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Preparation of a polyvinylidene fluoride tree-like nanofiber mat loaded with manganese dioxide for highly efficient lead adsorption

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A novel polyvinylidene fluoride/tetrabutylammonium chloride (PVDF/TBAC) tree-like nanofiber mat loaded with manganese dioxide (MnO₂) as a highly efficient lead adsorbent was successfully fabricated. The adsorbent was prepared by *in situ* polymerization of pyrrole monomer on the surface of the PVDF/TBAC tree-like nanofiber mat, and subsequently reacted with KMnO₄ solution to deposit MnO₂. The morphology and structure of the as-prepared adsorbent were measured by field emission scanning electron microscopy (FE-SEM) and the tree-like structures can be clearly seen from the FE-SEM images. Fourier transform infrared spectroscopy (FT-IR) results confirmed the presence of PPy and MnO₂ layers on the surface of PVDF/TBAC tree-like nanofibers. Thermo-gravimetric analysis (TGA) results exhibited that MnO₂ accounted for about 43.27% in the PVDF/TBAC–polypyrrole–MnO₂ (PVDF/TBAC–PPy–MnO₂) nanofiber mat. The kinetics of Pb²⁺ adsorption was found to follow a pseudo-second-order rate model. The adsorption isotherms were fitted best with the Langmuir isotherm model. The thermodynamic analysis confirmed that the adsorption process was endothermic and spontaneous. The regeneration experiments showed that the obtained tree-like PVDF/TBAC–PPy–MnO₂ nanofiber mat also exhibited high recyclable removal efficiency. XPS analysis showed that ion exchange was the main mechanism for Pb²⁺ adsorption.

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1. Introduction

Recently, heavy metal contamination in wastewater with the rapid increase in global industrial activities has received considerable attention due to the high toxicity and non-biodegradability of the heavy metal ions.^{1–3} Pb²⁺ is one of the most common heavy metal ions in polluted water bodies due to its widespread use in industry, especially in battery manufacturing and metal plating.^{4,5} Besides, it is one of the most toxic heavy metal ions to humans due to its accumulation in the bone, brain, kidney and liver, which may cause many serious diseases.^{6,7} Hence, it is essential to remove Pb²⁺ from drinking water or lower the concentration of Pb²⁺.

Various techniques have been used for this aim, including chemical precipitation,^{8,9} chemical reduction,¹⁰ ion exchange,^{11,12} membrane separation,¹³ ultrafiltration,¹⁴ biological treatment,¹⁵ and adsorption, *etc.*^{16–18} Among them, adsorption is an effective and versatile method that is capable of removing trace levels of heavy metal ions from dilute solutions.

Currently, nano-sized metal oxide adsorbents were used for the removal of heavy metal ions, including nanoparticles of ferric oxides (Fe₂O₃),^{3,19,20} aluminium oxides (Al₂O₃),²¹ cerium oxides (CeO₂),²² zinc oxides (ZnO)²³ and manganese oxides (MnO₂).^{24–26} MnO₂ with advantages of high adsorption capacity and selectivity, have recently been receiving considerable attention.^{25–27} However, owing to the high surface energy, nano-sized MnO₂ particles are inclined to agglomerate during the adsorption process which limits their practical application.²⁸ In addition, the separation of nano-sized MnO₂ particles from solutions is still a challenging issue. To solve these problems, nano-sized MnO₂ particles can be loaded on supports. Brandão *et al.*²⁹ first fabricated a new composite material, cellulose acetate-supported MnO₂ particles (CA-MnO₂), which exhibited excellent Pb²⁺ adsorption performance. As an increasingly popular nanofabrication technique, electrospinning has emerged as a versatile and effective method for manufacturing long continuous fibers with diameters ranging from several micrometers down to a few nanometers.^{30,31} Owing to the large specific surface area, electrospun nanofibers become promising candidates for the load of MnO₂. As we all know, polyvinylidene fluoride (PVDF) is a material with extraordinary properties including easy mouldability, good toughness, flexibility, durability, excellent chemical and thermal resistance.^{32,33} Through physical or chemical modification, electrospun PVDF nanofibers can be used in the application of heavy metal ions

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adsorption.^{34–37} As reported by the literatures, polypyrrole (PPy) was widely used to reduce high valence of Mn compounds to prepare MnO₂.^{38,39} However, the uniform coating of PPy by *in situ* polymerization of pyrrole monomer usually occurred on the surface of polymer nanofiber with hydrophilicity.²⁷ Therefore, it was difficult to homogeneously load MnO₂ on the surface of PVDF nanofibers due to its hydrophobicity. In our previous work, a novel PVDF/TBAC tree-like nanofiber mat was fabricated *via* one-step electrospinning for the first time.⁴⁰ The addition of TBAC significantly enhanced the hydrophilicity of the PVDF/TBAC tree-like nanofiber mat.⁴¹ Meanwhile, the emergency of the tree-like branch fibers dramatically increased the specific surface area of nanofiber mat.⁴² Therefore, the PVDF/TBAC tree-like nanofiber mats are promising supports for the loading of MnO₂.

In this contribution, polyvinylidene fluoride/tetrabutylammonium chloride (PVDF/TBAC) tree-like nanofiber mat loading MnO₂ as highly efficient lead adsorbent was designed by *in situ* polymerization of pyrrole monomer, and subsequently reacted with KMnO₄ solution to deposit MnO₂. The obtained PVDF/TBAC-PPy-MnO₂ adsorbents were characterized, and their adsorption properties for removal of Pb²⁺ from aqueous solution under varied experimental conditions were also investigated. The adsorption kinetics, adsorption isotherm, adsorption thermodynamic, competitive adsorption and reusability of PVDF/TBAC-PPy-MnO₂ adsorbent were further studied.

2. Experimental

2.1. Materials

PVDF ($M_w = 520\,000$) was purchased from Shanghai 3F New Materials Co., Ltd., China. TBAC was purchased from Aladdin Co., Ltd., China. *N,N*-Dimethylformamide (DMF), acetone, pyrrole, potassium permanganate (KMnO₄), NaOH, manganese sulfate (MnSO₄), iron(III) chloride hexahydrate (FeCl₃·6H₂O), lead nitrate [Pb(NO₃)₂], copper nitrate [Cu(NO₃)₂] and cadmium nitrate [Cd(NO₃)₂] were purchased from Tianjin Kermel Co., Ltd., China. All of the materials were used as received without further purification.

2.2. Preparation of tree-like nanofiber mat

The PVDF/TBAC tree-like nanofiber mat was fabricated by adding certain amount of TBAC into PVDF solution *via* one-step electrospinning. The exhaustive methods were reported in our published paper.⁴⁰ The *in situ* polymerization of pyrrole monomer and redox reaction between the PPy and MnO₄⁻ was reported by Wang.^{27,43} Typically, 1.3 g FeCl₃·6H₂O was firstly dissolved in 30 mL deionized water and then a piece of PVDF/TBAC tree-like nanofiber mat was added into the solution. 3.2 g pyrrole was dissolved in 100 mL of deionized water. The polymerization was started by mixing the two solutions together. The mixed reaction solution was shaken in a thermostatic shaker bath, operating at 25 °C and 100 rpm for 2 h. Then the PVDF/TBAC-PPy nanofiber mat was taken out, and subsequently added to a 250 mL round bottomed flask which

contains 100 mL of KMnO₄ solutions. The KMnO₄ concentrations were 0.01, 0.05, 0.10 and 0.15 mol L⁻¹, respectively. The reaction solution was shaken in a thermostatic shaker bath, operating at 25 °C and 100 rpm for 10 h. MnO₂ was spontaneously deposited onto the surface of PVDF/TBAC-PPy nanofiber through a direct redox reaction between the PPy and MnO₄⁻. Finally, the PVDF/TBAC-PPy-MnO₂ nanofiber mat was taken out, washed with deionized water repeatedly to remove MnO₂ particles, and then dried in vacuum overnight. As a control, the common PVDF-PPy-MnO₂ nanofiber mat was prepared by the same process. Meanwhile, the PVDF-MnO₂ and PVDF/TBAC-MnO₂ nanofiber mats were prepared as the control samples. The exhaustive processes were as follows. A piece of nanofiber mat was added into 50 mL KMnO₄ solution (70 °C, 0.1 M) solution and then NaOH solution (0.5 M) and MnSO₄ solution (0.5 M) were added drop-wise to the KMnO₄ solution using two separate burettes until a brown precipitate formed. The reaction solution was stirred at 100 rpm by using a heater/magnetic stirrer at 70 °C. Finally, the PVDF-MnO₂ and PVDF/TBAC-MnO₂ nanofiber mats were taken out, washed with deionized water repeatedly to remove MnO₂ particles, and then dried in vacuum overnight.

2.3. Adsorption experiments

Pb(NO₃)₂ was used as the source of Pb²⁺, 10 mg dry PVDF/TBAC-PPy-MnO₂ nanofiber mat was directly added into a beaker containing 30 mL 100 mg L⁻¹ Pb²⁺ solutions with pH value of 6.0 for 24 h. During this process, the beaker was shaken in a thermostatic shaker bath, operating at 25 °C and 100 rpm. 0.1 M HNO₃ and 0.1 M NaOH were used to adjust the pH value of the Pb²⁺ solution. The removal percentage of Pb²⁺ can be calculated by the following equation:

$$\% \text{ removal} = \frac{C_0 - C_e}{C_0} \times 100 \quad (1)$$

where C_0 is the initial concentration of Pb²⁺ in solution (mg L⁻¹) and C_e is the equilibrium concentration (mg L⁻¹).

The adsorption isotherms for Pb²⁺ were established by batch adsorption experiments. 10 mg dry PVDF/TBAC-PPy-MnO₂ nanofiber mat was immersed into 30 mL Pb²⁺ solutions with different initial concentrations ranging from 50 mg L⁻¹ to 250 mg L⁻¹. The initial pH value of the Pb²⁺ solution was adjusted to 6.0. The solution was shaken in a thermostatic shaker bath at different temperature (25 °C, 35 °C and 45 °C) for 24 h to establish adsorption equilibrium. The equilibrium adsorption amount was calculated by the following equation:

$$q_e = \frac{(C_0 - C_e)V}{m} \quad (2)$$

where q_e is the equilibrium adsorption amount (mg L⁻¹), C_0 is the initial concentration of Pb²⁺ in solution (mg L⁻¹), C_e is the equilibrium concentration (mg L⁻¹), m is the mass of adsorbent (g), and V is the volume of solution (L).

The kinetic adsorption performance was studied by adding 10 mg of PVDF/TBAC-PPy-MnO₂ nanofiber mat into Pb²⁺ solutions of different initial concentrations. The initial pH value



of the Pb^{2+} solution is 6.0. The adsorption was carried out with constant shaking at 25 °C. Samples were taken out of the solution at different time, each time 50 μL . The adsorption amount was calculated by the following equation:

$$q_t = \frac{(C_0 - C_t)V}{m} \quad (3)$$

where q_t is the adsorption amount at time t (mg L^{-1}), C_0 is the initial concentration of Pb^{2+} in solution (mg L^{-1}), C_t is the Pb^{2+} concentration at time t (mg L^{-1}), m is the mass of adsorbent (g), and V is the volume of solution (L).

2.4. Desorption experiment

Reusability is one of the important aspects for the practical application of the adsorbent. In order to study the reusability of the adsorbents, 10 mg dry PVDF/TBAC-PPy-MnO₂ nanofiber mat was first contacted with 30 mL 100 mg L^{-1} Pb^{2+} solution for 24 h at 25 °C to reach adsorption equilibrium. The adsorbents were immersed into 50 mL 0.05 M, 0.10 M, 0.20 M, 0.30 M and 0.50 M HCl solution for 5 h at 25 °C, respectively, and then washed thoroughly with deionized water for using in next cycle of adsorption. The above procedure was repeated for 5 times to test the reusability of the adsorbents.

2.5. Characterization

The morphology of nanofiber mats was observed by field emission scanning electron microscopy (FE-SEM) (S-4800, Hitachi Ltd., Japan). The structural information of nanofiber was characterized by Fourier transform infrared spectroscopy (FT-IR) (TENSOR37, BRUKER, Germany). The Thermogravimetric analysis (TGA) (NETZSCH, Germany) for temperatures ranging from 30 to 900 °C was performed at a heating rate of 10 °C min^{-1} and an air flow of 60 mL min^{-1} . The adsorption mechanism was studied by X-ray photoelectron spectroscopy (XPS) (ThermoFisher K-alpha, England). Ions concentration was measured with an inductively coupled plasma mass spectrometry (Varian 715-ES, USA).

3. Results and discussions

3.1. Characterization of the PVDF/TBAC-PPy-MnO₂ nanofiber mat

The FE-SEM images of the pure PVDF, PVDF-PPy, PVDF-PPy-MnO₂, PVDF-MnO₂, PVDF/TBAC, PVDF/TBAC-PPy, PVDF/TBAC-PPy-MnO₂ and PVDF/TBAC-MnO₂ nanofiber mat were shown in Fig. 1. It was clear that the surface of the pure PVDF nanofiber and PVDF/TBAC tree-like nanofiber became much coarser after the coating of PPy and MnO₂. It was worth noting that the hydrophilicity of nanofiber mat reduced after coating with PPy, whereas significantly enhanced after the coating of MnO₂. Meanwhile, the tree-like branch fibers provided larger specific surface area for the coating of PPy and MnO₂, which was beneficial for the adsorption performance of nanofiber mat. In addition, due to the hydrophobicity of pure PVDF nanofiber mat, PPy couldn't be uniformly coated on the surface of PVDF nanofiber, which led to the lower MnO₂ loading amount. For the PVDF-MnO₂ and PVDF/TBAC-MnO₂ nanofiber mat, a large area of MnO₂ agglomerate could be observed and the MnO₂ couldn't be homogeneously coated on the surface of nanofiber.

Fig. 2 shows the FT-IR spectroscopy results for the PVDF/TBAC, PVDF/TBAC-PPy, PVDF/TBAC-PPy-MnO₂ and PVDF/TBAC-MnO₂ nanofiber mats. By comparing the spectrums of PVDF/TBAC and PVDF/TBAC-PPy nanofiber mats, the new band at 1534 cm^{-1} was assigned to the pyrrole ring vibration. The bands located at 1124 cm^{-1} was ascribed to the C-N stretching vibrations. The new peaks at 958 and 896 were attributed to C-H stretching vibrations and C-H deformations, respectively.⁴⁴ In the spectrum of PVDF/TBAC-PPy-MnO₂ nanofiber mat, a new broad band at 1628 cm^{-1} and 3401 cm^{-1} was corresponded to the stretching vibrations of -OH, which indicate that MnO₂ was coated on the surface of PVDF/TBAC-PPy nanofiber mat, and in the presence of MnOOH.²⁷ In the spectra of PVDF/TBAC-PPy-MnO₂ and PVDF/TBAC-MnO₂ nanofiber mats, the expected characteristic absorption peaks of Mn-O at 450 cm^{-1} , 520 cm^{-1} and 720 cm^{-1} did not emerge because of the amorphous state of MnO₂, which was consistent with other researches.²⁶

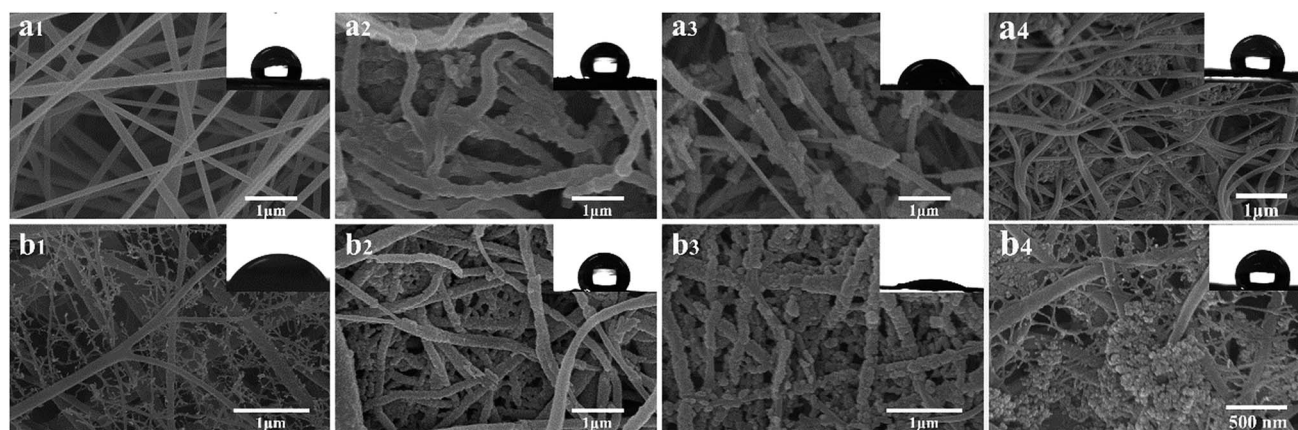


Fig. 1 FE-SEM images of (a1) PVDF, (a2) PVDF-PPy, (a3) PVDF-PPy-MnO₂, (a4) PVDF-MnO₂, (b1) PVDF/TBAC, (b2) PVDF/TBAC-PPy, (b3) PVDF/TBAC-PPy-MnO₂ and (b4) PVDF/TBAC-MnO₂ nanofiber mat (insert: contact angle of nanofiber mat).



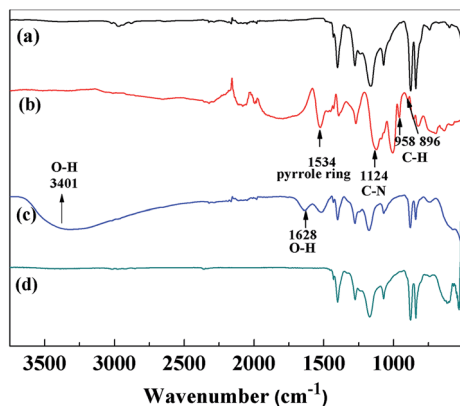


Fig. 2 FT-IR spectra of (a) PVDF/TBAC, (b) PVDF/TBAC-PPy, (c) PVDF/TBAC-PPy-MnO₂ and (d) PVDF/TBAC-MnO₂ nanofiber mat.

The thermal stabilities of PVDF/TBAC, PVDF/TBAC-PPy and PVDF/TBAC-PPy-MnO₂ nanofiber mats were studied by TGA, and the results were shown in Fig. 3. It was clear that degradations of the three materials finished (almost 0, 10.30% and 53.57% remained) at around 550, 480 and 500 °C for PVDF/TBAC, PVDF/TBAC-PPy and PVDF/TBAC-PPy-MnO₂ nanofiber mat, respectively. In PVDF/TBAC-PPy nanofiber mat, the residues were the catalyst particles and impurities came from the preparation of PVDF/TBAC-PPy nanofiber mat.²⁷ According to researches, MnO₂ particles were stable up to 800 °C. Therefore, the loading amount of MnO₂ in PVDF/TBAC-PPy-MnO₂ nanofiber mat can be estimated to be about 43.27%.

3.2. Adsorption study

3.2.1. Effect of MnO₂ loading ratios. To show the superiority of PVDF/TBAC-PPy-MnO₂ nanofiber mat on Pb²⁺ removal, contrast tests were done between PVDF/TBAC, PVDF/TBAC-PPy, PVDF-PPy-MnO₂, PVDF/TBAC-MnO₂ and PVDF/TBAC-PPy-MnO₂ nanofiber mats, and the results were shown in Fig. 4. It was clear to see that the Pb²⁺ adsorption amount of PVDF/TBAC-PPy-MnO₂ nanofiber mat was significantly higher than that of other adsorbents. This was caused by the fact that the tree-like branch fibers could offer large specific surface area for the loading of MnO₂. Meanwhile, the improvement of the hydrophilicity caused

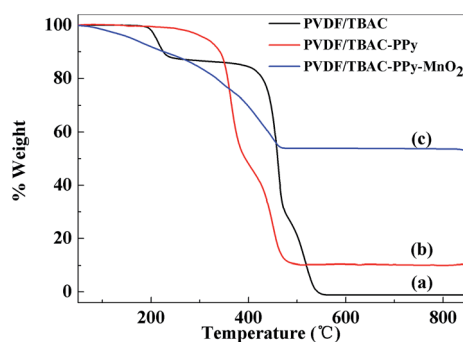


Fig. 3 Thermo-gravimetric curves for (a) PVDF/TBAC, (b) PVDF/TBAC-PPy and (c) PVDF/TBAC-PPy-MnO₂ nanofiber mat.

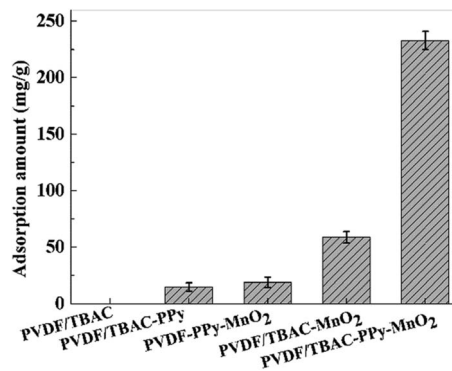


Fig. 4 The adsorption amount of Pb²⁺ by different adsorbents.

by the addition of TBAC enabled the PPy to uniform load on the surface of nanofiber, and then promoted the uniform loading of MnO₂. The combination of both gave the PVDF/TBAC-PPy-MnO₂ nanofiber mat excellent Pb²⁺ adsorption performance. However, the PVDF-PPy-MnO₂ nanofiber mat showed lower Pb²⁺ adsorption amount due to the less MnO₂ loading amounts caused by its poor hydrophilicity (Fig. 1). According to literature research, the chemical mechanism for the adsorption of polypyrrole (PPy) to metal ions included two steps. Firstly, the N atoms in the PPy chains could be partly changed to N⁺ after the treatment with HCl solution, therefore Cl⁻ can be recovered *via* the electrostatic interaction between N⁺ and Cl⁻ ions. Secondly, the doped Cl⁻ ions were replaced by metal ions through ion exchange.⁴⁵ However, lead ions are cations, and therefore this apparent selectivity is not related to the anion exchange process as mentioned above. Then, the adsorption capacity of PVDF/TBAC-PPy adsorbent for the lead ion was very low. For the PVDF/TBAC-MnO₂ nanofiber mat, the large area of agglomerate of MnO₂ led to its low adsorption capacity.

In order to investigate the optimal loading ratios of MnO₂, relevant experiments were performed. Adsorbents loaded with different amounts of MnO₂ were obtained by regulating the concentration of KMnO₄ solutions. The morphologies of PVDF/TBAC-PPy-MnO₂ nanofiber mats obtained from different

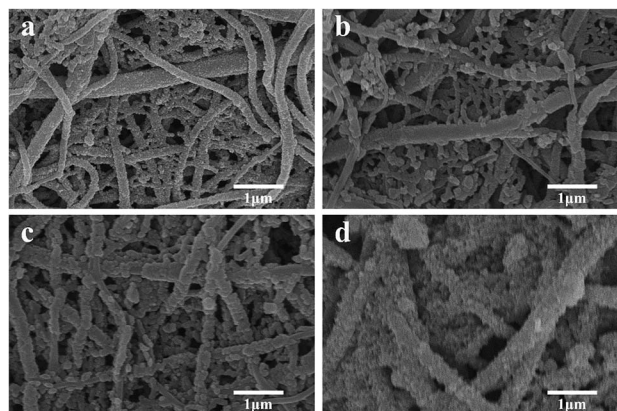


Fig. 5 FE-SEM images of PVDF/TBAC-PPy-MnO₂ nanofiber mat obtained from different KMnO₄ concentration: (a) 0.01 M, (b) 0.05 M, (c) 0.10 M and (d) 0.15 M.



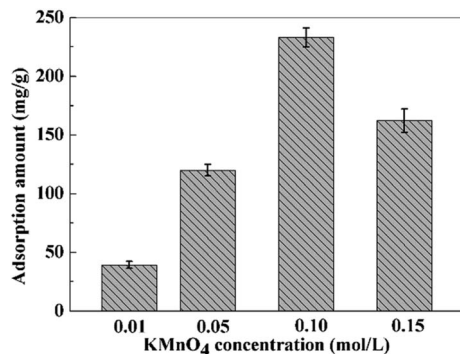


Fig. 6 The adsorption amount of PVDF/TBAC-PPy-MnO₂ nanofiber mat obtained from different KMnO₄ concentrations.

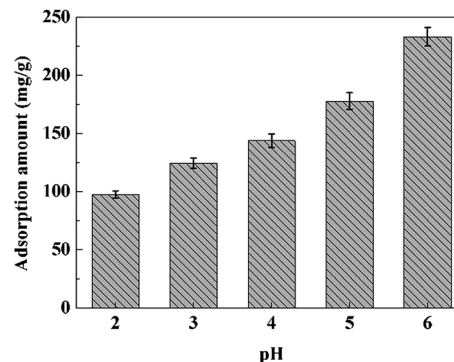


Fig. 8 Effect of pH on the removal of Pb²⁺ by the PVDF/TBAC-PPy-MnO₂ nanofiber mat.

KMnO₄ concentrations were shown in Fig. 5 and the effect of KMnO₄ concentrations on Pb²⁺ removal was presented in Fig. 6. It can be clearly seen that the adsorption amount was correlated to the KMnO₄ concentration to some extent. This was because that the MnO₂ content increased with the increase of KMnO₄ concentration to some extent and the coarser surface could provide more effective adsorption sites. However, further increase the KMnO₄ concentration, the adsorption amount decreased, which may be attributed to that the dense MnO₂ reduced the effective adsorption sites of the PVDF/TBAC-PPy-MnO₂ nanofiber mats. It can be concluded that the large specific surface area and the high MnO₂ content contributed to the good adsorption performance.

3.2.2. Effect of adsorbent dose. The dosage of adsorbent used in the adsorption process determines the adsorption performance of adsorbent for a given initial concentration of heavy metal ions solution. To study the effect of adsorbent dosage on the Pb²⁺ removal performance, different amounts of PVDF/TBAC-PPy-MnO₂ nanofiber mats were added into 30 mL 100 mg L⁻¹ Pb²⁺ solution for 24 h. The results were shown in Fig. 7. The results indicated that the Pb²⁺ removal rate increased slowly (from 85.50% to 99.04%) with the increase of adsorbent dosage (from 10 to 20 mg), which meant that extra PVDF/TBAC-PPy-MnO₂ nanofiber mat was waste for 30 mL 100 mg L⁻¹ Pb²⁺ solution. So we use 10 mg PVDF/TBAC-PPy-MnO₂ nanofiber mat to study the effect of PH on the Pb²⁺ adsorption.

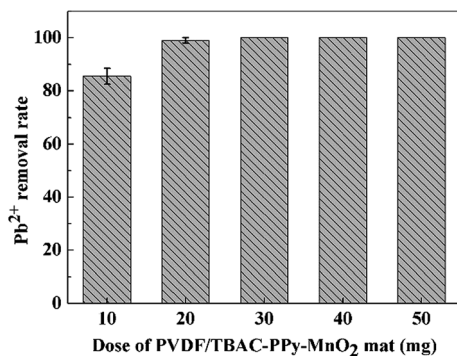


Fig. 7 Effect of PVDF/TBAC-PPy-MnO₂ nanofiber mat dose on the Pb²⁺ removal performance.

3.2.3. PH effect on the Pb²⁺ adsorption. Aqueous solution pH value plays a key role on the Pb²⁺ adsorption behavior because it affects both adsorbent and adsorbate.²⁷ According to the literature, lead ions will precipitate at alkaline solution.⁴⁶ Therefore, the effect of solution pH on the Pb²⁺ adsorption of was studied in the pH range of 2.0–6.0, and the results were shown in Fig. 8. It was clear that when the pH of the Pb²⁺ solution changed from 2.0 to 6.0, the adsorption performance of adsorbent increased from 97.57 mg g⁻¹ to 233.69 mg g⁻¹. The possible mechanism was reported by the literature.²⁷ The increase of pH may cause the less competitive interaction between the hydronium (H₃O⁺) ions and Pb²⁺ for the same sorption sites on the adsorbent surface. Additionally, the elevated levels of hydrogen ions may damage the surface structure of PVDF/TBAC-PPy-MnO₂ nanofiber mat due to the reaction between the metal oxide and hydrogen ions. Therefore, we choose pH value of 6.0 for the remaining experiments.

3.2.4. Adsorption kinetics. Fig. 9 shows the results of adsorption kinetics experiments within 24 h at 298 K. During the first 2 h, the adsorption capacity increased rapidly. But the adsorption rate became slow for the next 10 h and adsorption equilibrium was obtained after approximately 12 h. The adsorption equilibrium adsorption amounts of PVDF/TBAC-PPy-MnO₂ nanofiber mat for Pb²⁺ were 233.69 mg g⁻¹ and 277.26 mg g⁻¹ for 100 mg L⁻¹ and 200 mg L⁻¹ initial Pb²⁺ solution, respectively.

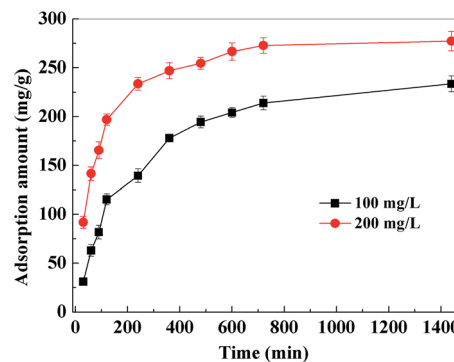


Fig. 9 Effect of contact time and initial concentration on the adsorption of Pb²⁺ by PVDF/TBAC-PPy-MnO₂ nanofiber mat.



In order to better understand the adsorption behaviors, adsorption kinetic data are often analyzed using two commonly used kinetic models, namely, the pseudo-first-order kinetic model and the pseudo-second-order kinetic model.^{47,48} These two kinetic models are used to describe the adsorption of solid/liquid systems, which can be expressed in the linear forms as eqn (4) and (5), respectively:

$$\log(q_e - q_t) = \log q_e - \frac{K_1}{2.303} t \quad (4)$$

$$\frac{t}{q_t} = \frac{1}{K_2 q_e^2} + \frac{1}{q_e} t \quad (5)$$

where K_1 and K_2 are the pseudo-first-order and the pseudo-second-order rate constants, respectively.

The adsorption kinetic plots for the adsorption of Pb^{2+} were shown in Fig. 10 and the obtained kinetic parameters were summarized in Table 1. High correlation coefficients ($R^2 > 0.998$) and the proximity between the experimental adsorption amount ($q_{e,\text{exp}}$) and the calculated equilibrium adsorption amount ($q_{e,\text{cal}}$) indicated that heavy metal ions adsorption onto PVDF/TBAC-PPy-MnO₂ nanofiber mat could be approximated favorably by the pseudo-second-order model. And pseudo-first-order model did not show goodness of fit. Therefore, the adsorption on PVDF/TBAC-PPy-MnO₂ nanofiber mat may be a chemical process through sharing or exchange of electrons between PVDF/TBAC-PPy-MnO₂ nanofiber mat and metal ions.⁴⁹

3.2.5. Adsorption isotherm and thermodynamic study. The results of adsorption isotherms of Pb^{2+} removal by PVDF/TBAC-PPy-MnO₂ nanofiber mat at temperature of 25 °C, 35 °C and 45 °C were shown in Fig. 11. The results indicated that there was an increase of the adsorption amount with an increase of solution temperature. Meanwhile, the increase of adsorption

amount with the increase of initial solution concentration was slightly faster at higher solution temperature than at low solution temperature. This may be due to an increase in thermal energy of the adsorbing species, which leads to higher adsorption amount and faster adsorption rate. The maximum adsorption capacities of the PVDF/TBAC-PPy-MnO₂ adsorbent for heavy metals were evaluated using the two well-known models of Langmuir and Freundlich isotherms.^{19,50,51} The Langmuir model is based in the assumption of adsorption homogeneity, representing equally available adsorption sites, monolayer surface coverage, and no interaction between adsorbed species. The linear form of Langmuir isotherm equation is given by eqn (6):

$$\frac{C_e}{q_e} = \frac{1}{q_m b} + \frac{C_e}{q_m} \quad (6)$$

where C_e (mg g^{-1}) is the equilibrium concentration on heavy metal ions in solution, q_e is the adsorption amount at equilibrium concentration (mg g^{-1}), q_m (mg g^{-1}) is the maximum adsorption capacity and b (L mg^{-1}) is the Langmuir constant relate to the energy of adsorption.

The Freundlich isotherm describes reversible adsorption and is not restricted to the formation of the monolayer. The Freundlich model takes the following form:

$$\ln q_e = \ln K_f + \frac{\ln C_e}{n} \quad (7)$$

where K_f and $1/n$ constants are related to the adsorption capacity and intensity of adsorption.

The values of these parameters, as analyzed from the plots shown in Fig. 12, were summarized in Table 2. The higher values of correlation coefficient revealed that Langmuir model better fitted the isotherm data well compared to the Freundlich model. The results indicated that this adsorption process took

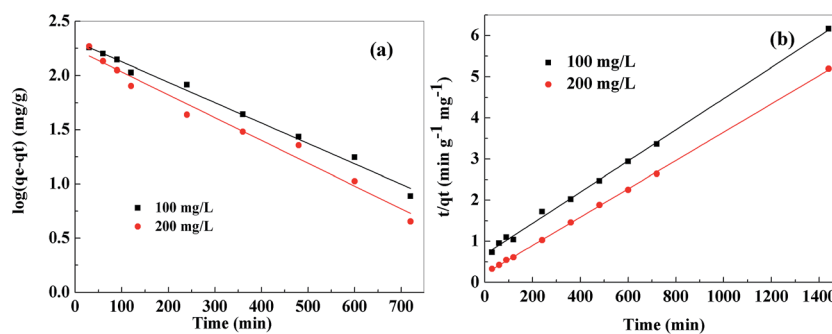


Fig. 10 Pseudo-first-order kinetic model (a), pseudo-second-order kinetic model (b) for adsorption of Pb^{2+} onto PVDF/TBAC-PPy-MnO₂ nanofiber mat.

Table 1 Kinetics parameters for Pb^{2+} adsorption onto the PVDF/TBAC-PPy-MnO₂ nanofiber mat

C_0 (mg L^{-1})	$q_{e,\text{exp}}$ (mg g^{-1})	Pseudo-first-model			Pseudo-second-model		
		K_1 (min^{-1})	$q_{e,\text{cal}}$ (mg g^{-1})	R^2	K_2 ($\text{min g}^{-1} \text{mg}^{-1}$)	$q_{e,\text{cal}}$ (mg g^{-1})	R^2
100	233.69	0.0043	206.50	0.9900	2.11×10^{-5}	263.85	0.9980
200	277.26	0.0049	175.18	0.9763	5.62×10^{-5}	290.70	0.9997



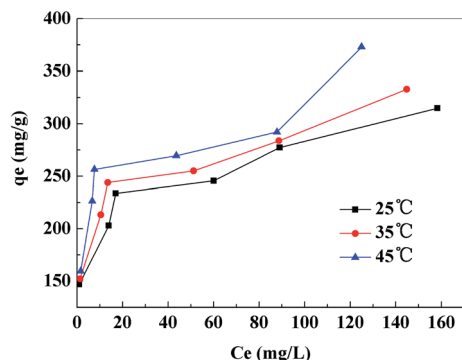


Fig. 11 Equilibrium isotherms of Pb^{2+} onto PVDF/TBAC-PPy- MnO_2 nanofiber mat.

place on the adsorption sites acted as monolayer adsorption on the surface of the PVDF/TBAC-PPy- MnO_2 adsorbent under the applied test conditions. The maximum adsorption capacity calculated from the Langmuir model increased from 318.47 to 355.87 mg g^{-1} as the temperature increased from 25 °C to 45 °C. This value was larger than those reported in literatures for other MnO_2 adsorbents, as shown in Table 3.

The thermodynamic parameters such as changes in standard Gibbs free energy change (ΔG^0), enthalpy change (ΔH^0) and entropy change (ΔS^0) for the adsorption of Pb^{2+} by PVDF/TBAC-PPy- MnO_2 adsorbent were determined by the following equations:^{58,59}

$$\ln K_d = \frac{\Delta S^0}{R} + \frac{-\Delta H^0}{RT} \quad (8)$$

$$K_d = \frac{(C_0 - C_e)V}{mC_e} \quad (9)$$

$$\Delta G^0 = -RT \ln K_d \quad (10)$$

Table 3 Comparison of the maximum adsorption capacities for Pb^{2+} onto PVDF/TBAC-PPy- MnO_2 nanofiber mat with other MnO_2 adsorbents reported in previous literatures

Adsorbent	Q_{max} (mg g^{-1})	References
Magnetic halloysite nanotubes@manganese oxide nanocomposite	59.9	52
Hydrous MnO_2 -poly(<i>N</i> -hydroxymethyl acrylamide/2-hydroxyethyl acrylate) hydrogel	100.18	24
MnO_2 -coated SD300 resin	140.9	53
Triethylenetetramine functionalized cellulose acetate grafted with the copolymer- MnO_2 composite	196.84	54
γ - MnO_2	200	55
Magnetic Fe_3O_4 - MnO_2 nanoplates	208.17	56
Nano manganese oxide	230	57
Polyacrylonitrile/PPy/ MnO_2	251.89	27
PVDF/TBAC-PPy- MnO_2	318.47	This study

where R ($\text{J mol}^{-1} \text{K}^{-1}$) is the gas constant, m (g) is the adsorbent dose and T (K) is the absolute solution temperature.

The value of ΔH^0 and ΔS^0 were obtained from the slope and intercept of the plots of $\ln K_d$ versus $1/T$, as shown in Fig. 13. The related parameters were calculated and presented in Table 4. The positive value of ΔH^0 confirmed that the adsorption process was endothermic in nature, while the positive entropy change (ΔS^0) suggested an increase in randomness at the PVDF/TBAC-PPy- MnO_2 adsorbent-solution interface. And the decrease in ΔG^0 values with the increasing temperature indicated that the adsorption process was spontaneous.

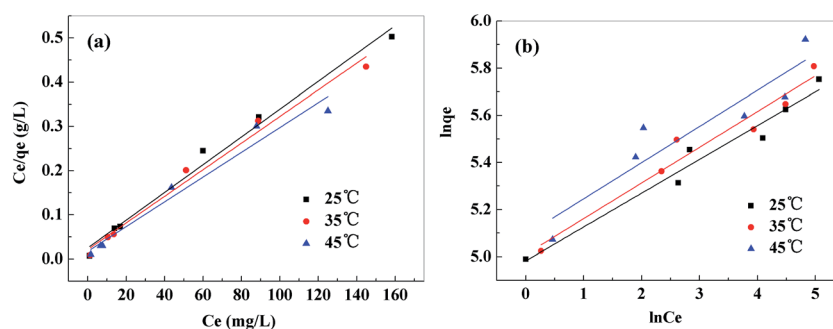


Fig. 12 Langmuir isotherm model (a), Freundlich isotherm model (b) for adsorption of Pb^{2+} onto PVDF/TBAC-PPy- MnO_2 nanofiber mat.

Table 2 Langmuir and Freundlich isotherm parameters for Pb^{2+} adsorption onto the PVDF/TBAC-PPy- MnO_2 nanofiber mat

Temperature (°C)	Langmuir model			Freundlich model		
	q_m (mg g^{-1})	b (L mg^{-1})	R^2	K_f (mg g^{-1})	$1/n$	R^2
25	318.47	0.1260	0.9853	145.84	0.1434	0.9547
35	332.22	0.1420	0.9822	149.88	0.1516	0.9412
45	355.87	0.1648	0.9621	162.70	0.1537	0.8314



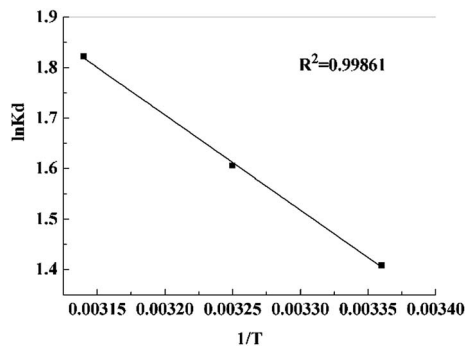


Fig. 13 Plots to determine thermodynamic parameters of Pb^{2+} onto PVDF/TBAC-PPy- MnO_2 nanofiber mat.

Table 4 Thermodynamic parameters for Pb^{2+} uptake by the PVDF/TBAC-PPy- MnO_2 nanofiber mat

Temperature (K)	ΔG^0 (kJ mol^{-1})	ΔH^0 (kJ mol^{-1})	ΔS^0 ($\text{J mol}^{-1} \text{K}^{-1}$)
298	-3.4882	—	—
308	-4.1118	15.6655	64.3151
318	-4.8182	—	—

3.3. Coexisting metal ions

It is very significant to investigate the adsorption process in multicomponent heavy metal ion system since the natural and industrial effluents rarely contain one kind of heavy metal. In this study, Cu^{2+} and Cd^{2+} were used as the coexisting metal ions. 10 mg PVDF/TBAC-PPy- MnO_2 nanofiber mats were immersed into 30 mL mixed solution, which contained Pb^{2+} , Cu^{2+} and Cd^{2+} . The initial concentration of each metal ion was 100 mg L^{-1} , and the pH value of the solution was 6.0. Fig. 14 shows the effects of coexisting metal ions on Pb^{2+} adsorption performance. As observed, the Pb^{2+} adsorption performance was slightly decreased due to the presence of Cu^{2+} and Cd^{2+} in the solution. However, the adsorption amounts of Cu^{2+} and Cd^{2+} were negligibly small. So the selectivity of PVDF/TBAC-PPy- MnO_2 nanofiber mat to Pb^{2+} was very good.

3.4. Regeneration study

In order to improve sustainable use of the adsorbent for subsequent application and reduce the overall cost treatment,

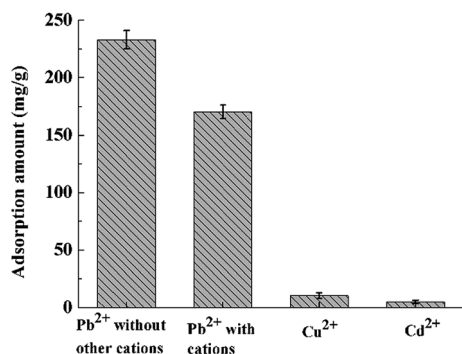


Fig. 14 Effect of coexisting metal ions on the Pb^{2+} adsorption.

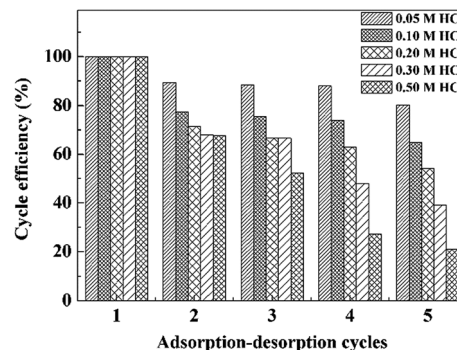


Fig. 15 Effects of acid concentration on the regeneration performance.

adsorbent regeneration study was performed. Since the adsorption of Pb^{2+} onto the PVDF/TBAC-PPy- MnO_2 nanofiber mat was pH-dependent and the higher pH was beneficial for the Pb^{2+} adsorption. Desorption of Pb^{2+} from the adsorbent can be achieved by decreasing the pH values of the system. Different concentrations of HCl were investigated to determine the conditions for unloading the Pb^{2+} from the metal-adsorbed PVDF/TBAC-PPy- MnO_2 nanofiber mat, and the results for four cycles were exhibited in Fig. 15. It was clear that the adsorption efficiency of the Pb^{2+} decreased as the concentration of HCl increased from 0.05 M to 0.50 M. When 0.05 M HCl was used for the regeneration, the Pb^{2+} adsorption efficiency still remained more than 80% of its original adsorption capacity after five times usage. However, when the concentration of HCl was higher than 0.20 M, the adsorption capacity of the adsorbent dropped dramatically after three adsorption/desorption cycles. This was attributed to the instability of MnO_2 at a high acid concentration and the MnO_2 was partially off from the surface of PVDF/TBAC-PPy- MnO_2 adsorbents.²⁷ Thus 0.05 M HCl was selected to be eluent for desorption of Pb^{2+} from the surface of adsorbent.

To investigate the desorption efficiency of the lead-loaded adsorbents, the adsorbents were desorbed with 50 mL 0.05 M HCl solution for 5 h, 24 h and 48 h, respectively, and the results were shown in Fig. 16. It could be seen that the desorption ratio of adsorbents desorbed for 5 h were above 90%. When the desorption time increased to 24 h and 48 h, both the desorption

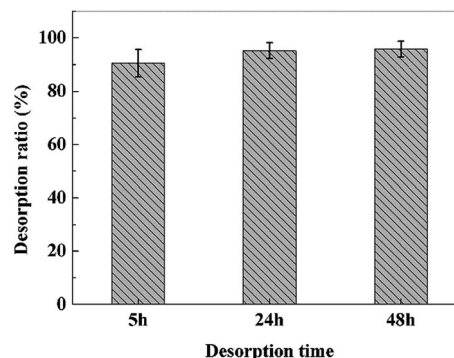


Fig. 16 Desorption efficiency of different desorption time for lead-loaded PVDF/TBAC-PPy- MnO_2 nanofiber mat.



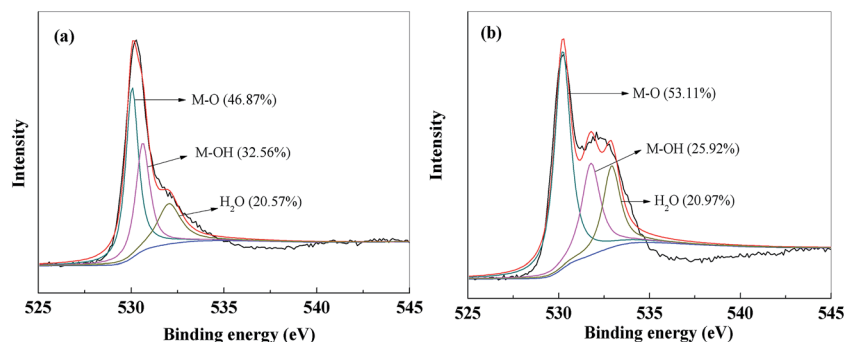


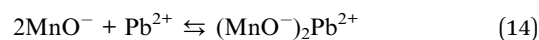
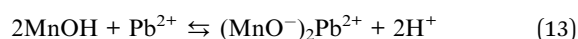
Fig. 17 XPS spectra of O 1s of PVDF/TBAC-PPy-MnO₂ nanofiber mat before (a) and after (b) Pb²⁺ adsorption.

ratios were above 95%. This results indicated that desorption equilibrium was obtained after approximately 12 h. However, considering the practical application, the desorption time was selected to be 5 h.

The regeneration experiments in this study indicated that the PVDF/TBAC-PPy-MnO₂ nanofiber mat can be used repeatedly as efficient adsorbents for practical wastewater treatment at an optimal concentration of acid. Furthermore, only a tiny amount of MnO₂ might fall off from the surface of the adsorbents during the multiple desorption and washing steps. Due to the low-toxicity of MnO₂,⁶⁰ there was almost no risk of secondary contaminants. Therefore, the prepared PVDF/TBAC-PPy-MnO₂ nanofiber mat in the present study will be an excellent candidate for the Pb²⁺ removal from the contaminated water based on the high adsorption capacity, good regeneration together with convenient separation.

3.5. Adsorption mechanism

The XPS spectra of O 1s before and after Pb²⁺ adsorption were compared to better understand the adsorption process, and the results were shown in Fig. 17. Before the Pb²⁺ adsorption, the O 1s spectrum of PVDF/TBAC-PPy-MnO₂ nanofiber mat can be divided into three peaks at around 530.0 eV, 531.3 eV and 533.1 eV, which can be assigned to metal oxide (M-O), hydroxyl bonded to metal (M-OH) and adsorbed H₂O in the adsorbent, respectively.^{25,27} After Pb²⁺ adsorption, it was clear that the area ratio of the peak at 530.0 eV assigned to M-O increased by 6.24% (from 46.87% to 53.11%). While the area ratio of the peak at 531.3 eV assigned to M-OH decreased by 6.64% (from 32.56% to 25.92%). It was obvious that the decreased M-OH area ratio equaled to that of M-O, which meant that the M-OH and M-O groups on the surface of PVDF/TBAC-PPy-MnO₂ adsorbent participated in the Pb²⁺ adsorption process. The main adsorption mechanism maybe an exchange reaction between Pb²⁺ and protons existed on the surface of manganese oxide,²⁴ which may be written as eqn (11)–(14):



4. Conclusions

Tree-like PVDF/TBAC-PPy-MnO₂ nanofiber mat with high adsorption performance towards Pb²⁺ was successfully fabricated by electrospinning and followed by *in situ* polymerization of pyrrole monomer, and subsequently reacted with KMnO₄ solution. FT-IR results confirmed the presence of MnO₂ layer on the surface of the PVDF/TBAC-PPy-MnO₂ nanofiber mat. Kinetics of the Pb²⁺ adsorptions was found to follow pseudo-second-order rate model. The adsorption isotherms were fitted better with Langmuir isotherm model. The thermodynamic analysis confirmed that the adsorption process was endothermic and spontaneous. The regeneration experiments showed that the obtained tree-like PVDF/TBAC-PPy-MnO₂ nanofiber mat also exhibited high recyclable removal efficiency. XPS analysis showed that ion exchange was the main mechanism for the Pb²⁺ adsorption. Therefore, the tree-like PVDF/TBAC-PPy-MnO₂ nanofiber mat may be applied as a promising adsorbent in heavy metal ions waste water treatment.

Acknowledgements

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References

- 1 R. Zhao, X. Li, B. L. Sun, M. Q. Shen, X. C. Tan, Y. Ding, Z. Q. Jiang and C. Wang, *Chem. Eng. J.*, 2015, **268**, 290–299.
- 2 R. Rojas, M. R. Perez, E. M. Erro, P. I. Ortiz, M. A. Ulibarri and C. E. Giacomelli, *J. Colloid Interface Sci.*, 2009, **331**, 425–431.
- 3 S. Rajput, C. U. Pittman and D. Mohan, *J. Colloid Interface Sci.*, 2016, **468**, 334–346.
- 4 D. Bulgariu and L. Bulgariu, *Bioresour. Technol.*, 2013, **129**, 374–380.



- 5 Y. Y. Xu, Q. F. Dang, C. S. Liu, J. Q. Yan, B. Fan, J. P. Cai and J. J. Li, *Colloids Surf., A*, 2015, **482**, 353–364.
- 6 S. H. Jang, B. G. Min, Y. G. Jeong, W. S. Lyoo and S. C. Lee, *J. Hazard. Mater.*, 2008, **152**, 1285–1292.
- 7 H. Yin, X. H. Feng, G. H. Qiu, W. F. Tan and F. Liu, *J. Hazard. Mater.*, 2011, **188**, 341–349.
- 8 T. A. Kurniawan, G. Y. S. Chan, W. H. Lo and S. Babel, *Chem. Eng. J.*, 2006, **118**, 83–98.
- 9 Q. Y. Chen, Z. Luo, C. Hills, G. Xue and M. Tyrer, *Water Res.*, 2009, **43**, 2605–2614.
- 10 V. Pena-Caballero, R. Aguilar-Lopez, P. A. Lopez-Perez and M. I. Neria-Gonzalez, *Desalin. Water Treat.*, 2016, **57**, 13056–13065.
- 11 S. A. Cavaco, S. Fernandes, M. M. Quina and L. M. Ferreira, *J. Hazard. Mater.*, 2007, **144**, 634–638.
- 12 B. Shah and U. Chudasama, *J. Hazard. Mater.*, 2014, **276**, 138–148.
- 13 S. M. Doke and G. D. Yadav, *Chem. Eng. J.*, 2014, **255**, 483–491.
- 14 M. Schwarze, M. Gross, M. Moritz, G. Buchner, L. Kapitzki, L. Chiappisi and M. Gradzielski, *J. Membr. Sci.*, 2015, **478**, 140–147.
- 15 S. Hena, *J. Hazard. Mater.*, 2010, **181**, 474–479.
- 16 M. Kumari, C. U. Pittman and D. Mohan, *J. Colloid Interface Sci.*, 2015, **442**, 120–132.
- 17 X. Z. Wu, L. L. Luo, Z. Y. Chen and K. L. Liang, *Appl. Surf. Sci.*, 2016, **364**, 86–95.
- 18 S. Ghoohestani and H. Faghian, *Desalin. Water Treat.*, 2016, **57**, 4049–4058.
- 19 Y. Ren, H. A. Abbood, F. B. He, H. Peng and K. X. Huang, *Chem. Eng. J.*, 2013, **226**, 300–311.
- 20 K. Li, Y. W. Wang, M. Huang, H. Yan, H. Yang, S. J. Xiao and A. M. Li, *J. Colloid Interface Sci.*, 2015, **455**, 261–270.
- 21 Y. X. Zhang, Y. J. Ye, Z. L. Liu, B. Li, Q. Z. Liu, Q. C. Liu and X. H. Li, *J. Alloys Compd.*, 2016, **662**, 421–430.
- 22 A. R. Contreras, E. Casals, V. Puentes, D. Komilis, A. Sanchez and X. Font, *Global NEST J.*, 2015, **17**, 536–543.
- 23 X. Zhang, Y. Wang, Y. F. Liu, J. L. Xu, Y. D. Han and X. X. Xu, *Appl. Surf. Sci.*, 2014, **316**, 333–340.
- 24 Q. Zhu and Z. K. Li, *Chem. Eng. J.*, 2015, **281**, 69–80.
- 25 C. Z. Hu, F. Y. Liu, H. C. Lan, H. J. Liu and J. H. Qu, *J. Colloid Interface Sci.*, 2015, **446**, 359–365.
- 26 J. Zhu, S. A. Baig, T. T. Sheng, Z. M. Lou, Z. X. Wang and X. H. Xu, *J. Hazard. Mater.*, 2015, **286**, 220–228.
- 27 C. Luo, J. Q. Wang, P. Jia, Y. X. Liu, J. H. An, B. Cao and K. Pan, *Chem. Eng. J.*, 2015, **262**, 775–784.
- 28 Y. T. Zhou, H. L. Nie, C. Branford-White, Z. Y. He and L. M. Zhu, *J. Colloid Interface Sci.*, 2009, **330**, 29–37.
- 29 M. S. B. Brandão and F. Galembeck, *Colloids Surf.*, 1990, **48**, 351–362.
- 30 S. Y. Xie, X. Y. Liu, B. W. Zhang, H. J. Ma, C. J. Ling, M. Yu, L. F. Li and J. Y. Li, *J. Mater. Chem. A*, 2015, **3**, 2552–2558.
- 31 Y. Li, J. Zhang, C. Xu and Y. F. Zhou, *Sci. China: Chem.*, 2016, **59**, 95–105.
- 32 R. Gopal, S. Kaur, Z. W. Ma, C. Chan, S. Ramakrishna and T. Matsuura, *J. Membr. Sci.*, 2006, **281**, 581–586.
- 33 X. Zhuang, L. Shi, K. Jia, B. Cheng and W. Kang, *J. Membr. Sci.*, 2012, **429**, 66–70.
- 34 Z. J. Zhang and D. L. Hao, *Adv. Mater. Res.*, 2014, **898**, 465–469.
- 35 M. Y. Bai and J. C. Tsai, *Fibers Polym.*, 2014, **15**, 2265–2271.
- 36 M. S. Birajdar, S. D. Wanjale and S. P. Lonkar, *J. Appl. Polym. Sci.*, 2013, **130**, 4508–4515.
- 37 X. Zhao, L. Song, J. Fu, P. Tang and F. Liu, *J. Hazard. Mater.*, 2011, **189**, 732–740.
- 38 J. G. Wang, Y. Yang, Z. H. Huang and F. Kang, *J. Mater. Chem.*, 2012, **22**, 16943–16949.
- 39 J. Zang and X. Li, *J. Mater. Chem.*, 2011, **21**, 10965–10969.
- 40 Z. J. Li, Y. Z. Xu, L. L. Fan, W. M. Kang and B. W. Cheng, *Mater. Des.*, 2016, **92**, 95–101.
- 41 Z. Li, W. Kang, H. Zhao, M. Hu, N. Wei, J. Qiu and B. Cheng, *Nanomaterials*, 2016, **6**, 152–162.
- 42 Z. J. Li, W. M. Kang, H. H. Zhao, M. Hu, J. G. Ju, N. P. Denga and B. W. Cheng, *RSC Adv.*, 2016, **6**, 91243–91249.
- 43 J. Q. Wang, K. Pan, Q. W. He and B. Cao, *J. Hazard. Mater.*, 2013, **244**, 121–129.
- 44 J. F. Zang and X. D. Li, *J. Mater. Chem.*, 2011, **21**, 10965–10969.
- 45 J. Wang, C. Luo, G. Qi, K. Pan and B. Cao, *Appl. Surf. Sci.*, 2014, **316**, 245–250.
- 46 L. Lalhmunsiam, S. M. Lee and D. Tiwari, *Chem. Eng. J.*, 2013, **225**, 128–137.
- 47 Y. Jiang, H. Pang and B. Liao, *J. Hazard. Mater.*, 2009, **164**, 1–9.
- 48 M. Bhaumik, C. Noubactep, V. K. Gupta, R. I. McCrindle and A. Maity, *Chem. Eng. J.*, 2015, **271**, 135–146.
- 49 S. J. Allen, Q. Gan, R. Matthews and P. A. Johnson, *J. Colloid Interface Sci.*, 2005, **286**, 101–109.
- 50 L. Q. Guo, P. R. Ye, J. Wang, F. F. Fu and Z. J. Wu, *J. Hazard. Mater.*, 2015, **298**, 28–35.
- 51 Y. Yu, R. G. Peng, C. Yang and Y. H. Tang, *J. Mater. Sci.*, 2015, **50**, 5799–5808.
- 52 D. Afzali and M. Fayazi, *J. Taiwan Inst. Chem. Eng.*, 2016, **63**, 421–429.
- 53 Y. Xiong, X. M. Lu and H. C. Tao, *Desalin. Water Treat.*, 2016, **57**, 2018–2027.
- 54 A. A. Yakout, R. H. El-Sokkary, M. A. Shreadah and O. G. A. Hamid, *Carbohydr. Polym.*, 2016, **148**, 406–414.
- 55 N. Chung Le and D. Van Phuc, *Adv. Nat. Sci.: Nanosci. Nanotechnol.*, 2015, **6**, 025014–025021.
- 56 J. H. Zhao, J. Liu, N. Li, W. Wang, J. Nan, Z. W. Zhao and F. Y. Cui, *Chem. Eng. J.*, 2016, **304**, 737–746.
- 57 J. Al Abdullah, A. G. Al Lafi, W. Al Masri, Y. Amin and T. Alnama, *Water, Air, Soil Pollut.*, 2016, **227**, 1–11.
- 58 S. Deng, G. S. Zhang, X. Wang, T. Zheng and P. Wang, *Chem. Eng. J.*, 2015, **276**, 349–357.
- 59 N. Mohammadi, H. Khani, V. K. Gupta, E. Amereh and S. Agarwal, *J. Colloid Interface Sci.*, 2011, **362**, 457–462.
- 60 S. Hernández, C. Ottone, S. Varetti, M. Fontana, D. Pugliese, G. Saracco, B. Bonelli and M. Armandi, *Materials*, 2016, **9**, 296–309.

