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Characteristics of a CaSO_4 composite oxygen carrier supported with an active material for *in situ* gasification chemical looping combustion of coal†

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CaSO_4 is considered to be a potential oxygen carrier for chemical-looping combustion (CLC) due to its cheapness and high oxygen transport capacity. To improve the physicochemical stability of the CaSO_4 oxygen carrier, CaSO_4 composite oxygen carriers supported with clay, cement, and ash separately were prepared. It was found that the attrition resistance of the CaSO_4 oxygen carrier composed of clay and cement improved due to the bond action of clay and cement. The reactivity of the composite oxygen carrier with coal was investigated in a thermogravimetric analyser (TGA) and fluidised bed. Sulphurous gas products were analysed by mass spectrometry (TG-MS) and gas chromatography (GC). Based on the catalysis of the active components in clay, cement and ash, the reaction rate of CaSO_4 with coal was improved by the active materials. However, the side reaction generating the sulphurous gas was severe in both the reduction and oxidation stages, especially when using steam as the gasifying agent. To enhance the regeneration, the CaSO_4 /clay composite oxygen carrier was upgraded by adding CaO . It was demonstrated that SO_2 release can be restrained in both the reduction and oxidation stages when the mass ratio of CaO to the CaSO_4 /clay composite oxygen carrier was higher than 1. At this point, the corresponding oxygen transport capacity was about 14.1 wt%.

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1. Introduction

It has been acknowledged that the emission of greenhouse gas (CO_2), mainly caused by extensive utilization of fossil fuels, brings about global warming and a rise in sea level. To minimize CO_2 emission in a short-medium term, CO_2 capture and storage (CCS) technologies are practical options. Because of the inherent separation of CO_2 without the need for extra energy, chemical-looping combustion (CLC) was identified as a promising technology for fossil fuel combustion.^{1,2} Oxygen carrier particles were used to divide the traditional combustion reaction into two-step reactions occurring separately in two interconnected reactors, which can avoid direct contact between air and fuel.

CLC can be applied for a variety of fuels including gaseous fuels, liquid fuels,^{3,4} and even solid fuels such as coal and biomass.^{2,5} CLC of coal is undergoing a high degree of development, since coal is expected to continue being one of the main energy sources in the medium term.⁶ To direct the

chemical looping combustion of coal, *in situ* gasification chemical looping combustion (iG-CLC) has been proposed.^{1,2} Compared with CLC of gaseous fuels, iG-CLC of coal faces two challenges. On the one hand, the char gasification rate is the limiting step of reactions in the fuel reactor, which greatly determines the CO_2 capture efficiency. To maximize the conversion of char in the fuel reactor, high temperature and a highly reactive oxygen carrier have been suggested.⁷ On the other hand, the effect of ash on the reactivity of oxygen carrier and its separation from the oxygen carrier have yet to be further explored. Oxygen carrier particles are prone to deactivation because of particle sintering and pore blocking by fine ash.⁸ Additionally, it is inevitable that some oxygen carrier particles will be extracted with the coal ashes when they are removed from the system. Ultimately, both challenges depend on the preparation of an appropriate oxygen carrier.

An alternative route, the so-called chemical looping with oxygen uncoupling (CLOU), has been proposed to improve iG-CLC performance.⁹ Oxygen carrier materials, such as synthetic materials of CuO and Mn_2O_3 , with the capability to release gaseous oxygen in the fuel reactor have been used in CLOU. Excellent performance regarding combustion efficiency has been shown by these oxygen carriers.^{10,11}

To realize the CLC process, it is important to find an inexpensive oxygen carrier available in large quantities and with sufficient oxygen transport capacity. The oxygen carrier should possess the ability to maintain excellent reactivity during many

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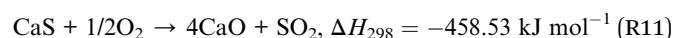
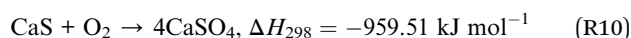
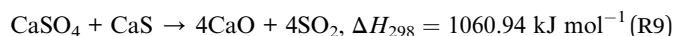
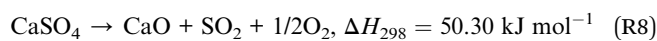
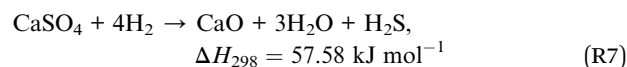
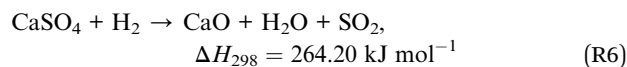
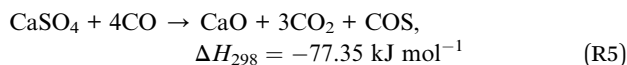
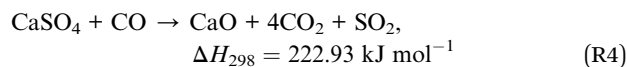
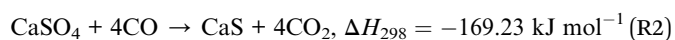
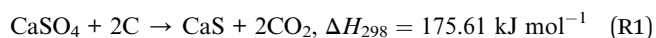
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reduction/oxidation cycles and withstand mechanical handling without severe physical degradation or loss in performance.¹² With respect to CLC of gaseous fuels, metal oxides such as Ni-based, Cu-based, Co-based, Fe-based, and Mn-based metal oxides and their metal blends have been regarded as oxygen carrier candidates and thus, they have been widely investigated. In the case of coal CLC, the inexpensive and widely available natural ore or by-products from industrial processes are promising options. Up to now, most studies have focused on iron-based materials from different sources such as ilmenite.^{13,14} In these cases, complete combustion has not been achieved, and unconverted products such as H₂, CO and CH₄ exit the fuel reactor with the CO₂ stream. Furthermore, complete combustion has still not been achieved using more reactive materials such as iron ore or the waste Fe-enriched sand fraction from alumina production.^{15–17}

In addition to these materials, CaSO₄ is deemed to be a promising oxygen carrier due to its abundance, low cost and high oxygen transport capacity; an oxygen transport capacity of 47 wt% has been reported, which is higher than those of any other proposed materials.¹⁸ The reactivity and kinetics of CaSO₄ as the oxygen carrier in CLC for both gases (CO and CH₄) and coal have been extensively investigated.^{19–24} Moreover, the feasibility and characteristics of the CaSO₄ oxygen carrier for coal chemical looping gasification (CLG) have also been demonstrated.^{25–27} Based on the desulphurization of limestone in a circulating fluidised bed combustion boiler, Alstom proposed a limestone chemical looping combustion process (LCL-C) in which CaCO₃ was continuously fed together with coal to produce CaSO₄ as an oxygen carrier *via* sulphur retention.²⁸ Subsequently, Alstom demonstrated good performance of LCL-C regarding CO₂ capture in a 3 MWth CLC unit.²⁹ However, there is no detailed information available in the literature regarding the LCL-C process. Recently, Abad *et al.*³⁰ verified the LCL-C process with CO₂ capture and *in situ* desulphurization using mass and energy balances, and they concluded that the oxygen transport capacity of sulphated limestone was about 16.7 wt%, corresponding to a fraction of 23 wt% CaSO₄. However, the CaSO₄ oxygen carrier formed *via* continuous sulphur retention of CaCO₃ can intensify the effect of coal ash on the oxygen carrier and reduce the lifespan of the oxygen carrier.

The main chemical reactions involved in the iG-CLC process using CaSO₄ as the oxygen carrier are listed in (R1)–(R11). In the fuel reactor, the reactions (R1)–(R3) are the desired reactions, whereas (R4)–(R8) are the main side reactions generating sulphurous gas, which diminish the oxygen transport capacity of CaSO₄. In the air reactor, (R10) is our desired reaction, whereas (R9) and (R11) are the main side reactions generating sulphurous gas.



The main drawback of the CaSO₄ oxygen carrier is its low reactivity compared to other materials and the release of SO₂ at temperatures of interest for CLC. Temperature greatly affects side reactions.^{18–22} Song *et al.*¹⁹ observed that CaS was the dominant compound at 900 °C, whereas the presence of CaO was more relevant at 950 °C in the fuel reactor. Shen *et al.*³¹ proposed 900–950 °C and 1050–1150 °C as the suitable ranges of temperatures for the fuel and air reactors, respectively. The CaSO₄–CaS–CaO phase diagram predicts that CaS is stable at the high partial pressure of sulphurous gas and reducing gas H₂/CO using gaseous fuel.^{32,33} Tian *et al.*³³ observed that the solid product was almost pure CaS when the partial pressure of CO or H₂ was higher than 40 kPa. However, the side reactions generating sulphurous gas are unavoidable in the fuel reactor due to the aim of minimizing the unconverted products such as H₂, CO and CH₄. Especially, ashes and a complex atmosphere with coal as the fuel exert tremendous influence on the side reactions emitting sulphurous gas. In addition, sulphur evolution is also a problem in the air reactor. The evolution of SO₂ can also occur *via* undesired reactions (R9) and (R11).

To improve the reactivity of CaSO₄ and avoid the generation of SO₂ from CaSO₄ in both reactors, several alternatives have been proposed. Some studies have reported that the addition of Fe₂O₃ to CaSO₄ by different preparation methods^{34–37} may help inhibit undesired side reactions and improve the reactivity of CaSO₄. Other authors have suggested introducing CaO or CaCO₃ together with CaSO₄ (ref. 33) to capture SO₂ released during CaSO₄ decomposition and inhibit the aggregation of CaSO₄ oxygen carrier particles at high temperature. Other studies have reported that the addition of K₂CO₃ *via* the impregnation method can improve the reactivity between the oxygen carrier and coal.^{38,39}

In open literature, the investigated CaSO₄ oxygen carriers are mostly pure CaSO₄, anhydrite ore having high CaSO₄ content (>94 wt%) or CaSO₄ modified by adding a small amount of active metal oxide.^{34–39} However, the conversion between the oxidized state (CaSO₄) and the reduced state (CaS) may significantly change the structure of the oxygen carrier due to its high



oxygen transport capacity. On the other hand, a large amount of heat generated by the oxide reaction of CaS can result in an increase in the localized temperature of the particles and eventually sintering of the oxygen carrier. As a consequence, the physicochemical stability of the CaSO₄ oxygen carrier can be degraded. The addition of bulk support materials such as Al₂O₃, SiO₂, and TiO₂ into the oxygen carrier can be an alternative approach to improve its physicochemical stability. Besides, the attrition resistance of the CaSO₄ oxygen carrier has been seldom investigated. Cabello *et al.*⁴⁰ evaluated the attrition resistance of 23 oxygen carriers and concluded that the oxygen carrier with a crushing strength higher than 1 N and an Air Jet Index (AJI) lower than 5% is eligible for the CLC process.

To improve the physicochemical stability and to avoid the release of SO₂ from the CaSO₄ oxygen carrier, three active support materials, *i.e.*, clay, cement and ash were employed in this study. The attrition resistance and the reactivity between the CaSO₄ composite oxygen carrier and coal were evaluated. Meanwhile, the generation of sulphurous gas during reduction and oxidation was analysed by TG-MS. Finally, the improvement in the CaSO₄/clay composite oxygen carrier through the addition of CaO was investigated.

2. Experimental section

2.1 Materials and characterization

The chemical composition analysis by XRF of the used clay, fly ash and cement is summarized in Table 1. These active support materials mainly contained inert supports such as Al₂O₃, SiO₂, catalysts for coal gasification, such as potassium and sodium salts, and the sulphur adsorption agent CaO. SiO₂ was selected for comparison. It was demonstrated that the main component of clay and fly ash was SiO₂ and that of cement was CaO. According to the pre-experiment, the mass ratio of CaSO₄ and support materials (clay, fly ash, cement and SiO₂) was 3 : 2. The corresponding oxygen transport capacity was about 28.2 wt%. The composite oxygen carriers were prepared by the mechanical mixing method. First, CaSO₄ (ACS reagent) and the support material were obtained by passing through a 200-mesh sieve separately. CaSO₄ and each support material having the desired mass ratios were placed in the beaker and stirred for 30 min. Then, enough deionized water was added into the beaker until the mixed sample was immersed in water. Subsequently, the mixed sample was stirred for 20 min and dried overnight. The dried samples were calcined at 800 °C for 4 hours in a muffle furnace. Then, for further use, the calcined oxygen carrier was crushed and sieved to obtain the determined particle size, *i.e.*,

98–180 μm. The proximate and ultimate analysis of the used coal is displayed in Table 2.

The microstructures of the oxygen carrier particles were observed by SEM (scanning electron microscope, JEOL JSM-6300). The element distribution in the solid particles was analysed by EDX (Energy-Dispersive X-ray Microanalyser, JEOL JED-2100). The crystalline phase of the fresh oxygen carrier was analysed using an X-ray diffractometer (Rigaku D/MAX-2500, Japan), with 2θ ranging from 10° to 90° with a step of 0.02° s⁻¹. All X-ray diffraction (XRD) patterns were analysed using Jade 6.0 of Material Data, Inc. (MDI).

The reactivities between coal and composite oxygen carriers were studied using a thermogravimetric analyser (TGA, Netzsch STA 409 PC, Germany). The desired mass ratio of the composite oxygen carrier to coal was determined by the fixed carbon in coal and by the transfer of oxygen in the composite oxygen carriers, according to the reaction $\text{CaSO}_4 + 2\text{C} \rightarrow \text{CaS} + \text{CO}_2$. To start the experiment using TGA, about 12 mg of the desired mass ratio mixture was put into the crucible and heated directly from ambient temperature to 900 °C at a heating rate of 30 °C min⁻¹ with N₂ as a purge gas, with or without H₂O as a gasifying gas. The flow rates of N₂ were determined to be approximately 20 ml min⁻¹, whereas that of H₂O was determined by its saturated vapor pressure in N₂ at 30 °C for TGA. The temperature was maintained at 900 °C for 40 min and then elevated to 1000 °C. At the same time, the purge gas was switched to O₂ to oxidize the oxygen carrier for 20 min.

To investigate the release behavior of sulphurous gas during the redox reaction, the sulphurous gas was analysed using TG-MS (STA449F1-QMS 403 D, Netzsch, Germany). To protect the device, the starting temperature was set as 150 °C with a heating rate of 15 °C min⁻¹, and the oxidation stage was conducted at 950 °C. The flow rate of water was determined to be 0.2 g min⁻¹. The other conditions were the same as the conditions of the former TGA analysis (Netzsch STA 409 PC, Germany).

2.2 Experimental setup and procedure

The schematic diagram of the fluidised-bed setup used in the experiments is presented in Fig. 1. The CLC system consisted of a high-temperature fluidised bed (stainless steel, Φ50 × 650 mm), temperature control unit, gas feeding system (steam generator, gas cylinders, and gas flow meter), gas analysis system, and data acquisition system. The reactor was electrically heated with a furnace. One K-type thermocouple was fixed in the section between the tube and the heater, and the other was fixed inside the tube.

Table 1 Component analysis of clay, fly ash and cement

materials	SiO ₂ , wt%	Al ₂ O ₃ , wt%	Fe ₂ O ₃ , wt%	CaO, wt%	Others, wt%
Clay	57.98	7.05	4.57	8.52	21.88
Fly ash	57.61	9.86	6.19	17.8	8.54
Cement	20.93	4.55	1.05	64.34	9.13



Table 2 Proximate and ultimate analysis of used coal

Proximate analysis/wt%, ad				
M	V	A	FC	
8.30	29.46	10.21	52.03	
Ultimate analysis/wt%, ad				
C	H	O	N	S
65.00	3.83	11.38	0.88	0.4

To set up the experiment, a mixture of coal (4 g) and composite oxygen carrier (50 g) particles was fed into the fluidised bed in advance. The mass ratio was determined to keep the oxygen carrier in excess, which has been used in previously reported methods. During the reduction process, argon (0.4 L min^{-1}) and steam (0.5 g min^{-1}) were used as the purge gas and gasifying agent, respectively. After the temperature reached the given reduction temperature (900°C), steam was introduced from the bottom of the reactor. The reduction reaction lasted for 40 min, during which coal was mostly converted into CO_2 and water vapour. Gaseous products were sampled by employing gas bags every 2 min for analysing by gas chromatograph. After the reduction reaction was complete, the inlet gas was switched to air (0.2 L min^{-1}) to oxidize the oxygen carrier for 1 hour. Subsequently, oxygen carrier particles were cooled to ambient temperature in an argon atmosphere and then collected for further analysis.

The gas in the gas bag was analysed directly by a gas chromatograph analyser (PE Clarus 500) using a TDX-01 packed column/thermal conductivity detector for syngas and a Propark QS packed column/flame photometric detector for sulphurous gas.

2.3 Data evaluation

(1) The gas concentration, C_i , is the ratio of the volume of the exit gases ($i = \text{CO}_2, \text{CO}, \text{CH}_4, \text{H}_2, \text{SO}_2$) to total volume and is given by the following formula:

$$C_i = \frac{P_i}{\sum P_i}$$

(2) The carbon conversion efficiency, X_c , is the ratio of the carbon mole of carbonaceous gases (CO, CO_2 , and CH_4) and the

total carbon mass contained in the coal, and it is calculated as follows:

$$X_c = \frac{12(P_{\text{CO}_2} + P_{\text{CO}} + P_{\text{CH}_4})}{22.4W_{\text{total}}}$$

(3) The CO_2 generation rate, R_{CO_2} , is calculated as follows:

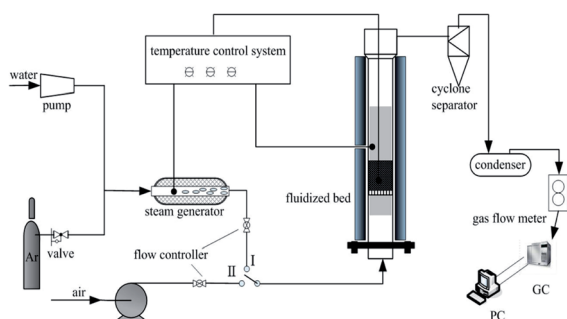
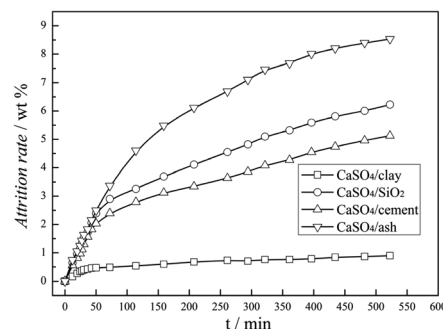
$$R_{\text{CO}_2} = \frac{VC_{\text{CO}_2}(t)}{\Delta t}$$

Here, V is the volume of generated gas in the time interval Δt .

3. Results and discussion

3.1 Attrition performance

The attrition resistance of three kinds of fresh composite oxygen carriers was determined using an attrition test system, configured according to the ASTM-D-5757 standard. As specified in the ASTM method, fifty grams particles was used for each sample tested. The samples were tested under 10 L min^{-1} of air flow, and the weight loss of fines was recorded at 1 and 5 h of time on stream. Particles with a size lower than $20 \mu\text{m}$ were considered as fines. The attrition rate was defined as the mass ratio of the accumulated fine particles to the total addition of oxygen carrier particles. As illustrated in Fig. 2, the attrition rate curves of three types of composite oxygen carriers *versus* time can be characterized by two stages: the rapid loss stage and the slow loss stage. The rapid loss stage was named as the fragmentation stage, which was mainly due to the surface asperities of the oxygen carrier particles, whereas the slow loss stage was named the abrasion stage, which reflected the attrition behaviour of the oxygen carriers for long-time operation in a fluidised bed. After a short time of the fragmentation stage, within 50 min, three kinds of composite oxygen carriers underwent the abrasion stage. It was demonstrated that the attrition rates of the $\text{CaSO}_4/\text{clay}$ and $\text{CaSO}_4/\text{cement}$ composite oxygen carriers were 1% and 5% within 500 min, respectively; thus they possessed excellent attrition resistance compared to $\text{CaSO}_4/\text{SiO}_2$. This resulted from the good bond action of cement and clay. On the contrary, the attrition rate of the CaSO_4/ash composite oxygen carrier was 8.5% within 500 min, showing high risk of particle integrity failure. However, it should be stressed that further study in the continuous CLC pilot plant is

**Fig. 1** Schematic diagram of the experimental fluidised-bed setup.**Fig. 2** Attrition rate of CaSO_4 composite oxygen carrier *versus* time.

necessary to determine whether the attrition resistance of the oxygen carrier is suitable for the CLC process.

3.2 Characterization of fresh composite oxygen carriers

XRD patterns of three kinds of fresh CaSO_4 composite oxygen carriers are shown in Fig. 3, which clearly demonstrates that the main crystalline phase in the fresh oxygen carrier was CaSO_4 . The main component SiO_2 in the clay and fly ash was observed in both the CaSO_4 /clay and CaSO_4 /ash composite oxygen carriers, whereas $\text{Ca}(\text{OH})_2$ was observed in the CaSO_4 /cement oxygen carrier. This was ascribed to the adsorption of water steam by CaO in the cement. Except for CaSiO_3 in the CaSO_4 /ash composite oxygen carrier, no new phases were observed after calcinations; this indicated that CaSO_4 did not significantly react with any component of clay, cement and ash. Al_2O_3 and Fe_2O_3 were not observed even though Fe and Al were present in clay, cement and ash. It should be pointed out that the peak corresponding to high content of CaSO_4 might overlap or cover the Al_2O_3 and Fe_2O_3 peaks because of their relatively lower contents.

3.3 Reactivity of composite oxygen carriers with coal

Investigations into the reactivity of composite oxygen carriers with coal were carried out using TGA. The mass loss (TG) and the mass loss rate (DTG) of the mixture of the composite oxygen carrier and coal with the increasing temperature are demonstrated in Fig. 4. It was clearly found that all three composite oxygen carriers showed the same variation trend with the increasing temperature. With N_2 as a purge gas, the mass loss at 100°C was due to the drying of the mixture of the composite oxygen carrier and coal. The mass loss at 400°C was caused mainly by coal pyrolysis. When the temperature rose to 850°C , the mass of the mixed samples began to decline rapidly. The temperature was maintained at 900°C for 40 min until the mass of the mixed samples began to level off. When the temperature was elevated to 1000°C , the purge gas was switched to O_2 . The mass of the mixed samples first decreased and then increased instantly to a steady value.

In the reduction stage, as was depicted in Fig. 4b, the mass loss rate of the mixture of composite oxygen carriers with coal was higher than that of $\text{CaSO}_4/\text{SiO}_2$; this indicated that the addition of clay, cement and ash could significantly improve the reaction rate between CaSO_4 and coal. The CaSO_4 /clay

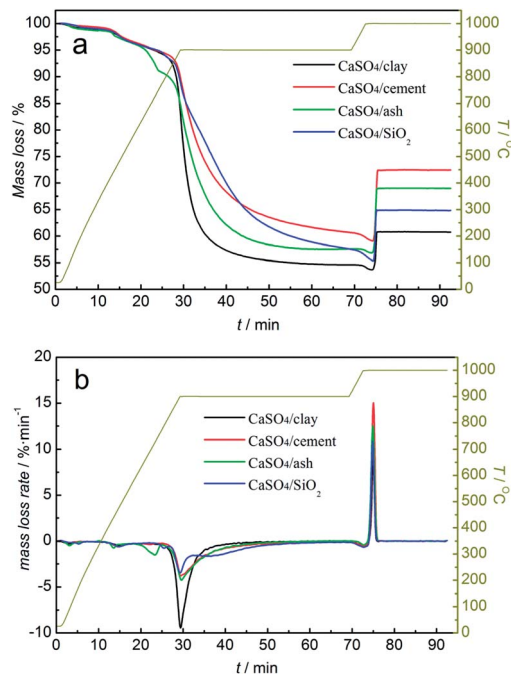


Fig. 4 TG (a) and DTG (b) curves of composition oxygen carriers undergoing one redox cycle with coal without a gasifying agent and air, alternately.

composite oxygen carriers had the highest reaction rate with coal. However, the mass losses of the three composite oxygen carriers differed from each other. The mass loss values were 45.38%, 39.35% and 42.49%, for CaSO_4 /clay, CaSO_4 /cement and CaSO_4 /ash, respectively. One reason for this observation was that the mass rate of CaSO_4 to active support materials changed during the preparation process, due to which the theoretical oxygen transport capacities of the composite oxygen carriers were different from each other. Another important reason was the generation of CaO from the side reaction, and its mass loss was different from that of CaS generated from the desired reaction. The occurrence of the side reaction was verified by the analysis of the generated sulphurous gas. The analysis of relative contents of the generated sulphurous gas with the increasing temperature using TG-MS will be discussed later.

In the oxidation stage, a mass decrease was observed in the mixed sample when the purge gas was switched to O_2 ; this indicated that a small amount of coal did not react with the oxygen carrier, because the solid products generated during the reactions prevented the further reaction between the oxygen carrier and coal. The significant mass gain within 2 min was due to the oxidation reaction of CaS . However, the mass gain values were 7.11%, 13.37% and 12.14% for CaSO_4 /clay, CaSO_4 /cement and CaSO_4 /ash, respectively; all of these values were lower than the theoretical value (23.4%). Generally, the difference between the mass gain value and the theoretical value is derived from the generation of CaO in one redox cycle. When the mass sum of CaSO_4 and CaO was 100% for the CaSO_4 /clay composite oxygen carrier, CaSO_4 accounted for 79.64%, whereas CaO accounted for 20.36% after one redox cycle; in other words, 38.3% of CaSO_4 was deactivated after one redox cycle. Similarly, it was calculated that 25.82% and

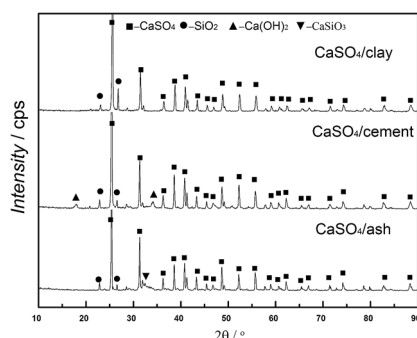


Fig. 3 XRD patterns of fresh composite oxygen carrier.



28.47% of CaSO_4 in the CaSO_4 /cement and CaSO_4 /ash composite oxygen carriers were deactivated after one redox cycle. Among the three types of composite oxygen carriers, the mass gain percent of CaSO_4 /cement was the highest, whereas that of CaSO_4 /clay was the lowest; this can be explained by the effect of different active components on the side reaction of CaSO_4 . Thermodynamically, the components SiO_2 , Al_2O_3 , and Fe_2O_3 were harmful to the regeneration of CaSO_4 due to the formation of a eutectic mixture such as CaSiO_3 , CaAl_2O_7 , CaAl_2O_4 and CaFe_2O_4 . On the other hand, the component CaO was favourable to the regeneration of CaSO_4 (please refer the supplement material). The higher content of CaO in cement gave rise to better regeneration of the CaSO_4 /cement composite oxygen carrier.

Solid products derived from composite oxygen carriers after one reduction with coal and one oxidation with air were analysed by SEM-EDS. As depicted by SEM in Fig. 5, no clear sintering was observed on the surface of the three composite oxygen carriers after one redox cycle; this was mostly due to the addition of active support materials. As shown by EDS analysis, the atomic percent of the sulphurous element in CaSO_4 decreased relative to the calcium element. The mole ratios of Ca to S were 2.69, 2.94 and 1.35 for the CaSO_4 /clay, CaSO_4 /cement and CaSO_4 /ash composite oxygen carriers, respectively. Theoretically, the mole ratio of Ca to S is equal to 1, considering that both Ca and S are obtained from CaSO_4 . The mole ratio of Ca to S in the CaSO_4 /ash composite oxygen carrier was 1.35, indicating that the sulphur release due to the side reaction was lower; this result was in accordance with the aforementioned TG results, where the mass gain value was high. Nevertheless, due to the high content of Ca in cement, the theoretical mole ratio of Ca to S in the CaSO_4 /cement oxygen carrier was higher than that in CaSO_4 /clay. Therefore, the relative mass loss of the sulphurous element of CaSO_4 /clay was highest, though the Ca/S mole ratio was relatively low; this result was also in accordance with the TG results, where the mass gain percent of CaSO_4 /cement was highest, whereas that of CaSO_4 /clay was lowest. Clearly, sulphur loss in the three composite oxygen carriers occurred due to side reactions. It should be pointed out that sulphur release from coal was not taken into account.

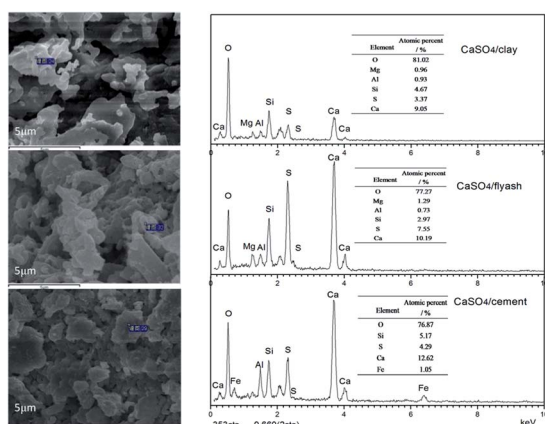


Fig. 5 SEM-EDS images of solid products after one reduction with coal and one oxidation with air, alternately.

Redox reactions between the three types of composite oxygen carriers and coal with steam as the gasifying agent were investigated. The mass loss curve of the mixture of composite oxygen carriers and coal is shown in Fig. 6. As depicted in Fig. 6b, the mass loss rates of all the three types of composite oxygen carriers were accelerated at the reduction stage. The reaction rates of the three types of composite oxygen carriers with coal were clearly higher than that of the CaSO_4 / SiO_2 oxygen carrier; this was ascribed to the higher gas–solid reaction rate between the gasified gas and the oxygen carrier compared to the solid–solid reaction rate between coal and the oxygen carrier. The coal gasification reaction generating gasified gases (H_2 and CO) could be catalysed by the active components Fe_2O_3 , CaO , potassium salt and sodium salt contained in clay, ash and cement. However, there was a slow decline in the mass of the mixed samples after a rapid decline, which was different from that without gasifying agent. According to our previous research,⁴¹ the reaction between CaSO_4 and coal with steam as the gasifying agent transformed from reaction control to diffusion control. Diffusion control deteriorated the subsequent reaction between the composite oxygen carriers and coal or gasified gas. The mass loss values were 51.39%, 48.35% and 45.15%, for CaSO_4 /clay, CaSO_4 /cement and CaSO_4 /ash, respectively, which were higher than that without gasifying agent.

At the oxidation stage, similar to the observations under inert atmosphere, when the purge gas was switched to O_2 at 1000°C , the mass of the mixed samples slightly decreased before increasing. It was demonstrated that the extent of the reaction of coal was still limited even after using steam as the gasifying agent. The mass gain values were 14.50%, 5.9% and 12.53% for CaSO_4 /clay, CaSO_4 /cement and CaSO_4 /ash, respectively, which were lower than the

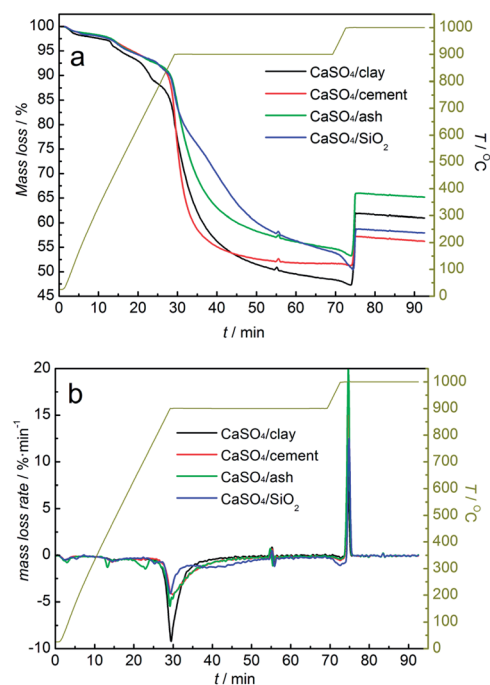


Fig. 6 TG (a) and DTG (b) curves of composite oxygen carriers undergoing one redox cycle with coal using steam as the gasifying agent and air, alternately.



theoretical value (23.4%). Side reactions generating sulphurous gas occurred in the redox reactions of the three types of composite oxygen carriers with coal. It was also calculated that 23.32%, 40.47% and 27.63% of CaSO_4 were deactivated in the CaSO_4 /clay, CaSO_4 /cement and CaSO_4 /ash composite oxygen carriers, respectively, after one redox cycle. Surprisingly, the mass gain percent of CaSO_4 /clay increased from 7.11% without gasifying agent to 14.50% with steam as the gasifying agent, whereas the mass gain percent of CaSO_4 /cement decreased from 13.37% to 5.90%. It can be inferred that the inhibition of CaO by the side reaction was greatly affected by steam. The mass gain percent of CaSO_4 /ash was 12.53%, and it was similar to that without gasifying agent. The positive effect of coal ash on the redox reaction of the CaSO_4 oxygen carrier was noted. However, further investigation into the CaSO_4 /ash composite oxygen carrier was not taken into account due to the poor attrition resistance.

3.4 Release of sulphurous gas from CaSO_4 /clay and CaSO_4 /clay reacted with coal

To study the release behaviour of sulphurous gas during the redox reaction, the redox reactions of the CaSO_4 /clay and CaSO_4 /cement composite oxygen carriers with/without a gasifying agent were analysed by TG-MS. The potential sulphurous gas release included SO_2 and COS without gasifying agent and SO_2 and H_2S with steam as the gasifying agent. The mass-to-charge ratios of SO_2 , COS and H_2S were represented by 64, 60, and 34, respectively, in MS. The relative amount of sulphurous gas was considered to be proportional to ionic strength.

MS analysis of sulphurous gas products from TGA *versus* time and temperature without gasifying agent is depicted in Fig. 7. The generation of SO_2 occurred in both the reduction and oxidation stages, whereas only a small amount of COS was released in the reduction stage. In the reduction stage, the generation of SO_2 started from 778 °C and 840 °C for the CaSO_4 /clay and CaSO_4 /cement composite oxygen carriers, respectively. The release amount of SO_2 with the CaSO_4 /cement composite oxygen carrier was far less than that with the CaSO_4 /clay composite oxygen carrier; this indicated that by restraining the side reactions (R4) and (R6), the CaSO_4 /cement composite oxygen carriers exhibited relatively high suppression ability of SO_2 release, mainly due to the high content of CaO in cement. It should be noted that the sulphur present in coal was also released at the same time in the form of SO_2 or COS. Besides, as shown in Fig. 7b, the initial temperature generating COS was relatively low, which was ascribed to the side reaction (R5).

In the oxidation stage, the release of SO_2 was very high for both the CaSO_4 /clay and CaSO_4 /cement composite oxygen carriers. The SO_2 peak area was larger for the CaSO_4 /cement composite oxygen carrier. The release of SO_2 was most likely due to low oxygen partial pressure or the reaction of the generated CaSO_4 with CaS, according to reaction equations (R11) and (R9), respectively.

As illustrated in Fig. 8, the SO_2 release occurred in both the reduction and oxidation stages, whereas the H_2S release occurred in the reduction stage using steam as the gasifying agent. The release amounts of SO_2 for both composite oxygen

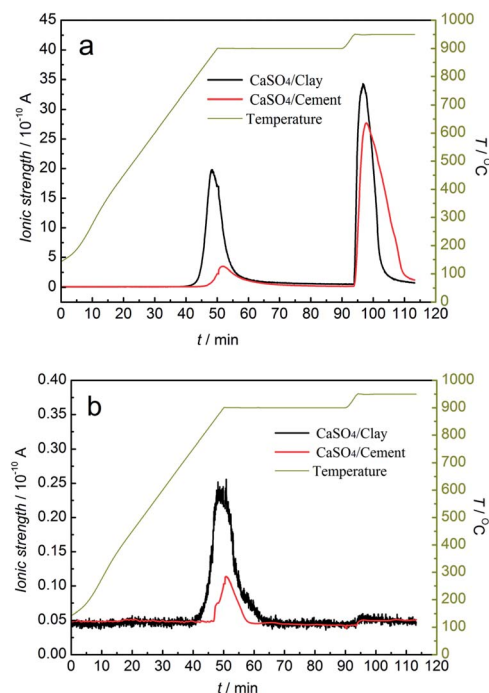


Fig. 7 Mass spectrometric analysis of sulphurous gas products from TGA without the gasifying agent (a) SO_2 ; (b) COS.

carriers were greater than that under the inert atmosphere. At the same time, the release behaviour lasted for the entire reduction stage. Especially for CaSO_4 /cement, the release amount of SO_2 and H_2S was always high during the entire reduction stage, and it was higher than that for the CaSO_4 /clay composite oxygen carrier. This resulted in lower mass gain at the oxidation stage, as seen in Fig. 6. With steam as the gasifying agent, the gasification reaction of coal generated syngas (CO and H_2). Thus, the generated syngas enhanced the side reactions (R4) and (R6) for releasing SO_2 and the side reaction (R7) for releasing H_2S . Considering the poor regeneration of CaSO_4 /cement with steam as the gasifying agent, it seemed that the steam reduced the suppression of sulphurous gas by CaO; this could be due to the effect of catalysis of other components in cement on the side reactions or the deactivation of CaO under steam atmosphere.

In summary, we infer that the addition of active support materials cannot hinder the release of sulphurous gas from side reactions in both the reduction and oxidation stages; however, the additions can improve the reaction rate. Compared with CaSO_4 /ash and CaSO_4 /cement, the CaSO_4 /clay composite oxygen carrier exhibited excellent attrition resistance and reactivity with steam as the gasifying agent. Hence, the CaSO_4 /clay composite oxygen carrier can be a potential oxygen carrier for coal CLC after being upgraded.

3.5 Upgrading the CaSO_4 /clay composite oxygen Carrier using CaO

Thermodynamic analysis and thermogravimetric experiments indicated that the side reaction emitting sulphurous gas was unavoidable, even without the addition of inert materials to



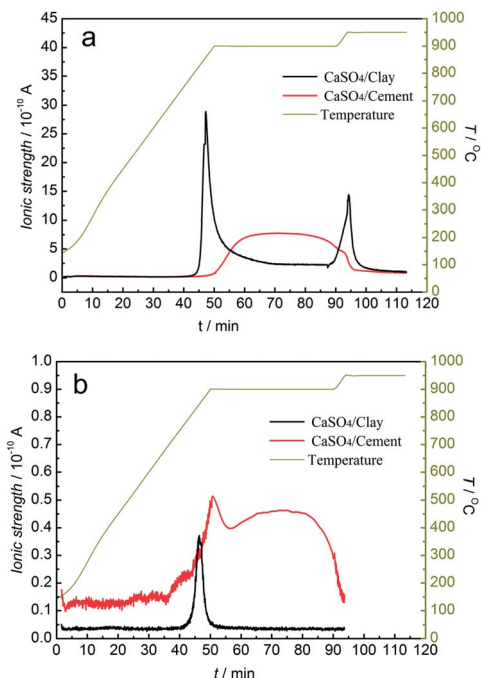


Fig. 8 Mass spectrometric analysis of sulphurous gas products from TGA with steam as the gasifying agent. (a) SO_2 ; (b) H_2S .

CaSO_4 . It seemed that the inhibition of sulphurous gas emission was vital for using CaSO_4 as an oxygen carrier for coal CLC. Because CaO is deemed as a sulphur adsorption agent in flue gas desulphurization, $\text{CaSO}_4/\text{clay}$ was upgraded by adding CaO . $\text{CaSO}_4/\text{clay}$ composite oxygen carriers with four mass ratios of CaO to CaSO_4 , i.e., 0.25, 0.5, 1 and 1.25, were prepared. Investigations into the reactivities of the four upgraded $\text{CaSO}_4/\text{clay}$ composite oxygen carriers with coal were carried out in a fluidised bed.

The results for carbon conversion efficiency and CO_2 generation rate with different CaO additions *versus* time are shown in Fig. 9. As displayed in Fig. 9a, the carbon conversion efficiency increased with the increasing time. Moreover, the carbon conversion efficiency increased with the increasing amounts of CaO . The time to achieve 95% conversion efficiency was shortened from 40 min without the addition of CaO to 28 min with 1.25 mass ratio additions. As was depicted in Fig. 9b, the time of the maximum CO_2 generation rate advanced with the increasing addition mass ratio. It shortened from 12 min to 5 min. When the mass ratio of CaSO_4 to CaO was equal to 1, the maximum value of the CO_2 generation rate was 0.43 L min^{-1} . The addition of CaO enhanced the reaction between the $\text{CaSO}_4/\text{clay}$ composite oxygen carrier and coal with steam as the gasifying agent.

The effect of CaO addition on the concentration of sulphurous gas is demonstrated in Fig. 10. In the reduction stage, the SO_2 volume concentration increased first and then decreased, with the maximum value of 0.24% without the addition of CaO . With the increase in the added amounts of CaO , both the duration of SO_2 emission and the maximum value of SO_2 volume concentration decreased. When the mass ratio was higher than 1, the SO_2 volume concentration was restrained. In

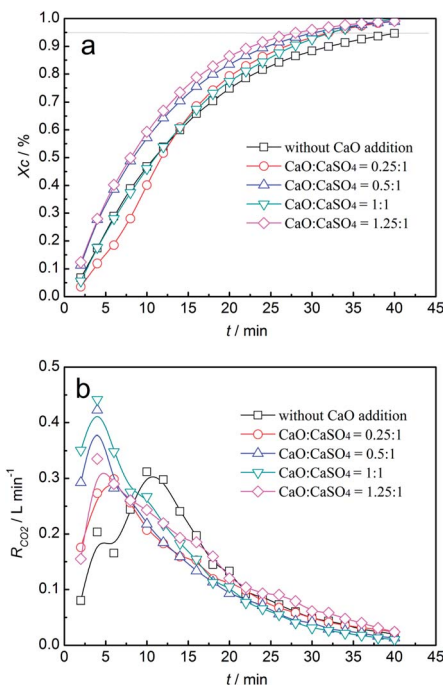


Fig. 9 Variation of carbon conversion efficiency (a) and CO_2 generation rate (b) with time for the $\text{CaSO}_4/\text{clay}$ composite oxygen carrier with different CaO additions.

the oxidation stage, the SO_2 volume concentration exhibited a similar trend with that in the reduction stage. The maximum value was 0.54% without the addition of CaO , which revealed that SO_2 emission during oxidation was significant. With the increasing mass ratio of CaO , both the duration of SO_2 emission

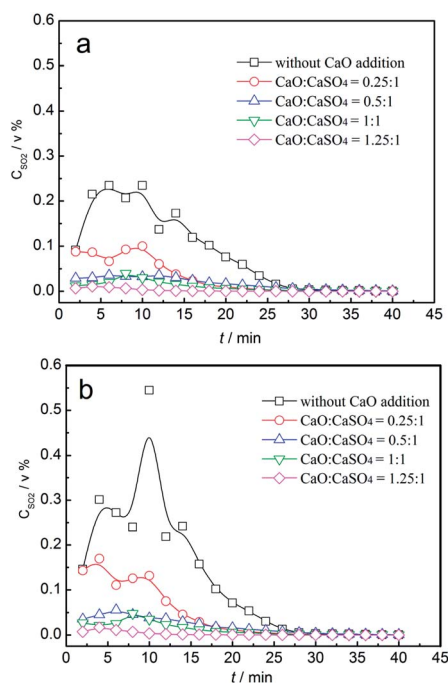


Fig. 10 Variation of gas concentrations of SO_2 with time for the $\text{CaSO}_4/\text{clay}$ composite oxygen carrier with different CaO additions; (a) reduction stage, (b) oxidation stage.

and the maximum value of SO₂ volume concentration decreased. In summary, SO₂ release was restrained until the mass ratio of CaO to CaSO₄ was higher than 1 at both the reduction and oxidation stages. The corresponding oxygen transport capacity of the composite oxygen carriers was about 14.1 wt%.

4. Conclusions

From the viewpoint of chemical reaction engineering, the addition of active supports can inhibit the sintering of oxygen carrier particles and enhance the reactivity and structural stability of the oxygen carriers. CaSO₄ composite oxygen carriers composed of clay, cement, and ash were prepared separately. The following conclusions were obtained: (1) the attrition resistance of CaSO₄ oxygen carriers composed of clay and cement was improved due to bond action of cement and clay. The reaction rate of CaSO₄ with coal could be improved by adding active support materials. However, due to side reactions, the regeneration of CaSO₄ deteriorated. (2) Sulphurous gas generation from three types of CaSO₄ composite oxygen carriers occurred in both the reduction and oxidation stages. The use of a gasifying agent greatly affected the release behaviour of sulphurous gas. Compared with CaSO₄/ash and CaSO₄/cement, the CaSO₄/clay composite oxygen carrier exhibited excellent reactivity with steam as the gasifying agent. (3) Addition of CaO to CaSO₄/clay could enhance its reaction rate with coal and restrain the release of SO₂. The SO₂ release was restrained until the mass ratio of CaO to CaSO₄ was higher than 1 in both the reduction and oxidation stages. At this point, the corresponding oxygen transport capacity of composite oxygen carriers was about 14.1 wt%. The addition of a large amount of CaO to the CaSO₄ oxygen carrier is a good strategy to prevent the release of sulphurous gas.

Conflicts of interest

There are no conflicts to declare.

Acknowledgements

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References

- 1 J. Adanez, A. Abad, F. Garcia-Labiano, P. Gayan and L. F. de Diego, *Prog. Energy Combust. Sci.*, 2012, **38**, 215–282.
- 2 J. Adánez, A. Abad, T. Mendiara, P. Gayán, L. F. de Diego and F. García-Labiano, *Prog. Energy Combust. Sci.*, 2018, **65**, 6–66.
- 3 A. Serrano, F. Garcia-Labiano, L. F. de Diego, P. Gayan, A. Abad and J. Adanez, *Fuel Process. Technol.*, 2017, **160**, 47–54.
- 4 P. Moldenhauer, M. Ryden, T. Mattisson, A. Jamal and A. Lyngfelt, *Fuel Process. Technol.*, 2017, **156**, 124–137.
- 5 X. Zhao, H. Zhou, V. Singh Sikarwar, M. Zhao, Ah-H. A. Park, P. S. Fennell, L. Shen and L. -Shih Fan, *Energy Environ. Sci.*, 2017, **10**, 1885–1910.
- 6 International Energy Agency, *World Energy Outlook*, 2013.
- 7 J. Wang and E. J. Anthony, *Appl. Energy*, 2008, **85**, 73–79.
- 8 A. Y. Ilyushechkin, M. Kochanek and S. Lim, *Fuel Process. Technol.*, 2016, **147**, 71–82.
- 9 T. Mattisson, A. Lyngfelt and H. Leion, *Int. J. Greenhouse Gas Control*, 2009, **3**, 11–19.
- 10 A. Abad, I. Adánez-Rubio, P. Gayán, F. García-Labiano, L. F. de Diego and J. Adánez, *Int. J. Greenhouse Gas Control*, 2012, **6**, 189–200.
- 11 M. Schmitz, C. Linderholm, P. Hallberg, S. Sundqvist and A. Lyngfelt, *Energy Fuels*, 2016, **30**, 1204–1216.
- 12 A. Fossdal, E. Bakken, B. A. Øye, C. Schøning, I. Kaus, T. Møkkelbost and Y. Larring, *Int. J. Greenhouse Gas Control*, 2011, **5**, 483–488.
- 13 A. Thon, M. Kramp, E. U. Hartge, S. Heinrich and J. Werther, *Appl. Energy*, 2014, **118**, 309–317.
- 14 J. Ströhle, M. Orth and B. Epple, *Appl. Energy*, 2014, **157**, 288–294.
- 15 T. Mendiara, A. Abad, L. F. de Diego, F. García-Labiano, P. Gayán and J. Adánez, *Energy Fuels*, 2012, **26**, 1420–1431.
- 16 T. Mendiara, L. F. de Diego, F. García-Labiano, P. Gayán, A. Abad and J. Adánez, *Int. J. Greenhouse Gas Control*, 2013, **17**, 170–182.
- 17 S. Zhang, C. Saha, Y. Yang, S. Bhattacharya and R. Xiao, *Energy Fuels*, 2011, **25**, 4357–4366.
- 18 Q. J. Guo, J. S. Zhang and H. J. Tian, *Chem. Eng. Commun.*, 2012, **199**, 1463–1491.
- 19 Q. Song, R. Xiao, Z. Deng, L. Shen, J. Xiao and M. Zhang, *Ind. Eng. Chem. Res.*, 2008, **47**, 8148–8159.
- 20 Q. Song, R. Xiao, Z. Deng, W. Zheng, L. Shen and J. Xiao, *Energy Fuels*, 2008, **22**, 3661–3672.
- 21 Q. Song, R. Xiao, Z. Deng, H. Zhang, L. Shen, J. Xiao and M. Zhang, *Energy Convers. Manage.*, 2008, **49**, 3178–3187.
- 22 M. Zheng, L. H. Shen and J. Xiao, *Int. J. Greenhouse Gas Control*, 2010, **4**, 716–728.
- 23 W. Z. Bi, T. J. Chen, R. D. Zhao, Z. Q. Wang, J. L. Wu and J. H. Wu, *RSC Adv.*, 2015, **5**, 34913–34920.
- 24 R. Xiao and Q. L. Song, *Combust. Flame*, 2011, **158**, 2524–2539.
- 25 Y. Liu and Q. Guo, *Chin. J. Chem. Eng.*, 2013, **21**, 127–134.
- 26 Y. Liu, Q. Guo, Y. Chen and H. -Jung Ryu, *Ind. Eng. Chem. Res.*, 2012, **51**, 10364–10373.
- 27 J. Yang, L. P. Ma, S. L. Dong, H. P. Liu, S. Q. Zhao, X. J. Cui, D. L. Zheng and J. Yang, *Fuel*, 2017, **194**, 448–459.
- 28 J. Chiu and H. Andrus, *Alstom's Chemical Looping Prototypes, CO2 Capture Technology Meeting*, Pittsburg, PA, 2014.
- 29 A. Levasseur, *Alstom's Limestone-based Chemical Looping Development For Advanced Gasification, DOE Workshop: Gasification Systems and Coal & Biomass to Liquids*, Morgantown, WV, August, 2015.



- 30 A. Abad, M. de las Obras-Loscertales, F. García-Labiano, L. F. de Diego, P. Gayán and J. Adánez, *Chem. Eng. J.*, 2017, **310**, 226–239.
- 31 L. Shen, M. Zheng, J. Xiao and R. Xiao, *Combust. Flame*, 2008, **154**, 489–506.
- 32 M. Zheng, Y. B. Xing, K. Z. Li, S. M. Zhong, H. Wang and B. L. Zhao, *Energy & Fuels*, 2017, **31**, 5255–5265.
- 33 H. Tian, Q. Guo, X. Yue and Y. Liu, *Fuel Process. Technol.*, 2010, **91**, 1640–1649.
- 34 S. Zhang, R. Xiao, J. Liu and S. Bhattacharya, *Int. J. Greenhouse Gas Control*, 2013, **17**, 1–12.
- 35 T. Song, M. Zheng, L. Shen, T. Zhang, X. Niu and J. Xiao, *Ind. Eng. Chem. Res.*, 2013, **52**, 4059–4071.
- 36 N. Ding, C. Zhang, C. Luo, Y. Zheng and Z. Liu, Effect of hematite addition to CaSO₄ oxygen carrier in chemical, *RSC Adv.*, 2015, **5**, 56362–56376.
- 37 M. Zheng, L. Shen and X. Feng, *Energy Convers. Manage.*, 2014, **83**, 270–283.
- 38 P. F. Zhang, N. Ding, H. J. Li, M. M. Lv, N. Guan and Z. G. Liu, *Energy Technol.*, 2017, **5**, 469–480.
- 39 Q. J. Guo, Y. Z. Liu, W. H. Jia, M. M. Yang, X. D. Hu and H.-J. Ryu, *Energy & Fuels*, 2014, **28**, 7053–7060.
- 40 A. Cabello, P. Gayan, F. Garcia-Labiano, L. F. de Diego, A. Abad and J. Adanez, *Fuel Process. Technol.*, 2016, **148**, 188–197.
- 41 Y. Z. Liu, W. H. Jia, Q. J. Guo and H.-J. Ryu, *Chin. J. Chem. Eng.*, 2014, **22**, 1208–1214.

