrich as a commercial sample in the form of 99 wt% aqueous solution. The anionic dyes MO and CR were obtained from Tianjin Fuchen Co., Ltd., China. The parameters of the anionic dyes are shown in Table 1. $\text{FeCl}_3 \cdot 6\text{H}_2\text{O}$, ethylsilicate (TEOS), anhydrous sodium acetate (CH_3COONa), glutaraldehyde solution (50 wt%) and all other chemicals used in this study were purchased from Sinopharm Chemical Reagent Co., Ltd. (Shanghai, China). All the reagents used in the experiments were of analytic quality.

Preparation of $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ nanocomposite

$\text{Fe}_3\text{O}_4@\text{SiO}_2$ was prepared following the methods described in our previous paper.²³ The preparation process of $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ following the method reported in the literature.²⁴ Briefly,

Table 1 The physical properties of the investigated anionic dyes

Name	Methyl orange (MO)	Congo red (CR)
Molecular formula	$\text{C}_{14}\text{H}_{14}\text{N}_3\text{SO}_3\text{Na}$	$\text{C}_{32}\text{H}_{22}\text{N}_6\text{Na}_2\text{O}_6\text{S}_2$
Molar mass (g mol ⁻¹)	327.33	696.68
λ_{max} (nm)	464	497
CAS number	547-58-0	573-58-0

0.27 g of $\text{Fe}_3\text{O}_4@\text{SiO}_2$ was immersed in 63 mL of PEI in the methanol for 30 min by ultrasonication, then 100 mL glutaraldehyde solution (2% w/v) was added dropwise to the mixture solution and mechanically stirred for 30 min at room temperature. Next, the mixture was washed with distilled water several times until the supernatant became clear. The final washed $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ nanocomposite was dried in the refrigeration dryer for the following characterization and adsorption experiments.

Characterization techniques

The morphological investigations of the samples were observed by the transmission electron microscopy (TEM; Tecnai G2 F20, USA). Thermo Gravimetric Analyzer (TGA, Q500, USA) was utilized for the identification of the thermal

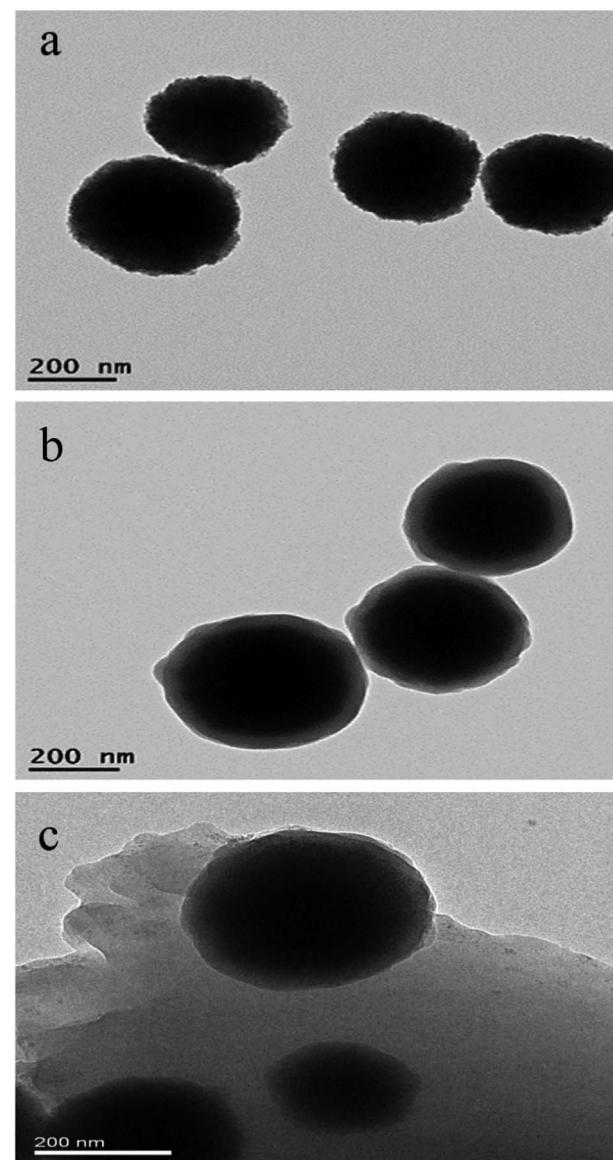


Fig. 1 The TEM images of (a) Fe_3O_4 , (b) $\text{Fe}_3\text{O}_4@\text{SiO}_2$ and (c) $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$.



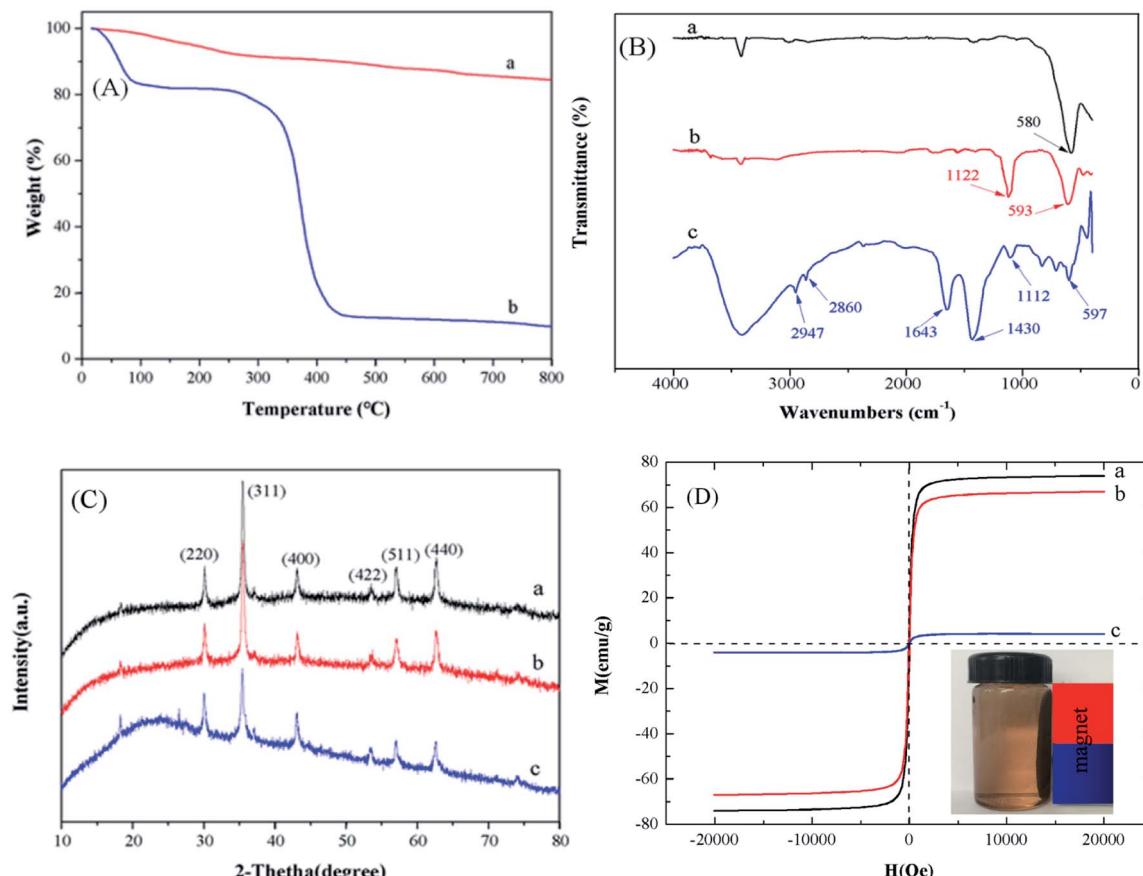


Fig. 2 (A) TGA analysis of (a) Fe₃O₄@SiO₂ and (b) Fe₃O₄@SiO₂/PEI. (B) FT-IR spectra of (a) Fe₃O₄, (b) Fe₃O₄@SiO₂, (c) Fe₃O₄@SiO₂/PEI. (C) XRD patterns of (a) Fe₃O₄, (b) Fe₃O₄@SiO₂, (c) Fe₃O₄@SiO₂/PEI. (D) Magnetization curve of (a) Fe₃O₄, (b) Fe₃O₄@SiO₂, (c) Fe₃O₄@SiO₂/PEI.

properties in the range of 0 to 800 °C (heating rate was 10 °C min⁻¹) under nitrogen atmosphere. The surface functional groups of the materials were characterized by Fourier transform infrared spectrum (FT-IR; Nicolet Magna-IR 750,

USA) at room temperature. Surface elemental composition of the samples were analyzed using a X-ray photoelectron spectroscopy (XPS; PerkinElmer PHI-5400, USA). The crystal structure of the samples were analyzed by X-ray diffraction

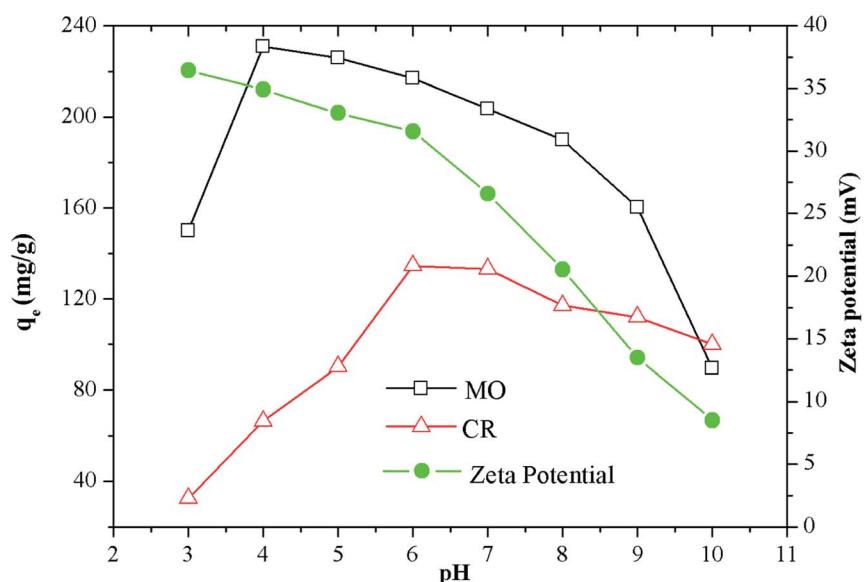
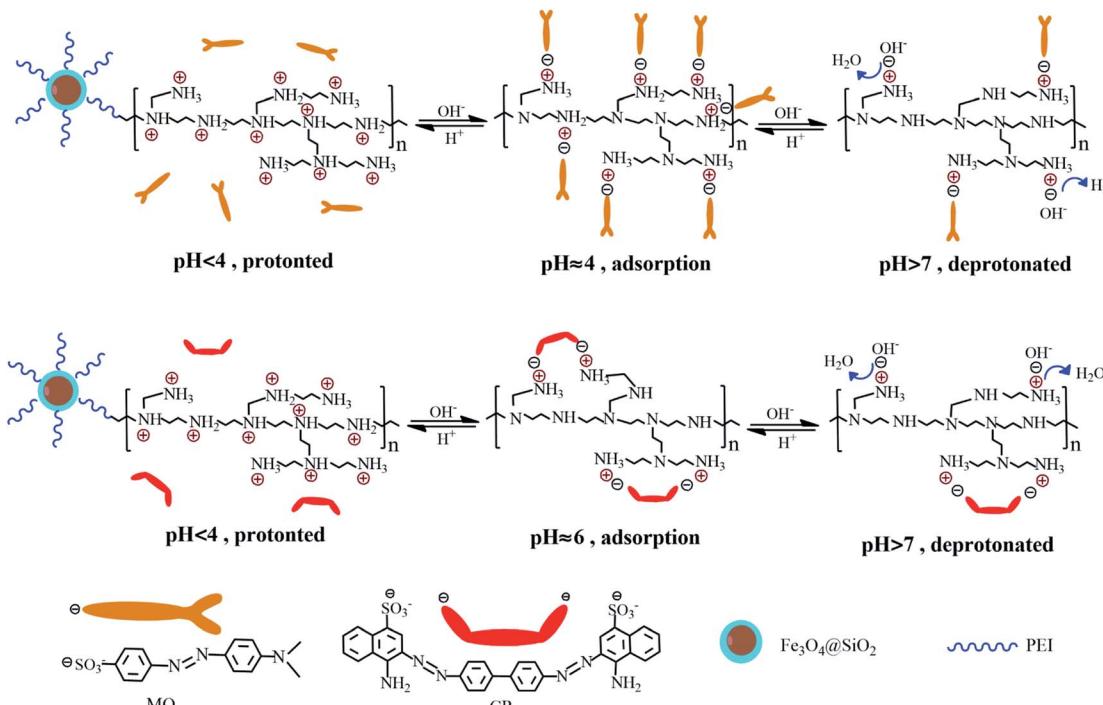


Fig. 3 Zeta potentials of Fe₃O₄@SiO₂/PEI at different pH and effect of pH on adsorption of MO and CR by Fe₃O₄@SiO₂/PEI.



Scheme 1 Schematic adsorption mechanism for the removal of MO and CR.

spectrometer (XRD; Bruker D8 ADVANCE, German). Magnetic property of the samples was analyzed using a vibration sample magnetometer (VSM; Quantum Design Instruments, USA) at room temperature. Zeta potentials of $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ was determined using a Zeta-sizer Nano-ZS (Malvern, UK) under different pH.

Batch adsorption tests

The batch experiments were performed to study the adsorption properties of the $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ nanocomposite for dyes at room temperature. Typically, 20 mg of the magnetic adsorbent ($\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$) was added into 50 mL of two kinds of anionic dye solutions (MO and CR were 100 mg L^{-1} , respectively) and one kind of cationic dye solution (MB was 100 mg L^{-1}). After the adsorption process, the mixture was drawn by a permanent magnet and separated immediately, and the absorbance value of MO, CR and MB was measured using a UV spectrophotometer (UV-2550, SHIMADZU, Japan) at 464, 497 and 664 nm, respectively. The solution pH was adjusted as necessary by adding a small amount of 1 mol L^{-1} HCl or 1 mol L^{-1} NaOH during the static adsorption. The adsorption capacity and the removal efficiency of dyes were calculated according to eqn (1) and (2):

Adsorption capacity

$$q_e = \frac{(C_0 - C_e)V}{m} \quad (1)$$

$$\text{Removal efficiency (\%)} = \frac{C_0 - C_e}{C_0} \times 100 \quad (2)$$

where C_0 and C_e denote the initial and equilibrium dye concentrations (mg L^{-1}), respectively. m is the mass of the sorbent (g), V is the volume of the solution (L).

All the experiments were replicated and only the mean values were presented. The relative errors were less than 5%.

Desorption and regeneration studies

The regeneration of $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ nanocomposite was carried out by desorbing MO and CR with 50 mL of 0.5 mol L^{-1} NaOH solution, respectively, and then sonicated for 20 min.

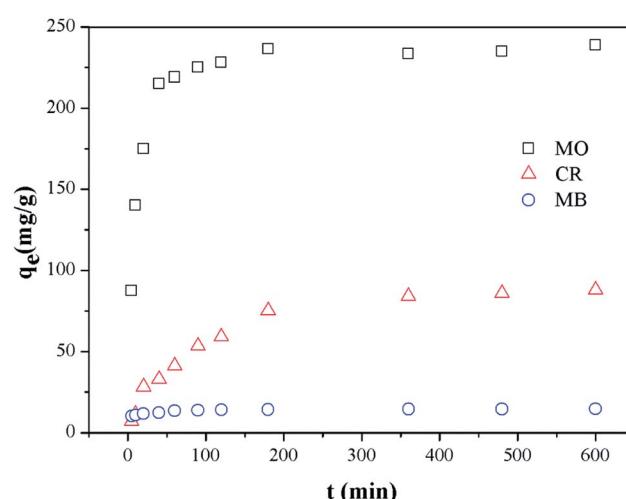


Fig. 4 Effect of contact time on adsorption of MO and CR by $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ ($m/V = 0.4 \text{ g L}^{-1}$, $C_0 = 100 \text{ mg L}^{-1}$, $T = 298 \text{ K}$, $t = 10 \text{ h}$).



Table 2 Parameters for different adsorption kinetics models

Dye	Pseudo-first-order kinetic			Pseudo-second-order kinetic			
	q_{exp} (mg g ⁻¹)	q_{cal} (mg g ⁻¹)	k_1 (min ⁻¹)	R^2	q_{cal} (mg g ⁻¹)	k_2 (g mg ⁻¹ min ⁻¹)	R^2
MO	238.8	47.25	7.07×10^{-3}	0.725	239.8	6.41×10^{-4}	0.999
CR	88.01	71.20	3.65×10^{-3}	0.903	98.91	1.43×10^{-4}	0.997

Finally, the $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ nanocomposites with desorbed anionic dyes were again collected onto the permanent magnet, washed with ultrapure water, and dried at 328 K for the next reusability cycle. The amount of residual MO and CR on the sorbent was analyzed.

Results and discussion

Characterization of $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$

TEM images shown in Fig. 1 perceptually reflect the successful preparation of the Fe_3O_4 , $\text{Fe}_3\text{O}_4@\text{SiO}_2$ and $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$. As shown in Fig. 1a and b, the average diameter of Fe_3O_4 and the thickness of silica were around 300 nm and 50 nm, the prepared $\text{Fe}_3\text{O}_4@\text{SiO}_2$ nanoparticles have rather obvious core-shell structure and relatively smooth surface. After the combination with $\text{Fe}_3\text{O}_4@\text{SiO}_2$ and PEI to form the $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ nanocomposite, the surface of the $\text{Fe}_3\text{O}_4@\text{SiO}_2$ nanoparticles were coated by a PEI layer as shown in Fig. 1c.

To examine the thermal properties of the $\text{Fe}_3\text{O}_4@\text{SiO}_2$ and $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$, TGA was carried out between 0 °C and 800 °C in a static nitrogen atmosphere (Fig. 2A). It can be seen from the TGA curve of $\text{Fe}_3\text{O}_4@\text{SiO}_2$ that the mass loss before 200 °C is might due to the releasing of the physically adsorbed water.²⁵ And the mass loss between 150 and 790 °C is estimated to be 9.9%, which reveals the good thermal property of $\text{Fe}_3\text{O}_4@\text{SiO}_2$. As for the $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$, it is clearly seen that the mass loss occurred in two steps. The first weight loss occurred below 100 °C involves the loss of adsorbed solvent or the trapped water from the adsorbent. The second weight loss (250–430 °C) can be attributed to the decomposition of the main organic moiety

loaded on the surface of $\text{Fe}_3\text{O}_4@\text{SiO}_2$. Similar observations were reported on the Fe_3O_4 MNPs grafted poly(methyl methacrylate), which obtained 43.1% to 82.5% weight loss after the increasing of temperature from 250 to 420 °C.²⁶ Based on the analysis of the weight loss, it can be concluded that the preparation of $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ is successful.

The FT-IR spectra analysis of Fe_3O_4 , $\text{Fe}_3\text{O}_4@\text{SiO}_2$ and $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ were shown in Fig. 2B. For Fe_3O_4 , the strong characteristic peaks at 580 cm⁻¹ was attributed to the absorption band of Fe–O bond vibration.²⁷ Apart from the peak at 593 cm⁻¹, the spectrum of $\text{Fe}_3\text{O}_4@\text{SiO}_2$ appeared a strong band at 1122 cm⁻¹ due to the Si–O stretching vibration,²⁸ indicating the successful coating of silica on the magnetite surface. After the functionalization of PEI, a less intense Si–O band at 1112 cm⁻¹ in $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ could be due to the loss of silanol groups upon functionalization. Moreover, absorption bands at 2947 and 2860 cm⁻¹ were indicative of the symmetric and asymmetric stretching vibration of C–H bonds.²⁹ One additional peak at 1643 cm⁻¹ is associated to the symmetric N–H band from the PEI, which gives strong evidence of $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$.¹⁷

XRD pattern of Fe_3O_4 , $\text{Fe}_3\text{O}_4@\text{SiO}_2$ and $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ were illustrated in Fig. 2C. The strong diffraction peaks indexed to (220), (311), (400), (422), (511) and (440) show the characteristics of Fe_3O_4 with a cubic inverse spinel structure. These characteristic peaks of Fe_3O_4 were also appeared in the $\text{Fe}_3\text{O}_4@\text{SiO}_2$ and $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$. In addition, the results have showed that the coating with SiO_2 and PEI process did not cause the crystalline state change of Fe_3O_4 . So, we can summarized from XRD that Fe_3O_4 nanoparticles in the systems are also of the spinel type structure.

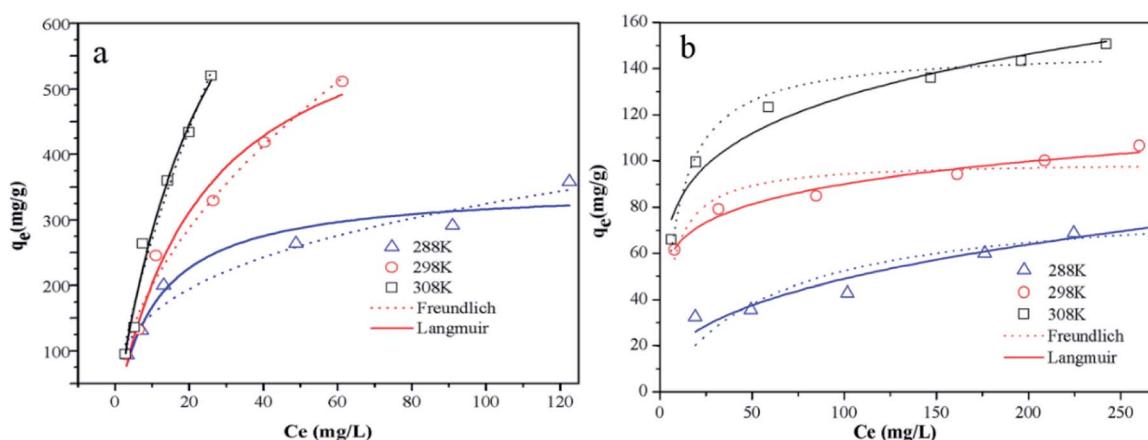


Fig. 5 The adsorption isotherms of (a) MO and (b) CR onto $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ at different temperatures ($m/V = 0.4 \text{ g L}^{-1}$, $t = 10 \text{ h}$).



Table 3 Adsorption isotherm parameters for the adsorption of MO and CR on the $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$

Dye	T (K)	Langmuir model		Freundlich model			
		q_{max} (mg g ⁻¹)	K_L (L mg ⁻¹)	R^2	K_F [(mg g ⁻¹)/(mg L ⁻¹) ^{1/n}]	1/n	R^2
MO	288	351.1	0.091	0.936	75.3	0.317	0.951
	298	681.2	0.042	0.974	59.7	0.525	0.981
	308	1058.4	0.036	0.968	54.4	0.697	0.977
CR	288	84.39	0.017	0.766	8.42	0.383	0.924
	298	99.75	0.173	0.802	45.8	0.147	0.969
	308	148.2	0.114	0.956	52.74	0.193	0.965

The magnetic properties of Fe_3O_4 , $\text{Fe}_3\text{O}_4@\text{SiO}_2$ and $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ were determined by VSM (Fig. 2D). The saturation magnetization is 74.0 emu g⁻¹ for Fe_3O_4 , after coating of silica, introduction of amount of PEI, it is reasonable that the saturation magnetization of $\text{Fe}_3\text{O}_4@\text{SiO}_2$ and $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ were decreased to 67.0 and 4.20, respectively. It should be noticed that near-zero remanence and coercivity were existed, indicating the superparamagnetic feature of the samples at room temperature.³⁰ After the adsorption process is completed, the $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ can be collected from a mixture solution after dye adsorption by a permanent magnet, which is very important for the convenient recycling.³¹

Effect of pH

As shown in Fig. 3, the pH_{PZC} of $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ estimated by the zeta potential was positive at pH 3.0–10.0, and the value was gradually decreased as the pH increased. It is believed that PEI is modeled at three protonation levels with three types of amino groups: at the high pH (>10), PEI is fully deprotonated; at neutral pH (~7), all the primary amines are protonated; at low pH (<4), most of the amines are protonated.³² MO ($pK_a = 3.4$) is an anionic dye with negatively charged when pH > 3.4. As can be seen from Fig. 3, the maximum adsorption of MO on the $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ surface was 231.0 mg g⁻¹ at pH 4, and the adsorption ability was gradually decreased as the pH increased, this might due to the electrostatic attractions between the positively charged groups of PEI and the negatively charged groups of MO. It can be interpreted as follows (Scheme 1). When the solution pH is below the pK_a value of MO, the surface of $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ is positively charged, while the MO was positively charged or neutral, indicating that there is no electrostatic attraction between $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ and MO. When the solution pH increases from 3.4, the anionic property of MO increase (formation of SO_3^- groups), there is electrostatic attractions between the positively charged groups of PEI and the negatively charged groups of MO. In particular, because the pH_{PZC} of $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ is higher at pH 4. This makes MO easier to combine with the composite. While the adsorption capacity was gradually decreased as the pH increased. However, when pH > 7, the OH^- increase gradually and compete with anionic MO for the positive charge site of $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ surface, which made the q_e of MO decrease rapidly.

CR is a typical dipole molecule ($pK_a = 5.5$) with negatively charged when pH > 5.5. Because of the dissociation of SO_3^- groups is suppressed when pH < 5.5, the maximum adsorption capacity of CR adsorbed on the $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ surface was 134.6 mg g⁻¹ at pH 6 (Fig. 3). The interaction between $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ and CR is the electrostatic attraction, and the change of q_e is similar to that of MO. The electrostatic attraction between CR and $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ is weaker than that between MO and $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$, this could be attributed to two reasons. Firstly, CR dye molecule with two SO_3^- groups was capable of occupying more than one active site. Secondly, there are stronger electrostatic attraction due to the large amount of protonated amine group on the surface of the adsorbent at pH 4 than 6.

Adsorption kinetics

The effect of contact time on the adsorption of MO and CR by $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ is shown in Fig. 4. The adsorption kinetics processes of MO and CR were similar. It is obvious that the adsorption capacities of $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ for MO and CR increased rapidly in the initial 40 min, subsequently the adsorption rate declined, and reached equilibrium in approximately 180 min. In addition, the adsorption of MO is found to be much higher than that of CR. The equilibrium adsorption capacities for MO and CR obtained after 180 min adsorption were 238 mg g⁻¹ and 88 mg g⁻¹, respectively. Nevertheless, the adsorption capacity for MB was basically constant, which is relative low even negligible. To further investigate the adsorption process of MO and CR by $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$, the pseudo-first-order model (eqn (3)) and pseudo-second-order model (eqn (4)) were used to fit the experimental data:

Table 4 Thermodynamic parameters for the adsorption of MO and CR on the $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$

Dye	T (K)	ΔG° (kJ mol ⁻¹)	ΔS° (J K ⁻¹ mol ⁻¹)	ΔH° (kJ mol ⁻¹)
MO	288	-21.69	223.18	42.59
	298	-23.96		
	308	-26.08		
CR	288	-14.92	229.71	51.24
	298	-17.45		
	308	-19.34		



$$\ln(q_e - q_t) = \ln q_e - k_1 t \quad (3)$$

$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{t}{q_e} \quad (4)$$

where q_e and q_t (mg g^{-1}) is the amount of dye adsorbed at equilibrium time and at time t , and k_1 (min^{-1}) is the rate constant of pseudo-first-order model, while k_2 ($\text{g mg}^{-1} \text{min}^{-1}$) is the rate constant of pseudo-second-order model.

The values of k_1 , k_2 , and q_e can be calculated from eqn (3) and (4), and the kinetic parameters obtained are presented with the linear regression values in Table 2. As shown in Table 2, the adsorption of MO and CR on $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ was described well with the pseudo-second-order kinetics model due to the greater value of determination coefficient R^2 , suggesting that the diffusion is the rate determining step mechanism, and the overall rate of this adsorption process might be due to chemisorption process. The explanation is verified in the further sections.

Adsorption isotherms

The equilibrium relationships between anionic dyes and $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ were described by adsorption isotherm in Fig. 5, and the well-known Langmuir (eqn (5)) and Freundlich (eqn (6)) isotherm models were used to fit the equilibrium adsorption data in order to determine the best-fit model.

$$q_e = \frac{q_{\max} K_L C_e}{1 + K_L C_e} \quad (5)$$

$$q_e = K_F C_e^{1/n} \quad (6)$$

where q_e (mg g^{-1}) is the dye adsorption quantity at equilibrium, q_{\max} (mg g^{-1}) is the dye adsorption capacity representing a full maximum monolayer coverage of the adsorbent surface, C_e (mg

L^{-1}) is the concentration of dye at equilibrium, K_L (L mg^{-1}) represents the Langmuir constant related to the energy of adsorption, K_F is the Freundlich constant and is related to the adsorption capacity; $1/n$ is the factor related to the adsorption intensity.

The adsorption isotherms for MO and CR obtained at 288, 298, and 308 K were shown in Fig. 5a and b, and the isotherms parameters of Langmuir and Freundlich for MO and CR adsorption on the $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ were shown in Table 3. Compared with the Langmuir isotherm model, the Freundlich isotherm model shows a better fit with the adsorption isotherms and the higher correlation coefficients were observed for the Freundlich isotherm model than that of the Langmuir isotherm model, indicating that both of the uptake of MO and CR on the sorbent are preferably followed the multilayer adsorption process. The $1/n$ value of the Freundlich isotherm model was less than 1, indicating that MO and CR adsorption on $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ was a favourable adsorption process from aqueous solution.

Adsorption thermodynamics

The van't Hoff equation was used to evaluate the thermodynamic characteristics of the dye adsorption process on the $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ adsorbent surface and the data were analyzed by the eqn (7) and (9) to get the thermodynamic parameters of ΔG° , ΔH° and ΔS°

$$\Delta G^\circ = -RT \ln(K_d) \quad (7)$$

$$K_d = \frac{q_e}{C_e} \quad (8)$$

$$\ln K_d = \left(-\frac{\Delta H^\circ}{R} \right) \frac{1}{T} + \frac{\Delta S^\circ}{R} \quad (9)$$

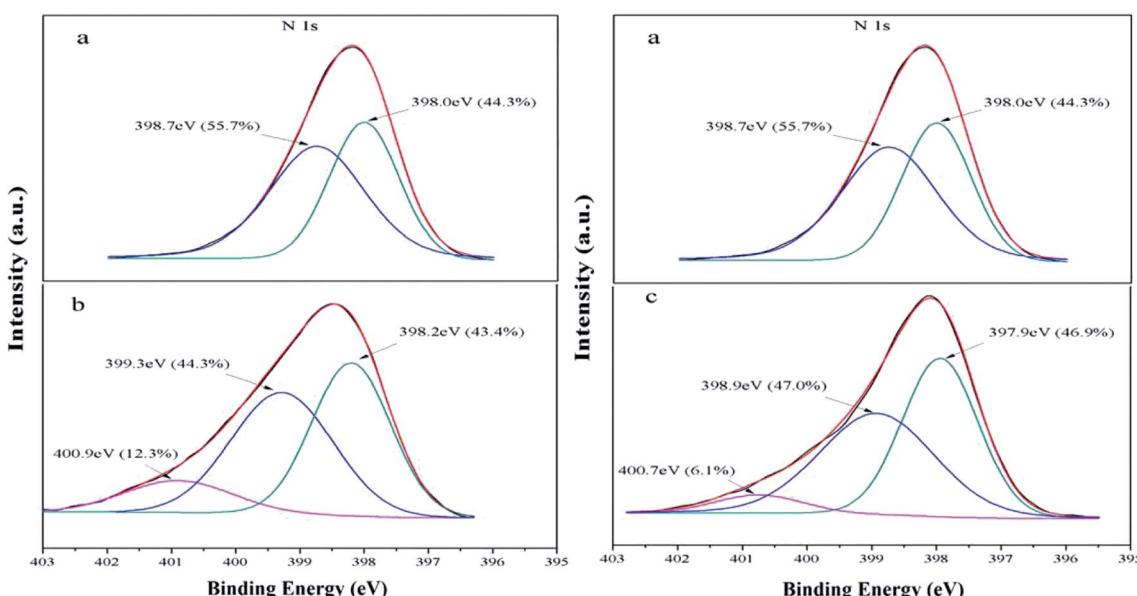


Fig. 6 The XPS spectra of N 1s (a) $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$, (b) after adsorption of MO, (c) after adsorption of CR.



where the ΔG° is the change of Gibbs free energy (J mol^{-1}), the ΔH° is the change of enthalpy (J mol^{-1}), and the ΔS° is the change of entropy ($\text{J mol}^{-1} \text{ K}^{-1}$), R ($8.314 \text{ J K}^{-1} \text{ mol}^{-1}$) and K_d are the universal gas constant and the equilibrium constant, respectively. The value of K_d at different temperatures was calculated according to eqn (8). The values of ΔS° and ΔH° were calculated from the intercept and slope of the plot of $\ln K_d$ versus $1/T$. All of the thermodynamic parameters were given in Table 4.³³ The negative values of ΔG° confirm the spontaneity of the adsorption process reaction, and the decrease of the ΔG° values with the temperature increased from 288 to 308 K, indicating that the adsorption process was more beneficial at higher temperatures, and it can be seen clearly from Fig. 5. In addition, positive value of ΔH° reveals the endothermic nature of the adsorption reaction, and the positive value of ΔS° indicates that during the adsorption of MO and CR dye on the surface of the adsorbent, there is an increase in randomness at the solid/solution interface.³⁴

Removal mechanism

Although the CR and MO are anionic dye, actually, the as-synthesized samples display better adsorption capacity for MO than CR. In order to better investigate the adsorption mechanism of MO and CR on $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$, XPS analysis of $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ before and after the adsorption of MO and CR have been carried out. The adsorption process was usually determined by the functional groups on the surface of adsorbent. The high-resolution N 1s XPS spectra of the $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ before and after adsorption of MO and CR are shown in Fig. 6. For $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$, N 1s peak was deconvoluted into two peaks, 44.3% of $-\text{N}=\text{}$ at 398.0 eV and 55.7% of $-\text{NH}_2$ at 398.7 eV, after MO adsorption, the new peak appeared at 400.9 eV corresponds to the protonated amine group (NH_3^+), and the molar ratio of NH_3^+ is 12.3%,

while the $-\text{NH}_2$ decreased from 55.7% to 44.3% and the $-\text{N}=\text{}$ decreased from 44.3% to 43.4%. As for the adsorption of CR, the new peak appeared at 400.7 eV corresponds to the NH_3^+ , and the molar ratio of NH_3^+ is 6.1%, while the $-\text{NH}_2$ decreased from 55.7% to 47.0% and the $-\text{N}=\text{}$ increased from 44.3% to 46.9%.

As we all known that with three types of amino groups, the PEI is modeled at three protonation levels: at the high pH (>10), PEI is fully deprotonated; at neutral pH (~ 7), all the primary amines are protonated; at low pH (<4), most of the amines are protonated. Meanwhile, when $\text{pH} < \text{p}K_a$, the dissociation of SO_3^- groups is suppressed, the MO and CR were positively charged or neutral. When $\text{pH} > \text{p}K_a$, the presence of NH_3^+ could absorb negatively charged SO_3^- groups in anionic dye through electrostatic attraction at pH 4–7. As we can seen from Fig. 6, the molar ratio of NH_3^+ is greater in MO than in CR, this is because of the large amount of protonated amine groups on the surface of the adsorbent at pH 4 than 6. It is also explained why the adsorption efficiency of MO is better than CR in the study of pH. Another important factor that affects the difference in adsorption between MO and CR is that CR dye molecule with two SO_3^- groups was capable of occupying more than one active site (Scheme 1). With the increase of pH, the extent of protonation of PEI decreased, along with the decrease in the adsorption ability of $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$. When $\text{pH} > 7$, the OH^- increase gradually and compete with anionic MO for the NH_3^+ of $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ surface, leading to a faster decrease in the adsorption ability of $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$. Based on the above analysis, the conceivable mechanism of anionic dye adsorption is the presence of electrostatic interaction between anionic dyes and the positively charged nitrogens on the surface of the $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$, and it's consistent with the pH study.

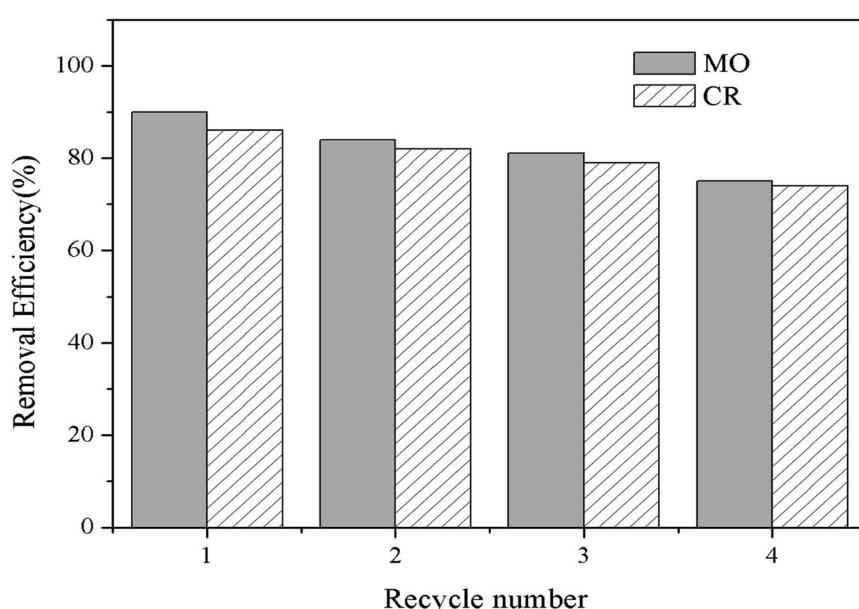


Fig. 7 Reusability of the $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ for MO and CR removal.



Reusability

From the viewpoint of the potential practical application, reusability test is one of the most important factors to be considered for a given adsorbent by the adsorption–desorption experiments. It was repeated four times through batch experiments, and the results are shown in Fig. 7. Compared to the virgin composite $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$, the regenerated $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ exhibited an excellent regeneration for the adsorption of MO and CR, and the removal efficiency of $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ adsorbent decreased to approximately 75.4% and 74.5% even after the fourth reusability cycle, indicating that the $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ could be a potential and efficient adsorbent for the MO and CR removal due to the excellent multiple reusability performance.

Conclusions

A eco-friendly and low-cost nitrogen-rich magnetic adsorbent ($\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$) was successfully synthesized by a cross-linking reaction. The results indicated that the equilibrium adsorption capacities for MO and CR obtained after 180 min adsorption were 238 mg g^{-1} and 88 mg g^{-1} , respectively, which was forceful and fast, while the adsorption capacity for MB was relative low even negligible, demonstrating the strong adsorptive affinity of $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ toward anionic compounds. The adsorption reaction was a spontaneous and endothermic process. Notably, the excellent adsorption characteristics for anionic dyes of the as-synthesized $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ mainly attribute to the electrostatic attraction between the negatively charged groups (SO_3^- groups) in anionic dyes and NH_3^+ of adsorbent surface. Furthermore, the mechanism of the adsorption process towards MO and CR share both commons and differences because CR with two SO_3^- groups was capable of occupying more than one active site. After the fourth adsorption–desorption cycle, the adsorbent still showed a high adsorption capacities for MO and CR in aqueous solution, indicating the potential of the $\text{Fe}_3\text{O}_4@\text{SiO}_2/\text{PEI}$ as a practical adsorbent for anionic dyes removal.

Conflicts of interest

There are no conflicts to declare.

Acknowledgements

This research was financially supported by the Natural Science Foundation of Hunan Province, China (Grant No. 2019JJ50015), the Project of Science Foundation of Hengyang Normal University, China (Grant No. 18D18) and the Aid Program of Research Center for Environment Education of Hunan Province, China (Grant No. HJ18K04).

References

- 1 A. Mirzaei, Z. Chen, F. Haghigat and L. Yerushalmi, *Chem. Eng. J.*, 2017, **330**, 407–418.
- 2 M. Falah, K. J. D. MacKenzie, R. Knibbe, S. J. Page and J. V. Hanna, *J. Hazard. Mater.*, 2016, **318**, 772–782.
- 3 X. Liu, W. Gong, J. Luo, C. Zou, Y. Yang and S. Yang, *Appl. Surf. Sci.*, 2016, **362**, 517–524.
- 4 L. Wang, C. Mao, N. Sui, M. Liu and W. W. Yu, *Environ. Technol.*, 2017, **38**, 996–1004.
- 5 R. M. M. Santos, J. Tronto, V. Briois and C. V. Santilli, *J. Mater. Chem. A*, 2017, **5**, 9998–10009.
- 6 U. Habiba, T. A. Siddique, T. C. Joo, A. Salleh, B. C. Ang and A. M. Afifi, *Carbohydr. Polym.*, 2017, **157**, 1568–1576.
- 7 Y. Jiang, B. Liu, J. Xu, K. Pan, H. Hou, J. Hu and J. Yang, *Carbohydr. Polym.*, 2018, **182**, 106–114.
- 8 C. Lei, M. Pi, C. Jiang, B. Cheng and J. Yu, *J. Colloid Interface Sci.*, 2017, **490**, 242–251.
- 9 S. Dawood and T. K. Sen, *Water Res.*, 2012, **46**, 1933–1946.
- 10 N. Nasuha, B. H. Hameed and A. T. Din, *J. Hazard. Mater.*, 2010, **175**, 126–132.
- 11 A. S. Patra, S. Ghorai, S. Ghosh, B. Mandal and S. Pal, *J. Hazard. Mater.*, 2016, **301**, 127–136.
- 12 A. K. Abay, D. H. Kuo, X. Chen and A. D. Saragih, *Chemosphere*, 2017, **189**, 21–31.
- 13 Z. Zhu, P. Wu, G. Liu, X. He, B. Qi, G. Zeng, W. Wang, Y. Sun and F. Cui, *Chem. Eng. J.*, 2017, **313**, 957–966.
- 14 K. Bello, B. K. Sarojini, B. Narayana, A. Rao and K. Byrappa, *Carbohydr. Polym.*, 2018, **181**, 605–615.
- 15 G. Bayramoglu and M. Yakuparica, *Chem. Eng. J.*, 2008, **139**, 20–28.
- 16 N. Sui, L. Wang, X. Wu, X. Li, J. Sui, H. Xiao, M. Liu, J. Wan and W. W. Yu, *RSC Adv.*, 2015, **5**, 746–752.
- 17 Y. Z. Zhang, J. Li, J. Zhao, W. Bian, Y. Li and X. J. Wang, *Bioresour. Technol.*, 2016, **222**, 285–293.
- 18 B. Huang, Y. Liu, B. Li, G. Zeng, X. Hu, B. Zheng, T. Li, L. Jiang, X. Tan and L. Zhou, *RSC Adv.*, 2015, **5**, 106339–106349.
- 19 R.-r. Shan, L.-g. Yan, K. Yang, Y.-f. Hao and B. Du, *J. Hazard. Mater.*, 2015, **299**, 42–49.
- 20 J. Ma, Y. Wu, Y. Zeng, Y. Li and D. Wu, *J. Mater. Chem. A*, 2015, **3**, 16762–16773.
- 21 C. Shuai, Y. Yaxian, Y. Jianlin, H. Sanyang and G. Renao, *Chem. Commun.*, 2011, **47**, 4225–4227.
- 22 R. Fan, H. Min, X. Hong, Q. Yi, W. Liu, Q. Zhang and Z. Luo, *J. Hazard. Mater.*, 2019, **364**, 780–790.
- 23 B. Huang, Y. Liu, B. Li, S. Liu, G. Zeng, Z. Zeng, X. Wang, Q. Ning, B. Zheng and C. Yang, *Carbohydr. Polym.*, 2017, **157**, 576–585.
- 24 B. Chen, X. Zhao, Y. Liu, B. Xu and X. Pan, *RSC Adv.*, 2015, **5**, 1398–1405.
- 25 X. J. Kong, C. Zheng, Y. H. Lan, S. S. Chi, Q. Dong, H. L. Liu, C. Peng, L. Y. Dong, L. Xu and X. H. Wang, *Anal. Bioanal. Chem.*, 2018, **410**, 247–257.
- 26 L. G. Bach, M. R. Islam, J. T. Kim, S. Seo and K. T. Lim, *Appl. Surf. Sci.*, 2012, **258**, 2959–2966.
- 27 S. Singh, K. C. Barick and D. Bahadur, *J. Hazard. Mater.*, 2011, **192**, 1539–1547.
- 28 S. Zhang, Y. Zhang, J. Liu, Q. Xu, H. Xiao, X. Wang, H. Xu and J. Zhou, *Chem. Eng. J.*, 2013, **226**, 30–38.



29 S. Dash, H. Chaudhuri, G. Udayabhanu and A. Sarkar, *Energy Fuels*, 2016, **30**, 6646–6653.

30 C. Jin, J. Han, F. Chu, X. Wang and R. Guo, *Langmuir*, 2017, **33**, 4520–4527.

31 Y. Zhou, T. Ping, I. Maitlo, B. Wang, M. Y. Akram, J. Nie and X. Zhu, *Nanotechnology*, 2016, **27**, 215301.

32 F. Qu, N. B. Li and H. Q. Luo, *Langmuir*, 2013, **29**, 1199–1205.

33 L. Sharma and R. Kakkar, *ACS Appl. Mater. Interfaces*, 2017, **9**, 38629–38642.

34 S. Ghorai, A. K. Sarkar, A. B. Panda and S. Pal, *Bioresour. Technol.*, 2013, **144**, 485–491.

