


 Cite this: *RSC Adv.*, 2019, **9**, 24386

## Evaluation of Fe(III)EDTA reduction with ascorbic acid in a wet denitrification system<sup>†</sup>

Xinyu Zhu, Feiqiang He, Meng Xia, Honggen Liu and Jianhua Ding\*

The reduction of Fe(III)EDTA to Fe(II)EDTA is the core process in a wet flue gas system with simultaneous desulfurization and denitrification. Herein, at first, the reductant ascorbic acid (VC) was used for reducing Fe(III)EDTA. The feasibility of Fe(II)EDTA reduction with ascorbic acid was investigated at different Fe(III)EDTA concentrations, various pH values, diverse temperatures, and different molar ratios of VC to Fe(III)EDTA. The results showed that the Fe(II)EDTA concentration increased with an increase in the initial Fe(III)EDTA concentration. Furthermore, the reduction efficiency increased as the mole ratio of VC to Fe(III)EDTA was increased, and all the Fe(III)EDTA reduction efficiencies were close to 100% when the mole ratio was more than 0.5. On the other hand, an alkaline environment did not favor the conversion of Fe(III)EDTA by VC. The Fe(III)EDTA conversion slightly increased as the temperature was increased. Moreover, compared with other reduction systems, ascorbic acid (VC) was found to be more powerful in reducing Fe(III)EDTA, especially in air. In addition, VC only exhibited powerful ability in the conversion of Fe(III)EDTA to Fe(II)EDTA and hardly reduced Fe(II)EDTA-NO. Finally, the stoichiometry of Fe(III)EDTA reduction by ascorbic acid was derived. Thus, our study would offer a bridge between foundational research and industrial denitrification using the combination of Fe(II)EDTA and VC.

 Received 10th April 2019  
 Accepted 4th July 2019

DOI: 10.1039/c9ra02678a

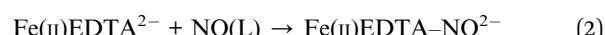
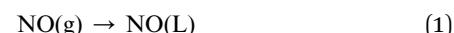
[rsc.li/rsc-advances](http://rsc.li/rsc-advances)

## 1. Introduction

The extensive utilization of fossil fuels has caused a series of serious environmental problems including acidification, photochemical smog, eutrophication, and ozone depletion.<sup>1</sup> It is generally considered that nitrogen oxide (NO<sub>x</sub>) and sulfur dioxide (SO<sub>2</sub>) are the main pollutants that originate abundantly from fuel gases.<sup>2</sup> Therefore, the emissions of NO<sub>x</sub> and SO<sub>2</sub> are strictly controlled by governments all over the world. To date, SO<sub>2</sub> can be easily removed by wet flue gas desulfurization (WFGD) because of its high solubility in water.<sup>3</sup> However, it is difficult to remove NO (accounting for 95% in total NO<sub>x</sub>) *via* wet flue gas desulfurization (WFGD) due to low solubility of NO in water. For this reason, several approaches, including the addition of an oxidizing agent, such as KMnO<sub>4</sub>,<sup>4</sup> NaClO<sub>2</sub>,<sup>5</sup> and H<sub>2</sub>O<sub>2</sub>,<sup>6</sup> to oxidize NO to the more soluble NO<sub>2</sub> and the use of a chelating agent to unite NO to form nitrosyl complexes,<sup>7</sup> have been refined to improve NO absorption.

Fe(II)EDTA as an absorbent of nitric oxide (NO) from the flue gas has been widely investigated since the 1970s by many research groups because of the high complexing ability of NO; Fe(II)EDTA can rapidly capture NO to produce the ferrous-nitrosyl complexes Fe(II)EDTA-NO, whose equilibrium constant is about 10<sup>7</sup> M<sup>-1</sup>. However, a notable shortcoming of Fe(II)EDTA is

that it can be readily oxidized with dissolved oxygen to produce inanimate Fe(III)EDTA.<sup>8</sup> The reactions are shown as follows:



Therefore, to maintain outstanding NO removal, it is necessary to transform Fe(III)EDTA to Fe(II)EDTA *via* a reduction reaction.

Sulfur-bearing derivatives (sulfite, bisulfite, dithionite, Na<sub>2</sub>S<sub>2</sub>O<sub>8</sub>, *etc.*) have been extensively associated with Fe(II)EDTA to improve NO absorption.<sup>9-12</sup> Specifically, SO<sub>3</sub><sup>2-</sup>, HSO<sub>3</sub><sup>-</sup>, and S<sub>2</sub>O<sub>4</sub><sup>2-</sup> have been used as reducing agents in the transformation of Fe(III)EDTA to Fe(II)EDTA, among which the reducibility of sulfite and bisulfite is relatively weak, and the dithionite ion is easily deteriorated in air, losing the strong reducing power. Therefore, they can hardly maintain high removal efficiency of NO for a long time. An enhanced absorption of NO using Na<sub>2</sub>S<sub>2</sub>O<sub>8</sub> has been accomplished *via* a synergistic reaction. Under the catalytic action of Fe(II)EDTA, Na<sub>2</sub>S<sub>2</sub>O<sub>8</sub> was decomposed to the sulfate radical and hydroxyl radical for enhancing the oxidation and absorption of NO; moreover, the derived ions HSO<sub>4</sub><sup>-</sup> were produced, which could reduce Fe(III)EDTA to Fe(II)EDTA. However, the synergistic reaction between Na<sub>2</sub>S<sub>2</sub>O<sub>8</sub> and Fe(II)EDTA was limited. Moreover, different bacteria, such as *Escherichia coli* FR-2 and *Klebsiella* sp. strain FD-3, have been employed to regenerate Fe(II)EDTA.<sup>13,14</sup> However, low Fe(III)EDTA reduction

State Key Laboratory for Nuclear Resources and Environment, School of Chemistry, Biology and Materials Science, East China University of Technology, Nanchang 330013, P. R. China. E-mail: [he\\_feiqiang@ecit.cn](mailto:he_feiqiang@ecit.cn); [dingjianh2004@126.com](mailto:dingjianh2004@126.com)

† Electronic supplementary information (ESI) available. See DOI: [10.1039/c9ra02678a](https://doi.org/10.1039/c9ra02678a)



efficiency and the isolation and culture of bacteria, whose requirements are complicated, would result in high operating costs. In addition, activated carbon,<sup>15</sup> selenium,<sup>16</sup> hydrazine ( $N_2H_4$ ),<sup>17</sup> metal powders (Zn, Fe, and Al),<sup>18</sup> polyphenolic compounds,<sup>19</sup> and other reducing agents have been found to be efficient for the reduction of Fe(III)EDTA at ambient temperature. However, due to the complicated process, environmental pollution, and high cost, the commercialization of the above-mentioned methods does not appear to be very successful.

Ascorbic acid (vitamin C) is an important biological reductant that is ubiquitous in humans and animals and non-toxic.<sup>20</sup> In addition, it shows significant reducibility in the reduction of methylene blue,<sup>21</sup> hexavalent chromium,<sup>22</sup> and preparation of nanomaterials. Inspired by this, herein, we proposed a strategy of applying ascorbic acid as a reducing agent to enhance the regeneration of Fe(II)EDTA.

## 2. Materials and methods

### 2.1. Chemicals

Ascorbic acid (VC) was supplied by Tianjin Kermel Chemical Reagent Co., LVC., China, and used without additional purification. Na<sub>2</sub>EDTA (99.0%), FeSO<sub>4</sub>·7H<sub>2</sub>O (99.0%), NH<sub>4</sub>Fe(SO<sub>4</sub>)<sub>2</sub>·12H<sub>2</sub>O (99.0%), Zn (99.0%), and Na<sub>2</sub>S<sub>2</sub>O<sub>4</sub> (99.0%) were purchased from Tianjin Kermel Chemical Reagent Co., LVC., China. N<sub>2</sub> (99.999%) and NO (99.99%) were obtained from Guangzhou Yuejia Gas Co., China. All chemicals were analytical grade reagents.

### 2.2. Fe(III)EDTA reduction by VC under anoxic conditions with different parameters

A 250 mL flask with a pH electrode was set on a heat collection-constant temperature-type magnetic stirrer (Fig. 1). Then, 250 mL aqueous Fe(III)EDTA solution was obtained by blending Na<sub>2</sub>EDTA and NH<sub>4</sub>Fe(SO<sub>4</sub>)<sub>2</sub>·12H<sub>2</sub>O in the flask. Then, the solution was separated from atmospheric oxygen using a 3 L min<sup>-1</sup> continuous N<sub>2</sub> (99.999%) stream. Then, the pH value was adjusted using the H<sub>2</sub>SO<sub>4</sub> and NaOH solution. The Fe(III)

EDTA reduction was initiated after the addition of VC to the Fe(III)EDTA solution at the stirring speed of 2500 rpm. Finally, the Fe(II)EDTA concentration of the solution was measured at specific time intervals. The effects of the initial Fe(III)EDTA concentration, the mole ratio of VC to Fe(III)EDTA, pH value, and temperature on the Fe(III)EDTA reduction with VC were systematically investigated (the experimental conditions can be seen in Table 1).

### 2.3. Fe(III)EDTA reduction by VC in air

To investigate the Fe(III)EDTA reduction by VC in air, the reactions were conducted in air at  $[Fe(III)EDTA]_0 = 0.002 \text{ mol L}^{-1}$ ,  $T = 30^\circ\text{C}$ , and  $\text{pH} = 6.0$  using VC at the initial concentrations of 0.002, 0.004, 0.008, and 0.016 M. The reduction operation was similar to that performed under anaerobic conditions except that the flask was not stoppered and the solution was directly exposed to air. Finally, the Fe(II)EDTA concentration was measured at specific time intervals. In addition, to highlight the advantages of VC in the Fe(III)EDTA reduction, Zn and Na<sub>2</sub>S<sub>2</sub>O<sub>4</sub> were used to reduce Fe(III)EDTA in air at  $[Fe(III)EDTA]_0 = 0.002 \text{ mol L}^{-1}$ ,  $T = 30^\circ\text{C}$ , and  $\text{pH} = 6.0$ .

### 2.4. Relevance for Fe(II)EDTA-NO reduction

According to a method reported in literature,<sup>23</sup> in this study, 250 mL 0.01 mol L<sup>-1</sup> Fe(III)EDTA was prepared in a 250 mL flask and maintained under N<sub>2</sub> positive pressure. Then, 0.005 M VC was added to the flask to reduce Fe(III)EDTA to Fe(II)EDTA at  $30^\circ\text{C}$  and  $\text{pH} = 6.0$ . After the reduction reaction was completed, the Fe(II)EDTA-NO complex solution was prepared by bubbling a known mixture of NO and N<sub>2</sub> through the produced Fe(II)EDTA solution until the NO concentration (determined using a NO analyzer, KM950) in the outlet was equal to that in the inlet. Then, 0.005 M VC was again added to reduce Fe(II)EDTA-NO. Moreover, the changes in the concentrations of Fe(III)EDTA and Fe(II)EDTA-NO were measured at specific time intervals, and color change during the reduction process was analyzed.

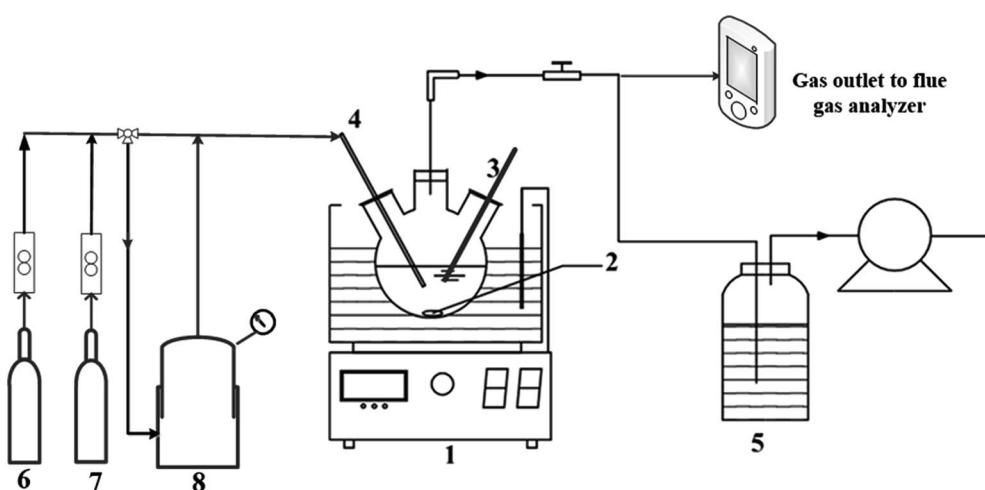


Fig. 1 Experimental apparatus for Fe(III)EDTA reduction by ascorbic acid ((1) heat collection-constant temperature type magnetic stirrer; (2) magnetic stirrer; (3) the pH electrode; (4) inlet for nitrogen or a mixture of NO and N<sub>2</sub>; (5) buffer tank; (6) N<sub>2</sub> cylinder; (7) NO cylinder; and (8) cylinder containing a mixture of NO and N<sub>2</sub>).



Table 1 Initial experimental conditions employed for Fe(III)EDTA reduction by VC

Serial no.	Parameter	[Fe(III)EDTA] <sub>0</sub>	[VC]/[Fe(III)] <sub>0</sub>	pH	Temperature
1	[Fe(III)EDTA] <sub>0</sub>	0.002 M	0.5	6.0	60 °C
		0.004 M			
		0.008 M			
		0.01 M			
2	[VC]/[Fe(III)] <sub>0</sub>	0.002 M	0.25	6.0	60 °C
			0.5		
			0.75		
			1		
3	pH	0.002 M	0.5	4.3 6 7 9.3	60 °C
4	Temperature	0.002 M	0.5	6.0	30 °C 40 °C 50 °C 60 °C

## 2.5. Analytical methods

The Fe(II)EDTA concentration was determined by *o*-phenanthroline colorimetric method at the absorbance of 510 nm<sup>9</sup>. On the other hand, the concentration of Fe(II)EDTA-NO was detected using a UV-vis spectrophotometer at the absorbance of 434 nm.<sup>24</sup> The Fe(III) concentration could be calculated from the difference between the total concentration of Fe and the concentration of Fe(II).<sup>25</sup> The dehydroascorbic acid concentration was measured by the 2,4-dinitrophenylhydrazine method.<sup>26</sup> The experimental data are the mean values of triplicate experiments along with the standard deviation.

The concentration of NO was analyzed by a flue gas analyzer (KM950, KANE International Ltd) after the gas stream passed through the drying tube filled with anhydrous calcium chloride.

The NO removal efficiency was calculated using the following equation:

$$\eta = \frac{c_{\text{in}} - c_{\text{out}}}{c_{\text{in}}} \times 100\% \quad (4)$$

where  $\eta$  is the NO removal efficiency with the unit of %;  $c_{\text{in}}$  is the inlet concentration of NO in ppm; and  $c_{\text{out}}$  is the outlet concentration of NO in ppm.

## 3. Results and discussion

### 3.1. Effect of the initial concentration of Fe(III)EDTA on the Fe(III)EDTA reduction

Experiments were carried out to investigate the effect of Fe(III)EDTA concentration on the Fe(III)EDTA reduction at pH 6.0, temperature = 60 °C, and [VC]/[Fe(III)] = 0.5. The experimental results are shown in Fig. 2. It was observed that the Fe(II)EDTA concentration increased with an increase in the initial Fe(III)EDTA concentration. For example, the produced Fe(II) concentrations were 0.00197, 0.00405, 0.00763, and 0.00948 M, corresponding to the initial Fe(III)EDTA concentrations of 0.002, 0.004, 0.008, and 0.01 M. The explanation is that the mass transfer impetus between the surface of VC and the liquid bulk gets enhanced as the Fe(III)EDTA concentration increases. The increase in the Fe(III)EDTA concentration on the surface of VC is

favorable for the Fe(II)EDTA reduction. In addition, the reduction efficiencies in all the cases exceeded 95%, showing that VC had powerful Fe(III)EDTA reduction ability.

### 3.2. Effect of the mole ratio of VC to Fe(III)EDTA on the Fe(III)EDTA reduction

The percentage of Fe(III) converted to Fe(II) significantly depends on the initial mole ratio of VC to Fe(III)EDTA. The effect of this mole ratio on the Fe(III)EDTA conversion was investigated in a 250 mL solution by varying the ratio of VC to Fe(III)EDTA from 0.25 to 1 at pH 6, [Fe(III)EDTA] = 0.002 M, and temperature = 60 °C. Fig. 3 indicates that the reduction rate increases as the molar ratio increases. For example, the produced Fe(II)EDTA concentrations were 0.00106, 0.00197, 0.00198, and 0.00198 M, corresponding to the mole ratio of 0.25, 0.5, 0.75, and 1. The reason is that more active sites can be obtained at higher initial molar ratios that will lead to higher Fe(II)EDTA regeneration rates. Note that the reduction efficiency at the mole ratio of 0.25 was about 50%; however, all the Fe(III)EDTA reduction efficiencies were close to 100% at the mole ratio of 0.5, 0.75, and 1.

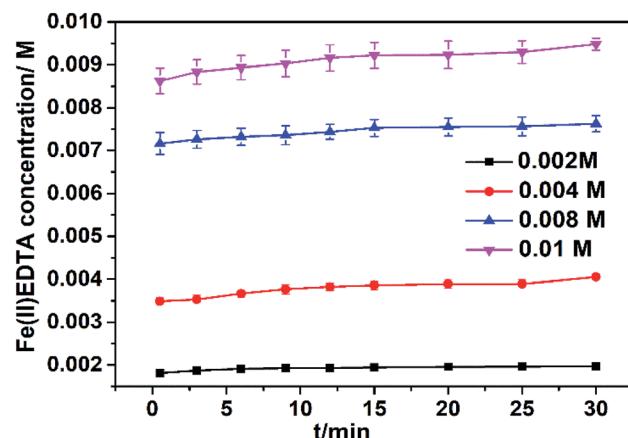


Fig. 2 Fe(II)EDTA concentration change with time at different Fe(III)EDTA concentrations. Conditions: temperature = 60 °C, original mole ratio of VC to Fe(III)EDTA = 0.5; and original pH value = 6.0.



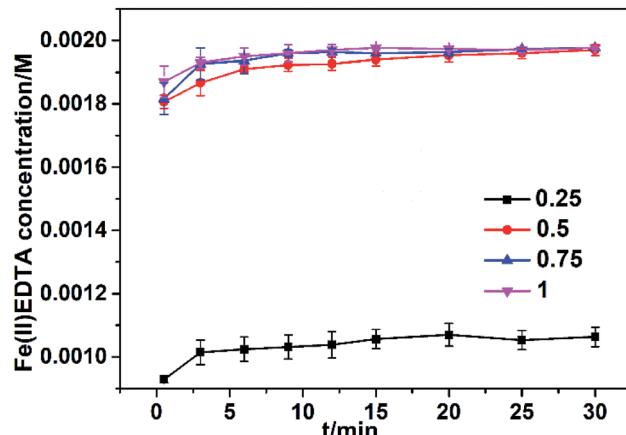


Fig. 3 Fe(II)EDTA concentration change at different original molar ratios of VC to Fe(III)EDTA. Conditions: temperature = 60 °C, Fe(III)EDTA concentration = 0.002 M; and original pH value = 6.0.

### 3.3. Effect of pH on the Fe(III)EDTA reduction

The effect of pH on the Fe(III)EDTA conversion by VC was studied in the pH range of 4.3–9.3 while keeping the other parameters unchanged. The Fe(III)EDTA conversion within 30 min is presented in Fig. 4. In Fig. 4, it was observed that initial pH had a significant impact on the Fe(III)EDTA reduction, and the reduction rate increased as the pH value decreased. Furthermore, lower Fe(III)EDTA conversion was found in an alkaline solution (pH 9.3), whereas the Fe(III)EDTA conversion increased rapidly to 0.00191 M when the initial pH of the solution was 4.3. Thus, an alkaline environment is unfavorable for the Fe(III)EDTA conversion by VC; the reason is that VC itself is acidic and becomes unstable in an alkaline solution.

### 3.4. Effect of temperature on the Fe(III)EDTA reduction

The effect of temperature on the Fe(III)EDTA conversion was explored at pH 6.0. The experimental results depicted in Fig. 5 demonstrate that the Fe(III)EDTA conversion slightly increases

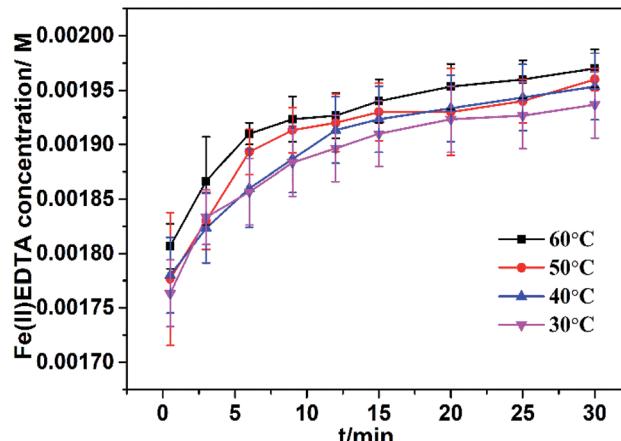


Fig. 5 Fe(II)EDTA concentration variation at different temperatures. The reaction parameters: pH = 6; original VC concentration = 0.001 M; and  $[Fe(III)EDTA]_0 = 0.002$  M.

as the temperature increases. However, 30 min after the start of the reaction, all the Fe(III)EDTA conversion efficiencies obtained were up to about 98%. On the one hand, high temperatures can accelerate the molecule movement and increase the chemical rate. On the other hand, high temperatures can also promote the decomposition of VC. Therefore, the slight increase in the Fe(III)EDTA conversion with the increasing temperature was the result of the interaction between the abovementioned two effects.

### 3.5. Fe(III)EDTA reduction by VC in air

Industrial flue gases in China generally contain 3–9% (v/v) oxygen, which imposes a significant effect on the Fe(III)EDTA reduction. Therefore, it is necessary to investigate the Fe(III)EDTA reduction by VC in air. In Fig. 6, the results show that VC also exhibits excellent Fe(III)EDTA reduction ability in air, and the Fe(III)EDTA reduction is improved at higher VC concentrations; for example, more than 95% Fe(III)EDTA reduction efficiency can be maintained for 25 min at 0.002 M VC; moreover, more than 90% Fe(III)EDTA reduction efficiency can be maintained for 240 min at 0.016 M VC.

### 3.6. Comparison between the Fe(III)EDTA reduction efficiency obtained herein and those obtained in some previously reported studies

Via our related experiments, it has been found that ascorbic acid (VC) possesses high Fe(III)EDTA reduction efficiency. However, it is still uncertain whether VC has an advantage over other systems in the Fe(III)EDTA reduction. For this reason, the results of Fe(III)EDTA reduction by VC obtained in our study were compared with those of some other systems reported in the literature; however, comparisons between efficiencies of Fe(III)EDTA reduction achieved by various systems were difficult to make due to dissimilarities in the reaction conditions used by different research teams. The comparisons of the Fe(III)EDTA reduction efficiency obtained herein with those obtained using different systems are presented in Table 2; the selected data in

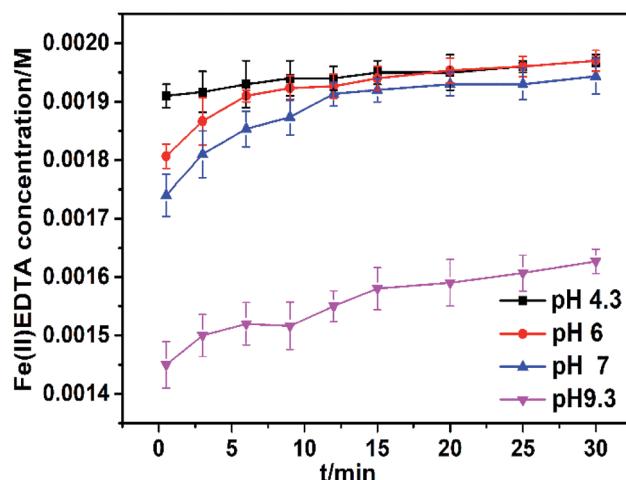


Fig. 4 Fe(II)EDTA concentration variation at different pH values in the range of 4.3–9.3. The reaction parameters: temperature = 60 °C; 0.001 M VC; and  $[Fe(III)EDTA]_0 = 0.002$  M.



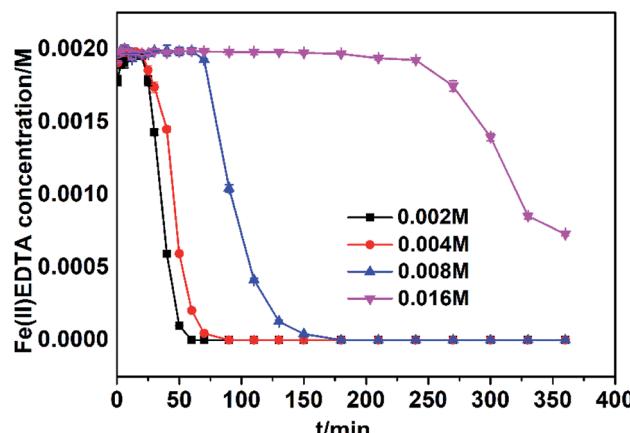


Fig. 6 Fe(II)EDTA concentration variation at different VC concentrations in air. Other reaction parameters: pH = 6; temperature = 30 °C; and  $[Fe(III)EDTA]_0 = 0.002$  M.

Table 2 show the best performance in terms of Fe(III)EDTA reduction reported in the respective references. Further, from Table 2, it can be observed that only Zn and Na<sub>2</sub>S<sub>2</sub>O<sub>4</sub> are close to VC in terms of the Fe(III)EDTA reduction efficiency in an anaerobic environment. All three showed very high reduction rates. However, the mole ratio of VC to Fe(III)EDTA was much lower than that of Zn to Fe(III)EDTA. Thus, the amount of VC required to obtain similar reduction efficiency was much less than that of Zn. In addition, Zn and Na<sub>2</sub>S<sub>2</sub>O<sub>4</sub> get easily oxidized in air; thus, to date, it is unknown whether Zn and Na<sub>2</sub>S<sub>2</sub>O<sub>4</sub> show a remarkable effect on the Fe(III)EDTA reduction in air. To further explore the advantage of VC, the Fe(III)EDTA reduction in air was conducted using 0.062 M Zn, 0.070 M Na<sub>2</sub>S<sub>2</sub>O<sub>4</sub>, and 0.016 M VC, having the same price, to achieve economic performances, and the reduction efficiencies were compared; it was observed that the times for which more than 90% reduction efficiency was maintained were 3 min, 15 min, and 240 min in an aerobic environment, corresponding to 0.062 M Zn, 0.070 M Na<sub>2</sub>S<sub>2</sub>O<sub>4</sub>, and 0.016 M VC (Table 3). In other words, with the same cost, the Fe(III)EDTA reduction efficiency of VC was best.

Therefore, VC is a better reducing agent for Fe(III)EDTA reduction in terms of economy and efficiency.

### 3.7. Relevance for Fe(II)EDTA-NO reduction

Via the abovementioned analysis and discussions, VC was found to be a powerful reducing agent for the Fe(III)EDTA reduction. In addition, color change can indicate this conclusion. It was found from Fig. 7, (A → B → C) that the solution varied from light yellow Fe(III)EDTA to colorless Fe(II)EDTA; however, when the nitric oxide and nitrogen mixture was bubbled into the colorless Fe(II)EDTA solution, black-brown Fe(II)EDTA-NO appeared, revealing that the product of Fe(III)EDTA reduction with VC could effectively capture NO. These color changes were in accordance with the results reported in relevant literature.<sup>16,33</sup> It was further confirmed that VC could transform Fe(III)EDTA into Fe(II)EDTA.

Fe(II)EDTA-NO is a by-product of flue gas denitrification, which uses Fe(II)EDTA. Thus, the conversion of Fe(II)EDTA-NO to Fe(II)EDTA was also very significant for the NO removal process. However, it is unknown whether VC can transform Fe(II)EDTA-NO to Fe(II)EDTA. Therefore, the color change of the Fe(II)EDTA-NO solution during the reduction process was investigated (Fig. 7, C → D). It was found that the black-brown solution remain unchanged after VC was added to Fe(II)EDTA-NO, and the Fe(II)EDTA-NO concentration was almost unchanged. The color variation illustrates that VC only exhibits powerful conversion efficiency for reducing Fe(III)EDTA to Fe(II)EDTA.

### 3.8. Reaction mechanism

According to a previous study,<sup>34</sup> VC as a reducing agent can be oxidized to dehydroascorbic acid. Consequently, during the Fe(III)EDTA reduction by VC, dehydroascorbic acid is speculated as the main product of VC oxidation.

For verifying our conjecture, the changes occurring in dehydroascorbic acid during the reduction process were investigated. It was found that dehydroascorbic acid was produced during the process, and its concentration increased with an increase in Fe(III)EDTA concentration (Fig. 8). Specifically, the dehydroascorbic acid concentrations were increased by

Table 2 Comparisons between the simultaneous Fe(III)EDTA reduction in different systems (the details of the experimental conditions could be obtained from respective references)

Reference	Reduction material	Reduction efficiency (initial concentration)	Mole ratio of reducing agent to Fe(III)EDTA	Required time
14	<i>Escherichia coli</i> FR-2	87% (6 mM)	—	9 h
13	<i>Klebsiella</i> sp. strain FD-3	52.4% (10 mM)	—	12 h
18	Zn	100% (50 mM)	10	Instantaneous
	Al	60% (100 mM)	10	40 min
	Sn	<20% (50 mM)	10	60 min
27	Fe	100% (10 mM)	10	~40 min
28	Na <sub>2</sub> S <sub>2</sub> O <sub>4</sub>	99.5% (50 mM)	0.7	3 min
29	Na <sub>2</sub> SO <sub>3</sub>	About 65% (10 mM)	10	120 min
30	Na <sub>2</sub> SO <sub>3</sub> /Na <sub>2</sub> SeSO <sub>3</sub> mixture	69.4% (10 mM)	100	—
31	Activated carbon/Na <sub>2</sub> SO <sub>3</sub>	About 70% (10 mM)	20	200 min
32	Thiourea dioxide	100% (2 mM)	5	30 min
This work	VC	100%	0.5	0.5 min



Table 3 Economic analysis of Zn,  $\text{Na}_2\text{S}_2\text{O}_4$  and VC

Reduction material	Zn	$\text{Na}_2\text{S}_2\text{O}_4$	VC
Approximate price <sup>a</sup>	\$2857 per ton	\$928 per ton	\$4142 per ton
Unit-price by mole	\$186 per kmol	\$163 per kmol	\$726 per kmol
Concentration corresponding to the same price	0.062 M	0.070 M	0.016 M
Time of maintaining 90% efficiency in air	3 min	15 min	240 min

<sup>a</sup> Obtained by the Alibaba internet website.

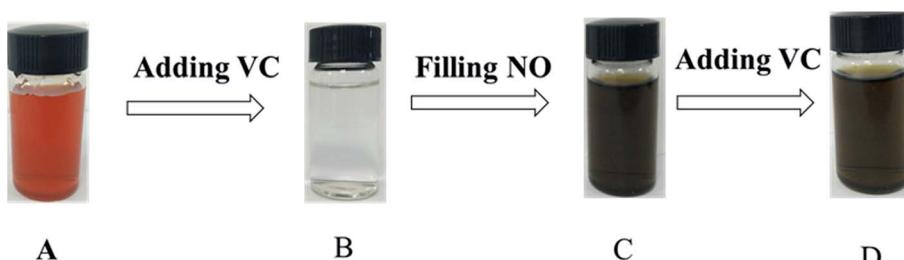
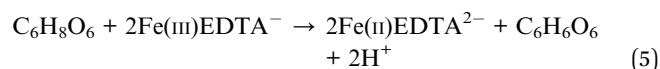


Fig. 7 Color change during the  $\text{Fe}(\text{III})\text{EDTA}$  reduction by VC ((A) color of  $0.01 \text{ mol L}^{-1}$   $\text{Fe}(\text{III})\text{EDTA}$  with pH 6.0; (B) color of the  $\text{Fe}(\text{III})\text{EDTA}$  solution after reduction by  $0.005 \text{ mol L}^{-1}$  VC at  $30^\circ\text{C}$  for 30 min; (C) color of the solution after a known mixed gas of NO and  $\text{N}_2$  was bubbled into (B); and (D) color of the  $\text{Fe}(\text{II})\text{EDTA}$ –NO solution after reduction by  $0.005 \text{ mol L}^{-1}$  VC at  $30^\circ\text{C}$  for 30 min).

0.00091, 0.00186, 0.00361, and 0.00457 M during 30 min corresponding to 0.002, 0.004, 0.008, and 0.01 M initial  $\text{Fe}(\text{III})\text{EDTA}$  concentration. Moreover, it was found that the ratio of produced  $\text{Fe}(\text{II})\text{EDTA}$  to generated dehydroascorbic acid was close to 2 : 1 at four different initial concentrations of  $\text{Fe}(\text{III})\text{EDTA}$ . Thus, *via* the abovementioned observations, it is proved that dehydroascorbic acid can be considered as the main product of VC oxidation.

Furthermore, the pH value variation is closely connected to  $\text{Fe}(\text{II})\text{EDTA}$  reduction by VC. We found that the pH values of the  $\text{Fe}(\text{III})\text{EDTA}$  solution after its reduction by VC dropped; this revealed that the  $\text{H}^+$  was produced when  $\text{Fe}(\text{III})\text{EDTA}$  was reduced by VC.

Based on the charge and element conservation principle, a credible stoichiometry of  $\text{Fe}(\text{III})\text{EDTA}$  reduction with VC was achieved:



In addition, in Fig. 3, we found that when the mole ratio of VC to  $\text{Fe}(\text{III})\text{EDTA}$  was 0.5 : 2, about 50% of  $\text{Fe}(\text{III})\text{EDTA}$  was reduced to  $\text{Fe}(\text{II})\text{EDTA}$ ; however, about 100%  $\text{Fe}(\text{III})\text{EDTA}$  was reduced to  $\text{Fe}(\text{II})\text{EDTA}$  when the molar ratios of VC to  $\text{Fe}(\text{III})\text{EDTA}$  were 1 : 2, 1.5 : 2, and 2 : 2. Thus, the reduction efficiency at different molar ratios of VC to  $\text{Fe}(\text{III})\text{EDTA}$  further confirmed that the stoichiometric ratio of VC to  $\text{Fe}(\text{III})\text{EDTA}$  was 1 : 2, and the chemical reaction equation of the  $\text{Fe}(\text{III})\text{EDTA}$  reduction by VC was determined as eqn (5).

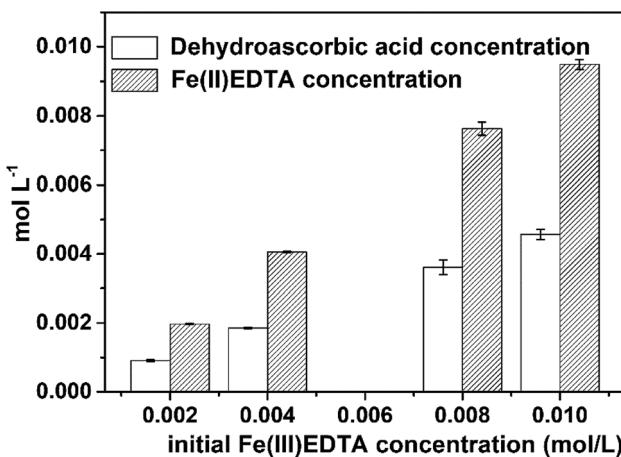


Fig. 8 The concentrations of the produced dehydroascorbic acid and  $\text{Fe}(\text{II})\text{EDTA}$  during 30 min. Conditions: original mole ratio of VC to  $\text{Fe}(\text{III})\text{EDTA}$  = 0.5; initial pH = 6.0; and  $T = 60^\circ\text{C}$ .

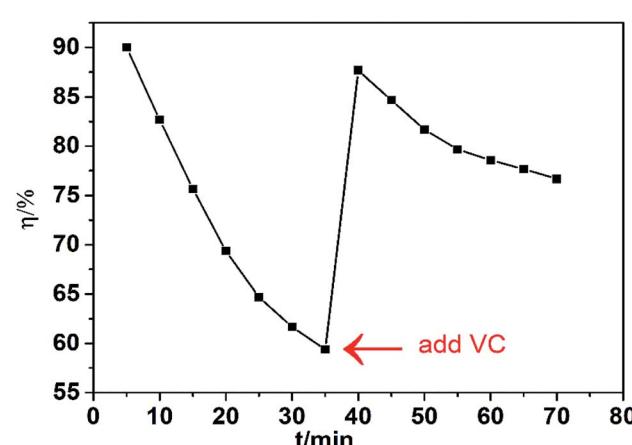


Fig. 9 The NO removal efficiency with time.



### 3.9. NO removal

The denitrification experiment using the Fe(II)EDTA solution with VC was conducted in a lab-scale bubbling reactor, as shown in Fig. S1.† The absorption liquid was 500 mL aqueous solution with the initial Fe(II)EDTA concentration of 0.002 M. The inlet NO concentration was 500 ppm. The oxygen concentration in the simulated flue gas was 7% (volume). The gas flow rate was 0.4 L min<sup>-1</sup>. The experimental results are shown in Fig. 9. We can observe that the NO removal efficiency decreases sharply from 90% to 59.36% in the first 35 min, during which the absorption liquid only contains Fe(II)EDTA. The decrease observed in the NO removal efficiency was mainly due to the oxidation of Fe(II)EDTA by dissolved oxygen present in the flue gas. Then, 0.1761 g VC (equivalent to 0.002 mol L<sup>-1</sup> of VC in 500 mL absorbent solution) was added to the absorption liquid at the 35th min. The NO removal efficiency increased to 87.67% and could remain above 76% for 35 min. Thus, the experimental results illustrate that NO removal with Fe(II)EDTA can be improved by the addition of VC.

## 4. Conclusion

The specific conclusions are as follows:

(1) Ascorbic acid (VC) was found to be a powerful reducing agent for the Fe(III)EDTA reduction. In addition, the stoichiometry of Fe(III)EDTA reduction with ascorbic acid was derived as follows:



(2) Fe(II)EDTA concentration increased as the initial Fe(III)EDTA concentration increased. Furthermore, the reduction efficiency increased with an increase in the mole ratio of VC to Fe(III)EDTA, and all the Fe(III)EDTA reduction efficiencies were close to 100% when the mole ratio was more than 0.5. Thus, an alkaline environment was unfavorable for Fe(III)EDTA conversion by VC. Moreover, the Fe(III)EDTA conversion slightly increased as the temperature increased.

(3) Upon comparing the present reduction system with other reduction systems, it was established that ascorbic acid (VC) was more powerful in reducing Fe(III)EDTA, especially in air.

(4) VC exhibited powerful ability only in the conversion of Fe(III)EDTA to Fe(II)EDTA and hardly reduced Fe(II)EDTA-NO.

(5) Thus, these findings could provide constructive information for the removal of NO in industrial applications using a mixed Fe(II)EDTA and ascorbic acid system.

## Conflicts of interest

There are no conflicts to declare.

## Acknowledgements

The project was supported by National Natural Science Foundation of China (No. 21808031).

## References

- 1 F. He, X. Deng and J. Ding, *RSC Adv.*, 2018, **8**, 32138–32145.
- 2 K. Liu, Q. Yu, J. Liu, K. Wang, Z. Han, Y. Xuan and Q. Qin, *New J. Chem.*, 2017, **41**, 13993–13999.
- 3 M. Chen, B. Xie, F. He and X. Deng, *Energy Fuels*, 2019, **33**, 1383–1391.
- 4 H. Chu, T.-W. Chien and S. Li, *Sci. Total Environ.*, 2001, **275**, 127–135.
- 5 Y. Zhao, T.-x. Guo, Z.-y. Chen and Y.-r. Du, *Chem. Eng. J.*, 2010, **160**, 42–47.
- 6 Y. Liu, J. Zhang, C. Sheng, Y. Zhang and L. Zhao, *Chem. Eng. J.*, 2010, **162**, 1006–1011.
- 7 F. He, X. Deng and M. Chen, *Chemosphere*, 2017, **168**, 623–629.
- 8 S. Wang, Z. Qi, Z. Gu, Z. Wang and Z. Ping, *J. Energy Inst.*, 2016, **90**, 522–527.
- 9 T. T. Susecki, B. Mathews and H. Kumazawa, *Ind. Eng. Chem. Res.*, 2005, **44**, 4249–4253.
- 10 E. Sada, H. Kumazawa, I. Kudo and T. Kondo, *Ind. Eng. Chem. Process Des. Dev.*, 1980, **19**, 377–382.
- 11 E. Sada, H. Kumazawa and Y. Takada, *Ind. Eng. Chem. Fundam.*, 1984, **23**, 60–64.
- 12 Y. G. Adewuyi and M. A. Khan, *Chem. Eng. J.*, 2015, **281**, 575–587.
- 13 Z. Zhou, G. Jing and X. Zheng, *Bioresour. Technol.*, 2013, **132**, 210–216.
- 14 W. Li, C.-Z. Wu, S.-H. Zhang, K. Shao and Y. Shi, *Environ. Sci. Technol.*, 2007, **41**, 639–644.
- 15 X. Yang, L. Yang, L. Dong, X. Long and W. Yuan, *Energy Fuels*, 2015, **25**, 4248–4255.
- 16 K. Xiang, H. Liu, B. Yang, C. Zhang, S. Yang, Z. Liu, C. Liu, X. Xie, L. Chai and X. Min, *Environ. Sci. Pollut. Res.*, 2016, **23**, 8113–8119.
- 17 T. T. Susecki and H. Kumazawa, *Org. Lett.*, 1998, **14**, 1206–1209.
- 18 T. T. Susecki, B. Mathews, A. W. Augustyniak and H. Kumazawa, *Ind. Eng. Chem. Res.*, 2014, **53**, 14234–14240.
- 19 M. H. Mendelsohn and J. B. L. Harkness, *Energy Fuels*, 1991, **5**, 244–248.
- 20 J. Hou, R. Lu, M. Sun, S. A. Baig, T. Tang, L. Cheng and X. Xu, *J. Hazard. Mater.*, 2012, **217–218**, 224–230.
- 21 S. Mowry and P. J. Ogren, *J. Chem. Educ.*, 1999, **76**, 970–974.
- 22 X. Xiang-Rong, L. Hua-Bin, L. Xiao-Yan and G. Ji-Dong, *Chemosphere*, 2004, **57**, 609–613.
- 23 X. Dong, Z. Yu, J. Zhou, H. Li, X. Wang and M. Chen, *J. Chem. Technol. Biotechnol.*, 2013, **89**, 111–116.
- 24 V. Zang and R. Van Eldik, *Inorg. Chem.*, 1990, **29**, 4462–4468.
- 25 A. E. Harvey, J. A. Smart and E. S. Amis, *Anal. Chem.*, 1955, **27**, 26–29.
- 26 J. H. Roe and M. J. Oesterling, *J. Biol. Chem.*, 1944, **152**, 511–517.
- 27 F. He, X. Deng and M. Chen, *Fuel*, 2016, **186**, 605–612.
- 28 T. T. Susecki, A. Barbara Mathews and H. Kumazawa, *Ind. Eng. Chem. Res.*, 2005, **44**, 4249–4253.



29 T. T. Suchecki, E. Sada and H. Kumazawa, *Ind. Eng. Chem. Res.*, 1991, **30**, 2201–2204.

30 K. Xiang, L. Hui, B. Yang, Z. Cong, Y. Shu, Z. Liu, L. Cao, X. Xie, L. Chai and X. Min, *Environ. Sci. Pollut. Res.*, 2016, **23**, 8113–8119.

31 X. J. Yang, Y. Lin, D. Li, X. L. Long and W. K. Yuan, *Energy Fuels*, 2011, **25**, 4248–4255.

32 F. He, Y. Qian and J. Xu, *Energy Fuels*, 2019, **33**(4), 3331–3338.

33 S.-H. Zhang, W. Li, C.-Z. Wu, H. Chen and Y. Shi, *Appl. Microbiol. Biotechnol.*, 2007, **76**, 1181–1187.

34 X. R. Xu, H. B. Li, X. Y. Li and J. D. Gu, *Chemosphere*, 2004, **57**, 609–613.

