Nitrogen oxide removal by non-thermal plasma for marine diesel engines

Zongyu Wang, a Hailang Kuang, a Jifeng Zhang, ab Lilin Chu a and Yulong Ji c, a

The transportation industry plays an important role in the world economy. Diesel engines are still widely used as the main power generator for trucks, heavy machinery and ships. Removal technology for nitrogen oxides in diesel exhaust are of great concern. In this paper, a gas supply system for simulating the marine diesel engine exhaust is set up. An experimental study on exhaust denitration is carried out by using a dielectric barrier discharge (DBD) reactor to generate non-thermal plasma (NTP). The power efficiency and the denitration efficiency of different gas components by NTP are discussed. The exhaust gas reaction mechanism is analyzed. The application prospects of NTP are explored in the field of diesel exhaust treatment. The experimental results show that the power efficiency and energy density (ED) increase with the input voltage for this system, and the power efficiency is above 80% when the input voltage is higher than 60 V. The removal efficiency of NO is close to 100% by NTP in the NO/N2 system. For the NO/O2/N2 system, the critical oxygen concentration (COC) increases with NO concentration. The O2 concentration plays a decisive role in the denitration performance of the NTP. H2O contributes to the oxidative removal of NO, and NH3 improves the removal efficiency at low ED while slightly reducing the denitration performance at high ED. CO2 has little effect on NTP denitration performance, but as the ED increases, the generated CO gradually increases. When simulating typical diesel engine exhaust conditions, the removal efficiency increases first and then decreases with the increase of ED in the NO/O2/CO2/H2O/N2 system. After adding NH3, the removal efficiency of NOx reaches up to 40.6%. It is necessary to add reducing gas, or to combine the NTP technology with other post treatment technologies such as SCR catalysts or wet scrubbing, to further increase the NTP denitration efficiency.

1. Introduction

Today, diesel engines are still widely used in the fields of road and non-road machinery and are dominant as the main power and generation power units in the marine sector. Exhaust pollutants such as hydrocarbons (HC), particulate matter (PM), SO2 and NOx are inevitably generated during the operation of the diesel engine. These exhaust pollutants not only cause acid rain, building and soil corrosion, but also endanger human health and lead to diseases such as cancer and respiratory diseases. The SO2 and NOx emission of ocean-faring vessels has been limited by IMO’s MARPOL73/78 Annex VI.1 The removal technologies of diesel exhaust pollutants include pre-treatment, in-machine treatment and post-treatment.2,3 The post-treatment technologies that can remove a variety of diesel exhaust pollutants at the same time have gained higher attention.

Plasma, the fourth form of matter, was discovered in the mid-18th century. It can generally be classified into high temperature plasma, thermal equilibrium plasma, and non-thermal equilibrium plasma depending on the temperature of the particles. It is used in a variety of fields such as welding and cutting,7 surface modification of materials,5–7 and the removal of contaminants.8–10 At present, the research on non-thermal equilibrium plasma (NTP) is the most extensive. The methods for generating NTP mainly include electron beam method,11,12 microwave irradiation method,5–7,13–14 high voltage discharge method (including DC, AC and pulse power),15–17 etc. The method of pulse power combined with the dielectric barrier discharge (DBD) reactor to generate NTP has many advantages, such as higher power efficiency, uniform and silent discharge.18–21 So it receives more attention. NTP used for exhaust post-treatment began in the 1970s and is one of the hotspots. It is almost no secondary pollution and has a good application prospect.

There are two ways for NTP to remove NO. One is to form N2 and O2 by reduction, and the other is to form higher valence oxides by oxidation, such as NO2, N2O5, etc.12,22 Moreover, diesel exhaust contains a large amount of N2 (≈76%) and O2 (≈14%),23 and because the dissociation energy of O2 (5.2 eV mol−1) is smaller than N2 (9.8 eV mol−1),24 O2 will be converted to stronger oxidizing substance such as oxygen radical (·O) and O3. These reasons make it more unfavorable to remove NO by...
the reduction route. NO which is difficult to remove in the exhaust gas is converted into other components, whether it is reduced to N$_2$ or oxidized to high-valent nitrogen oxides, the denitrification efficiency can be increased. We mainly focus on the denitrification efficiency, and the exact percentage of NO reduction to N$_2$ will no long be discussed in the paper.

Zhang and Zhou$^{25,26}$ used pulse corona discharge plasma combined with lye absorption for desulfurization and denitrification simultaneously. They studied NO oxidation efficiency and removal efficiency by parameters such as gas flow, discharge current, NO concentration and SO$_2$ concentration. But they did not consider the effect of O$_2$ concentration on NTP NO removal efficiency. Some studies$^{24,27}$ show that with the increasing of O$_2$ concentration, the removal efficiency of NO gradually decreases, and there is so called critical oxygen concentration (COC) at which the NOx removal efficiency is zero, which means the reduction rate of converting NO$_x$ to N$_2$ and O$_2$ is equal to the oxidation rate of converting N$_2$ to NO$_x$. And critical O$_2$ concentration may change with initial NO concentration. Mok et al.$^{16}$ used pulsed corona discharge NTP to study the effect of O$_2$ concentration, humidity and peak voltage on the removal efficiency, but the NO concentration was only 210 ppm (1 ppm = 1 µL$^{-1}$) in their experiment. Zhao et al.$^{24}$ studied the removal efficiency of corona discharge NTP at different O$_2$ concentrations (0–13.6%), but the NO concentration in this study was only 437 ppm, which was much smaller than that of typical real ship diesel exhaust. What’s more, both H$_2$O and CO$_2$ are inevitable components in diesel exhaust. These components may affect the NTP denitrification reaction, while Zhao$^{24}$ and Aritoshi$^{27}$ did not consider the effect of H$_2$O on NTP denitrification performance. And most researchers have not considered the effect of CO$_2$ on NTP denitrification performance.

In addition, because the exhaust gas temperature changes with the diesel engine power, and engine exhaust gas generally contains SO$_2$ (the concentration varies with the sulfur content of the fuel used, such as the concentration of SO$_2$ in the exhaust gas is about 600 µL$^{-1}$ for marine large low-speed two-stroke diesel engine when using heavy fuel oil of 3.5% sulfur content$^{28}$), both temperature and SO$_2$ will have a certain impact on the NTP denitrification efficiency. Chmielewski et al.$^{29}$ studied the effects of different temperatures (70 °C and 90 °C) and the SO$_2$ concentration (0–2000 µL$^{-1}$) on the NTP denitrification efficiency. Their experimental results show that the plasma denitrification efficiency will increase with the temperature, so as the SO$_2$ concentration. We want to investigate the NTP denitrification efficiency in the poor cases and the application prospect of NTP. Therefore, the influence of temperature and SO$_2$ concentration on the NTP denitrification efficiency is not considered in this paper. All experiments are carried out at room temperature of 25 °C and the concentration of SO$_2$ is zero.

In summary, in order to study the effects of energy density, different initial NO, O$_2$, NH$_3$, H$_2$O and CO$_2$ on the NTP denitrification performance, a simulating diesel exhaust supply system was set up in this paper. A coaxial cylindrical DBD reactor was designed and fabricated with quartz glass. Non-thermal plasma was generated by pulsed power. NO$_x$ removal mechanism was proposed and analyzed by experimental results. The application prospects of NTP were explored in the field of diesel exhaust treatment.

2. Approach

The exhaust composition of a typical large-scale low-speed two-stroke diesel engine is shown in Table 1. NO accounts for more than 90% of NO$_x$ in the diesel exhaust, the rest mainly being NO$_2$, and NO is relatively more difficult to remove. Therefore, this paper mainly focuses on the removal of NO. The concentration of NO in the diesel exhaust will change with the engine load, and generally does not exceed 1500 µL$^{-1}$. Therefore, three NO concentrations indicating the emissions at different engine loads are studied in the paper, including the low NO concentration (500 µL$^{-1}$), the medium NO concentration (1000 µL$^{-1}$) and the high NO concentration (1500 µL$^{-1}$). When NH$_3$ is needed, it is added at an ammonia–nitrogen ratio of 1. The highest O$_2$ concentration in engine exhaust gas is 14%, and in order to study the changes of critical O$_2$ concentrations with initial NO concentrations, the final selected O$_2$ concentrations are 1%, 5%, 8%, 10% and 14%, respectively. H$_2$O can generate a variety of strong oxidizing free radicals in NTP, including HO$_2$, OH and H$_2$O$_2$. They will have a certain influence on the NTP denitrification performance. Therefore, three groups of experiments with H$_2$O concentrations of 0%, 3.5% and 5.1% are selected for comparison. Because the experimental results show that CO$_2$ has little effect on the NTP denitrification performance, 0% and 4.5% CO$_2$ concentration are selected for comparative analysis in this paper. Considering the different components in the diesel exhaust and the performance of the power source, the gas components of each experiment vary from simple to complex. Because N$_2$ accounts for the largest proportion in the exhaust gas, it is used as carrier gas in our experiments. In addition, O$_2$ accounts for the second proportion in the exhaust gas and has the fundamental effect on the NTP denitrification performance, the experiments of single component gas in NTP are carried both in NO/N$_2$ system and O$_2$/N$_2$ system. Other gases do not seem to be major players in this case and given the length of the article, they are not included in this paper so far. The following research is carried out:

(1) The denitrification performance of NTP is strongly related to the energy density in DBD, and the output efficiency of the power source is an important factor to be considered in future industrial applications. Therefore, the energy density and power supply efficiency varying with different input voltages and currents were studied firstly under a typical diesel exhaust condition, i.e., 1500 µL$^{-1}$ NO + 14% O$_2$ + 5.1% H$_2$O + 4.5% CO$_2$ + 76.2% N$_2$. Secondly, to understand the removal mechanism of the single-component NO exhaust, and the effect of the NO concentration on the power supply characteristics under a specific ideal condition, we carried out a group of experiments.

<table>
<thead>
<tr>
<th>Table 1</th>
<th>Low-speed two-stroke diesel engine exhaust composition (volume fraction)$^{23}$</th>
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<tbody>
<tr>
<td>Components</td>
<td>N$_2$ (%)</td>
</tr>
<tr>
<td>Volume fraction</td>
<td>76.2</td>
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</tbody>
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in NO/N₂ system, where initial NO concentrations were 500 µL L⁻¹, 1000 µL L⁻¹ and 1500 µL L⁻¹, respectively.

(2) Since N₂ and O₂ are the main components of diesel exhaust, O₂ will be converted to a large amount of oxygen active particles with strong oxidative in NTP system, and N₂ will also be converted to nitrogen active particles. These active particles will recombine to a certain concentration of NOₓ, which will partially offset the NOₓ removal effect of NTP. Therefore, the oxidative formation of NOₓ at different energy densities was investigated in O₂/N₂ system with the O₂ concentration of 1%, 5%, 10% and 14%, respectively.

(3) It is expected that NO is converted to N₂ and O₂ through the reduction route. And the mechanism of NO removal by reduction way needs further study. Therefore, we carried out these experiments in NO/N₂ system with the initial NO concentrations of 500 µL L⁻¹, 1000 µL L⁻¹ and 1500 µL L⁻¹, respectively.

(4) As mentioned above, there is a critical oxygen concentration (COC) that makes the NOₓ removal efficiency zero. When the O₂ concentration is higher than the COC, the NTP will not have denitration ability with the ED increasing. It even leads to the increase of the NOₓ. While when the O₂ concentration is lower than the COC, the removal efficiency increases with the ED. What's more, the COC is also related to the initial NO concentration. Therefore, we investigated the range of the COC at initial NO concentrations of 500 µL L⁻¹, 1000 µL L⁻¹ and 1500 µL L⁻¹, respectively.

(5) NH₃ is often used as reducing agent in traditional selective catalytic reduction (SCR) denitration, while the NTP denitration performance and mechanism are not well understood when NH₃ exists. To this end, the effect of NH₃ in NO/O₂/N₂ system was studied at O₂ concentrations of 1%, 5%, 8%, 10% and 14% with initial NO concentrations of 500 µL L⁻¹, 1000 µL L⁻¹ and 1500 µL L⁻¹, respectively.

(6) H₂O is an inevitable component in diesel exhaust. H₂O will generate strong oxidizing hydroxyl radical (·OH) by the action of NTP, which will further inhibit NO reduction removal. When NH₃ is added to the system, the NO removal reactions will be more complicated. In this paper, the NTP denitration performance was studied in NO/O₂/H₂O/N₂ system with initial NO concentrations of 500 µL L⁻¹, 1000 µL L⁻¹ and 1500 µL L⁻¹, O₂ concentration of 14% and H₂O concentrations of 0%, 3.5% and 5.1%, respectively. These results were also compared with that when NH₃ was added.

(7) CO₂ is the fourth major component of diesel exhaust. At present, most studies ignore the effect of CO₂ on NTP denitration performance. However, CO₂ may convert to CO in NTP, and CO has been used as the reducing agent for denitration in some reports. Therefore, in order to study the possible effects of CO₂ on NTP denitration, we added 4.5% CO₂ in NO/O₂/H₂O/N₂ system with initial NO concentrations of 500 µL L⁻¹, 1000 µL L⁻¹ and 1500 µL L⁻¹, O₂ concentration of 14% and H₂O concentration of 5.1%, respectively. Then NH₃ was added in NO/O₂/H₂O/CO₂/N₂ system to investigate the possible changes.

(8) Finally, based on the above experimental results, the NTP denitration mechanism of simulating diesel exhaust is proposed in the paper, and the application prospect of NTP denitration in the field of marine diesel exhaust is discussed.

3. Experimental system and data processing

3.1 Experimental system

The experimental system is shown in Fig. 1. It mainly includes gas supply unit, pulse plasma power unit, DBD reactor, flue gas analyzer (Testo350, Germany), and exhaust absorption device. The gas supply unit mainly includes gas cylinder, pressure reducing valve, mass flow controller (Beijing Sevenstar CS200A,
D, China), mixing chamber. The pulse plasma power unit mainly includes plasma power source (Nanjing Suman CTP-2000K, China), pulse modulator (Nanjing Suman PC-07, China), voltage regulator (Zhejiang Chint TDGC2-1, China), digital oscilloscope (Agilent MSO7104B, USA). N₂, O₂ and CO₂ are high-purity standard gas. NO is 10% standard gas with N₂ as the carrier gas, and so as NH₃. H₂O is added by the N₂ bubbling method with constant temperature water bath. The DBD reactor is made of quartz glass. The structure and dimensions are shown in Fig. 2. The left end of the reactor is inlet, and the right end is outlet and detection port. The inner diameter of the reactor body is 24 mm, and the outer diameter is 27 mm. A copper rod is inserted in the reactor body as high-voltage electrode which is 550 mm long and 15 mm in diameter. The surface of the copper rod is machined with thread which is 2 mm in pitch and 1 mm in depth. The outer surface of the reactor is coated with 60 mesh copper net connected with the low voltage electrode of the power source. The discharge space is 4.5 mm.

3.2 Experimental method
The total gas flow is kept constant at 2 L min⁻¹ with N₂ as carrier gas in all experiments. The concentrations of N₂, NO, CO₂, O₂, and NH₃ in the DBD reactor are adjusted by the mass flow controllers. The H₂O is added in by bubbling method, and its concentration is controlled by the temperature of the water bath. Before each group of experiments, the plasma power source is kept off and the initial gas concentration is adjusted to the demanded level according to the flue gas analyzer’s measurement. Then the plasma power source is turned on to generate NTP in the DBD reactor. The output power of the power source is adjusted by input voltage and input current. And the concentrations of the DBD reactor outlet gas at different powers are monitored continuously. The input voltage of the plasma power source is controlled by the voltage regulator, and the input current is controlled by the frequency adjustment knob of the plasma power source. When the input voltage is constant, the input current can adjust to the maximum value by the frequency knob. It is also the maximum input power at the voltage, and the frequency of the output sine wave is generally 6–8 kHz.

The initial output of the plasma power source is sine waves with certain frequency, as shown in Fig. 3. The sine waves with certain frequency can be modulated into pulse waves with different duty cycles and pulse frequencies by the pulse modulator, as shown in Fig. 4. The attenuation ratio K of output voltage U₀ is 1000 : 1 in Fig. 3 and 4, which means the true coordinate unit of U₀ is kV (in order to be consistent with the calculation process below, the unit of U₀ is still set as V in the Fig. 3, 4 and 6). In all these experiments, the pulse frequency and the duty cycle are kept at 200 Hz and 50%, respectively.

3.3 Data processing
The NOₓ removal efficiency is calculated according to the imported and exported NOₓ concentration of DBD recorded by
Here $P_{in}$ is the input power of the plasma power source, W; $U_{in}$ is the input voltage of the power source, V; $I_{in}$ is the input current of the power source, A.

The output power of the plasma power source is calculated by the voltage-electric charge Lissajous method.23–34 Its equivalent circuit is shown in Fig. 5. The measurement principle is that a 0.141 μF capacitor $C_M$ is connected in series with the low-voltage end of the DBD reactor. The voltage $U_M$ of the capacitor $C_M$ is measured by the digital oscilloscope. The electric charge $Q$ of the DBD discharge is equal to the capacitor $C_M$. The current flowing through the loop is calculated as:

$$I_{out} = \frac{dQ}{dt} = \frac{d(C_M U_M)}{dt} = \frac{C_M dU_M}{dt}$$

(3)

The discharge power is defined as:

$$P = \frac{1}{T} \int_0^T U_{out} I_{out} dT = \frac{C_M}{T} \int_0^T U_{out} \frac{dU_M}{dt} dT = f C_M \int U_{out} dU_M$$

(4)

The $U_M$ and $U_{out}$ are measured by the digital oscilloscope. When the input voltage is 100 V and the input current is 0.8 A, the curves of the high-voltage signal $U_{out}$ and $U_M$ are shown in Fig. 4. The waveforms during 0.0150–0.0155 seconds of Fig. 4 are shown in Fig. 3. The Lissajous figure is shown in Fig. 6 with $U_{out}$ as the X-axis and $U_M$ as the Y-axis. The Lissajous figure area A of a single pulse period is calculated by integration tool of origin. The sensitivity of the oscilloscope in the X and Y directions is $K_X$ and $K_Y$, respectively. The output power by voltage-electric charge Lissajous method can be defined as:

$$P_{out} = f K C_M K_X K_Y A$$

(5)

The power efficiency of NTP can be expressed as:

$$\eta_P = \frac{P_{out}}{P_{in}} \times 100\%$$

(6)

The output power of the plasma power source is equal to the power loaded in the DBD reactor. The energy density (ED) is usually used to represent the amount of energy that the DBD reactor acting on the flue gas. The calculation formula can be expressed as:

$$ED = \frac{60 \times P_{out}}{1000 \times Q_{exhaust}}$$

(7)

Here ED is the energy density of the DBD reactor, kJ L⁻¹; $P_{out}$ is the output power of the plasma power source, W; $Q_{exhaust}$ is the volume flow rate of the exhaust gas, L min⁻¹.

4. Results and discussion

4.1 The changes of ED and power efficiency

Gas breakdown will occur when the voltage between the two electrodes of the DBD reactor is higher than a certain value. The breakdown voltage is closely related to the discharge space. And it is also related to the gas composition and the ED in DBD
When the typical marine diesel engine exhaust is simulated, that is, 1500 µL L\(^{-1}\) NO + 14% O\(_2\) + 5.1% H\(_2\)O + 4.5% CO\(_2\) + 76.2% N\(_2\), the changes of ED and power efficiency with input voltage are shown in Fig. 7. When the input voltage is low (<20 V), the discharge is extremely unstable, and the input current is also small, so as the output power and power efficiency. As the input voltage increases (20–60 V), the DBD discharge becomes more and more stable, and the input current also increases gradually, so as the output power and power efficiency. When the input voltage is higher than 60 V, the input current is also small, so as the output power and power efficiency.

In order to improve the power efficiency, the input voltage should be higher than 60 V. If the system is used in practice, it is recommended that the input voltage be higher than 60 V. However, when the system is scaled up, the optimal efficiency zone and the corresponding ED may be different, and should be determined in actual tests.

### 4.2 Effect of NTP on single component O\(_2\)

As mentioned in the introduction, the main components in diesel exhaust are N\(_2\) and O\(_2\). They will have a great influence on NTP denitration. Therefore, we first study the effect of NTP in the N\(_2\)/O\(_2\) system. The possible reactions may include R1–R18 (refer to the end of this section) at least. N\(_2\) and O\(_2\) will be converted into a variety of active particles in NTP. Herron\(^{38}\) and Fernandez’s\(^{39}\) results show that only N(3D) and N(5S) active particles participate in the NO generation process. Zhao,\(^{44}\) Herron,\(^{37}\) Atkinson\(^{48}\) and others have shown that mainly O(3P) and O(1D) active particles participate in the NO\(_2\) reaction process. Because of the high concentration of N\(_2\) and O\(_2\), and the reaction of R5 and R6 are faster, O(1D) will be quenched very quickly, and finally only O(3P) participates in the NO\(_2\) reaction process. The formations of active particles are shown as R1–R6. N(3D) and N(5S) are no longer distinguished below, so as O(3P) and O(1D). They are denoted by ·N and ·O, respectively.

The changes of DBD outlet gas concentration with ED in O\(_2\)/N\(_2\) system are shown in Fig. 9. Fig. 9(a) and (b) show the trend of NO and NO\(_2\) with ED at different O\(_2\) concentrations, respectively. From Fig. 9(a), it can be seen that NO will not be generated at low ED regardless of the O\(_2\) concentration. However as the O\(_2\) concentration and the ED increase, the NO concentration also increases. There will be no NO generated in the ED range of 0–7.6 kJ L\(^{-1}\) at 1% O\(_2\) concentration. But when the O\(_2\) concentration is 14% and the SIE is 7.3 kJ L\(^{-1}\), the NO concentration is as high as 1018 µL L\(^{-1}\). The reactions R7 and R8 may take place in the process. From Fig. 9(b), it can be seen that NO\(_2\) is more easily generated at lower ED in the N\(_2\)/O\(_2\) system. The concentration of NO\(_2\) increases first and then decreases with the increase of O\(_2\) concentration when the ED is higher than 2.1 kJ L\(^{-1}\), indicating that the excessive O\(_2\) will inhibit the generation of NO\(_2\). When O\(_2\) concentration is 5%, it is most beneficial to the formation of NO\(_2\). The concentration of NO\(_2\) is up to 380 µL L\(^{-1}\) when the O\(_2\) concentration is 5% and the ED is 0.9 kJ L\(^{-1}\).

Based on the above results, reactions R7–R16 may take place under our experimental condition. For R16, O\(_3\) will be formed by NTP in the presence of O\(_2\). Because O\(_3\) is more oxidative than O\(_2\), NO will be oxidized to higher valence oxides. The possible reactions are mainly R16–R18.\(^{29}\) When the ED is 4 kJ L\(^{-1}\) and the O\(_2\) concentration is 14%, the NO\(_2\) concentration is 750 µL L\(^{-1}\). According to the above experimental results, the generated NO\(_2\) concentration by NTP gradually increases with the O\(_2\) concentration and ED in O\(_2\)/N\(_2\) system.

\[
e + N_2 \rightarrow e + N(4S) + N(5S) \quad \text{(R1)}
\]
4.3 Effect of NTP on single component NO

Fig. 10 shows the denitration performance of NTP in NO/N₂ system. Under this condition, NO is converted to N₂ by the reduction way, and the reactions mainly are R12 and R14. It can be seen from Fig. 10(a) that the removal efficiency slightly decreases with the increase of the NO concentration under the same ED. However, the NO is almost completely removed when the ED reaches a certain value. The removal efficiency increases with the ED when the NO concentration is constant.
removal efficiencies are all above 95% when the NO concentrations and the EDs are 500 µL L⁻¹ 0.99 kJ L⁻¹, 1000 µL L⁻¹ 1.26 kJ L⁻¹ and 1500 µL L⁻¹ 1.68 kJ L⁻¹, respectively. The reason for the above experimental result is mainly that the collisions between molecules are more intense with the increase of ED, and R14 is more likely to take place. Also, the reaction R1 and R2 will produce more ·N active particles, and then the reaction R12 will increase the removal efficiency. Sun et al. got the similar results with us when the NO concentration was 2000 µL L⁻¹. It can be seen from Fig. 10(b) that while NO is removed by the reduction way, a small part of NO is also oxidized to NO₂, which means that the reactions of R9 and R14 take place. And the higher the initial NO concentration, the more NO₂ generates. However, with the increase of the ED, the concentration of NO₂ increases first and then decreases. When the ED reaches a certain value, NO₂ is completely removed mainly due to the reactions of R10, R11 and R15. Under the experimental condition, the highest removal efficiency of 99% is obtained in this paper.

Zhao’s results show that with the increase of ED, the removal efficiency of NOₓ will reach 98.5% with no 5%, and could not be further increased. It is mainly because that N₂O which generated by the reaction of R19 is difficult to convert. When NH₃ is added, R20 may also take place, but R21 is more difficult to take place. However, the speculation cannot be confirmed in this paper due to that the flue gas analyzer we use could not measure the N₂O concentration.

Based on the above results, NO tends to reduce to N₂ by NTP in NO/N₂ system, and the removal efficiency is higher than 95%.

\[
\begin{align*}
\text{NO} + \cdot \text{N} & \rightarrow \text{N}_2\text{O} \quad \text{(R19)} \\
\text{NO}_2 + \cdot \text{NH}_2 & \rightarrow \text{N}_2\text{O} + \text{H}_2\text{O} \quad \text{(R20)} \\
\text{N}_2\text{O} + \cdot \text{O} & \rightarrow 2\text{NO} \quad \text{(R21)}
\end{align*}
\]

### 4.4 Effect of O₂ on NTP denitration performance

Tokunaga and Zhao believe that the removal efficiency of NOₓ will decrease with the increase of O₂ concentration, and there is a critical O₂ concentration to make the NOₓ removal efficiency zero. Tokunaga’s results show that the critical O₂ concentration is about 3.6% when the initial NO concentration is 500 µL L⁻¹. Zhao’s results show that the critical O₂ concentration is about 2.5% when the initial NO concentration is 350 µL L⁻¹. The NO concentration of the typical marine diesel engine exhaust is generally much higher than 500 µL L⁻¹. When the initial NO concentration is higher, the changes of the critical O₂ concentration are not very clear. Based on this, the NOₓ removal performance of NTP is studied when the initial NO concentrations, O₂ concentrations and the EDs change. Fig. 11 shows the NTP denitration performance under different conditions of NO and O₂. Since the remove rate is close to 100% at lower energy density when O₂ concentration is zero, the experiments at higher energy densities are no longer performed.

It can be seen that the higher the NO concentration, the higher the removal efficiency at the same O₂ concentration. The reason for this phenomenon is that the probability of NO collision with high-energy particles is higher in the reaction system with the increase of NO concentration. The greater the likelihood, the more NO will be removed by converting to N₂. While it may be helpful to use the isotope labeling method to detect the migration of nitrogen atoms in NO to confirm the speculation, it is not included in this study due to experimental conditions.
Secondly, it can be seen that the removal efficiency decreases with the increase of O₂ concentration when the initial NO concentration is constant. The removal efficiency is more than 90% at low O₂ concentration (less than 1%) when the ED is bigger enough. When the O₂ concentration is more than 14%, the removal efficiency is negative. And with the increase of the ED, the removal efficiency is further reduced. The critical O₂ concentration gradually increases with the initial NO concentration. The critical O₂ concentration range is 5–8% when the initial NO concentration 500 µL L⁻¹. Tokunaga [41] believes that the critical O₂ concentration is about 3.6% when the initial O₂ concentration is 500 µL L⁻¹. But our experimental results are not agreed with his. The possible reason is that the NTP generation methods are different. Tokunaga uses electron beam method and we use the DBD method. It is reported that the average free energy of electrons generated by the electron beam method is much higher than that of the DBD method. [42] Therefore, more N₂ which is difficult to excite is converted to \( \cdot \) N by the electron beam method. So the reactions of R1–R9 are easier by electron beam method even at lower O₂ concentration. And then the chemical reaction direction is toward to NO₃ generation.

The critical O₂ concentration range is 5–8%, 8–10% and 10–14% corresponding to the initial NO concentration 500 µL L⁻¹, 1000 µL L⁻¹ and 1500 µL L⁻¹, respectively. The COC gradually increases with the initial NO concentration. The removal efficiency increases with the ED when the O₂ concentration is lower than the COC, and decreases with the ED when the O₂ concentration is higher than the COC.

4.5 **Effect of NH₃ on NTP denitration performance**

NH₃ is often used as reducing agent for the removal of NOₓ in the conventional SCR method. At present, there are few studies on the NOₓ removal by NH₃ in NTP system, and the NOₓ removal mechanism by NTP + NH₃ is even more unclear. It is generally believed that the ammonia radical (·NH₃) plays a major role in the NOₓ removal in NTP system when NH₃ is added. The reactions of the ·NH₃ formation are R22–R25 (refer to the end of this section).

Fig. 12 shows the NTP denitration performance of NH₃ under different conditions of NO and O₂. Taking the initial NO concentration of 1000 µL L⁻¹ as an example, it can be seen from the comparison with Fig. 11 that when the O₂ concentration is 1% and the energy density is greater than 2.25 kJ L⁻¹, the NOₓ removal efficiency changes from 90% (Fig. 11(b) shown) to 80% (as shown in Fig. 12(b)) when NH₃ is added. When the O₂ concentration is greater than 10%, the removal efficiency remains positive at low ED after adding NH₃. But with the increase of the ED, the removal efficiency gradually decreases and eventually becomes negative. When the O₂ concentration is 14% and the ED is 7.8 kJ L⁻¹, the lowest removal efficiency changes from −49.2% (as shown in Fig. 11(b)) to −62.4% (as shown in Fig. 12(b)) after NH₃ is added. When the O₂ concentration is less than 5%, NO is removed by the reduction way and R13 is supposed to be the main reaction. The generated NO₂ is removed by the reaction R11. After the addition of NH₃, the reactions of R22–R28 mainly take place. At low O₂ concentration, the addition of NH₃ leads the removal efficiency decreasing. The experimental result may be caused by the fact that in this experimental condition the removal of NOₓ is mainly caused by the collision, which breaks the molecular bonds between high-energy particles and molecules generated by NTP. And then the molecular bonds recombine to generate bigger bond energy molecules, such as N₂. However, after the addition of NH₃, the probability of collision between the high energy...
particles and the NO molecules is lowered, thereby the removal efficiency decreases. When the O₂ concentration is greater than 10%, the NTP removal efficiency becomes negative. The reason for this phenomenon may be that the O₂ concentration is much higher than NO, and the oxygen-active particles generated by collisions between high-energy particles and O₂ molecules are much more, making the reactions R7 and R8 more likely to react. Although the NO₂ can be decreased after the addition of NH₃ at the low ED, the concentration of oxygen-activated particles increases with the ED. And the NO generation rate by oxidation reaction is higher than the removal efficiency by NH₃ reduction reaction, which eventually leads to the negative removal efficiency.

Secondly, the experimental results show that the NO₂ concentration will be slightly reduced at different O₂ concentrations when NH₃ is added. Mizuno⁴⁴ believes that NO does not react with NH₃ at room temperature, and the conversion of NO to NO₂ is determined by the concentration of O instead of the NH₃. In the range of room temperature to 150 °C, NH₃ only affects the removal efficiency of NO₂ instead of NO.

Based on the findings above, NH₃ has little effect on the critical O₂ concentration. However, when other initial conditions are constant, the removal efficiency will be greatly improved at the low energy density and further reduced at the high energy density after adding NH₃. Therefore, it should be avoided the energy density is too much when NH₃ is added, which will result in the oxidation of NH₃ and the decrease of the removal efficiency.

\[
e + \text{NH}_3 \rightarrow \cdot\text{NH}_2 + \cdot\text{H} \quad (R22)
\]
\[
\cdot\text{O} + \text{NH}_3 \rightarrow \cdot\text{NH}_2 + \cdot\text{OH} \quad (R23)
\]
\[
\cdot\text{OH} + \text{NH}_3 \rightarrow \cdot\text{NH}_2 + \text{H}_2\text{O} \quad (R24)
\]
\[
\cdot\text{H} + \text{NH}_3 \rightarrow \cdot\text{NH}_2 + \text{H}_2 \quad (R25)
\]
\[
\cdot\text{NH}_2 + \text{NO} \rightarrow \text{N}_2 + \text{H}_2\text{O} \quad (R26)
\]
\[
\cdot\text{NH}_2 + \text{NO}_2 \rightarrow \text{N}_2 + \text{H}_2\text{O} \quad (R27)
\]
\[
\text{NH}_3 + \text{NO}_2 \rightarrow \text{N}_2 + \text{H}_2\text{O} \quad (R28)
\]

### 4.6 Effect of H₂O on NTP denitration performance

H₂O is an inevitable combustion product in diesel exhaust. H₂O can generate a variety of strong oxidizing free radicals in NTP, including -HO₂, ·OH and H₂O₂. These strong oxidizing active particles will contribute to the oxidative removal of NO. The possible reactions are R29–R33 (M is a third inert body in the reaction system).⁴⁴

Fig. 13 shows the changes of the NO concentrations at the outlet when the inlet gases are 1500 µL L⁻¹ NO + 14% O₂ + 5.1% H₂O + N₂, i.e., H₂O is added. Fig. 14 shows the denitration performance of NH₃ at different H₂O concentrations in the system of 1500 µL L⁻¹ NO + 14% O₂ + H₂O + N₂, i.e., both H₂O and NH₃ are added. It can be seen from Fig. 13 that H₂O nearly has no effect on the concentration of NO₂. H₂O can reduce the NO concentration a little bit at the low ED and increase the NO concentration at the high ED, which means H₂O can increase the removal efficiency at the low ED and lower the removal efficiency at the high ED. However, there is no obvious regular pattern between H₂O concentration and NTP removal efficiency with the increase of the ED (Fig. 14). The reactions mainly consist of R29–R38. When NH₃ is added, the removal efficiency increases first and then decreases with the increase of the ED. When both NH₃ and H₂O are present, the removal efficiency of NO can be greatly improved.

Fig. 15 shows the changes of NTP removal efficiency at different initial NO concentrations in the system of 14% O₂ + 5.1% H₂O + NO + NH₃ + N₂. Whether NH₃ is added or not, the removal efficiency of NO gradually increases with the initial NO concentration. At constant initial NO concentration, the removal efficiency increases first and then decreases with the increase of the ED. This conclusion has been shown in Section 3.5. The addition of NH₃ will contribute to the increase of removal efficiency in the whole ED range in the system of 1500 µL L⁻¹ NO + 14% O₂ + 5.1% H₂O + NH₃ + N₂,
and the removal efficiency is up to 40.6% when the ED is 1.57 kJ L\(^{-1}\).

In general, H\(_2\)O can increase the NTP removal efficiency at the low ED and reduce the removal efficiency at the high ED. When H\(_2\)O and NH\(_3\) coexist, the NTP removal efficiency is greatly improved. Therefore, the system consists of the following reactions:

\[ e + H_2O \rightarrow e + H + \cdot OH \] (R29)
\[ O(\text{i}) + H_2O \rightarrow \cdot OH + \cdot OH \] (R30)
\[ \cdot OH + \cdot OH + M \rightarrow H_2O_2 + M \] (R31)
\[ O(\text{i}) + \cdot OH \rightarrow O_2 + \cdot H \] (R32)
\[ O_2 + \cdot H \rightarrow \cdot HO_2 \] (R33)
\[ NO + 2 \cdot OH \rightarrow NO_2 + H_2O \] (R34)
\[ NO + \cdot HO_2 \rightarrow NO_2 + \cdot OH \] (R35)
\[ NO + \cdot OH \rightarrow HNO_2 \] (R36)
\[ HNO_2 + \cdot OH \rightarrow NO_2 + H_2O \] (R37)
\[ NO_2 + \cdot OH \rightarrow HNO_3 \] (R38)

4.7 Effect of CO\(_2\) on NTP denitration performance

CO\(_2\) is also an inevitable combustion product in diesel exhaust, and most researchers have not explored the effects of CO\(_2\) on the NTP denitration. Therefore, in this section we investigate the impact of CO\(_2\) in the NTP system for different initial NO concentrations and the presence of NH\(_3\).

The effects of 4.5% CO\(_2\) on the DBD outlet gas concentration in the system of 1500 \(\mu\)L L\(^{-1}\) NO + 14% O\(_2\) + 5.1% H\(_2\)O + N\(_2\) are shown in Fig. 16(a). The effect of 4.5% CO\(_2\) on the NTP removal efficiency at different initial NO concentration in 14% O\(_2\) + 5.1% H\(_2\)O + 76.2% N\(_2\) + NO is shown in Fig. 16(b). It can be seen from Fig. 15(a) that CO\(_2\) almost has no effect on the concentration of NO and NO\(_2\) at the DBD outlet. But when 4.5% CO\(_2\) is added, a certain concentration of CO will be produced, and the outlet CO concentration will gradually increase with the ED. When the ED is 7.64 kJ L\(^{-1}\), the CO concentration is up to 1920 \(\mu\)L L\(^{-1}\). The possible reactions are R39–R43. And R40 is the main reaction because CO\(_2\) concentration is much higher than the CO. It can be seen from Fig. 16(b) that CO\(_2\) basically does not have any influence on the NTP removal efficiency at different initial NO concentrations.

The changes of NTP removal efficiency at different initial NO concentrations are shown in Fig. 17 when NH\(_3\) is added in the system of 14% O\(_2\) + 5.1% H\(_2\)O + 4.5% CO\(_2\) + 76.2% N\(_2\) + NO. The CO concentration in the outlet increases with the ED instead of the initial NO concentration. The NTP removal efficiency increases significantly when NH\(_3\) is added. However, CO\(_2\) has no effect on the NTP removal efficiency in the presence of NH\(_3\) when compared with Fig. 15. In addition, we opened the DBD reactor after the experiment of adding CO\(_2\), and a layer of black carbon was observed on the surface of high-pressure copper. 

Fig. 15 Effect of H\(_2\)O on NTP denitration performance at different initial NO concentrations.

Fig. 16 Effects of CO\(_2\) on the NTP denitration performance, (a) effects of CO\(_2\) on DBD outlet gas concentration in the system of 1500 \(\mu\)L L\(^{-1}\) NO + 14% O\(_2\) + 5.1% H\(_2\)O + 76.2% N\(_2\). (b) Effects of CO\(_2\) on the NTP removal efficiency at different initial NO concentration in 14% O\(_2\) + 5.1% H\(_2\)O + 76.2% N\(_2\) + NO.
rod. This may be due to the further decomposition of CO, which means reaction R44 occurred.

In summary, CO2 nearly has no effect on the removal efficiency of NOx under different experimental conditions. But CO will be detected in the outlet and black carbon will be generated on the surface of high-voltage electrode after adding CO2, and the CO concentration will gradually increase with the ED.

\[
\begin{align*}
e + \text{CO}_2 & \rightarrow e + \cdot \text{O} + \text{CO} \quad \text{(R39)} \\
\text{CO} + \cdot \text{O} & \rightarrow \text{CO}_2 \quad \text{(R40)} \\
\text{CO} + 2 \cdot \text{OH} & \rightarrow \text{CO}_2 + \text{H}_2\text{O} \quad \text{(R41)} \\
\text{CO} + \cdot \text{HO}_2 & \rightarrow \text{CO}_2 + \cdot \text{OH} \quad \text{(R42)} \\
3\text{CO} + \text{O}_3 & \rightarrow 3\text{CO}_2 \quad \text{(R43)} \\
e + \text{CO} & \rightarrow e + \cdot \text{O} + C \\ 
\end{align*}
\]

5. Reaction mechanism of NOx and application prospect of NTP

5.1 Reaction mechanism of NOx in NTP

When the inlet gases are N2, O2, CO2, H2O, NO and NH3, NTP denitration can be roughly divided into three stages.16,22,40,45

(1) The first stage is the discharge. At this stage gas molecules are mainly bombarded by high-energy electrons. It breaks the molecular covalent bonds, changes the gas molecules into free radicals and excites some decomposed atoms to the unstable excited state. The following reactions mainly take place at this stage: R1–R4, R14 and R15, R22, R29, R39 and R44.

(2) The second stage is the post-discharge. At this stage the excited-state atoms generated in the first stage collide with the gas molecules to generate secondary radicals. Then the radicals collide with other particles leading to quench or new radicals' generation. The reactions are mainly R5 and R6, R16, R23–R25, R30–R33.

(3) The third stage is that the free radicals react with NOx. The reactions mainly include R7–R13, R17–R21, R26–R28, R34–R38. The reactions of free radicals with CO mainly include R40–R43.

In general, under our experimental conditions, the reaction mechanism of diesel engine exhaust in the NTP system is shown in Fig. 18. It mainly consists of two parts. The first is that gas molecules generate various free radicals under the bombardment of high-energy electrons. The second is the radicals react with other particles. The reaction process of NOx with free radicals is described in the right of Fig. 18. Other researchers16,22,40,45 get the similar results with us. While our experimental results show that CO2 has little effect on the removal efficiency of NOx. CO2 is converted to CO and even further converted to black carbon. Therefore, the reaction process of CO2 in NTP is separately listed in the left of Fig. 18.

5.2 Analysis of NTP application prospect

As NO can be removed by oxidation or reduction, when the O2 concentration is low, NO is mainly removed by the reduction route. When the O2 concentration is high, many strong oxidative constituents are generated such as ·OH, ·O and O3. And then the NO is oxidized to NO2. But N2 will be converted to NOx at high O2 concentration and high energy density, causing the NTP removal efficiency to become negative. Although it is beneficial to the NTP removal efficiency improvement when the initial NO concentration is higher, it is difficult to further improve removal efficiency only by NTP technology because of
the high O₂ concentration in the marine diesel exhaust. In this paper, the highest removal efficiency is only 8.6% at 0.8 kJ L⁻¹ energy density when inlet gases are 1500 µL L⁻¹ NO + 14% O₂ + 5.1% H₂O + 4.5% CO₂ + 76.2% N₂. Chmielewski²⁵ simulated the diesel engine exhaust conditions when burning heavy oil, and the removal efficiency obtained was also very low. Therefore, in order to achieve high-efficiency NO₃ removal at high O₂ concentration, it is necessary to combine NTP denitration technology with other methods. The possible methods include the following:

(1) Adding reducing gas, such as H₂, NH₃ (ref. 11, 21 and 47–49) and HC₅₉–₅₂ (including CH₄, C₂H₂, C₂H₄, C₃H₆, etc.). In this paper, the highest removal efficiency of 40.6% is obtained at the energy density of 1.65 kJ L⁻¹ when NH₃ is added.

(2) Combining NTP with the denitration catalysts.¹⁵,²¹,⁴⁹,⁵¹,⁵₃–⁵₆ The catalysts commonly used include molecular sieves, activated carbon and metal oxides. For example, we can combine NTP with traditional vanadium-based SCR catalyst. Because NTP can convert NO to NO₂, which means increasing NO₂/NO ratio, it is beneficial to increase the reaction rate of SCR.²⁷ The removal efficiency is generally above 90% in this way, but it is also necessary to add reducing gas.

(3) Combining NTP with wet scrubbing technology.⁵₈–⁶⁰ This technology has matured on ships today. NTP can oxidize NO to NO₂, which means increasing NO₂/NO ratio. It is beneficial to increase the reaction rate of SCR. Therefore, the O₂ concentration plays a decisive role in NTP denitration performance. And the critical O₂ concentration will be above 90%. And NO is mainly removed by the reduction route. When low concentration O₂ exists, NTP removal efficiency at the high O₂ concentration above 90%. And NO is mainly removed by the reduction route. When high O₂ concentration, the removal efficiency becomes negative. And the higher the ED, the more NO₂ will generate. Therefore, the O₂ concentration plays a decisive role in NTP denitration performance. And the critical O₂ concentration increases with the initial NO concentration.

(4) Under typical diesel engine exhaust condition, H₂O has little effect on NO when NH₃ is not added, but it can increase NO removal efficiency at the low ED and the excessive ED causes the increase of NO at the outlet. The highest removal efficiency of 4.5% can be obtained in the system of 1500 µL L⁻¹ NO + 14% O₂ + 5.1% H₂O + N₂. The highest removal efficiency of 18.9% can be obtained in the system of 1500 µL L⁻¹ NO + 1500 µL L⁻¹ NH₃ + 14% O₂ + N₂. However, when H₂O and NH₃ are added, the removal efficiency is up to 40.6% in the system of 1500 µL L⁻¹ NO + 1500 µL L⁻¹ NH₃ + 5.1% H₂O +14% O₂ + N₂ because of the synergistic effect.

(5) CO₂ nearly has no effect on the removal efficiency of NO, but the concentration of CO will increase gradually with the ED. When the ED is 7.64 kJ L⁻¹, the CO concentration will up to 1920 µL L⁻¹. The reaction process of CO₂ is added in the reaction mechanism of diesel engine exhaust in NTP system.

(6) Because of the high concentration of O₂ in the marine diesel engine exhaust, in order to further improve the removal efficiency of NO, it is necessary to add reducing gas to the NTP reactor, or combine the NTP technology with SCR catalysts or other technologies such as wet scrubbing.

6. Conclusions

A non-thermal plasma denitration system based on simulated diesel engine exhaust was set up in this paper. The NTP was generated by dielectric barrier discharge reactor. The NO removal performance by NTP under different O₂, H₂O, CO₂, NO, NH₃ concentrations and energy densities conditions were studied. The reaction mechanism of NOx in NTP system was proposed. The application prospect of NTP technology was analyzed. In the end we get the following conclusions:

(1) For the experimental system, the power source efficiency gradually increases with the input voltage. When the input voltage is greater than 60 V, the power supply efficiency is basically maintained above 80%.

(2) NO concentration increases gradually with the O₂ concentration and the ED in the N₂/O₂ system. When the O₂ concentration is 14% and the ED is 7.3 kJ L⁻¹, the concentration of NO is up to 1018 µL L⁻¹. The amount of NO₂ increases and then stabilizes with the ED. When the O₂ concentration is 5% and the ED is 0.9 kJ L⁻¹, the concentration of NO₂ is up to 380 µL L⁻¹. Therefore, a certain concentration of NOₓ will be generated when introducing NTP in the air.

(3) NTP has a high removal efficiency for NO/N₂ system. When low concentration O₂ exists, NTP removal efficiency is above 90%. And NO is mainly removed by the reduction route. NTP will have no denitration performance at high concentration of O₂. However, NH₃ can inhibit the formation of NO₂ and improve the removal efficiency. At low O₂ concentration, the removal efficiency of NOx gradually increases with the ED; at high O₂ concentration, the removal efficiency becomes negative. And the higher the ED, the more NO₂ will generate. Therefore, the O₂ concentration plays a decisive role in NTP denitration performance. And the critical O₂ concentration increases with the initial NO concentration.

Conflicts of interest

There are no conflicts to declare.

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Notes and references


