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Preparation of magnetic mesoporous carbon from polystyrene-grafted magnetic nanoparticles for rapid extraction of chlorophenols from water samples†

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A magnetic mesoporous carbon material ($\text{Fe}_3\text{O}_4@\text{C}$) was fabricated by carbonizing polystyrene grafted polydopamine-coated magnetic nanoparticles. The chemical composition, morphology and magnetism of the mesoporous carbon materials were characterized. The as-prepared magnetic mesoporous carbon was employed for enrichment of chlorophenols from water samples coupled with high-performance liquid chromatography with UV detection (HPLC-UV). Various parameters affecting the enrichment were investigated. A good linearity was obtained in the range of $10\text{--}200\text{ ng mL}^{-1}$ for 2-CP, 4-CP and 2,4-DCP, and $5\text{--}200\text{ ng mL}^{-1}$ for 2,4,6-TCP. The recoveries of CPs were in the range of 84.2–120% with intra- and inter-relative standard deviations (RSD) lower than 11.0%. The proposed method is fast, convenient and environmentally friendly. The results revealed the suitability of $\text{Fe}_3\text{O}_4@\text{C}$ nanoparticles as adsorbents for enrichment of CPs from environmental water samples.

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1. Introduction

Chlorophenols (CPs) as important raw materials are widely used in the production of petrochemicals, resins, pharmaceuticals, paints, herbicides and insecticides.¹ During the production and application procedures, chlorophenols are introduced into the environment, inevitably. Owing to its toxic, biorefractory, and bioaccumulative characteristics, the United States Environmental Protection Agency has listed most CPs as priority pollutants.^{2,3} Therefore, it is very necessary to develop convenient and effective analytical methods for determination of these compounds in environment samples.

Because of the low concentrations of CPs in environmental samples, extraction and preconcentration processes are usually required. Several pre-treatment methods for enrichment of CPs have been reported, including solid-phase-extraction (SPE),^{4–6} solid phase microextraction (SPME),^{7–9} liquid–liquid–liquid–microextraction (LLME),^{10,11} and stir bar sorptive extraction (SBSE).¹² Among them, the sorbent-based techniques, mainly SPE, SPME and magnetic solid-phase-extraction (MSPE), have become widely used for the enrichment of chlorophenols from environmental samples owing to the variety of the adsorbent

materials.^{13,14} Compared with traditional SPE, MSPE can greatly simplify the phase separation procedure and increase the contact areas between analytes and adsorbents, which increase the mass transfer rate and extraction equilibrium. In the MSPE procedure, the magnetic sorbent is essential, which determines the extraction efficiencies of analytes. So, the exploration of new types of the MSPE adsorbents to improve the enrichment capability for different kinds of analytes has become an active field in analytical chemistry.¹⁵

With the properties of high surface area, tunable surface chemistry, and high chemical and physical stability,¹⁶ nanoporous carbon materials are currently finding its wide application in many fields including adsorbents, drug deliver, electrode materials, and catalyst supports.^{17–20} To prepare the nanoporous carbon materials, the most common and effective routes are template methods, including soft- and hard-template.²¹ There are some drawbacks with both methods. The hard-template method often involves silica as hard-template, which requires harsh chemical treatment, such as HF etch, to remove the sacrificial silica after carbonization.²² By contrast, the soft-template method which use of block copolymers as precursor, remove the template by thermal decomposition prior to carbonization. But the interactions between soft templates and carbon sources are usually too weak to prepare well-defined mesostructures.²³ And both methods are complicated, expensive.²⁴

Dopamine molecules (DA) can self-polymerize to form polydopamine (PDA) under alkaline conditions.²⁵ Owing to the adhesion of PDA, PDA can be easily coated on the surface of

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substrates. Such properties can allow PDA to serve as functional layer, which contains catechol and amine function groups. This provides a route to modify the surface chemistry of materials. In addition, PDA has been proved as promising carbon precursors, owing to the nontoxic and good carbon yield.²⁶ Surface-initiated atom transfer radical polymerization (SI-ATRP) has been explored as one of the most successful controlled/living radical polymerization techniques due to its mild reaction conditions, excellent tolerance to functional groups and impurities, and precise control on molecular weight.²⁷ It also has been proved to be an excellent tool to introduce organic (co) polymers with precise molar mass, composition, and functionality onto nanoparticles.^{28,29} Recently, organic polymer chains grafted from particles by using SI-ATRP have been used as carbon precursors to prepare mesocarbon materials.^{22,23}

Herein, we fabricated carbon precursors by using ATRP technology to graft polystyrene from polydopamine-encapsulated magnetic nanoparticles. After carbonization, the magnetic mesoporous carbon materials were obtained. The prepared magnetic carbon material was applied as adsorbent for the enrichment and analysis of chlorophenols from environmental aqueous samples.

2. Materials and methods

2.1 Chemical and materials

Anhydrous ferric chloride, anhydrous sodium acetate, trisodium citrate dihydrate, 2-chlorophenol (2-CP), 4-chlorophenol (4-CP) were obtained from Sinopharm Chemical Reagent Co., Ltd. (Shanghai, China). Dopamine hydrochloride (98%) was purchased from Macklin Biochemical Co., Ltd. (Shanghai, China). 2-Bromoisobutyl bromide (2-BiBB), 2,4-dichlorophenol (2,4-DCP), 2,4,6-trichlorophenol (2,4,6-TCP), copper(i) bromide (>98%) and 2,2-bipyridyl (Bpy) (>99.5%) were all supplied from Aladdin Chemistry Co., Ltd (Shanghai, China). Styrene (St) was obtained from Tianli Chemical Reagent Co., Ltd. (Tianjin, China), and was passed over a column of basic alumina prior to use. Acetonitrile (HPLC grade) was purchased from Fisher (USA). HPLC grade methanol and acetic acid were obtained from Kermel Chemical Reagent Co. (Tianjin, China). Other reagents were of analytical grade.

Stock solutions of CPs were prepared in methanol, with concentration levels of 1 mg mL⁻¹ for each compound, and were stored in a freezer at 4 °C. Working solutions were freshly diluted with 20 mM NaAC-HAC solution at given concentrations. The tap water was collected from our laboratory, the river water from Ju River in Chang'an District, Xi'an. The water samples were stored at 4 °C and filtered using 0.45 µm Nylon membranes before use.

2.2 Instrumentation

The Fourier-transform infrared spectrometer (FT-IR, TENSOR27, Bruker, Germany) was used to determine the functional group of nanoparticles surface. The X-ray photoelectron spectroscopy (XPS) analysis (K-Alpha Thermo Fisher Scientific) was used to determine the contents of the elements

and chemical state in the nanospheres. The morphology and size were observed by scanning electron microscopy (SEM, Quanta 600FEG, America) and transmission electron microscopy (TEM, H-600, 75 kV, Hitachi, Japan). The magnetic properties were determined *via* a SQUID magnetometer (MPMS-XL-7, Quantum Design, USA). Powder X-ray diffraction (XRD) measurements were made with CuKα radiation on a D8 ADVANC (Bruker, Germany). Brunauer-Emmett-Teller (BET) specific surface measurements were performed using ASAP 2460 (Micromeritics, USA).

2.3 Synthesis of Fe₃O₄@PDA

Magnetic Fe₃O₄ nanoparticles were prepared *via* solvothermal method following the reference with minor modification.³⁰ Briefly, 1.35 g of FeCl₃ and 0.45 g of trisodium citrate dehydrate were added in 30 mL of ethylene glycol. The mixture was heated to dissolve the solid. After cooling to room temperature, 2.40 g of sodium acetate was added with magnetic stirring for 30 min. The mixture was sealed in a Teflon-lined stainless-steel autoclave. After reaction for 12 h at 200 °C, the autoclave was cooled down to room temperature. The resulting Fe₃O₄ nanoparticles (NPs) were washed with ethanol and distilled water several times respectively, and then dried under vacuum at 45 °C.

The Fe₃O₄@PDA core-shell NPs were fabricated according to the literature with minor modification.³¹ 200 mg Fe₃O₄ was dispersed in 100 mL of Tris-HCl solution (10 mM, pH 8.5) under ultrasonication for 30 min, followed by addition of dopamine hydrochloride (200 mg). After mechanical stirring at room temperature for 24 h, Fe₃O₄@PDA NPs were collected by magnetic separation and washed with distilled water and ethanol respectively. The product was dried under vacuum.

2.4 Immobilization of initiator on Fe₃O₄@PDA NPs

1.0 g of Fe₃O₄@PDA NPs were dispersed in 20 mL of anhydrous tetrahydrofuran (THF) and sonicated for 5 min. After stirred the mixture for 30 min in an ice bath, 2 mL of triethylamine (TEA) and 2 mL of 2-BiBB were drop-wise added into the solution and maintained in an ice bath for 3 h under vigorous stirring. The reaction was left at 35 °C for another 12 h. Afterwards, the obtained initiator-functionalized nanospheres, denoted as Fe₃O₄@PDA-Br, were extensively rinsed with THF, methanol and water in sequence.

2.5 Synthesis of Fe₃O₄@PDA@St by ATRP

0.5 g of Fe₃O₄@PDA-Br nanospheres were dispersed in a mixture of 2.4 mL of styrene (21 mmol), 66.5 mg of 2,2-bipyridyl (0.42 mmol) and 10 mL anisole in a 50 mL three-neck flask. After the mixture was deoxygenated *via* two freeze-pump-thaw cycles, CuBr (30.5 mg, 0.21 mmol) was quickly added to the flask under an nitrogen atmosphere. After two freeze-pump-thaw cycles, the polymerization was allowed to carry out at 90 °C for 16 h with continuous stirring. The resulting nanospheres were extensively washed with anisole, methanol and water in sequence. To remove the residual catalyst completely, the obtained nanoparticles were re-dispersed into a mixture of methanol and 0.25 M EDTANa₂ solution (1/1, v/v), and the



mixture was stirred at 40 °C for 4 h. After magnetic separation, the nanoparticles were washed with water, methanol and dried at 45 °C under vacuum.

2.6 Synthesis of Fe₃O₄@C

Fe₃O₄@C was obtained by carbonizing the as-synthesized Fe₃O₄@PDA@St in an N₂ atmosphere. The Fe₃O₄@PDA@St NPs were heated from ambient temperature to 600 °C with a heating rate of 3 °C min⁻¹, and was maintained at this temperature for 1 h. The whole preparation process is shown in Fig. 1.

2.7 MSPE procedures

20 mg of Fe₃O₄@C was added to a 20 mL sample solution. The mixture was shaken for 1 min. Subsequently, the adsorbent was separated from the sample solution using a magnet, and the supernatant was discarded. Next, 0.6 mL of alkaline methanol was added into the centrifuge tube to elute the chlorophenols by shaking for 1 min. The desorption solution was also separated under a magnet, then the elute solution was adjusted to neutral by HCl.

2.8 Analytical conditions

The HPLC analyses were conducted on a LC-20A system from Shimadzu (Tokyo, Japan), equipped with two LC-20AD pumps, a SPD-20A ultraviolet detector, an injector with a 20 µL sample loop and a CTO-20AC column oven. Chromatographic separations were performed on an Agilent TC-C18 (150 mm × 4.6 mm i.d., 5 µm) column. The mobile phase consisted of acetonitrile/water containing 1% acetic acid (20/80, v/v, solvent A) and acetonitrile containing 1% acetic acid (solvent B). The gradient elution program was as follows: started at 20% B and kept for 10 min, then increased B to 70% in 10 min and kept for 5 min. Finally the mobile phase was returned to initial condition. The flow rate was set at 1 mL min⁻¹. The UV monitoring wavelength was chosen at 285 nm and the column oven temperature was maintained at 25 °C. The injection volume was 20 µL.

3. Results and discussion

3.1 Synthesis and characterizations of material

In this study, polystyrene chains were grafted from magnetic nanoparticles *via* SI-ATRP, and the polymer chains were then taken as carbon resource to fabricate the Fe₃O₄@C at a carbonization temperature of 600 °C.

The morphology of the adsorbent was observed by SEM and TEM. As can be seen from the SEM (Fig. 2A) and TEM (Fig. 2C) images, the Fe₃O₄ NPs are spherical with an average diameter of 220 nm and have a coarse surface. The SEM (Fig. 2B) and TEM (Fig. 2D) images of Fe₃O₄@C show that the resulting microspheres are clearly core-shell structure, with a thickness of carbon coating ~40 nm.

The functional group of Fe₃O₄, Fe₃O₄@PDA, Fe₃O₄@PDA@St, and Fe₃O₄@C were investigated with FT-IR. As shown in Fig. 3A, the peak at 590 cm⁻¹ (Fig. 3A(a-d)) is assigned to the vibration of Fe-O. Compared with Fe₃O₄ (Fig. 3A(a)), the IR spectrum of Fe₃O₄@PDA (Fig. 3A(b)) shows the characteristic adsorption of the stretching of C-O bond at 1291 cm⁻¹, the stretching vibration of the aromatic rings at 1620 cm⁻¹ and the O-H stretching vibration and N-H stretching vibration at 3380 cm⁻¹ from PDA. After grafting St (curve c in Fig. 3A), new bands appeared at 1442 cm⁻¹ and 1110 cm⁻¹, which are ascribed to the stretching vibration of C=C and in-plane bending vibrations of C-H from aromatic rings. After carbonization (Fig. 3A(d)), most of the characteristic peaks of organic groups are greatly weakened, which confirm that the polymer is successfully carbonized.

Surface chemical compositions of the Fe₃O₄@PDA, Fe₃O₄@PDA@St, and Fe₃O₄@C are further investigated by XPS. The survey spectrum (Fig. 3B) shows the two main peaks for C 1s and O 1s, and the content of the each element is showed in the Table 1. From Table 1, the content of C element increases after carbonization, while the content of O element decreases. For high-resolution C 1s spectra of Fe₃O₄@C in Fig. 3C, there are three peaks at around 284.4 eV, 285.8 eV, 289.1 eV, which are

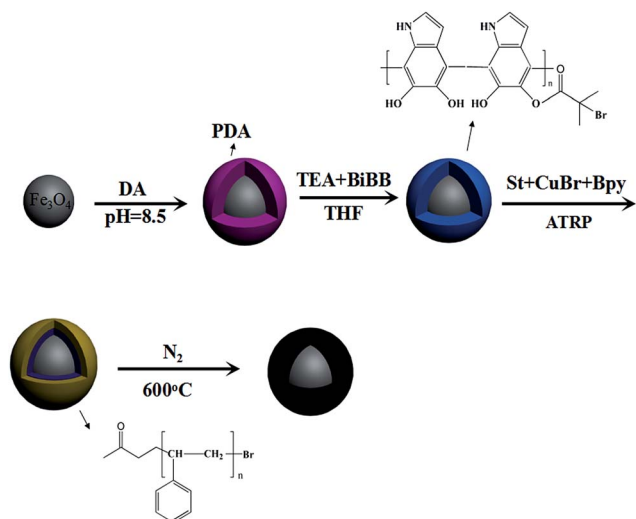


Fig. 1 Schematic illustration of the preparation of Fe₃O₄@C.

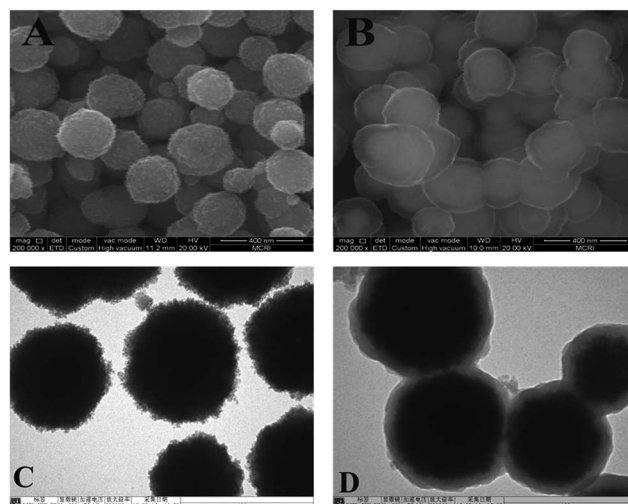


Fig. 2 SEM images of Fe₃O₄ (A), Fe₃O₄@C (B); TEM images of Fe₃O₄ (C), Fe₃O₄@C (D).



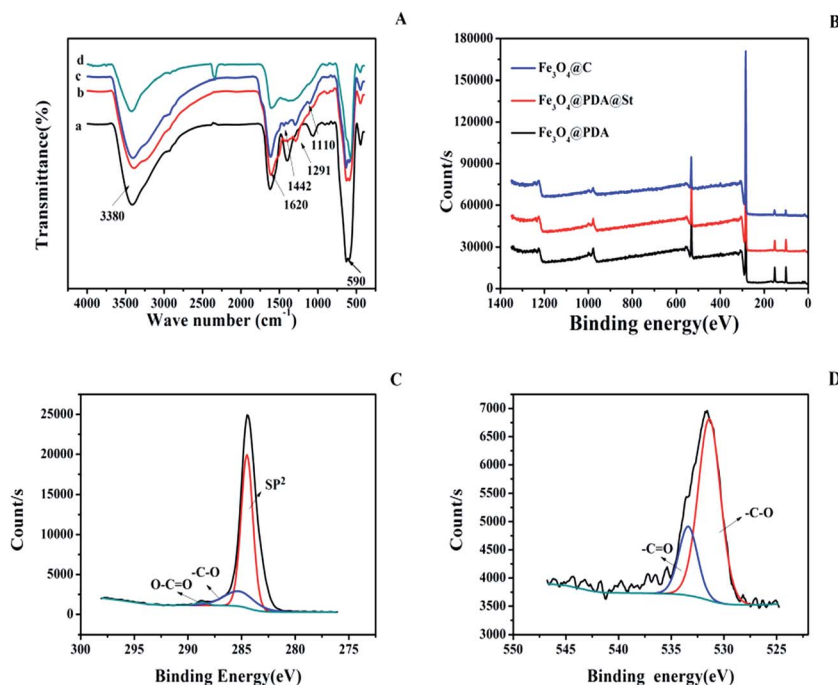


Fig. 3 FT-IR spectrum (A) of Fe_3O_4 (a), $\text{Fe}_3\text{O}_4\text{@PDA}$ (b), $\text{Fe}_3\text{O}_4\text{@PDA@St}$ (c) and $\text{Fe}_3\text{O}_4\text{@C}$ (d); XPS survey spectrum (B) of $\text{Fe}_3\text{O}_4\text{@PDA}$, $\text{Fe}_3\text{O}_4\text{@PDA@St}$, and $\text{Fe}_3\text{O}_4\text{@C}$; high-resolution XPS spectrum of C 1s (C), and O 1s (D) of $\text{Fe}_3\text{O}_4\text{@C}$.

Table 1 Elemental content from XPS analysis

	C 1s (at%)	N 1s (at%)	O 1s (at%)
$\text{Fe}_3\text{O}_4\text{@PDA}$	75.18	1.32	14.40
$\text{Fe}_3\text{O}_4\text{@PDA@St}$	80.82	1.22	11.60
$\text{Fe}_3\text{O}_4\text{@C}$	87.24	1.49	7.73

ascribed to $-\text{C}=\text{C}$, $-\text{C}-\text{O}$ and $-\text{C}=\text{O}$, respectively.³² From the high-resolution XPS spectra of O 1s for $\text{Fe}_3\text{O}_4\text{@C}$ (Fig. 3D), the binding energy peak at 531.6 eV belongs to $-\text{C}-\text{O}$ and the peak at 533.7 eV belongs to $-\text{C}=\text{O}$ can be found.³³

The magnetic properties of Fe_3O_4 and $\text{Fe}_3\text{O}_4\text{@C}$ were measured *via* SQUID magnetometer at 300 K (Fig. S1A†). The saturated magnetization values of Fe_3O_4 , $\text{Fe}_3\text{O}_4\text{@PDA}$, $\text{Fe}_3\text{O}_4\text{@PDA@St}$ and $\text{Fe}_3\text{O}_4\text{@C}$ are 26.7, 21.2, 18.5 and 26.5 emu g^{-1} , respectively. It indicated that after polydopamine coating and polystyrene grafting, the magnetism of the nanoparticles decreased gradually, and then increased after carbonization. The fact that the $\text{Fe}_3\text{O}_4\text{@C}$ has stronger magnetism than $\text{Fe}_3\text{O}_4\text{@PDA@St}$ maybe due to the $\text{Fe}_3\text{O}_4\text{@C}$ nanoparticles prepared at 600 °C exhibit ferromagnetic properties (Fig. S1A,† insert).³⁴ Fig. S1B† shows the XRD pattern of Fe_3O_4 and $\text{Fe}_3\text{O}_4\text{@C}$. All of the diffraction peaks of $\text{Fe}_3\text{O}_4\text{@C}$ are the same as Fe_3O_4 , indicating that no new phase formed during the subsequent modification process.

The nitrogen adsorption/desorption isotherm (77 K) and pore size distribution curve (Fig. S2†) indicate that the BET surface area of $\text{Fe}_3\text{O}_4\text{@C}$ is $8.78 \text{ m}^2 \text{ g}^{-1}$; the pore volume is $0.0171 \text{ cm}^3 \text{ g}^{-1}$ and the pore size is 7.73 nm. The adsorption/desorption isotherms are type IV³⁵ with H4 hysteresis loops, which indicate the carbon materials is mesoporous.

3.2 Optimization of MSPE conditions

The prepared magnetic carbon materials were then employed as an adsorbent for enrichment of four chlorophenols (2-CP, 4-CP, 2,4-DCP and 2,4,6-TCP). Principally, hydrophobic interactions and $\pi-\pi$ interactions should be involved during adsorption. To achieve the maximal extraction efficiency, several important parameters, such as solution pH, type and volume of elution solvent, amounts of adsorbent, salt concentration, adsorption time and elution time were optimized.

3.2.1 Effect of pH. The pH value is a key factory in the extraction process because it affects the existing form of target analytes. In this study, the pH values of sample solution (3.0–9.0) were tested. As shown in Fig. 4A, the extraction efficiencies for all four CPs were increased slightly when the pH values were increased from 3.0 to 5.0. And then declined when the pH values was further increased. The reason for this can be explained as follows. The chlorophenols are acidic compounds, and the existing form of the CPs in low pH values is molecule form, while in high pH values the existing form of CPs is ionic form. On one hand, the $\text{Fe}_3\text{O}_4\text{@C}$ is a hydrophobic material, which has a high affinity toward the chlorophenols in their molecule forms. On the other hand, the surface of $\text{Fe}_3\text{O}_4\text{@C}$ has negative charge with the increase of the pH values.³⁶ As the result of electrostatic repulsion, the interactions between the adsorbent and analytes are greatly weakened. Finally, the solution pH at 5.0 was chosen for subsequent experiments.

3.2.2 Effect of salt concentration. In order to study the effect of the ionic strength on the extraction efficiency, a variable amount of NaCl [0% to 20% (w/v)] was added to the sample solution. As shown in Fig. S3,† the best extraction efficiency could be attained at 10% NaCl. The reason can be explained



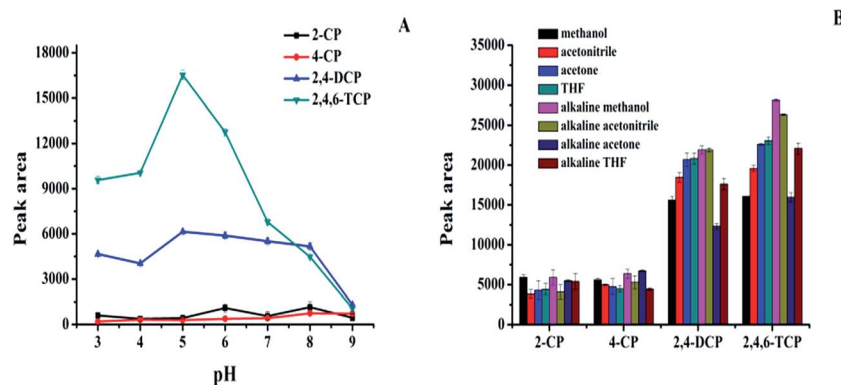


Fig. 4 Effect of several parameters for adsorption of chlorophenols. Sample solution pH (A); desorption solvents (B); CPs concentrations are all at $1.0 \mu\text{g mL}^{-1}$. Sample volume, 10 mL; amount of adsorbent, 5 mg.

that the addition of NaCl could reduce the solubility of chlorophenols and then improve the extraction efficiency. Thus, 10% (m/v) NaCl was added in sample solution for the extraction of target chlorophenols.

3.2.3 Effect of type of desorption solution. To ensure the effective desorption of the chlorophenols analytes from sorbent, in this work, eight kinds of solvents were investigated as desorption solution, including pure methanol, acetonitrile, acetone, THF, and alkaline methanol, alkaline acetonitrile, alkaline acetone, alkaline THF. All the volume of elution solutions was fixed at 1 mL. The results showed that the best desorption efficiency was achieved when alkaline methanol was used (Fig. 4B), which could be attributed to that in alkaline conditions, the chlorophenols could be converted into ionized form and the ionized chlorophenols would have weakened affinity to the adsorbent. Therefore alkaline methanol was used to elute the four CPs.

3.2.4 Effect of extraction time and desorption time. The effect of extraction time (Fig. S4A†) was investigated in the range of 1–60 min. The results showed that the extraction efficiencies for 2-CP and 4-CP kept almost unchanged. However, the extraction efficiencies for 2,4-DCP and 2,4,6-TCP were decreased with the time increase from 1 to 30 min, and

remained constant after that. The reason for this may be attributed to the competing mass-transfer process of the analytes.³⁷ Thus, the extraction time was 1 min. Desorption time was also studied. The obtained results were shown in Fig. S4B,† it can be found that 1 min was enough to elute all CPs.

3.2.5 Effect of the adsorbent amount. The adsorbent amount had a significant effect on extraction efficiency. In this work, the effect of adsorbent amounts was investigated in the range of 2–30 mg. The results shown in Fig. 5A indicated that the extraction efficiencies of the four CPs increased with the increase of the adsorbent dosage from 2 to 20 mg, and then remained almost unchanged with the further increase of the amount of the adsorbent. Hence, 20 mg of $\text{Fe}_3\text{O}_4@\text{C}$ was employed in the following experiment.

3.2.6 Effect of the volume of elution solution. The volume of desorption solution was optimized in the range of 0.2 to 1.0 mL. As shown in Fig. 5B, relatively good elution results could be achieved when 0.6 mL of desorption solvent was used.

3.3 Validations of the method

Under the optimal experimental conditions, the linearity, limit of detection (LOD) and limit of quotation (LOQ) were evaluated. All the parallel experiments were repeated three times and the

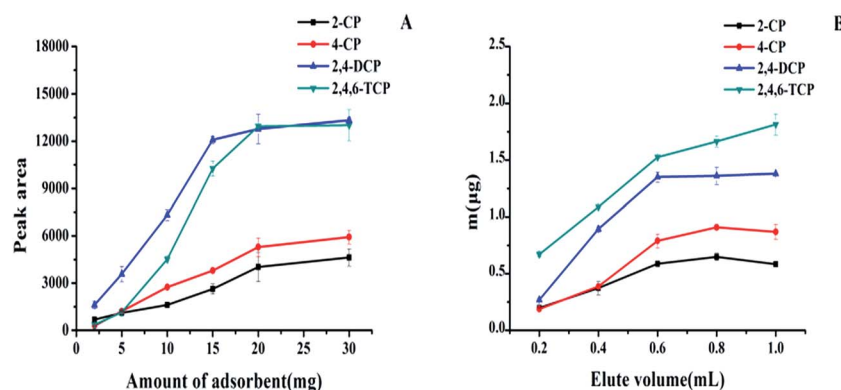
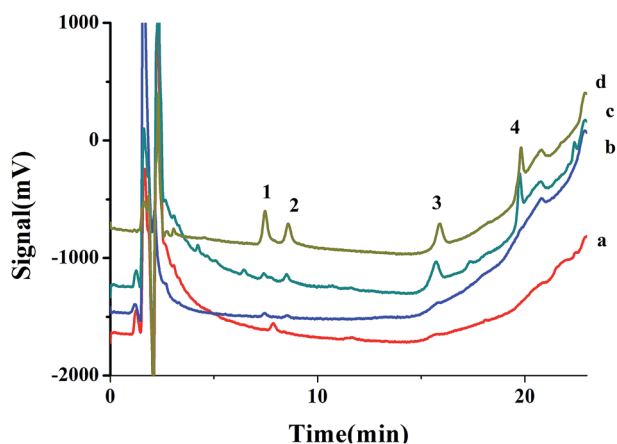


Fig. 5 Effect of the amount of adsorbent (A) and volume of elution solution (B) on the extraction of CPs. CPs concentrations are all at 200 ng mL^{-1} . Salt concentration, 10% (m/v); pH values, 5; extraction and desorption time, 1 min; sample volume, 20 mL; elution solution, alkaline methanol.



Table 2 The linearity, LOD and LOQ of the method for the determination of CPs

Analytes	Calibration curve	<i>r</i>	Linear range (ng mL ⁻¹)	LOD (ng mL ⁻¹)	LOQ (ng mL ⁻¹)
2-CP	$y = 15.951x + 223.39$	0.9935	10–200	1.41	4.70
4-CP	$y = 21.836x + 107.7$	0.9913	10–200	0.88	2.95
2,4-DCP	$y = 64.155x + 775.43$	0.9971	10–200	2.50	8.34
2,4,6-TCP	$y = 66.827x + 11.061$	0.9987	5–200	1.32	4.40

**Fig. 6** Chromatograms of (a) blank river water after extraction, (b) river water spike with CPs (50 ng mL⁻¹ for each CPs), (c) river water spike with CPs at 500 ng mL⁻¹ after extraction, and (d) standard solution of CPs at 500 ng mL⁻¹. Peaks: (1) 2-CP, (2) 4-CP, (3) 2,4-DCP, (4) 2,4,6-TCP.

results were listed in Table 2. A good linearity was obtained in the range of 10–200 ng mL⁻¹ for 2-CP, 4-CP, and 2,4-DCP, and 5–200 ng mL⁻¹ for 2,4,6-TCP, with correlation coefficient (*r*) ranging from 0.9913 to 0.9987. The LODs were ranging from

0.88–2.50 ng mL⁻¹ (LOD = $3S_b/S$, S_b is the SD of three replicated runs of spiked samples at the lowest concentration, S is the slope of calibration curve). The LOQs were ranging from 2.95–8.34 ng mL⁻¹ (LOQ = $10 S_b/S$).

3.4 Real sample analysis

The method was applied for extraction chlorophenols from tap water and river water under optimized conditions. The chromatograms of river water sample are shown in Fig. 6. The analytical figure merits are shown in Table 3, none of four CPs was found in tap and river water. The recoveries of chlorophenols spiked at different concentration were ranging from 84.2–120% with RSDs less than 11.0%, which indicated that the method is applicable for real sample analyses.

3.5 Compare with other methods

The comparison of different methods for the determination of CPs is displayed in Table 4. Compared with other methods, the recovery of current method is comparable, while the sample preparation time (only 2 min) of the proposed method are greatly shorter than the previous method,^{6,38–41} which indicates that the extraction procedure is greatly fast. In addition, the

Table 3 Recovery and precision of four CPs in real water sample^a

Compound	Added (ng mL ⁻¹)	Recovery (%)		Precision (RSD, %)			
				Tap water		River water	
		Tap water	River water	Tap water	River water	Tap water	River water
		(<i>n</i> = 4)	(<i>n</i> = 4)	Intra-day (<i>n</i> = 4)	Inter-day (<i>n</i> = 3)	Intra-day (<i>n</i> = 4)	Inter-day (<i>n</i> = 3)
2-CP	0	nd	nd				
	10	107	84.2	4.85	4.48	7.52	9.84
	20	107	117	6.79	3.94	8.28	0.39
	200	117	95.5	4.14	1.52	2.82	6.32
4-CP	0	nd	nd				
	10	113	107	5.22	2.88	6.00	1.20
	20	110	117	5.75	0.67	5.94	2.77
	200	111	95.7	6.35	2.92	6.39	2.28
2,4-DCP	0	nd	nd				
	10	84.2	86.1	7.82	6.61	10.1	6.10
	20	96.3	101	10.9	4.47	9.28	6.27
	200	120	99.2	2.80	1.62	6.93	4.49
2,4,6-TCP	0	nd	nd				
	10	113	112	3.34	1.13	4.51	3.81
	20	115	112	3.08	2.51	6.64	1.68
	200	114	103	0.84	2.13	3.32	5.70

^a nd: not detect.

Table 4 Comparison of the current method with other reported method for determination of chlorophenols^a

Method	Adsorbent	Amount of sorbent (mg)	Sample preparation time (min)	LOD (ng mL ⁻¹)	Recovery	Ref.
MSPE-GC-MS	Fe ₃ O ₄ @C@PANI	40	25	3.22–5.55	87.7–103.1	38
MSPE-HPLC-UV	OMMT-Fe ₃ O ₄ @PSF	150	35	0.17–0.22	90.9–115	39
SPE-HPLC-UV	Si–Ti@CN/IL	100	10	0.83–0.95	73.9–105.54	40
SPE-HPLC-UV	Cyano-functionalized MWCNT	100	>67	0.45–3	82.53–102.13	41
SPE-HPLC-UV	Graphene	20	>35	0.1–0.4	77.2–116.6	6
MSPE-HPLC-UV	Fe ₃ O ₄ @C	20	2	0.88–2.50	84.2–120	This work

^a SPME: solid phase microextraction. SPE: solid phase extraction. GC-MS: gas chromatography-tandem mass spectrometry. MWCNT: multiwalled carbon nanotube.

amount of the sorbent (20 mg for 20 mL sample) is much lower than reported methods (40 mg for 10 mL sample,³⁸ 150 mg for 80 mL sample,³⁹ 100 mg for 10 mL,⁴⁰ 100 mg for 70 mL sample⁴¹), meaning that the adsorbent has a higher extraction capability for the CPs. Compared with other carbon materials,^{6,41} the as-prepared sorbent shows short sample preparation time, less adsorbent dosage, and good recovery; in addition, the as-prepared carbon possess magnetism, which makes the extraction process greatly simple, avoid the need of centrifugation or filtration. These results suggested that the proposed method is very rapid, sensitive, and convenient method to enrichment of CPs in environmental water samples.

4. Conclusion

In this study, magnetic mesoporous carbon materials were fabricated by carbonization of polystyrene-grafted magnetic nanoparticles. The Fe₃O₄@C NPs were used as adsorbents to determination of chlorophenols from environmental water. The as-prepared adsorbent showed a good recovery for extraction chlorophenols from water samples because of hydrophobic interactions and π - π interactions. In addition, this method not only proposed a greatly fast, convenient and simple extraction process, but was environmental friendly because of using less organic solvent. These advantages reveal that the proposed method is applicable for real water sample analysis.

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