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A hybrid magnetic core–shell fibrous silica nanocomposite for a chemosensor-based highly effective fluorescent detection of Cu(II)

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Herein, a novel hybrid magnetic core–shell fibrous silica nanocomposite (RhB–Fe₃O₄/MnO₂/SiO₂/KCC-1) probe-based chemosensor was designed and its behaviour towards Cu(II) metal ion was investigated using a fluorescence spectrometer. The organic receptor rhodamine B (RhB) fluorophore derivative was covalently grafted onto the surface of the magnetic core–shell fibrous silica nanocomposite. This sensing probe achieved the selectivity towards Cu(II) in an aqueous solution, and other competing metal ions basically induced no spectral change. Thus, this sensing probe can work as a Cu(II) selective fluorescent sensor. The synthesized material was characterized using transmission electron microscopy (TEM), scanning electron microscopy-energy dispersive spectroscopy (SEM-EDS), surface analysis (with BET), Fourier transform infrared spectroscopy (FT-IR), X-ray diffractometry, and thermogravimetric analysis (TGA). The rhodamine B derivative has a unique signalling probe that exhibits a turn-on fluorescence enhancement upon the recognition of Cu(II) ion with an excitation at 569 nm. The linearity of the Stern–Volmer plot ($R^2 = 0.9882$) and the detection limit of 12.3×10^{-8} M were achieved. Finally, the sensor was tested to detect Cu(II) ion in different real water samples.

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1. Introduction

Cu(II) is one of the most toxic heavy metals. Even at low concentrations, it is a threat to human health and environment due to its high toxicity, mobility, and ability to accumulate through food chains or atmosphere in the ecological system.¹ Copper(II), as the third most abundant essential trace metal ion in the human body, also plays an important role in various environmental, chemical, and physiological systems. There are many reactions in the biological process that involve the electron transfer that is catalyzed by copper-containing enzymes.² Copper poisoning induces a wide variety of serious complications, such as vomiting, lethargy, increased blood pressure, acute haemolytic anaemia, Mekeles and Wilson diseases, neurotoxicity, neurodegenerative, and Alzheimer's and prion diseases, in humans.^{3–5} Additionally, recently, copper has been suspected to cause infant liver damage. Both Indian childhood cirrhosis (ICC) and non-Indian childhood cirrhosis (NICC) have been proven to be associated with an excessive intake of copper ions.⁶ The permissible limit of copper in drinking water set by the US Environmental Protection Agency (EPA) is 1.3 mg l^{-1} ($20 \text{ }\mu\text{M}$) and that set by the World Health Organization (WHO) is 1.0 mg l^{-1} .⁷ Thus, detection of copper in water is of great

significance due to the wide distribution of copper in the environment and its bioaccumulation through food chain.

Therefore, the interest in developing rapid, specific, and cost-effective tools to detect Cu(II) is steadily growing.⁸ The traditional quantitative methods to develop a Cu(II) sensor include inductively coupled plasma mass spectroscopy (ICPMS), atomic absorption/emission spectroscopy (AAS/AES), electrochemical detection, and non-material-based probes.^{9–13} However, these methods are time-consuming and require expensive instrumentation, complicated equipment operation *etc.* Thus, significant efforts have been dedicated towards the development of selective and sensitive detection methods. Among them, fluorescent and colorimetric sensors are the most convenient techniques to monitor the level of copper. The paramagnetic behaviour of Cu(II), with a $3d^9$ outer shell electron configuration, attributes to the quenching of the emission intensity.¹⁴ With regard to the detection of copper *via* a fluorescent chemosensor, many previous reports have been focused on the fluorescence quenching process through energy or electron transfer processes. In term of the sensitivity, fluorescence enhancement (turn-on) signals are much easier to detect than fluorescence quenching (turn-off) signals. The rhodamine receptor-based fluorescent chemosensor for the detection of metal cations has received significant attention in recent years by virtue of its excellent spectroscopic properties and good water solubility. It is an ideal moiety to construct turn-on fluorescence chemosensor in term of metal ion triggering the spirolactam ring opening mechanism along with change in chromaticity and fluorescence

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at the same time. Considering the high affinity of Cu(II) to N and O atoms, rhodamine hydrazone is an excellent platform to design a novel fluorescent chemosensor for specially sensing Cu(II) with high sensitivity in an aqueous solvent.^{15–19}

Mesoporous silica materials provide a good support for fluorescence probes due to their unique characteristics such as excellent solvent dispersibility, good optical transparency in the visible region, favourable biocompatibility, large surface area, open pore structure, and abundant functionalization sites in the silica network.^{20,21} Some silica-based mesoporous materials such as SBA-15, MCM-41, and MCM-48 have been reported as inorganic carriers for chemosensor.^{22–25} However, most of these optical nanoprobe have some limitations, which include the requirement of organic or aqueous organic solvent mixtures, poor performance in the removal of metal ions, and poor accessibility of analytes to sensing sites inside the probe.²⁶ Therefore, recently, the most promising fibrous silica spheres (KCC-1) have been developed, and they exhibit excellent physical properties such as fibrous surface morphology, thermal and mechanical stability, better sorption capacity, and ease of functionalization of the organic molecule. This valuable material is also used as a catalyst, drug delivery vehicle, chromatography support, and chemosensor.²⁷

Herein, we explored a new hybrid nonmaterial that was formed by grafting the functionalized rhodamine B-based receptor onto the inner surface of magnetic core-shell fibrous silica KCC-1. The new hybrid sensor displayed an extremely high selectivity and sensitivity towards Cu(II) ion as compared to that towards other metal ions. The prepared iron oxide nanoparticles acted as a centre core, and they were then coated with manganese oxide over the surface of silica shell. The Fe₃O₄ core was used to saturate the magnetic behaviour, to protect the fluorescence quenching, and to avoid the aggregation of the nanocomposite. This approach has improved the selectivity and enhancement of fluorescence signal, as depicted in Scheme 1. Finally, the developed sensing system was successfully utilized for the detection of Cu(II) ion in environmental water samples. The design strategy of the hybrid sensor I is such that the recovery and reusability of the sensing material can be achieved. To the best of our knowledge, this is the first time that the detection of Cu(II) ion using an Fe₃O₄/MnO₂ magnetic nanocomposite on a fibrous silica core functionalized with a rhodamine-based derivative has been reported.

2. Experimental

2.1. Reagents and instruments

All chemicals were obtained from the commercial source and used without further purification. Unless otherwise stated, all chemicals were of analytical reagent grade. Rhodamine B, 5-chlorosalicylic aldehyde, ethylenediamine, cyclohexane, *n*-pentanol, tetraethoxysilane (TEOS), cetylpyridinium bromide (CPB), and 3-aminopropyltriethoxysilane (APTES) were purchased from Alfa-Aesar. Salts used in the stock solutions of metal ions were KNO₃, AgNO₃, Ca(NO₃)₂·4H₂O, Ni(NO₃)₂·6H₂O, Hg(NO₃)₂·H₂O, Pb(NO₃)₂, Zn(NO₃)₂·6H₂O, Co(NO₃)₂·6H₂O, Al(NO₃)₃·9H₂O, and Fe(NO₃)₃·9H₂O. They were purchased from Sigma-Aldrich (India). All other organic solvents such as ethanol,

toluene, dichloromethane, and ethyl acetate used in this study were obtained from Siscon research laboratories Pvt. Ltd (India) and purified by a standard procedure.

Transmission electron microscopy (TEM) analysis was performed using JEOL/JEM 2100. For the sample preparation, the powders were dispersed in ethanol, and the suspension was dropped onto a 200 mesh carbon-coated TEM grid. The morphologies of the as-prepared samples were investigated using the field emission scanning electron microscope (FE-SEM) FEI Quanta FEG200. X-Ray diffraction (XRD) measurements were carried out *via* a PAN analytical X'pert powder diffractometer using Cu-K α radiation ($\lambda = 0.15405$ Å). Fourier transform infrared (FT-IR) spectra were obtained using an Agilent Resolution Pro FT-IR spectrometer. The surface area was measured using a smart sorbs 92 surface area analyser, and N₂ was used as an adsorbate. Thermogravimetric analysis (TGA, STA 2500 Regulus, NETZSCH) was performed at a heating rate of 10 °C min⁻¹, and the nitrogen flow was 10 ml min⁻¹. The magnetization property of the sample was measured using a vibrating sample magnetometer (Lakeshore VSM 7410). Fluorescence spectra of the sample were obtained using a HORIBA JOBIN YVON Fluoromax-4 spectrofluorometer with a xenon lamp excitation source. The UV-vis absorption spectra were obtained using a Shimadzu UV-2600 spectrophotometer. All pH values were obtained using a PHS-3C digital pH meter.

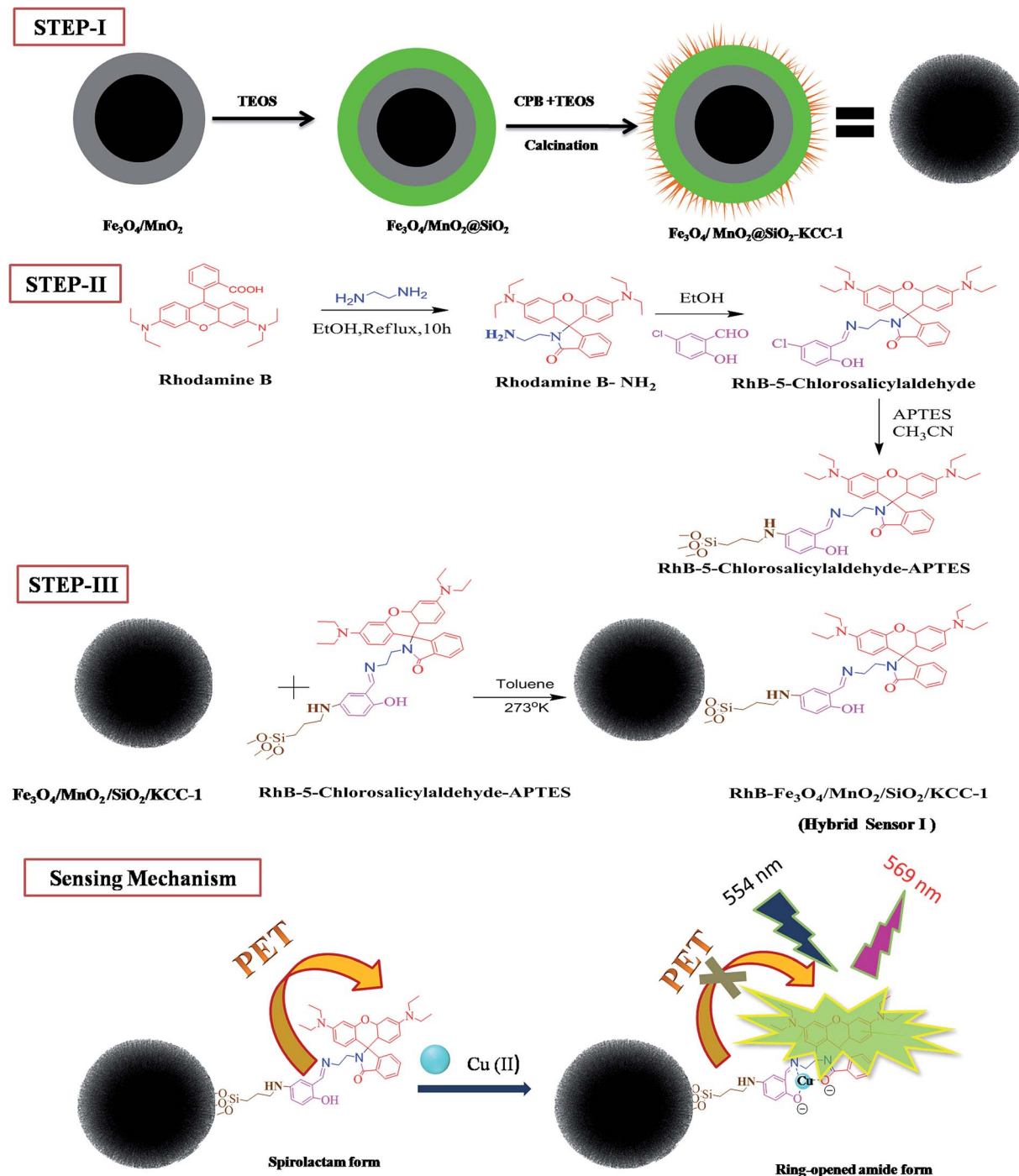
2.2. Preparation of the Fe₃O₄/MnO₂/SiO₂ nanocomposite

The Fe₃O₄/MnO₂ nanocomposite was prepared by a simple hydrothermal process following the procedure reported in the literature.²⁸ In a typical preparation, 0.51 g of KMnO₄ was dissolved in 35 ml of deionized water and 0.7 ml of HCl (37 wt%) under stirring for 15 minutes. To this mixture, 0.3 g of Fe₃O₄ was added, and the mixture was continuously stirred for another 30 min. Then, the mixture was transferred to a 50 ml Teflon-lined autoclave that was kept in an oven at 120 °C for 6 h for hydrothermal treatment under a static condition. The required product was obtained by applying an external magnet, washed several times with ethanol and deionized water, and then dried at 60 °C. The abovementioned prepared Fe₃O₄/MnO₂ nanocomposite was dispersed in 30 ml ethanol and deionized water, and the mixture was stirred with 0.7 ml of concentrated ammonia solution using a magnetic stirrer. After this, 0.2 g of TEOS was added dropwise, and the mixture was continuously stirred at room temperature for 6 h. Finally, the resulting product Fe₃O₄/MnO₂/SiO₂ nanocomposite was separated *via* an external magnetic field and washed several times with deionized water.

2.3. Preparation of Fe₃O₄/MnO₂/SiO₂/KCC-1

For the facile synthesis of Fe₃O₄/MnO₂/SiO₂/KCC-1, 0.3 g of Fe₃O₄/MnO₂/SiO₂ was first mixed with distilled water and ultrasonicated for 30 min. Then, 0.3 g of urea dissolved in 20 ml of water was added dropwise to the solution and ultrasonicated for 20 min. Subsequently, 0.5 g of CPB dissolved in 0.7 ml of *n*-pentanol and 30 ml of cyclohexane was added slowly under stirring at room temperature. TEOS (1.25 g, 6.0 mmol) was added dropwise to the mixture and it was stirred for 1 h. Then,





Scheme 1 Schematic of the design and turn-on fluorescence mechanism for Cu(II) sensing.

the reaction mixture was kept in an autoclave at 120 °C for 6 h. The obtained nanocomposite was separated by a strong magnet, washed sequentially with water and acetone, and then dried overnight at 60 °C. Finally, the prepared hybrid silica nanocomposite material was calcined at 550 °C for 6 h.

2.4. Preparation of RhB-NH₂

The preparation of RhB-NH₂ was carried out using the previously reported method.²⁹ First, rhodamine B (3.82 g, 8 mmol) and ethane-1,2-diamine (2.6 ml, 40 mmol) were dissolved in

100 ml of ethanol and refluxed for 8 h. After this, the reaction mixture was cooled down to room temperature, and then, the solvent was removed under vacuum. Finally, the resultant product was obtained and washed with CH₂Cl. The organic layer was dried over anhydrous magnesium sulphate.

2.5. Preparation of RhB-5-chlorosalicylaldehyde-APTES

Typically, 0.46 g of RhB-NH₂ was dissolved in 30 ml of ethanol and 4 mmol of 5-chlorosalicylaldehyde was added dropwise to the solution. The mixture was refluxed for 8 h and allowed to



cool down to room temperature. To the mixture, 10 ml of ethyl acetate was added, and then, the mixture was kept in a refrigerator overnight. The material was washed with distilled water followed by ethanol and dried at 65 °C. After the resulting material RhB-5-chlorosalicylaldehyde (0.270 g, 1.0 mmol) was obtained and reacted with (0.241 g, 0.1 mmol) APTES, K₂CO₃ (0.138 g, 1.0 mmol) in 30 ml acetonitrile was slowly added to the solution mixture that was then refluxed for 8 h. Finally, the product was purified by column chromatography with ethyl acetate : hexane solvent.

2.6. Synthesis of RhB-Fe₃O₄/MnO₂/SiO₂/KCC-1

In a typical synthesis, 200 mg of Fe₃O₄/MnO₂/SiO₂/KCC-1 was dissolved in 40 ml of toluene, and 200 mg of RhB-Si was added. The mixture was stirred for 12 h under a nitrogen atmosphere. The resulting product was separated by a strong magnet and washed several times with 30 ml of different solvents, including toluene, dichloromethane, and ethanol, each time. The purified product was dried at 60 °C and then used for further analysis.

3. Results and discussion

3.1. Design and analysis of the hybrid sensor

The synthesis of rhodamine-based receptor and functionalized mesoporous magnetic core-shell fibrous silica KCC-1 involves several steps, as outlined in Scheme 1. The main aim of this study was to identify the ability of the hybrid sensor I towards the detection of Cu(II) ion over other metal ions and the recovery of this sensor. To achieve an effective sensing system, the design of multicore-shell is architected in a way that paramagnetic microspheres of iron oxide core (Fe₃O₄) are deposited with nanosized manganese oxide (MnO₂) to prevent aggregation and dipole interaction between magnetic nanoparticles. It is

found that Fe₃O₄ provides an effortless magnetic separation and increases the durability of the material. The modified deposited MnO₂ nanoparticles have a high acid etching resistance, are stable in organic solvents such as methanol, and also closely interact with Fe₃O₄.^{30,31} To avoid the solubility of the core-shell Fe₃O₄/MnO₂ in water, a multistep strategy is engaged to deposit dense silica on the surface. SiO₂ provides stability to the inner core, and its uniform surface supports the growth of fibrous silica. Very importantly, the thickness of the fabricated silica prevents the fluorescence energy transfer process (FRET) between the magnetic core-shell and the fluorophore.³² Finally, the nanocomposite (Fe₃O₄/MnO₂/SiO₂) is modified with fibrous silica on the outer surface. The fibrous mesoporous silica surface could be easily functionalized with a silane reagent (APTES), on the surface moieties, such that sensing molecules of rhodamine B derivatives are uniformly covered on the silica surface. The functionalized modified nanocomposite is a suitable solid base material with a better dispersibility. The importance of this hybrid sensor I is that it has electron rich species, such as carbonyl (O), imino (N), and phenol (O) atoms, which effectively and selectively binding with Cu(II) over other metal ions. The spirolactam ring-opening upon binding with cation to enhance the fluorescence is a special feature of the hybrid sensor I to create a barrier for quenching and improving the sensing performance of Cu(II).

3.2. Characterization

The morphology of the hybrid magnetic core-shell fibrous silica nanocomposite was observed by SEM and TEM, as shown in Fig. 1a–f. The Fe₃O₄ nanoparticles exhibited a spherical shape with a diameter of about 480 nm (Fig. 1a). After coating by amorphous MnO₂, the average diameter of Fe₃O₄/MnO₂

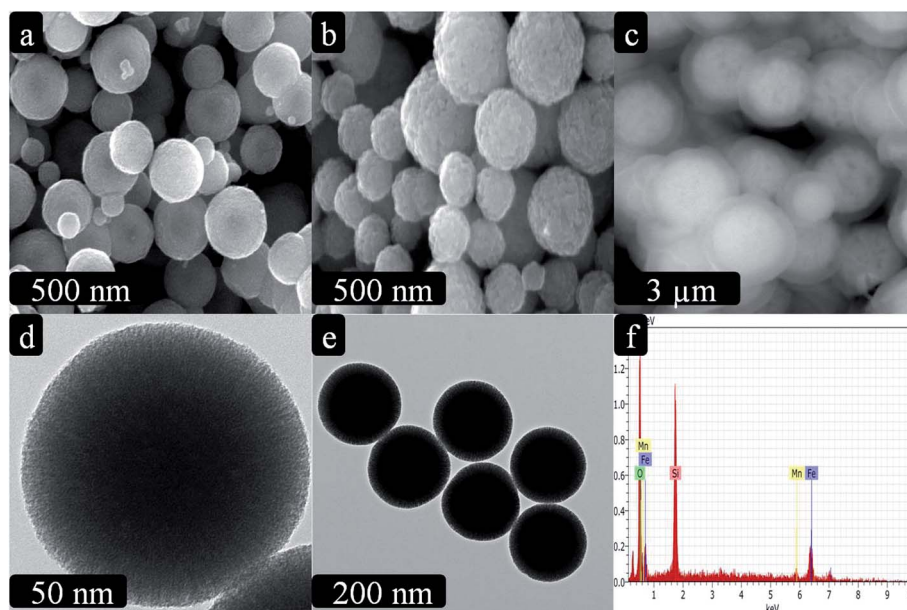


Fig. 1 SEM images of (a) Fe₃O₄, (b) Fe₃O₄/MnO₂, and (c) Fe₃O₄/MnO₂/SiO₂/KCC-1. TEM images of (d) KCC-1 and (e) Fe₃O₄/MnO₂/SiO₂/KCC-1. (f) Elemental analysis of the Fe₃O₄/MnO₂/SiO₂/KCC-1 nanocomposite.



nanocomposite increased to 620 nm. This clearly indicates that uniform and well-dispersed spherical structures were formed without aggregation (Fig. 1b). After this, successful coating of magnetic nanocomposite with fibrous silica was clearly observed in Fig. 1c. Furthermore, in the TEM images (Fig. 1d and e), the fibrous structure of the parent KCC-1 clearly shows the unique fibrous surface with a spherical shape (Fig. 1d). As shown in Fig. 1e, the inner surface clearly shows the core shell modification ($\text{Fe}_3\text{O}_4/\text{MnO}_2/\text{SiO}_2/\text{KCC-1}$) compared to that of parent KCC-1. In Fig. 1e, the black inner surface in the core-shell is represented by $\text{Fe}_3\text{O}_4/\text{MnO}_2$ and gray-coloured outer surface is represented by fibrous silica that is covered all over the surface. The EDS elemental analysis of fibrous silica $\text{Fe}_3\text{O}_4/\text{MnO}_2/\text{SiO}_2/\text{KCC-1}$ exhibited the presence of Fe, Mn, Si, and O (Fig. 1f), further indicating the formation of the hybrid core-shell magnetic fibrous silica nanocomposite.

The X-ray diffraction (XRD) pattern was obtained to investigate the formation of $\text{Fe}_3\text{O}_4/\text{MnO}_2$ and $\text{Fe}_3\text{O}_4/\text{MnO}_2/\text{SiO}_2/\text{KCC-1}$. As shown in Fig. 2a, the characteristic diffraction peaks of the magnetic nanocomposite located at 30.17° , 35.51° , 43.10° , 53.75° , 57.32° , and 62.89° can be attributed to the (220), (311), (400), (422), (511), and (440) crystal planes, respectively, which represent the inverse spinel structure of Fe_3O_4 nanoparticles (JCPDS card no. 19-0629).³³ Moreover, the well-resolved diffraction peaks at around 28.6° and 37.3° indicated the presence of MnO_2 .³⁴ As displayed in Fig. 2b, the broad diffraction peak at around 22.48° indicated the growth of fibrous silica on the surface of $\text{Fe}_3\text{O}_4/\text{MnO}_2$ nanocomposite.³⁵

FT-IR analysis has been performed on $\text{Fe}_3\text{O}_4/\text{MnO}_2/\text{SiO}_2/\text{KCC-1}$ and the hybrid sensor I. As shown in Fig. 3a, the absorption band at 563.2 cm^{-1} is assigned to the Fe–O–Fe stretching and bending mode, and no characteristic peaks for the amorphous MnO_2 are observed. The vibration peaks of fibrous mesoporous silica at 467.4 cm^{-1} , 809.3 cm^{-1} , and 1092.5 cm^{-1} are assigned to the symmetric and asymmetric stretching of the Si–O–Si bonds. The strong absorption peak at 3426.5 cm^{-1} is assigned to the hydroxyl groups. As shown in Fig. 3b, new bands at 2856.4 cm^{-1} , 2925.9 cm^{-1} , and

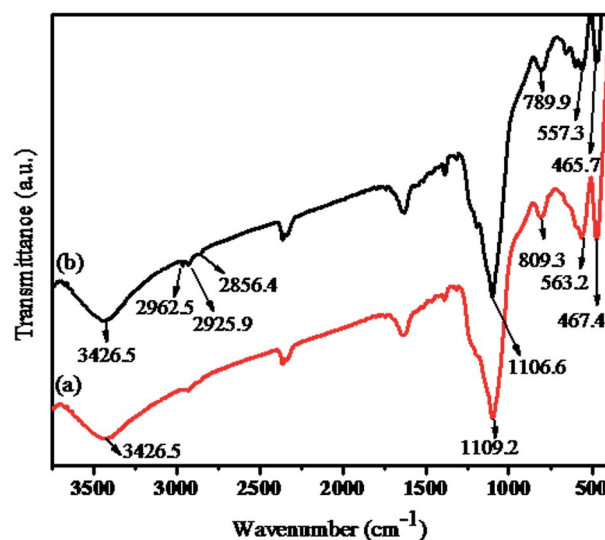


Fig. 3 FT-IR spectra of (a) $\text{Fe}_3\text{O}_4/\text{MnO}_2/\text{SiO}_2/\text{KCC-1}$ and (b) $\text{RhB-Fe}_3\text{O}_4/\text{MnO}_2/\text{SiO}_2/\text{KCC-1}$.

2962.5 cm^{-1} can be assigned to the stretching vibration of methylene and amide group,^{36,37} which confirm that rhodamine B derivatives have been functionalized with hybrid core-shell magnetic fibrous silica nanocomposite.

Nitrogen adsorption/desorption isotherm reveals the surface modification for the hybrid sensor I. Upon investigation, it was observed that the magnetic nanocomposite and rhodamine-functionalized nanocomposite showed a modified textural property and surface area, as shown in Fig. 4. For the $\text{Fe}_3\text{O}_4/\text{MnO}_2/\text{SiO}_2/\text{KCC-1}$ and $\text{RhB-Fe}_3\text{O}_4/\text{MnO}_2/\text{SiO}_2/\text{KCC-1}$ nanocomposite, the BET surface area was 225.08 and $69.25\text{ m}^2\text{ g}^{-1}$, the pore diameter was 7.96 and 6.74 nm , and the pore volume was 0.56 and $0.13\text{ cm}^3\text{ g}^{-1}$, respectively. These results describe the effective functionalization of the organic probe on the surface of the fibrous nanocomposite.

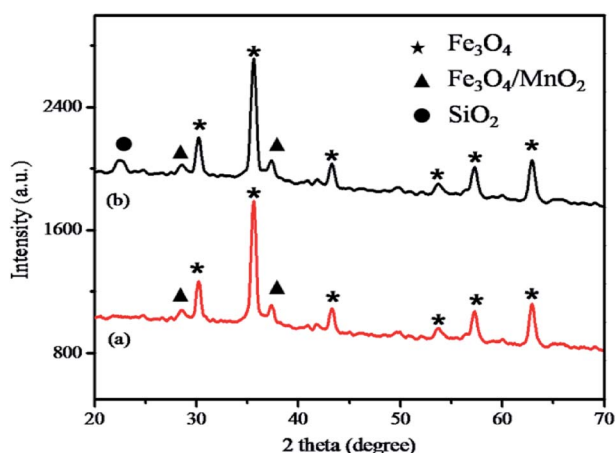


Fig. 2 X-ray diffraction patterns of (a) $\text{Fe}_3\text{O}_4/\text{MnO}_2$ and (b) $\text{Fe}_3\text{O}_4/\text{MnO}_2/\text{SiO}_2/\text{KCC-1}$.

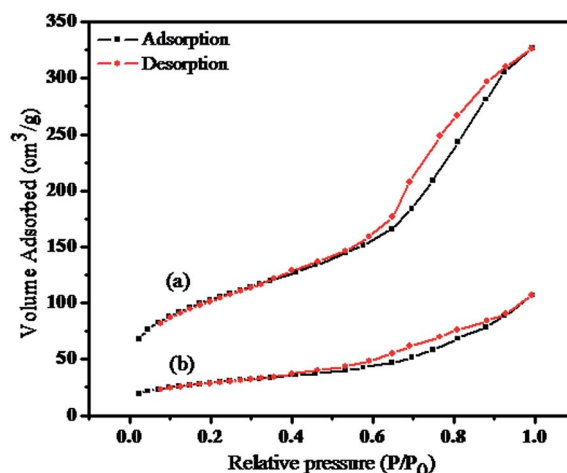


Fig. 4 The nitrogen adsorption-desorption isotherms of (a) $\text{Fe}_3\text{O}_4/\text{MnO}_2/\text{SiO}_2/\text{KCC-1}$ and (b) $\text{RhB-Fe}_3\text{O}_4/\text{MnO}_2/\text{SiO}_2/\text{KCC-1}$.



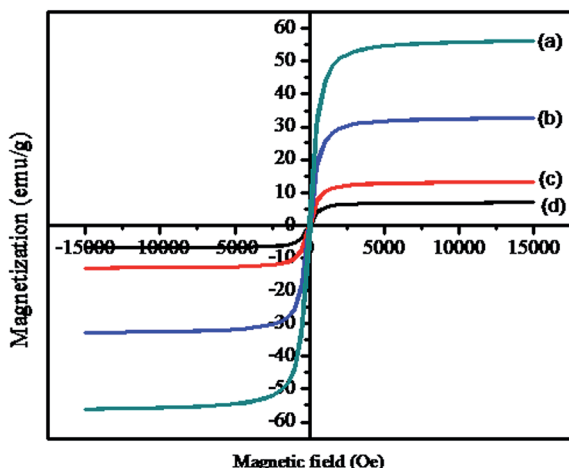


Fig. 5 Magnetization curves of (a) Fe_3O_4 , (b) $\text{Fe}_3\text{O}_4/\text{MnO}_2$, (c) $\text{Fe}_3\text{O}_4/\text{MnO}_2/\text{SiO}_2/\text{KCC-1}$, and (d) $\text{RhB-Fe}_3\text{O}_4/\text{MnO}_2/\text{SiO}_2/\text{KCC-1}$. The inset image shows the magnetic separation behaviour of the hybrid sensor I.

As shown in Fig. 5a–d, the magnetic measurements of pristine Fe_3O_4 , $\text{Fe}_3\text{O}_4/\text{MnO}_2$, $\text{Fe}_3\text{O}_4/\text{MnO}_2/\text{SiO}_2/\text{KCC-1}$, and the hybrid sensor I were 56.09, 32.6, 13.0, and 6.9 emu g^{-1} , respectively. All samples exhibited super paramagnetic behaviour at room temperature. The magnetic saturation value decreased from Fig. 5a to d due to the increase in the diameter of the outer surface of Fe_3O_4 nanoparticles and the presence of the organic molecule. As a result, the functionalized core-shell nanocomposite can rapidly redisperse and prevent the aggregation of nanoparticles. It can be observed that the hybrid sensor I can be homogeneously dispersed in different solvents and rapidly separated within 1 min by an external magnet, as illustrated in the inset of Fig. 5. In addition, the saturated magnetic behaviour of the hybrid sensor I acted as an effective barrier to the quenching fluorophore.

The stability of the hybrid sensor I was determined by TGA measurements in the temperature range from 50 to 615 $^{\circ}\text{C}$ at

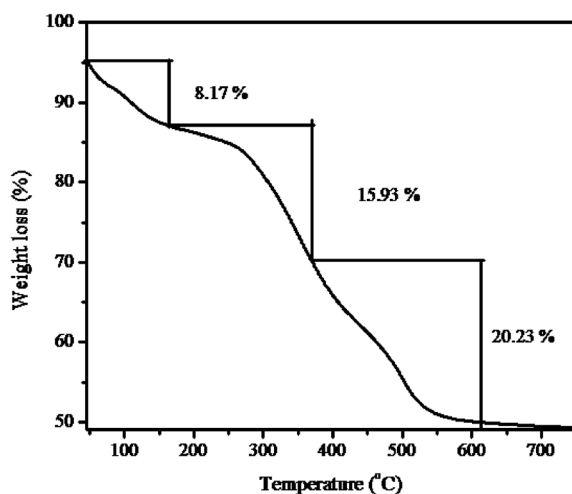


Fig. 6 TGA data of $\text{RhB-Fe}_3\text{O}_4/\text{MnO}_2/\text{SiO}_2/\text{KCC-1}$.

a heating rate of 10 $^{\circ}\text{C min}^{-1}$ under a nitrogen atmosphere. As shown in Fig. 6, the curve exhibits three-step weight loss over the temperature ranges of 50–165 $^{\circ}\text{C}$, 166–367 $^{\circ}\text{C}$, and 368–615 $^{\circ}\text{C}$. The first weight loss of 8.17% was predominantly due to the removal of physically absorbed water and organic solvent. The second weight loss of 15.93% was attributed to the thermal decomposition of functionalized organic compounds. The final weight loss of 20.23% was due to the complete decomposition of the whole nanocomposite. Total weight loss was about 44.86% in the temperature range from 50 to 615 $^{\circ}\text{C}$. The TGA results of the hybrid sensor I demonstrate that it has a superior thermal stability at high temperatures.

3.3. Sensing performance

3.3.1. Absorption response for the detection of Cu(II) ion.

To optimize the sensing ability of the hybrid sensor I, $\text{CH}_3\text{CN}/\text{water}$ solution (1 : 1, v/v) was preferred for UV-vis detection process. After the addition of various metal ions (K(I) , Ag(I) , Ca(II) , Ni(II) , Hg(II) , Pb(II) , Zn(II) , Co(II) , Ba(II) , Fe(II) , and Cd(II)) in the form of their nitrates, only copper ion showed a higher response than other metal ions such as Fe(II) and Co(II) , as shown in Fig. 7c. Upon increasing the concentration of Cu(II) ion in the solution in the range of $0\text{--}5 \times 10^{-4}$ M, the hybrid sensor I turned from colourless to pink with a significant enhancement of a new absorbance peak centred at 556 nm, and a smaller peak at 522 nm was obtained. This proves the selectivity of Cu(II) ion towards the binding site. The significant appearance of colour and the enhancement of absorbance peak indicate the ring-opened amide form of the probe.³⁸ These results indicate that the hybrid sensor I can be used for the rapid determination of Cu(II) by the naked eye. The titration for Cu(II) concentration variation showed a better linear relationship ($R^2 = 0.9838$), as shown in Fig. 7b. The detection limit was found to be 26.40 nM. The data obtained from the Stern–Volmer plot suggested that the synthesized material had a good selectivity for Cu(II) ion.

3.3.2. Emission response for the detection of Cu(II) . The fluorescence spectra of the hybrid sensor I have been used to recognize the sensing behaviour and the molecular interaction towards Cu(II) . To identify the sensitivity of the process, inter-related parameters in the sensing ability were analyzed. The fluorescence titration of the hybrid sensor I (0.1 mg ml^{-1}) in $\text{CH}_3\text{CN}/\text{water}$ (1 : 1, v/v) system was carried out with Cu(II) in the concentration range of $0\text{--}5 \times 10^{-4}$ M. As shown in Fig. 8a, the emission band of the hybrid sensor I (5 μM) exhibited a weak fluorescence at 569 nm (λ_{ex} 520 nm), which was attributed to the close ring spirolactam form of the probe with a photo induced electron-transfer (PET) process that was induced by nitrogen atoms bonded to a hydrogen atom. The fluorescence emission intensity was enhanced around 569 nm with the increase in Cu(II) ion concentration. The rhodamine derivative coordinated with Cu(II) ; this led to the spirolactam ring-untie mechanism,^{39,40} which was associated with a switch on UV-vis spectral response at 556 nm and the emission response at 569 nm.

The Stern–Volmer plot, which expresses the linear relationship between F_0/F values (where F_0 and F indicate the presence and absence of Cu(II) ion at 569 nm) and the Cu(II)



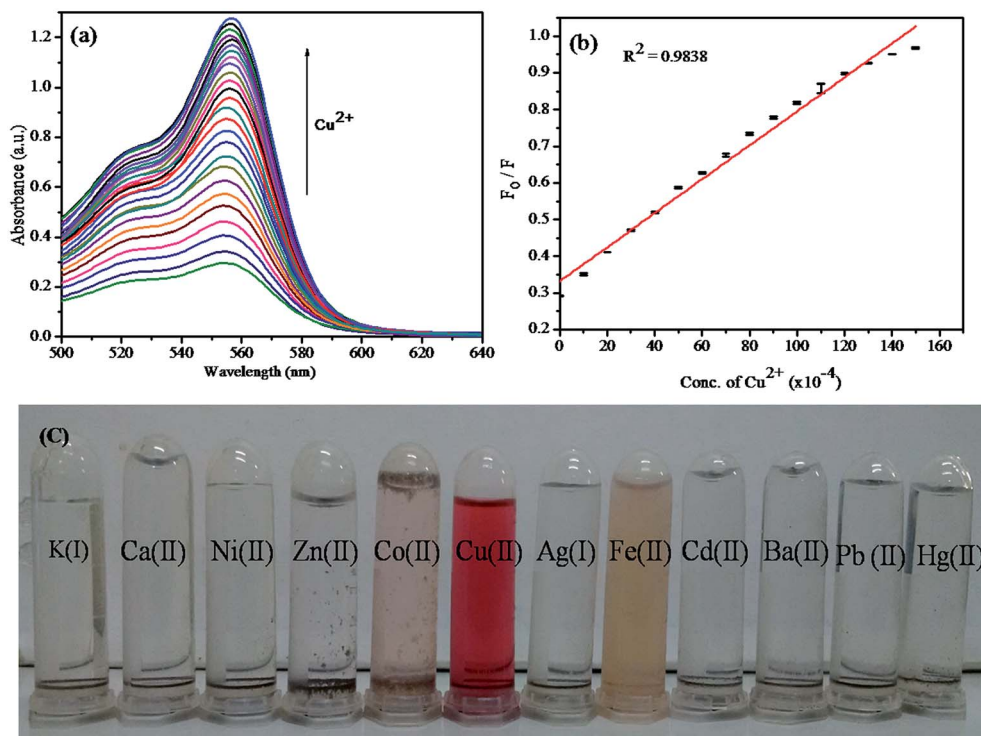


Fig. 7 (a) Absorption spectra of RhB-Fe₃O₄/MnO₂/SiO₂/KCC-1 (0.4 mg ml⁻¹) in CH₃CN/H₂O (1 : 1, v/v) and different concentrations of Cu(II) (0–5 × 10⁻⁴ M). (b) The Stern–Volmer plot between various concentrations of Cu(II) ion and F_0/F values. (c) Image of RhB-Fe₃O₄/MnO₂/SiO₂/KCC-1 with Cu(II) and different metal ions (K(I), Ag(I), Ca(II), Ni(II), Hg(II), Pb(II), Zn(II), Co(II), Al(II), and Fe(II)).

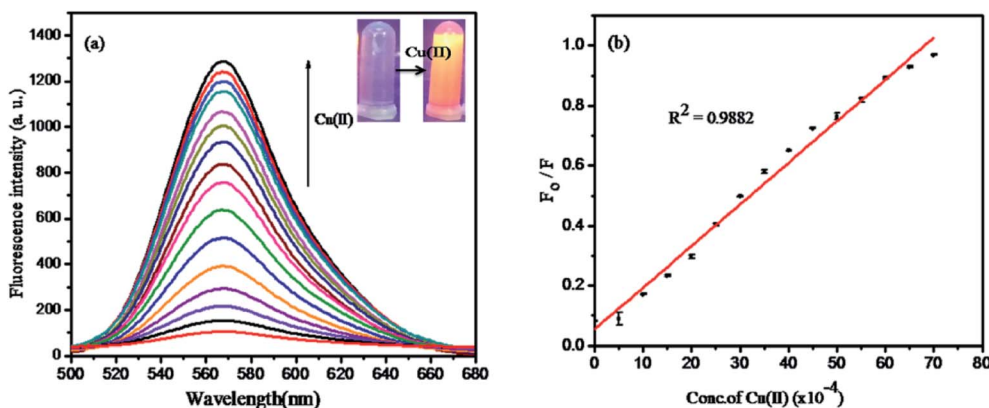


Fig. 8 (a) Fluorescence spectra of RhB-Fe₃O₄/MnO₂/SiO₂-KCC-1 (0.4 mg ml⁻¹) with different concentrations Cu(II) ion in the Tris–HCl solution (10 mM, pH = 7.20). (b) The Stern–Volmer plot between various concentrations of Cu(II) ion and F_0/F values. Error bars represents the standard deviations of the results from three independent experiments.

concentration in the range of 0–5 × 10⁻⁴ M with the correlation coefficient $R^2 = 0.9882$ is shown in Fig. 8b. The detection limit was calculated by the $3\sigma/k$ method (where σ is the standard deviation of the blank measurement and k is the slope) to be about 12.3 × 10⁻⁸ M, which was comparable to that of other methods (Table 1). Moreover, the linear relationship data fitting with Stern–Volmer plot expressed the 1 : 1 complex formation.

3.3.3. Selectivity and coexisting ion effect on the hybrid sensor I. The selectivity of the hybrid sensor I for Cu(II) ion was examined by testing the response of the system to various other

metal ions including K(I), Ag(I), Ca(II), Ni(II), Hg(II), Pb(II), Zn(II), Co(II), Ba(II), Fe(II), and Cd(II) *via* both UV-visible absorption and fluorescence emission spectra under identical condition as used with Cu(II) (Fig. 9a). Compared to the remarkable enhancement of absorption intensity in the fluorescent spectra of the hybrid sensor I induced by Cu(II), no significant change was caused by other metal ions.⁴¹ Consistent with the absorption enhancement, the hybrid sensor I system obviously changed from colourless to pink after the addition of Cu(II). This result clearly indicates that Cu(II) is one of the well-known



Table 1 Comparison between the proposed method and other methods for the detection of Cu(II)

Materials	Detection methods	Working range (M)	LOD	Ref. no.
SBA-15	Colorimetric sensor	$0.01\text{--}1.0 \times 10^{-6}$	0.015×10^{-6}	43
SBA-15	Electrochemical sensor	$8.0\text{--}100.0 \times 10^{-7}$	2.0×10^{-7}	44
SBA-15	Fluorescence chemosensor	$0\text{--}50 \times 10^{-6}$	0.1×10^{-6}	45
HOM-11	Colorimetric sensor	1.57×10^{-8} to 7.86×10^{-6}	3.1×10^{-8}	46
Silica nanoparticles	Colorimetric sensor	$0\text{--}6.0 \times 10^{-4}$	2.13×10^{-6}	47
SBA-15	Optical sensor	—	3.2×10^{-6}	48
KCC-1	Fluorescence chemosensor	$0\text{--}5 \times 10^{-4}$	12.3×10^{-8}	Present work

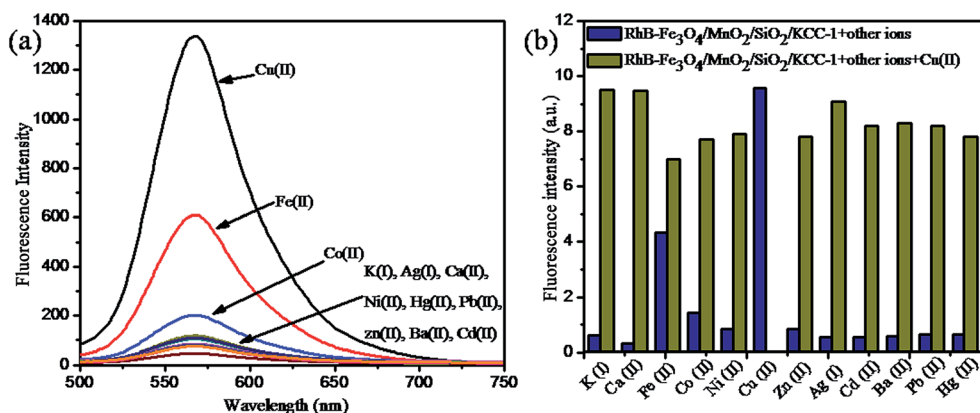


Fig. 9 Fluorescence spectra of (a) RhB-Fe₃O₄/MnO₂/SiO₂/KCC-1 (0.4 mg ml⁻¹) in the Tris-HCl solution (10 mM, pH = 7.2) with different metal ions (5.0×10^{-4} M), (b) the fluorescence responses of RhB-Fe₃O₄/MnO₂/SiO₂/KCC-1 (0.4 mg ml⁻¹) various cations in water (pH = 7.2). The blue bars represent the emission spectra of various interfering ions (5.0×10^{-4} M). The green bars represent the change in emission spectra of upon the addition of Cu(II) (4.0×10^{-4} M). The intensities were recorded at λ_{ex} = 520 nm and λ_{em} = 569 nm.

paramagnetic metal ions. This hybrid sensor I displayed the chelation-enhanced fluorescence (CEFE) that was attributed to the spiro-opening of the bis rhodamine Schiff-base derivative after the complexation with Cu(II) ion, which was induced by the intermolecular electron transfer from the rhodamine group to cationic metal ion.⁴² To further explore the selectivity of the hybrid sensor I towards Cu(II), a competition experiment was conducted *via* the additional mixing of Cu(II) with a solution containing other metal ions (Fig. 9b). The bar graph of the fluorescence emission intensity revealed that Cu(II) fluorescence enhancement of the hybrid sensor I almost remained unaffected in the presence of other metal ions at higher concentrations. Hence, the abovementioned results of the optical experiments indicated that the hybrid sensor I had a high sensitivity and selectivity towards Cu(II) with no interference by other metal ions even at higher concentrations. Thus, our result reveals that the hybrid sensor I is a selective chemosensor for the detection of Cu(II) ion.

3.3.4. Influence of pH values. Usually, the pH values of the solution have a remarkable influence on the fluorescence probe for the detection of target analytes. Thus, the Cu(II) sensing ability of the hybrid sensor I at different pH values was investigated. The pH value was adjusted in the range from 2 to 10, and the fluorescence emission intensities were measured, as shown in Fig. 10. We have observed that the fluorescence

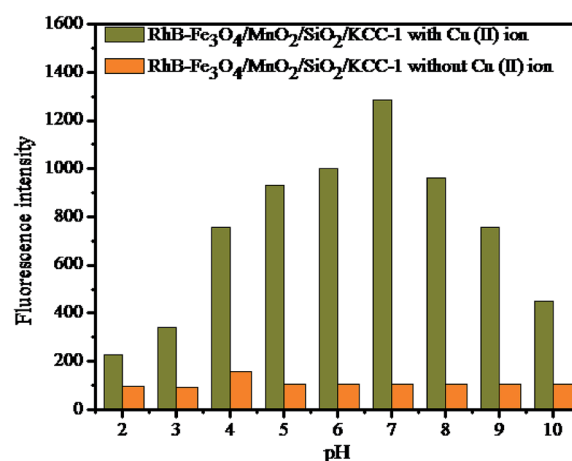


Fig. 10 The fluorescence response of RhB-Fe₃O₄/MnO₂/SiO₂/KCC-1 in CH₃CN/H₂O as a function of pH.

intensity of the hybrid sensor I increases with the increasing pH value because Cu(II) has a high affinity to chelate with fluorophoric carbonyl (O), imino (N), and phenol (O) atoms to form a ring-opened amide form to enhance the fluorescence intensity. When the pH values changed from 5 to 8, the dissociation reaction of amino groups played a key role on the surface of the



hybrid sensor I. Thus, the strong fluorescence response was observed when $\text{pH} < 5$ because the deprotonated imine, carbonyl, and phenol groups were protonated, and the resulting ligands were separated from the surface of the hybrid sensor I. The results indicate a weak emission response in an acidic medium that can be attributed to the poor binding response of Cu(II) because the sensing probes tend to protonate. At a higher pH in the range from 5 to 8, the amino groups tend to deprotonate; this create a feasible environment for the detection of Cu(II) ion. Taking this type of pH influence behaviour into account, the practical sensing in a $\text{CH}_3\text{CN}/\text{water}$ solution (1 : 1, v/v) with a buffer solution is highly accurate and reliable.

3.3.5. Recycling of the hybrid sensor I. The recycling performance of the hybrid sensor I after the addition of Cu(II) (5.0×10^{-4} M) ions was investigated. As shown in Fig. 11A–D, the absence of Cu(II) ion in the hybrid sensor I caused no change in the fluorescence enhancement (bar graph A). After the addition of Cu(II) ion to the abovementioned sensing system, the fluorescence intensity dramatically increased (bar graph B) and fully recovered upon the addition of EDTA (5.0×10^{-4} M) (bar graph C). When Cu(II) was added in excess, the fluorescence emission intensity reached the saturation level again (bar graph D). Further subsequent addition of Cu(II) ion showed no

obvious change in the fluorescence emission. These results demonstrated that the hybrid sensor I had a superiority of recycling performance for the detection of Cu(II) ion.

3.3.6. Analysis of the environmental water sample. To identify the performance of the abovementioned chemosensor, we attempted to detect the concentration of Cu(II) in the environmental water samples. The water samples (including tap water, pond water, and industrial waste water) obtained near the SRM University campus were filtered to remove the solid impurity, and pH was adjusted to 7.2 using the Tris–HCl buffer for effective analysis in a neutral medium. The desired amount of Cu(II) ion was spiked (2, 4, 6, and 8 μM) into water samples followed by the addition of the hybrid sensor I to each sample. The obtained results are summarized in Table 2. They prove that this sensing system is able to determine Cu(II) in environmental water samples.

4. Conclusion

We have successfully prepared a hybrid magnetic core–shell fibrous silica nanocomposite and covalently functionalized it with a rhodamine-based derivative. The rhodamine-based derivative was modified with nitrogen and oxygen atoms, which strongly bonded to Cu(II) ion, playing a critical role in the selective detection of Cu(II) ion. This hybrid sensor exhibited significantly improved sensing ability towards Cu(II) ion with high sensitivity and selective fluorescence responses in an aqueous solution. The detection limit was found to be 12.3×10^{-8} M, which was superior to that of the previously reported Cu(II) detection methods. In addition, this hybrid sensing system recovered 95% original fluorescence sensing signal for the detection of Cu(II) ion. The abovementioned results demonstrated the successful use of this nanocomposite for the detection of Cu(II) in the environmental real samples. Thus, this simple and reliable hybrid sensor can be used for Cu(II) ion detection in aqueous solutions.

Conflicts of interest

There are no conflicts to declare.

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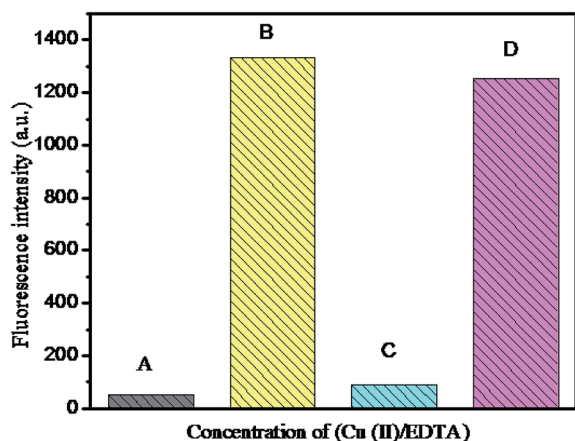


Fig. 11 Fluorescence spectra of $\text{RhB-Fe}_3\text{O}_4/\text{MnO}_2/\text{SiO}_2/\text{KCC-1}$ (0.4 mg mL^{-1}) in the Tris–HCl solution (10 mM , $\text{pH} = 7.2$). In bar graph (A) $\text{RhB-Fe}_3\text{O}_4/\text{MnO}_2/\text{SiO}_2/\text{KCC-1}$, (B) $\text{RhB-Fe}_3\text{O}_4/\text{MnO}_2/\text{SiO}_2/\text{KCC-1} + \text{Cu(II)}$ (5.0×10^{-4}), (C) $\text{RhB-Fe}_3\text{O}_4/\text{MnO}_2/\text{SiO}_2/\text{KCC-1} + \text{Cu(II)}$ (5.0×10^{-4}) + EDTA (5.0×10^{-4} M), and (D) $\text{RhB-Fe}_3\text{O}_4/\text{MnO}_2/\text{SiO}_2/\text{KCC-1} + \text{Cu(II)}$ (10×10^{-4}) + EDTA.

Table 2 Determination of Cu(II) ion in the environmental water samples

Samples	Spiked $[\text{Cu(II)}]/(\mu\text{M})$	Found $\text{Cu(II)}/(\mu\text{M})$	Recovery (%)	RSD (%)
Tap water	2	1.99	99.5	1.24
River water	4	3.98	99.5	2.23
Pond water	6	6.05	101	1.03
Industrial waste water	8	3.99	99.8	1.99



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