



Cite this: RSC Adv., 2019, 9, 7747

## A microwave-activated coal fly ash catalyst for the oxidative elimination of organic pollutants in a Fenton-like process†

 Nannan Wang, \*ab Han Xu<sup>a</sup> and Shuo Li\*<sup>c</sup>

Raw coal fly ash was first activated by microwave irradiation to promote its catalytic potential and then used as a Fenton-like catalyst to treat polyacrylamide-contaminated wastewater. The optimal activation conditions of the raw coal fly ash (microwave power = 700 W, irradiation time = 10 min, mixing speed = 120 rpm, and raw coal fly ash loading = 20 g L<sup>-1</sup>) were determined by the orthogonal test. The significance of each effective parameter follows the order: raw coal fly ash loading > microwave power > irradiation time > mixing speed. Microwave irradiation can change the surface morphology and remarkably increase the specific surface area and pore volume. More than 75% of the TOC in the polyacrylamide-contaminated wastewater can be removed under the optimized treatment conditions ([H<sub>2</sub>O<sub>2</sub>] = 12 mg L<sup>-1</sup>, catalyst loading = 10 g L<sup>-1</sup>, [polyacrylamide] = 200 mg L<sup>-1</sup>, T = 313 K). The kinetic study shows that the variation in the catalyst loading and the polyacrylamide concentration can change the degradation path of the polyacrylamide, whereas the variation in the H<sub>2</sub>O<sub>2</sub> dosage can accelerate the degradation of polyacrylamide. The Fenton-like process studied herein has a wider optimal pH range (2–5) than that of the classic Fenton process (3). The catalyst has weak catalytic capacity but better catalytic persistence than that of Fe<sup>2+</sup>. During the five runs of the catalyst, some heavy metallic and toxic elements (Fe, Al, Ti, Cr, Mn, etc.) can be detected but under the limits of the GB8978-1996 standard. Leaching can weaken the catalytic capacity (*i.e.*, stability) of the catalyst. The catalytic process is caused by the synergism of multiple metals and consists of heterogeneous and homogeneous processes.

Received 1st February 2019

Accepted 21st February 2019

DOI: 10.1039/c9ra00875f

rsc.li/rsc-advances

## Introduction

The Fenton process, as a wastewater treatment process, is well known because of the generation of ·OH (2.8 V of redox potential *versus* the normal hydrogen electrode) *via* the reaction (1)<sup>1</sup>



However, the optimal pH (approx. 3)<sup>2</sup> and the generation of sludge containing the Fe element<sup>3</sup> limit the application of the Fenton process.

Exploring a new catalyst to replace Fe<sup>2+</sup> is a widely recognized solution to remove this limitation.<sup>4</sup> Some catalysts have been synthesized such as Z-scheme Fe<sub>2</sub>O<sub>3</sub>-doped Cu<sub>2</sub>O, Fe<sup>3+</sup>-Al<sub>2</sub>O<sub>3</sub>, CoFe<sub>2</sub>O<sub>4</sub> powder, Pd-Fe<sub>3</sub>O<sub>4</sub>, CoM<sub>x</sub>Fe<sub>2-x</sub>O<sub>4</sub>, Fe-Ti pillared bentonite, Fe<sub>73.5</sub>Si<sub>13.5</sub>B<sub>9</sub>Cu<sub>1</sub>Nb<sub>3</sub>, and Fe-Mn-sepiolite.<sup>5–11</sup> All

these catalysts can effectively catalyze the above reaction. It is valuable to notice that these catalysts have an identical feature, *i.e.*, they contain the Fe element and some auxiliary metallic elements such as Cu, Al, Co, Pd, Mn, Ti, and/or Nb. However, the preparation cost of these catalysts is high and the source of raw materials is limited. Thus, exploring a cheaper catalyst is encouraged.

Coal fly ash (CFA), as a well-known environmental pollutant, could be an appropriate alternative because of its distinctive features: (1) contains Fe element and other elements that are always doped on the synthesized Fenton-like catalyst; (2) can cause serious environmental pollution and exploring new methods to reuse it are urgent; (3) exists widely, easy to purchase, and its price is low. In the scientific research history of CFA, it has been studied in some research fields such as soil improvement, metal extraction, and preparation of zeolite/adsorbent.<sup>12–15</sup> Relatively speaking, the use of a Fenton-like catalyst after an appropriate activation to treat organic wastewater has not attracted enough attention until now. CFA has been reported to absorb electromagnetic waves because of the existence of ferromagnetic material and other metallic oxides.<sup>16</sup> However, this property was never noticed in the preparation of Fenton-like catalysts based on CFA.

<sup>a</sup>Department of Environmental Engineering, Beijing Institute of Petrochemical Technology, Beijing 102617, PR China. E-mail: wnn\_flying@163.com

<sup>b</sup>Beijing Key Laboratory of Pipeline Critical Technology and Equipment for Deepwater Oil & Gas Development, Beijing 102617, PR China

<sup>c</sup>College of Chemistry and Chemical Engineering, Qiqihar University, Qiqihar, 161006, PR China. E-mail: shuo\_105@163.com

† Electronic supplementary information (ESI) available. See DOI: 10.1039/c9ra00875f



Microwave (MW) is a wavelength band (300 MHz to 300 GHz) of the electromagnetic spectrum. It has distinctive thermal and non-thermal effects.<sup>2</sup> As for the thermal effect, MW can penetrate into (or pierce through) some materials (such as metallic oxides) and heat materials throughout their volume simultaneously. This heating mode exhibits a distinction compared with the traditional heating. As for the non-thermal effect, MW can excite the reactant molecules to the higher vibrational and rotational energy levels. MW has been used widely in chemical reactions depending on the ability of accelerating the reaction rate, reducing the activation energy, and strengthening the vibration of chemical bonds.<sup>17</sup>

In recent years, polyacrylamide (PAM) has been used widely in the enhanced oil recovery process as it can increase the bulk viscosity and improve the oil-water mobility ratio.<sup>18</sup> However, the generated polymer-flooding wastewater (PFW) has caused serious production pressure and environmental issues in China. The PAM in the PFW can enhance the stability of the oil-water emulsion, which makes it difficult to treat the PFW by conventional methods.<sup>19</sup> Nowadays, part of the PFW is being injected back into the stratum and the rest is discharged directly into natural water and/or soil, thus causing water and soil pollution. Therefore, the removal of PAM from PFW is a critical step in the effective treatment of PFW before efflux.

In this work, raw CFA (CFA<sub>R</sub>) was chosen as the raw material of a Fenton-like catalyst, synthetic PAM wastewater was chosen as the target pollutant, and a Fenton-like process catalyzed by MW-activated CFA (denoted by CFA<sub>MW</sub>) was chosen to treat the PAM wastewater. Firstly, the activation condition of the CFA<sub>R</sub> was optimized and the CFA<sub>MW</sub> was characterized by SEM, BET, XRD, XRF, and FTIR. After that, the treatment conditions of the PAM wastewater were optimized and the degradation kinetics of PAM was analyzed. Subsequently, the features of the Fenton-like process (optimal pH range, leachability, and stability of CFA<sub>MW</sub>) were investigated and discussed. Finally, the catalytic mechanism of CFA<sub>MW</sub> was proposed. This work gives a new activation method for CFA<sub>R</sub> and provides an idea on how to minimize the pollution from CFA and PFW.

## Materials and methods

### Chemical reagents

PAM (analytical grade, molecular weight =  $3.0 \times 10^6$ ) and H<sub>2</sub>O<sub>2</sub> were purchased from Sigma-Aldrich, while HCl and NaOH were purchased from Honeywell. H<sub>2</sub>O<sub>2</sub> was stored in the dark at 277 K. All chemical reagents were prepared using 18 MΩ cm Milli-Q water (MQ). All used glassware were soaked in an alcohol solution for at least 24 h and washed with MQ before use.

The CFA<sub>R</sub> was obtained from a coal-fired power plant in China. The electromagnetic parameters of the CFA<sub>R</sub> are shown in Table 1. It was washed by MQ before the experiments to remove the impurities and dried at 378 K.

### Preparation of CFA<sub>MW</sub>

The MW activation process of CFA<sub>R</sub> was conducted in a retrofitted MW oven (EM-202MS1, SANYO, Japan, Fig. 1). A certain

Table 1 Electromagnetic parameters of CFA<sub>R</sub> used in this work<sup>a</sup>

Frequency/GHz	$\epsilon'$	$\epsilon''$	$\tan \delta_e$	$\mu'$	$\mu''$	$\tan \delta_m$
1	2.42	0.13	0.054	0.96	0.15	0.156
2	2.42	0.11	0.045	0.98	0.11	0.112
2.45	2.41	0.08	0.033	0.99	0.08	0.081
3	2.40	0.05	0.021	1.01	0.08	0.079
4	2.43	0.05	0.021	1.01	0.07	0.069

<sup>a</sup>  $\epsilon'$  and  $\epsilon''$  are the real part and imaginary part of complex permittivity, respectively;  $\tan \delta_e = \epsilon''/\epsilon'$  is the dielectric loss tangent;  $\mu'$  and  $\mu''$  are the real part and imaginary part of complex permeability, respectively;  $\tan \delta_m = \mu''/\mu'$  is the magnetic loss tangent.

amount of CFA<sub>R</sub> was tiled on the multi-layer plate in the preparation chamber. The MW and the vibratory rod were switched on simultaneously and the activation process was started. The vibration of the chamber can force the CFA<sub>R</sub> to disperse in the whole space of the preparation chamber and to receive adequate MW irradiation. After the activation, the prepared CFA catalyst (*i.e.*, CFA<sub>MW</sub>) was stored in a drying basin.

### Experimental process

In the treatment of the PAM wastewater, a certain amount of CFA<sub>MW</sub> and 300 mL PAM solution (200 mg L<sup>-1</sup>) with predetermined pH were added into a beaker (500 mL) successively. The mixture was mixed by mechanical agitation. The degradation of PAM commenced immediately after adding the H<sub>2</sub>O<sub>2</sub> solution.

All experiments were conducted at a constant temperature controlled by a water bath ( $\pm 1$  K). At every determined time interval, the mixture was sampled and every sample was pre-treated by adding NaOH solution and MnO<sub>2</sub> solid as soon as possible, before filtration. This can stop the generation of 'OH and decompose the residual H<sub>2</sub>O<sub>2</sub>. All the pretreated samples were filtered by vacuum filtration, and all the filtrates were analyzed to determine the TOC and concentration of the leached metallic ions.

In the study of the stability of CFA<sub>MW</sub>, the separated CFA<sub>MW</sub> from the treated PAM wastewater was washed by MQ water first

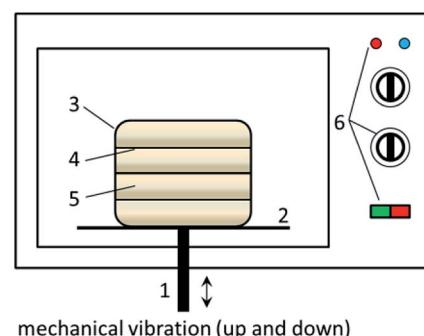


Fig. 1 Schematic for the preparation setup of CFA<sub>MW</sub> (1) vibratory rod; (2) platform; (3) preparation chamber with multilayers (material: glass; total volume = 1 L); (4) multi-layer plate (material: glass); (5) CFA<sub>R</sub>; (6) switch and control knob.



and then dried at 378 K until the weight did not change any more. The CFA<sub>MW</sub> was used repeatedly under the identical experimental conditions of the first run.

### Analytical methods

The TOC in the PAM wastewater was measured by a Rosemount Analytical Dohrmann DC-190 ASM TOC analyzer. The removal rate of TOC was calculated by eqn (2).

$$\text{Removal rate (\%)} = ((c_0 - c)/c_0) \times 100 \quad (2)$$

where,  $c_0$  and  $c$  mean the initial concentration of TOC (mg L<sup>-1</sup>) and the instantaneous concentration of TOC (mg L<sup>-1</sup>), respectively.

The utilization rate (UR, L mg<sup>-1</sup>) of H<sub>2</sub>O<sub>2</sub> was calculated by eqn (3):

$$\text{UR}_{n-m} = (\varphi_m - \varphi_n)/([H_2O_2]_m - [H_2O_2]_n) \quad (3)$$

where,  $[H_2O_2]_m$  and  $[H_2O_2]_n$  mean  $[H_2O_2] = m \text{ mg L}^{-1}$  and  $[H_2O_2] = n \text{ mg L}^{-1}$ , respectively.  $\varphi_m$  and  $\varphi_n$  are the removal rates of TOC corresponding  $[H_2O_2]_m$  and  $[H_2O_2]_n$ , respectively, where  $n = 0, 4, 8, 12, 16$  corresponding to  $m = 4, 8, 12, 16, 20$ , one to one.

The life span of ·OH in aqueous solution is always very short and it is, therefore, difficult to determine the specific concentration of ·OH directly. Thus, a terephthalic acid fluorescence probing technique was applied in this work. Terephthalic acid can capture ·OH effectively and generate 2-hydroxyterephthalic acid. The fluorescence intensity of 2-hydroxyterephthalic acid is in proportion to the amount of ·OH. Terephthalic acid solution was prepared using MQ water according to the literature.<sup>20</sup>

### Characterization methods

The surface morphology of CFA<sub>MW</sub> and CFA<sub>R</sub> was observed by scanning electron microscopy (SEM, S4800, Hitachi). The specific surface area and the features of the pores of CFA<sub>MW</sub> and CFA<sub>R</sub> were determined by Brunauer–Emmett–Teller automated analyzer (BET, ASAP 2420, Micromeritics). The crystal structure of CFA<sub>MW</sub> and CFA<sub>R</sub> was determined by X-ray diffraction (XRD, D/max-rB, Rigaku Corporation). The chemical composition of CFA<sub>MW</sub> and CFA<sub>R</sub> was analyzed quantitatively by X-ray fluorescence spectrometry (XRF, ARLADVANT XP+, Thermo electron corporation). The Fourier transform infrared spectrum of CFA<sub>MW</sub> was measured by FTIR spectrometer (FTIR-650). The concentration of metallic elements in the wastewater was measured by flame atomic absorption spectrometry (PinAAcle 900T, PerkinElmer). The electromagnetic property of CFA<sub>R</sub> was measured by Vector Network Analyzers (HP8722ES).

## Results and discussion

### Optimization of activation condition

The electromagnetic property of CFA<sub>R</sub> is given in Table 1, from which it can be seen that both of the imaginary parts ( $\epsilon''$  and  $\mu''$ ) of the complex permittivity and complex permeability are affected by the variation of MW frequency. They always decrease

with the increase of frequency. However, the real parts ( $\epsilon'$  and  $\mu'$ ) of the complex permittivity and complex permeability exhibit insensitivity to the variation of MW frequency. Therefore, the dielectric loss tangent ( $\tan \delta_e$ ) and magnetic loss tangent ( $\tan \delta_m$ ) have the same variation trend with the two imaginary parts. As both of the  $\tan \delta_e$  and  $\tan \delta_m$  are greater than  $1.0 \times 10^{-2}$  within the range of 1–4 GHz, the CFA<sub>R</sub> shows absorbability of MW by means of dielectric loss and magnetic loss.

In this section, the activation condition of CFA<sub>R</sub> by MW irradiation (2.45 GHz) was optimized based on the removal rate of TOC in the PAM wastewater. The experiment was arranged according to the orthogonal test ( $L_{25}(4^4)$ ) and the optimum experimental conditions were obtained by the analysis of the average value ( $I_{ij}$ , eqn (4)) and the extremum ( $R_{E,i}$ , eqn (5)) of the TOC removal rate,<sup>21</sup>

$$I_{ij} = \left( \sum_{m=1}^5 x_{ijm} \right) / 5 \quad (4)$$

$$R_{E,i} = I_{ij,\max} - I_{ij,\min} \quad (5)$$

where,  $i$ ,  $j$ , and  $m$  mean the number of factor, the number of level, and the number of occurrence of (level  $j$  of factor  $i$ ), respectively.

As shown in Table S1,<sup>†</sup> taking the factor 1 (power of MW) for example, the  $I_{14}$  (40.2) has the highest value, while  $I_{13}$  (26.0) has the lowest one. Therefore, it can be inferred that 700 W ( $j = 4$ ) of MW power is the optimal value. The same is applicable for each of the other factors (irradiation time, mixing speed, loading of CFA<sub>R</sub>). Thus, the optimal experimental conditions are: power of MW = 700 W, irradiation time = 10 min, mixing speed = 120 rpm, and loading of CFA<sub>R</sub> = 20 g L<sup>-1</sup>.

The order of the significance of the four factors can be determined by the comparison of extrema ( $R_E$ ). As shown in Table S1,<sup>†</sup> the loading of CFA<sub>R</sub> has the most pronounced effect on the removal rate of TOC because of the highest  $R_E$  (27.3), while mixing speed has the weakest effect (9.9). The effect of the power of MW and irradiation time occupy the second and third places, respectively. Therefore, the order of significance is: loading of CFA<sub>R</sub> > power of MW > irradiation time > mixing speed.

Actually, the above order is in accordance with the property of MW. For a given mixing speed, the MW irradiation can penetrate into the substrate and cause the simultaneous heating of a whole block of material. Therefore, the effect of mixing speed is relatively weaker. For a given irradiation time, the hot-spot effect of the MW irradiation can be produced quickly. Thus, the activation of CFA<sub>R</sub> occurs and finishes in a short time. This determines that the irradiation time is relatively less significant as well. As for the power of the MW and the loading of CFA<sub>R</sub>, the power of MW can determine the instantaneous temperature of the hot-spot, while more loading of CFA<sub>R</sub> will disperse the MW energy, thus causing the decrease of instantaneous temperature of the hot-spot. Therefore, they have an important effect compared with the irradiation time and mixing speed.



Characteristics of CFA<sub>MW</sub> and CFA<sub>R</sub>

**Surface morphology.** The surface morphology of CFA<sub>R</sub> and CFA<sub>MW</sub> was observed and the components in the special locations were measured by the combination of SEM and EDX. As shown in Fig. 2, the CFA always exhibits uneven surface (Fig. 2A1 and B1) and spheres (Fig. 2A2 and B2) regardless of whether the CFA<sub>R</sub> is activated or not. However, the surface morphology of CFA<sub>R</sub> has a pronounced variation after MW activation. On focusing the attention on the surface of CFA<sub>R</sub> and CFA<sub>MW</sub>, it can be observed that it is relatively smooth for CFA<sub>R</sub>, while many spherical particles are visible on the surface of CFA<sub>MW</sub>.

The ratio of the components of the elements on the smooth surface and the spherical particles were measured and compared. As shown in Fig. 2C1–C4, the smooth surface (P<sub>1</sub> and P<sub>2</sub>) of CFA<sub>R</sub> and CFA<sub>MW</sub> can have the similar ratio of elements, while the main catalytic component (*i.e.*, Fe element) is higher

**Table 2** Specific surface area and pore characteristics of CFA<sub>MW</sub> and CFA<sub>R</sub>

Catalyst	Specific surface area (m <sup>2</sup> g <sup>-1</sup> )	Pore volume (cm <sup>3</sup> g <sup>-1</sup> )	Average pore diameter (nm)
CFA <sub>MW</sub>	32.19	0.079	4.79
CFA <sub>R</sub>	17.68	0.058	5.76

in CFA<sub>MW</sub> (P<sub>2</sub>) than that in CFA<sub>R</sub> (P<sub>1</sub>). The components in the spherical particles (P<sub>3</sub> and P<sub>4</sub>) exhibit remarkable difference. The Fe element on the surface of the spherical particles is much higher than that on the smooth surface, while the Ca element is relatively low. Thus, it can be inferred that the active sites on the surface of the CFA<sub>MW</sub> are higher than those on CFA<sub>R</sub>, which is beneficial for the improvement of the catalytic capacity.

**Specific surface area and pore characteristics.** The specific surface areas and pore characteristics of CFA<sub>MW</sub> and CFA<sub>R</sub> were measured and are compared in Table 2. From Table 2, it can be seen that the specific surface area and pore characteristics are significantly different before and after activation. The specific surface area (32.19 m<sup>2</sup> g<sup>-1</sup>) and pore volume (0.079 cm<sup>3</sup> g<sup>-1</sup>) of CFA<sub>MW</sub> is higher than those (17.68 m<sup>2</sup> g<sup>-1</sup>, 0.058 cm<sup>3</sup> g<sup>-1</sup>) of CFA<sub>R</sub>, while the average pore diameter (4.79 nm) of CFA<sub>MW</sub> is lower than that of CFA<sub>R</sub> (5.76 nm).

**Crystalline phase and components of the metallic oxides.** The crystalline phase and components of CFA<sub>MW</sub> was compared with those of CFA<sub>R</sub>. The XRD pattern of CFA<sub>MW</sub> (not given) shows that the primary crystalline phase is mullite (Al<sub>6</sub>Si<sub>2</sub>O<sub>13</sub>), while quartz (SiO<sub>2</sub>), and albite ((Na, Ca)Al(Si, Al)<sub>3</sub>O<sub>8</sub>) occupy the minority. This result is similar with that of CFA<sub>R</sub>. Moreover, the components of CFA<sub>MW</sub> and CFA<sub>R</sub> are shown in Table S2,† from which it can be seen that the components are analogous to each other as well.

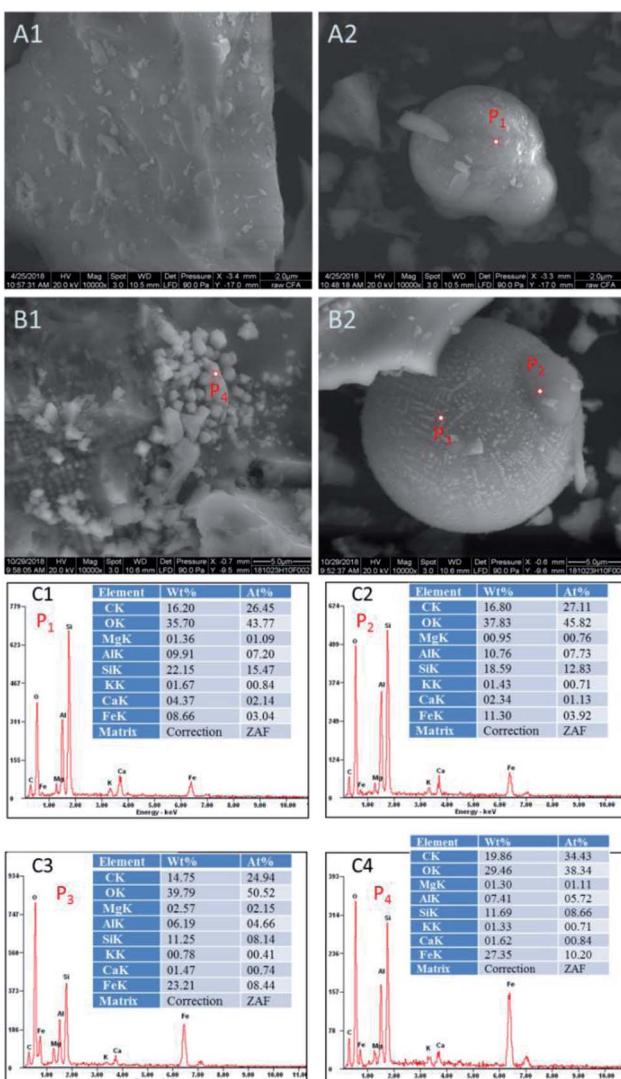
In essence, it is not difficult to understand this result. The actual temperature on the surface of CFA during the activation process is no more than 573 K, which is much lower than that in the circulating fluidized bed boiler and is not enough to change the crystalline phase. In addition, the variation of the components of the metallic elements will never occur under the MW irradiation. The fluctuation of data is caused by the measurement error.

**Functional groups of CFA<sub>MW</sub>.** The FTIR result in Fig. 3 echoes with the result of XRD. The peaks at 795 cm<sup>-1</sup> and 1086 cm<sup>-1</sup> represent the anti-symmetric stretching vibration of Si–O–Si, while the peak at 557 cm<sup>-1</sup> represents the stretching vibration of Al–O in the AlO<sub>6</sub> group.<sup>22,23</sup> The three peaks correspond to the existence of mullite and quartz. The peak at 457 cm<sup>-1</sup> is the characteristic peak of the bending vibrations of the [FeO<sub>6</sub>] group.<sup>24</sup>

It is valuable to notice the stretching vibrations of 'OH inferred from the peak at 3410 cm<sup>-1</sup>, which is critical to the catalytic activity of CFA<sub>MW</sub> due to the formation of ≡Fe–OH (see Fig. 7).

## Experimental parameters in the treatment of PAM wastewater

The optimization of the treatment conditions is a critical step in studying the CFA<sub>MW</sub> catalyzed Fenton-like process. In this work,



**Fig. 2** SEM images and EDX spectra of CFA<sub>MW</sub> and CFA<sub>R</sub> (A1 and A2: SEM of CFA<sub>R</sub>; B1 and B2: SEM of CFA<sub>MW</sub>; C1, C2, C3, and C4: EDX spectra of P<sub>1</sub>, P<sub>2</sub>, P<sub>3</sub>, and P<sub>4</sub>).



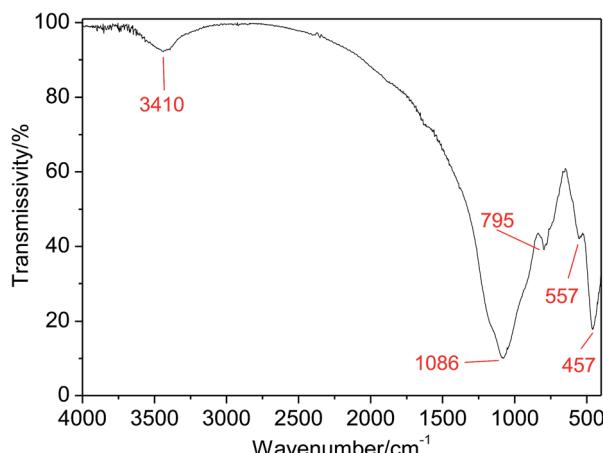


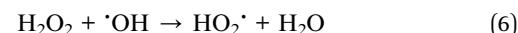
Fig. 3 FTIR spectrum of  $\text{CFA}_{\text{MW}}$ .

$\text{H}_2\text{O}_2$  dosage,  $\text{CFA}_{\text{MW}}$  loading, TOC in the PAM wastewater, and the treatment temperature were studied and optimized, respectively.

The data in Fig. 4A shows the effect of  $\text{H}_2\text{O}_2$  dosage on the TOC removal rate. The removal rate of TOC always increases with the increase of the  $\text{H}_2\text{O}_2$  dosage. However, it is important to notice that the increase of  $\text{H}_2\text{O}_2$  dosage from 12 to 20  $\text{mg L}^{-1}$  only results in a small growth in the TOC removal rate (from 63.8% to 70.6%) at 50 min. The UR of  $\text{H}_2\text{O}_2$  in the inset of Fig. 4A shows that more loading of  $\text{H}_2\text{O}_2$  can cause the rapid decrease of the UR of  $\text{H}_2\text{O}_2$ . When the loading of  $\text{H}_2\text{O}_2$  is 4  $\text{mg L}^{-1}$ , the UR is 10.3  $\text{L mg}^{-1}$ , while more loading of  $\text{H}_2\text{O}_2$

(8  $\text{mg L}^{-1}$ ) results in the decrease of UR from 10.3  $\text{L mg}^{-1}$  to 2.9  $\text{L mg}^{-1}$ . Considering the balance of the removal rate of TOC with the UR of  $\text{H}_2\text{O}_2$ , the optimal dosage of  $\text{H}_2\text{O}_2$  was determined to be 12  $\text{mg L}^{-1}$ .

The side reaction occurring in Fenton-like process is responsible for the decline of the UR of  $\text{H}_2\text{O}_2$ . As shown in reaction (6), the excessive loading of  $\text{H}_2\text{O}_2$  can hinder the reaction between PAM and 'OH by the mutual consumption of 'OH with  $\text{H}_2\text{O}_2$ .



The data in Fig. 4B shows the effect of  $\text{CFA}_{\text{MW}}$  loading on the TOC removal rate. Similar to the effect of  $\text{H}_2\text{O}_2$  dosage, with the increase of  $\text{CFA}_{\text{MW}}$  loading, more TOC can be eliminated but it encounters a bottleneck when the  $\text{CFA}_{\text{MW}}$  loading reaches 10  $\text{g L}^{-1}$ . More  $\text{CFA}_{\text{MW}}$  loading causes the rapid consumption of  $\text{H}_2\text{O}_2$  and the rapid generation of 'OH. Reaction (7) becomes more and more significant with the increase of 'OH concentration, *i.e.*, the internal consumption of 'OH becomes more severe.



The data in Fig. 4C shows the effect of PAM concentration on the TOC removal rate. It can be seen that the increase in PAM concentration has less effect in the range of 50–200  $\text{mg L}^{-1}$ , while the effect become substantial when the PAM concentration increases beyond 200  $\text{mg L}^{-1}$ . This result can be caused by two reasons. Firstly, the 'OH can degrade only a fixed amount of

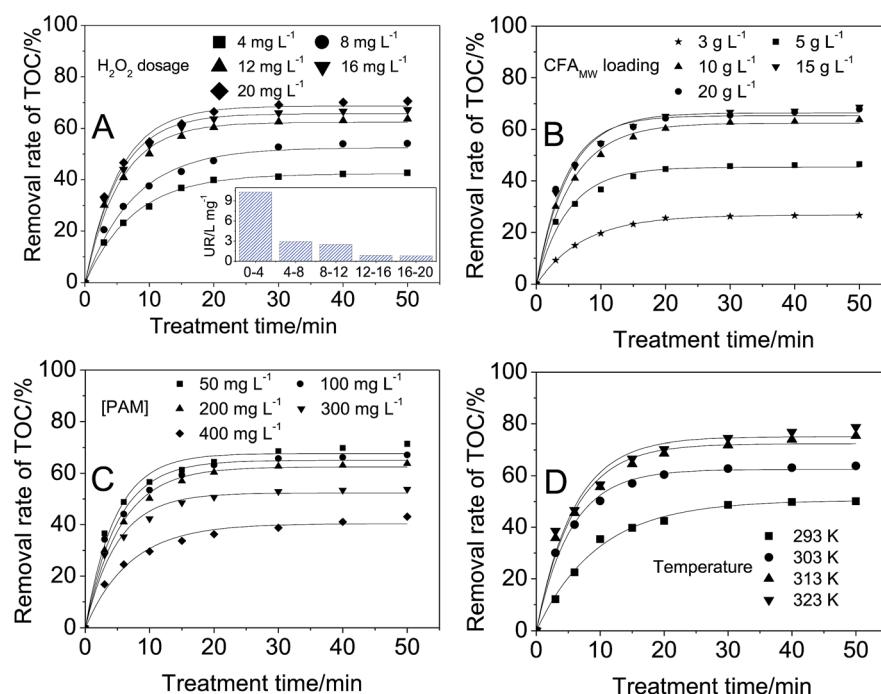


Fig. 4 The effect of experimental conditions on the removal rate of TOC in PAM wastewater. ((A)  $\text{CFA}_{\text{MW}}$  loading = 10  $\text{g L}^{-1}$ ,  $[\text{PAM}]$  = 200  $\text{mg L}^{-1}$ ,  $T$  = 303 K, pH = 3; (B)  $\text{H}_2\text{O}_2$  dosage = 12  $\text{mg L}^{-1}$ ,  $[\text{PAM}]$  = 200  $\text{mg L}^{-1}$ ,  $T$  = 303 K, pH = 3; (C)  $\text{H}_2\text{O}_2$  dosage = 12  $\text{mg L}^{-1}$ ,  $\text{CFA}_{\text{MW}}$  loading = 10  $\text{g L}^{-1}$ ,  $T$  = 303 K, pH = 3; (D)  $\text{H}_2\text{O}_2$  dosage = 12  $\text{mg L}^{-1}$ ,  $\text{CFA}_{\text{MW}}$  loading = 10  $\text{g L}^{-1}$ ,  $[\text{PAM}]$  = 200  $\text{mg L}^{-1}$ , pH = 3).

PAM, thus more PAM can decrease its removal rate. Secondly, more PAM implies an increase in the viscosity of the wastewater, thus the diffusion of the catalyst  $\text{H}_2\text{O}_2$  and  $\cdot\text{OH}$  is limited seriously even if adequately mixed.

The data in Fig. 4D shows the effect of the wastewater temperature, from which it can be seen that higher temperature encourages in the direction of increasing removal rate of TOC. The removal rate increases from 50.1% to 78.7% at 50 min when the temperature rises from 293 K to 323 K. In this work, 313 K was the optimal value as higher temperature cannot significantly increase the removal rate of TOC.

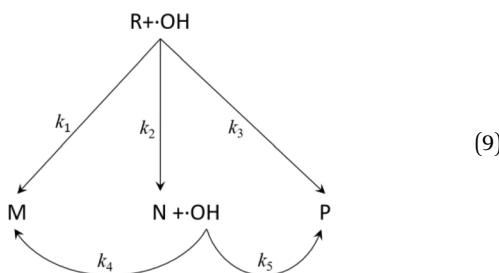
### Kinetic study

The frequently-used pseudo-first order kinetic model is<sup>25</sup>

$$-\frac{d[\text{TOC}_R]}{dt} = k_{ap} \cdot [\text{TOC}]_R \quad (8)$$

where R represents PAM, while  $\text{TOC}_R$  means the TOC ( $\text{mg L}^{-1}$ ) of the PAM wastewater,  $k_{ap}$  is the pseudo-first order kinetic constant.

However, eqn (8) does not consider the degradation process of PAM and the properties of the degradation products. In reality, the degradation products of organics (including PAM) can always be divided into three kinds, *i.e.*, inorganic matter ( $\text{H}_2\text{O}$ ,  $\text{CO}_2$ , *etc.*), degradable organic matter (intermediates), and non-degradable organic matter.<sup>26</sup> In this section, the generation of the three kinds of products was considered (see eqn (9)) and the corresponding detailed degradation kinetic (DDK) model was derived.



where  $k_1$ ,  $k_2$ ,  $k_3$ ,  $k_4$ , and  $k_5$  are the kinetic constants.

In the DDK model, the initial organic pollutant (R, *i.e.*, PAM) can be degraded by  $\cdot\text{OH}$  *via* three paths, *i.e.*, being degraded to non-degradable organic matters (M,  $k_1$ ), being degraded to intermediates (N,  $k_2$ ), and being degraded to inorganic matter (P,  $k_3$ ). Moreover, N can be degraded further by  $\cdot\text{OH}$  to M and P, corresponding to the kinetic constants  $k_4$  and  $k_5$ , respectively.

Assuming that the five degradation processes follow the pseudo-first kinetics, the DDK model can be described as:

$$-\frac{d[\text{TOC}_R]}{dt} = (k_1 + k_2 + k_3) \cdot [\text{TOC}]_R \quad (10)$$

$$-\frac{d[\text{TOC}_N]}{dt} = (k_4 + k_5) \cdot [\text{TOC}]_N - k_2 \cdot [\text{TOC}]_R \quad (11)$$

$$-\frac{d[\text{TOC}_M]}{dt} = -k_1 \cdot [\text{TOC}]_R - k_4 \cdot [\text{TOC}]_N \quad (12)$$

The TOC of the PAM wastewater can be formulated by the integration of eqn (10)–(12):

$$\begin{aligned} \frac{[\text{TOC}]}{[\text{TOC}]_0} &= \frac{[\text{TOC}]_R + [\text{TOC}]_N + [\text{TOC}]_M}{[\text{TOC}]_{R0} + [\text{TOC}]_{N0} + [\text{TOC}]_{M0}} \\ &= \frac{[\text{TOC}]_{R0}}{[\text{TOC}]_0} \left( \frac{k_3(k_3 + k_2 + k_1 - k_5 - k_4) - k_5 k_2}{(k_3 + k_2 + k_1)(k_3 + k_2 + k_1 - k_5 - k_4)} \right. \\ &\quad \times \exp(-(k_3 + k_2 + k_1)t) \\ &\quad \left. + \frac{k_5 k_2}{(k_2 + k_4)(k_3 + k_2 + k_1 - k_5 - k_4)} \right. \\ &\quad \times \exp(-(k_5 + k_4)t) + \frac{k_1(k_5 + k_4) + k_4 k_2}{(k_3 + k_2 + k_1)(k_5 + k_4)} \left. \right) \\ &\quad + \frac{[\text{TOC}]_{N0}}{[\text{TOC}]_0} \left( \frac{k_5}{k_5 + k_4} \exp(-(k_5 + k_4)t) + \frac{k_4}{k_5 + k_4} \right) \\ &\quad + \frac{[\text{TOC}]_{M0}}{[\text{TOC}]_0} \end{aligned} \quad (13)$$

$[\text{TOC}]_R$ ,  $[\text{TOC}]_N$ , and  $[\text{TOC}]_M$  represent the TOCs of R, N, and M at time  $t$ , respectively;  $[\text{TOC}]_{R0}$ ,  $[\text{TOC}]_{N0}$ , and  $[\text{TOC}]_{M0}$  are the initial TOCs of R, N, and M, respectively.

As there is no other chemical substance in the initial PAM wastewater, the initial TOC of M and N is zero. Thus, eqn (13) can be further simplified to eqn (14).

$$\begin{aligned} \frac{[\text{TOC}]}{[\text{TOC}]_0} &= \frac{[\text{TOC}]_R + [\text{TOC}]_N + [\text{TOC}]_M}{[\text{TOC}]_{R0} + [\text{TOC}]_{N0} + [\text{TOC}]_{M0}} \\ &= \frac{[\text{TOC}]_{R0}}{[\text{TOC}]_0} \left( \frac{k_3(k_3 + k_2 + k_1 - k_5 - k_4) - k_5 k_2}{(k_3 + k_2 + k_1)(k_3 + k_2 + k_1 - k_5 - k_4)} \right. \\ &\quad \times \exp(-(k_3 + k_2 + k_1)t) \\ &\quad \left. + \frac{k_5 k_2}{(k_2 + k_4)(k_3 + k_2 + k_1 - k_5 - k_4)} \right. \\ &\quad \times \exp(-(k_5 + k_4)t) + \left. \frac{k_1(k_5 + k_4) + k_4 k_2}{(k_3 + k_2 + k_1)(k_5 + k_4)} \right) \end{aligned} \quad (14)$$

The five kinetic parameters in eqn (14) can be obtained by the use of Matlab, based on the multi-variable least-squares method.<sup>2</sup>

The fitting results of  $k_1$ ,  $k_2$ ,  $k_3$ ,  $k_4$ , and  $k_5$  under different  $\text{H}_2\text{O}_2$  dosage,  $\text{CFA}_{\text{MW}}$  dosage, and PAM concentration are shown in Table 3, from which it can be seen that the regression coefficient ( $R^2$ ) is always higher than 0.998, thus indicating that the experimental data can follow the DDK model well.

With the increase of the  $\text{H}_2\text{O}_2$  dosage from 4 to 20  $\text{mg L}^{-1}$ , the apparent kinetic constant  $k_1$ ,  $k_2$ ,  $k_3$ , and  $k_5$  in the DDK model always increase from 0.0891, 0.0536, 0.0636, and 0.0556  $\text{min}^{-1}$  to 0.1072, 0.0650, 0.0704, and 0.0665  $\text{min}^{-1}$ , respectively. The increase of the above four kinetic constants confirms that PAM can be removed more rapidly due to the generation of more  $\cdot\text{OH}$  during the treatment time, while whether the degradation path can be changed cannot be inferred from this result. With the increase of  $\text{CFA}_{\text{MW}}$  loading from 3 to 20  $\text{g L}^{-1}$ ,  $k_1$  and  $k_2$  decrease while  $k_3$  and  $k_5$  increase, thus revealing that the primary degradation pathway of PAM shifts towards the direct formation of inorganic matter due to the attack of more  $\cdot\text{OH}$  in a short time under the catalysis by more  $\text{CFA}_{\text{MW}}$ . With the increase of PAM concentration from 50 to 400  $\text{mg L}^{-1}$ ,  $k_1$ ,  $k_2$ ,  $k_3$ , and  $k_5$  change as well, revealing that the degradation rate is also dependent on the PAM concentration. The increase of  $k_2$  (from 0.0601 to 0.0685  $\text{min}^{-1}$ ) and the decrease of  $k_1$ ,  $k_3$ , and  $k_5$



Table 3 DDK model parameters calculated from the degradation kinetic analysis of PAM in the  $\text{CFA}_{\text{MW}}$  catalyzed Fenton-like process

$\text{H}_2\text{O}_2$ dosage (mg L <sup>-1</sup> )	$\text{CFA}_{\text{MW}}$ loading (g L <sup>-1</sup> )	PAM concentration (mg L <sup>-1</sup> )	Kinetic constant of the DDK model (min <sup>-1</sup> )					
			$k_1$	$k_2$	$k_3$	$k_4$	$k_5$	
4	3	50	0.0891	0.0536	0.0636	0	0.0556	0.9986
12	3	50	0.0904	0.0603	0.0661	0	0.0613	0.9989
20	3	50	0.1072	0.0650	0.0704	0	0.0665	0.9991
20	10	50	0.0879	0.0626	0.0799	0	0.0701	0.9981
20	20	50	0.0753	0.0601	0.0856	0	0.0755	0.9992
20	20	200	0.0719	0.0644	0.0799	0	0.0638	0.9987
20	20	400	0.0678	0.0685	0.0728	0	0.0609	0.9996

implied that the primary degradation pathway of PAM is more towards the formation of intermediates (N), *i.e.*, more PAM is not degraded completely. The value of  $k_4$  is always kept at zero, showing that the degradation path of N to M was not manifested.

### Features of $\text{CFA}_{\text{MW}}$ catalyzed Fenton-like process

**Optimal pH range.** In the classic Fenton process, the optimum range of pH is always fixed at 3,<sup>27</sup> thus limiting the application of the Fenton process in industry. Thus, widening the pH range is important to break the limitation. In this section, the optimal pH range of the Fenton-like process was investigated to examine whether the  $\text{CFA}_{\text{MW}}$  has advantage in this aspect.

As shown in Fig. 5, when pH varies in the range of 2–5, the removal rate of TOC varies in a small range (64.6–70.5%) at 50 min. Raising the pH further causes a remarkable decrease of the removal rate of TOC, revealing that the optimal range of pH is 2–5. This range is not very wide but has shown progress compared with that (pH = 3) of the classic Fenton process.

The difference in the pH range could be caused by the different properties of the catalysts. The performance of  $\text{Fe}^{2+}$  in the Fenton process is affected significantly by pH. A pH higher than 3 can cause the oxidation of  $\text{Fe}^{2+}$  and precipitation in the form of  $\text{Fe}(\text{OH})_3$ , resulting in the deactivation of  $\text{Fe}^{2+}$ . Relatively speaking, the effect of pH on the surface of  $\text{CFA}_{\text{MW}}$  is weaker. The active sites on the surface of  $\text{CFA}_{\text{MW}}$  always contain  $\text{Fe}_2\text{O}_3$ ,

which have been oxidized to +3. The catalytic process of the active sites is different from  $\text{Fe}^{2+}$  and the  $\text{OH}^-$  has a relatively weak effect on the catalytic mechanism.

**Investigation of  $\cdot\text{OH}$  in different oxidation systems.** The relative amount of  $\cdot\text{OH}$  generated in different Fenton-like processes were compared. As shown in Fig. 6, it can be observed intuitively that  $\text{Fe}^{3+}\text{--H}_2\text{O}_2$  and  $\text{CFA}_{\text{R}}\text{--H}_2\text{O}_2$  systems generate a lesser amount of  $\cdot\text{OH}$  than that of another two systems. In the  $\text{Fe}^{3+}\text{--H}_2\text{O}_2$  system, the  $\text{Fe}^{3+}$  needs to be reduced to  $\text{Fe}^{2+}$  first before generating  $\cdot\text{OH}$ . However, the reduction reaction of  $\text{Fe}^{3+}$  (eqn (15)) is the rate controlled step ( $k = 0.01\text{--}0.02 \text{ L mol}^{-1} \text{ s}^{-1}$ ).<sup>28</sup>



In the  $\text{CFA}_{\text{R}}\text{--H}_2\text{O}_2$  system, the active sites are relatively lesser in number because they may be covered by other minerals or inactive oxides.

When focussing attention on the  $\text{Fe}^{2+}\text{--H}_2\text{O}_2$  and  $\text{CFA}_{\text{MW}}\text{--H}_2\text{O}_2$  systems, it is interesting that the amount of  $\cdot\text{OH}$  generated in the  $\text{Fe}^{2+}\text{--H}_2\text{O}_2$  system is more in the prophase of the reaction (at 5 min) but lesser in the later stage of the reaction (at 10 and 15 min) (see Fig. 6). This result indicates the different advantages reflected by the  $\text{Fe}^{2+}\text{--H}_2\text{O}_2$  system and the  $\text{CFA}_{\text{MW}}$  system. As for the  $\text{Fe}^{2+}\text{--H}_2\text{O}_2$  system,  $\text{Fe}^{2+}$  can have a better catalytic capacity than  $\text{CFA}_{\text{MW}}$ . This is the primary reason that more  $\cdot\text{OH}$  can be generated before 5 min. However, as the reaction proceeds,  $\text{Fe}^{2+}$  gets oxidized to  $\text{Fe}^{3+}$  rapidly but  $\text{Fe}^{3+}$  cannot be reduced to  $\text{Fe}^{2+}$  in time. Thus, the amount of  $\cdot\text{OH}$  generated reaches a platform with a gentle slope. As for the  $\text{CFA}_{\text{MW}}\text{--H}_2\text{O}_2$  system,  $\text{CFA}_{\text{MW}}$  can have a better catalytic persistence than  $\text{Fe}^{2+}$ , resulting in the amount of  $\cdot\text{OH}$  generated that catches up with and exceeds that of the  $\text{Fe}^{2+}\text{--H}_2\text{O}_2$  system at the later stage.

The importance of  $\cdot\text{OH}$  was compared with other free radicals (*i.e.*,  $\cdot\text{HO}_2/\text{O}_2^-$ ) in the  $\text{CFA}_{\text{MW}}\text{--H}_2\text{O}_2$  system. *n*-Butanol and benzoquinone were applied to scavenge  $\cdot\text{OH}$  and  $\cdot\text{HO}_2/\text{O}_2^-$ , respectively.<sup>29</sup> Under the optimal treatment conditions of the wastewater (75.3% of TOC removal rate), the removal rate of TOC (17.6%) has a remarkable decline in the presence of *n*-butanol, while 64.1% of TOC can still be removed in the presence of benzoquinone, revealing the  $\cdot\text{OH}$  play a primary role in the  $\text{CFA}_{\text{MW}}\text{--H}_2\text{O}_2$  system, while  $\cdot\text{HO}_2/\text{O}_2^-$  play the auxiliary role.

**Leachability and stability of  $\text{CFA}_{\text{MW}}$ .** From Table S2,<sup>†</sup> it can be inferred that the use of  $\text{CFA}_{\text{MW}}$  may introduce heavy metallic and toxic elements into treated wastewater such as Fe, Cr, Mn, Ni, and Cu. These elements may cause secondary water

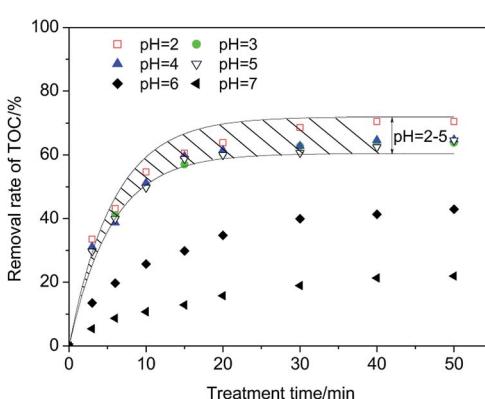


Fig. 5 Optimal range of pH in the  $\text{CFA}_{\text{MW}}$  catalyzed Fenton-like process ( $\text{H}_2\text{O}_2$  dosage = 12 mg L<sup>-1</sup>,  $\text{CFA}_{\text{MW}}$  loading = 10 g L<sup>-1</sup>, [PAM] = 200 mg L<sup>-1</sup>,  $T = 313 \text{ K}$ ).



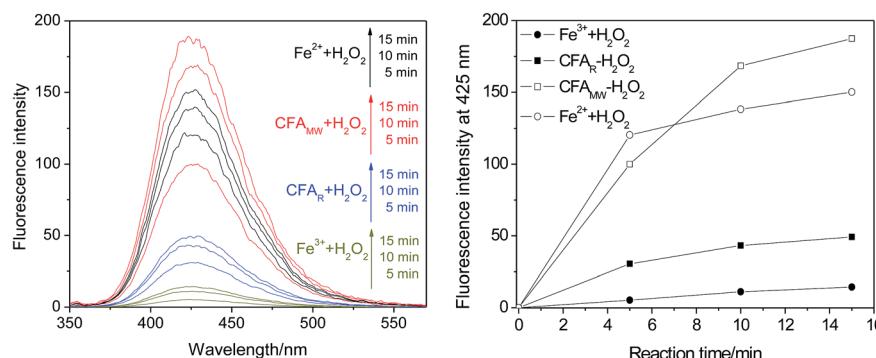


Fig. 6 Comparison of the relative generation amount of  $\cdot\text{OH}$  in different Fenton-like processes. ( $\text{H}_2\text{O}_2$  dosage =  $12 \text{ mg L}^{-1}$ ,  $[\text{PAM}] = 200 \text{ mg L}^{-1}$ ,  $T = 313 \text{ K}$ ,  $\text{pH} = 3$ . For the  $\text{CFA}_{\text{MW}}\text{-H}_2\text{O}_2$  system and  $\text{CFA}_{\text{R}}\text{-H}_2\text{O}_2$  system,  $\text{CFA}_{\text{MW}}$  loading =  $\text{CFA}_{\text{R}}$  loading =  $10 \text{ g L}^{-1}$ . For the  $\text{Fe}^{2+} + \text{H}_2\text{O}_2$  system and  $\text{Fe}^{3+}\text{-H}_2\text{O}_2$  system,  $[\text{Fe}^{2+}] = [\text{Fe}^{3+}] = 445 \text{ mg L}^{-1}$ , which is equivalent to the  $10 \text{ g L}^{-1}$   $\text{CFA}_{\text{MW}}$  loading).

pollution. Thus, the leachability of these elements from  $\text{CFA}_{\text{MW}}$  must be considered before industrial application. As shown in Table 4, nine primary elements were detected. Fortunately, the leaching amount of each element in the 1<sup>st</sup> run is lower than the Chinese standard (GB8978-1996). Meanwhile, the leaching amount of these elements always decreases in the successive five runs, revealing the possibility of direct use of  $\text{CFA}_{\text{MW}}$  in the treatment of PAM-contaminated wastewater.

While the test of the leachability can indicate the possibility of use of  $\text{CFA}_{\text{MW}}$  in the industry from the perspective of environmental protection, the test of stability indicates the economic feasibility. As shown in Fig. S1,† it can be observed that the removal rate of TOC always decreases with the increase in run time (from 63.8% to 52.8%), revealing the loss of the catalytic capacity of  $\text{CFA}_{\text{MW}}$ . This could be caused by the dissolution and loss of the active metallic elements.

This decline of catalytic capacity is common in these studies. As shown in Table S3,† all listed catalysts cannot be used continuously but have a limited run time. The comparison of these catalysts should be conducted on different aspects. The preparation condition of a catalyst is a significantly effective parameter, and a simple and cheap preparation method can decrease the cost of wastewater treatment remarkably. This is the advantage of  $\text{CFA}_{\text{MW}}$  because it was prepared by a single step physical method, *i.e.*, MW irradiation, while other catalysts were prepared by multi-step processes where calcination and chemical reagents were always used. As for the stability and the minimum treatment efficiency, they can mutually affect the efficiency of the catalyst. The catalyst may be reused more times when the minimum treatment efficiency is not limited strictly, whereas it needs to be replaced frequently in a harsher working condition (see Table S3†). Thus, it is more appropriate to compare the stability of different catalysts in identical conditions.

Table 4 Concentration ( $\text{mg L}^{-1}$ ) of the leached elements from  $\text{CFA}_{\text{MW}}$  in the Fenton-like process ( $\text{CFA}_{\text{MW}}$  loading =  $10 \text{ g L}^{-1}$ )

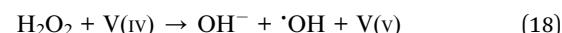
Metallic element	Fe	Al	Ti	Cr	Mn	Ni	Cu	Nb	Ca
1 <sup>st</sup> run	0.557	0.202	0.058	0.032	0.094	0.011	0.019	0.006	0.157
3 <sup>rd</sup> run	0.461	0.158	0.037	0.031	0.075	0.085	0.011	0.005	0.112
5 <sup>th</sup> run	0.409	0.126	0.021	0.023	0.068	0.059	0.007	—	0.067

### Proposed catalytic mechanism of $\text{CFA}_{\text{MW}}$

The catalytic mechanism associated with Fe element is shown in Fig. 7. It can be seen that the trivalent Fe on the surface of  $\text{CFA}_{\text{MW}}$  ( $\equiv\text{Fe}(\text{III})\text{-OH}$ ) can be reduced to divalent Fe ( $\equiv\text{Fe}(\text{II})$ ) through two paths, namely, on reacting with  $\text{H}_2\text{O}_2$  (path I) and reacting with  $\cdot\text{HO}_2/\text{O}_2^-$  (path II). Among the two paths, path I is the main one as  $\text{H}_2\text{O}_2$  has an external source (artificial input), while the generation of  $\cdot\text{HO}_2/\text{O}_2^-$  is accompanied by the consumption of  $\equiv\text{Fe}(\text{II})$  and  $(\text{H}_2\text{O}_2)_s$  (path III and IV).

Despite the mechanism in Fig. 7 that gives a detailed catalytic path, it is only associated with Fe catalysis and could only be a part of the complicated mechanism in the  $\text{CFA}_{\text{MW}}$ -catalyzed Fenton-like process. Some metallic elements existing in  $\text{CFA}_{\text{MW}}$  (besides Fe) have been reported as additives in the Fe-base Fenton-like catalysis, such as Mn, Ni, Co, Cu, Nb, and V.<sup>30-35</sup> Thus, the adjunct catalysis from these elements must be considered. In reality, these metallic elements have two adjunct catalytic paths, *i.e.*, (I) promoting the generation of active species that have oxidation capacity, and (II) promoting the transfer of electron.

As for the generation of the active species, beside the Fe element, the above listed metallic elements can generate  $\cdot\text{OH}$  *via* the Haber-Weiss path as well.<sup>36</sup> Taking the V element for instance, the catalytic process is shown in eqn (16)–(18).



As for the transfer of electron, different metallic oxides can affect each other by the oxidation-reduction reactions. This can



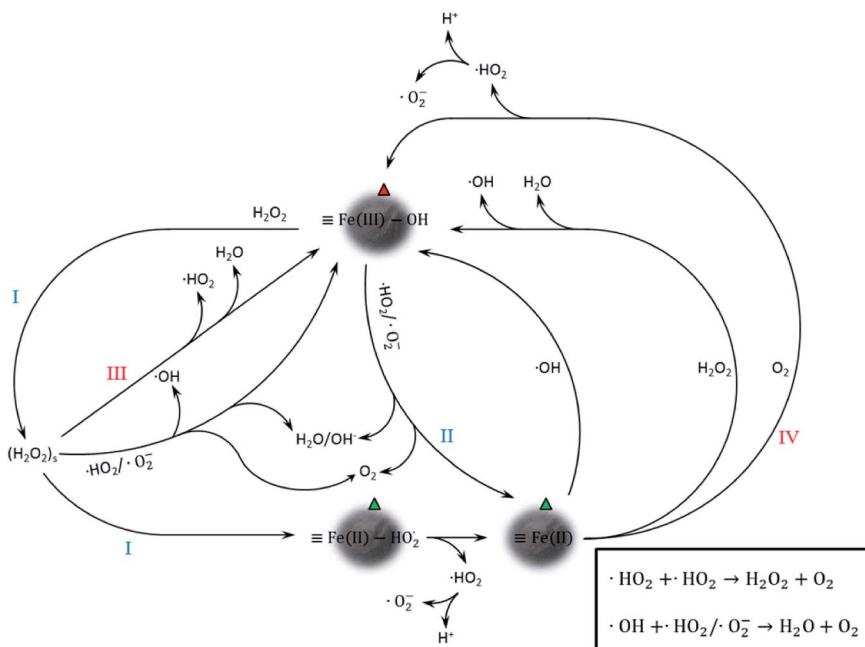


Fig. 7 Proposed catalytic mechanism of  $\text{CFA}_{\text{MW}}$  in the heterogeneous Fenton-like process.

promote the recycling of  $\text{B}^{n+x}/\text{B}^n$  ( $n \geq 1, x \geq 1$ ), which is critical for the Fenton-like process.

According to the statement above, the heterogeneous catalytic mechanism of  $\text{CFA}_{\text{MW}}$  is actually very complicated because of the mutual effect of the different metallic elements. The detailed mechanism still needs more profound study.

Finally, it has to be noted that the mechanism stated above is based on the surface catalysis, where  $\text{H}_2\text{O}_2$  must spread to the surface of  $\text{CFA}_{\text{MW}}$  first before the catalysis by the active sites. However, according to the results in the section "Investigation of ·OH in different oxidation systems", the catalysis by the dissolved metallic elements should be considered as well. Thus, the catalysis by  $\text{CFA}_{\text{MW}}$  is a multi-metal catalytic and two-phase process.

## Conclusions

$\text{MW}$  irradiation can activate the catalytic potential of the  $\text{CFA}_{\text{R}}$  effectively.

(I) The most significant factor in the activation process is the  $\text{CFA}_{\text{R}}$  loading, while the importance of the  $\text{MW}$  power, irradiation time, and mixing speed decline in this sequence. During the  $\text{MW}$  irradiation, the surface morphology of the  $\text{CFA}_{\text{R}}$  can be changed, and the specific surface area and the pore volume of  $\text{CFA}_{\text{R}}$  can be increased significantly.

(II) In the treatment of the PAM wastewater by  $\text{CFA}_{\text{MW}}$ -catalyzed Fenton-like process, more than 75% of the TOC can be eliminated under the optimal conditions ( $[\text{H}_2\text{O}_2] = 12 \text{ mg L}^{-1}$ ,  $\text{CFA}_{\text{MW}}$  loading =  $10 \text{ g L}^{-1}$ ,  $[\text{PAM}] = 200 \text{ mg L}^{-1}$ ,  $T = 313 \text{ K}$ ). The variation in the  $\text{CFA}_{\text{MW}}$  loading and PAM concentration can change the degradation path of PAM, while the increase in the  $\text{H}_2\text{O}_2$  dosage can only accelerate the degradation of PAM.

(III) The  $\text{CFA}_{\text{MW}}$  can widen the optimal pH range of the classic Fenton process from 3 to 2–5. Moreover, the  $\text{CFA}_{\text{MW}}$  has better catalytic persistence than that of  $\text{Fe}^{2+}$  but less catalytic capacity. The leaching of heavy metallic/toxic elements is lower than the limits of the GB8978-1996 standard but it still weakens the stability of  $\text{CFA}_{\text{MW}}$ . The anti-leaching study directed for  $\text{CFA}_{\text{MW}}$  could be promising from the standpoint of industrial application.

(IV) The  $\text{Fe}(\text{III})-\text{OH}$  on the surface of  $\text{CFA}_{\text{MW}}$  contributes to the primary heterogeneous catalytic capacity, while other metallic elements play a role by accelerating the generation of ·OH and the redox reaction of  $\text{Fe}(\text{III})/\text{Fe}(\text{II})$ . The homogeneous catalysis caused by the leached metallic elements cannot be overlooked as well.

## Conflicts of interest

There are no conflicts to declare.

## Acknowledgements

This work was supported by the National Natural Science Foundation of China (grant number 51808039); Science and Technology Projects of Beijing Municipal Education Commission (grant number KM201910017008); National Innovation and Entrepreneurship Training Program for students (grant number 18010282001/005).

## References

- 1 J. Anotal, N. Wasukran and N. Boonrattanakij, *Chem. Eng. J.*, 2018, **352**, 247–254.



2 N. N. Wang, T. Zheng, J. P. Jiang and P. Wang, *Chem. Eng. J.*, 2015, **260**, 386–392.

3 W. Leal, L. A. Lourenco, H. D. Brandao, A. da Silva, S. M. A. G. U. de Souza and A. A. U. de Souza, *J. Hazard. Mater.*, 2018, **359**, 96–103.

4 L. Clarizia, D. Russo, I. Di Somma, R. Marotta and R. Andreozzi, *Appl. Catal., B*, 2017, **209**, 358–371.

5 J. Zhang, S. Q. Wu, W. J. Bi, X. F. Zhao and G. D. Liu, *Mater. Lett.*, 2019, **234**, 13–16.

6 C. di Luca, P. Massa, J. M. Grau, S. G. Marchetti, R. Fenoglio and P. Haure, *Appl. Catal., B*, 2018, **237**, 110–1123.

7 T. P. H. Ngo and T. K. Le, *J. Sol-Gel Sci. Technol.*, 2018, **88**, 211–219.

8 Y. Y. Zhou, X. C. Liu, Y. Q. Zhao, S. Luo, L. L. Wang, Y. Yang, M. A. Oturan and Y. Mu, *J. Catal.*, 2018, **365**, 184–194.

9 C. Y. Su, W. G. Li, X. Z. Liu, X. F. Huang and X. D. Yu, *Front. Environ. Sci. Eng.*, 2016, **10**, 37–45.

10 S. Jauhar and S. Singhal, *Ceram. Int.*, 2014, **40**, 11845–11855.

11 J. Gao, L. Wu, S. J. Liang, S. L. Zhang, P. Liu and X. Z. Fu, *Chin. J. Catal.*, 2010, **31**, 317–321.

12 A. A. S. Tigue, R. A. J. Malenab, J. R. Dungca, D. E. C. Yu and M. A. B. Promentilla, *Minerals*, 2018, **8**, 411.

13 N. N. Wang, L. L. Hao, J. Q. Chen, Q. Zhao and H. Xu, *Environ. Sci. Pollut. Res.*, 2018, **25**, 12481–12490.

14 S. B. Yang, G. L. Song, Y. J. Na and Z. Yang, *Appl. Therm. Eng.*, 2018, **141**, 29–41.

15 O. Z. Ozdemir and S. Piskin, *Waste Biomass Valorization*, 2019, **10**, 143–154.

16 Y. B. Huang, J. S. Qian and J. Y. Zhang, *J. Build. Mater.*, 2010, **13**, 291–294.

17 S. Deng, P. Wang, G. S. Zhang and Y. Dou, *J. Hazard. Mater.*, 2016, **307**, 64–72.

18 G. C. Liu, L. Li, L. P. Qiu, S. L. Yu, P. Liu, Y. B. Zhu, J. Hu, Z. Y. Liu, D. S. Zhao and H. J. Yang, *J. Membr. Sci.*, 2018, **545**, 348–357.

19 T. Zheng, *J. Water Reuse Desalin.*, 2017, **7**, 378–386.

20 J. G. Yu, W. G. Wang, B. Cheng and B. L. Su, *J. Phys. Chem. C*, 2009, **113**, 6743–6750.

21 R. H. G. Ranil, H. M. L. Niran, M. Plazas, R. M. Fonseka, H. H. Fonseka, S. Vilanova, I. Andujar, P. Gramazio, A. Fita and J. Prohens, *Sci. Hortic.*, 2015, **193**, 174–181.

22 S. Sivalingam and S. Sen, *Appl. Surf. Sci.*, 2018, **455**, 903–910.

23 R. Kaur and D. Goyal, *J. Mater. Cycles Waste Manage.*, 2016, **18**, 186–200.

24 M. E. Kalaw, A. Culaba, H. Hinode, W. Kurniawan, S. Gallardo and M. A. Promentilla, *Materials*, 2016, **9**, 580.

25 L. M. Hu, G. S. Zhang, M. Liu, Q. Wang and P. Wang, *Chem. Eng. J.*, 2018, **338**, 300–310.

26 J. A. Giroto, A. C. S. C. Teixeira, C. A. O. Nascimento and R. Guardani, *Chem. Eng. Process.*, 2008, **47**, 2361–2369.

27 S. Li, G. S. Zhang, P. Wang, H. S. Zheng and Y. J. Zheng, *Chem. Eng. J.*, 2016, **294**, 371–379.

28 J. H. Ramirez, F. M. Duarte, F. G. Martins, C. A. Costa and L. M. Madeira, *Chem. Eng. J.*, 2009, **148**, 394–404.

29 H. Liwei, W. Liguo, R. Sébastien and Z. Hui, *J. Hazard. Mater.*, 2016, **302**, 458–467.

30 Q. Wang, X. F. Wang and B. Tian, *Water Sci. Technol.*, 2018, **77**, 2772–2780.

31 Y. J. Yao, H. Chen, C. Lian, F. Y. Wei, D. W. Zhang, G. D. Wu, B. J. Chen and S. B. Wang, *J. Hazard. Mater.*, 2016, **314**, 129–139.

32 Y. C. Wang, S. Shia, C. Wang and S. Fanga, *Environ. Prot. Eng.*, 2018, **44**, 131–145.

33 C. M. Zheng, C. W. Yang, X. Z. Cheng, S. C. Xu, Z. P. Fan, G. H. Wang, S. B. Wang, X. F. Guan and X. H. Sun, *Sep. Purif. Technol.*, 2017, **189**, 357–365.

34 Z. Jia, S. X. Liang, W. C. Zhang, W. M. Wang, C. Yang and L. C. Zhang, *J. Taiwan Inst. Chem. Eng.*, 2017, **71**, 128–136.

35 Y. Y. Zhang, J. H. Deng, C. He, S. S. Huang, S. H. Tian and Y. Xiong, *Environ. Technol.*, 2010, **31**, 145–154.

36 Z. J. Song, B. Wang, J. Yu, C. Ma, T. Chen, W. Yang, S. Liu and L. S. Sun, *Chem. Eng. J.*, 2018, **354**, 517–524.

