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A proposed industrial scale-up of circular bisphenol-A (BPA) production

Aswin Nair, Megan Novak,* Jonathan Ochoa, Dyron Powell, Josh Ramsey and Logan Stadtmueller

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To meet the evolving sustainability goals of the modern-day chemical industry, there is a demand for novel chemical processes that minimize environmental impact while also maximizing profitability. This paper proposes a novel commercial solution for producing bisphenol-A (BPA) from the advanced recycling of polycarbonate waste. Current BPA production methods have major environmental and safety concerns from the use of benzene, high temperatures and pressures, and strong acids, and the proposed novel method addresses all these issues. This advanced recycling process utilizes a base-catalyzed methanolysis reaction with a methanol/toluene solvent mixture to produce BPA and dimethyl carbonate (DMC), a sustainable fuel additive both at 99.99 wt% purity. A prototype process was simulated in Aspen Plus®, and a preliminary process flow diagram was developed. With a target production capacity of 200 000 metric tons of BPA per annum, the major processing equipment is one packed-bed reactor, two crystallizers, and three distillation columns. All required heat exchangers and pumps were integrated into the simulation and can be adjusted based on product specifications and processing capacity. Analysis of green metrics for the novel process demonstrated that the process minimizes waste from a mass standpoint, and a rigorous economic analysis showed that the process is highly profitable in several varied scenarios.

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1. Introduction: overview of traditional BPA manufacturing

There are several pathways to produce bisphenol-A (BPA), most of which use acetone and phenol as raw materials. The predominant method, the Hock process, allows for both phenol and acetone to be produced in a multi-step mechanism that uses propylene and benzene as primary reactants. The production of both raw materials in one process, although complex, is what makes the Hock process convenient in industry. However, it comes with several drawbacks. One of the main intermediates, cumene, is oxidized to cumene hydroperoxide (CHP) at high operating temperatures. CHP poses inherent safety risks as it is an unstable intermediate that undergoes a highly exothermic degradation. In addition to this, the degradation is performed in large quantities of strong acid, which needs to be neutralized and disposed of carefully to avoid harm to operators and the surrounding environment.[1,2]

The goal of this paper is to evaluate the feasibility of an alternative method of BPA production. The methanolysis of polycarbonate waste has the potential to produce BPA at high yields while also addressing the process safety risks associated with the Hock process. Methanol is a well-known and common industrial reagent, and the solvent mechanism is efficient at significantly lower operating temperatures than the traditional method. At a lab-scale, BPA was synthesized using only methanol and toluene as solvents and relatively low amounts of strong base to catalyze the reaction. If scaled up, this can remedy the need for a complex multi-stage process with extreme operating conditions, thus simultaneously reducing operating costs and improving operational, environmental, and safety concerns. The science behind the methanolysis of plastic waste is still in its infancy, and there are doubts over the feasibility of using waste plastic as a feedstock, but this paper provides a comprehensive review of the scale-up potential of this innovative process.

2. Advanced recycling of waste polycarbonate to BPA

This process chemically recycles polycarbonate waste (PC) to produce bisphenol-A (BPA) and dimethyl carbonate (DMC). PC is reacted with a stoichiometric equivalent of methanol in a solvent mixture of methanol and toluene, catalyzed by sodium hydroxide in a batch reactor for 15 minutes at 60 °C.

Artie McFerrin Department of Chemical Engineering, Texas A&M University, USA. E-mail: megann1718@gmail.com

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The reaction is quenched with toluene and then crystallized to produce the products of BPA and DMC. The solvents and catalyst are recovered with distillation.[3–5]

2.1 Reference data

The reagents, solvents, reaction conditions, and quantities of reagents were scaled up from Hu *et al.* In the experiment by Hu *et al.*, 1.27 g of polycarbonate was reacted in a solvent mixture of 1 mL of methanol in 1.5 mL toluene.[6] The kinetic data for the reaction was obtained from Kim *et al.*[3] The mechanism of the methanolysis reaction was obtained from Piñero *et al.*[4] The solubility of BPA in toluene was obtained from Little.[5]

2.2 Process assumptions & scale-up methodology

The lab-scale experiment conducted by Hu *et al.* was scaled up to an industrial-sized process of 200 000 metric tons of BPA produced per annum at 99.99 wt% purity. With an assumption of 80% on-stream efficiency, the process would stoichiometrically require 31 790 kg h^{-1} of PC to meet production requirements. The solubility of BPA in the reactor effluent was approximated to be the solubility of BPA in toluene. It is assumed that the catalytic NaOH is regenerated in the reactor and does not leave the reactor. A scale factor between the stoichiometric requirement of PC and the amount of PC used in the lab-scale experiment was calculated, and the amounts of respective solvent were scaled up with this value. These solvent amounts, along with the mathematical determination of the scale factor, are detailed in Appendix A. The scale factor used in this paper serves as a benchmark for initial pilot plant trials. In practice, the exact amounts of feedstock, reagents, and solvents will be determined after these trials are complete. It is assumed that the major contaminants in waste PC feed will be flame retardants and organic materials, as further discussed in section 2.3.1. While there will most likely be other contaminants present in the waste polycarbonate, further lab scale and pilot plant scale trials will need to be conducted to identify all contaminants. Detailed kinetics will also need to be obtained from pilot plant studies. Furthermore, this process design does not consider the production of utilities or disposal of waste products.

2.3 Process design and description

A potential process flow diagram for the advanced recycling of polycarbonate waste to BPA as simulated in Aspen Plus® is shown below in Fig. 1.

2.3.1 Polycarbonate washing. PC waste commonly contains flame-retardant impurities and organic material. This preliminary washing method removes the impurities expected to be associated with PC waste. Data on the exact contaminants in each batch of PC will have to be validated before PC is fed as contaminants can vary based on the source of the PC. Due to the limited availability of

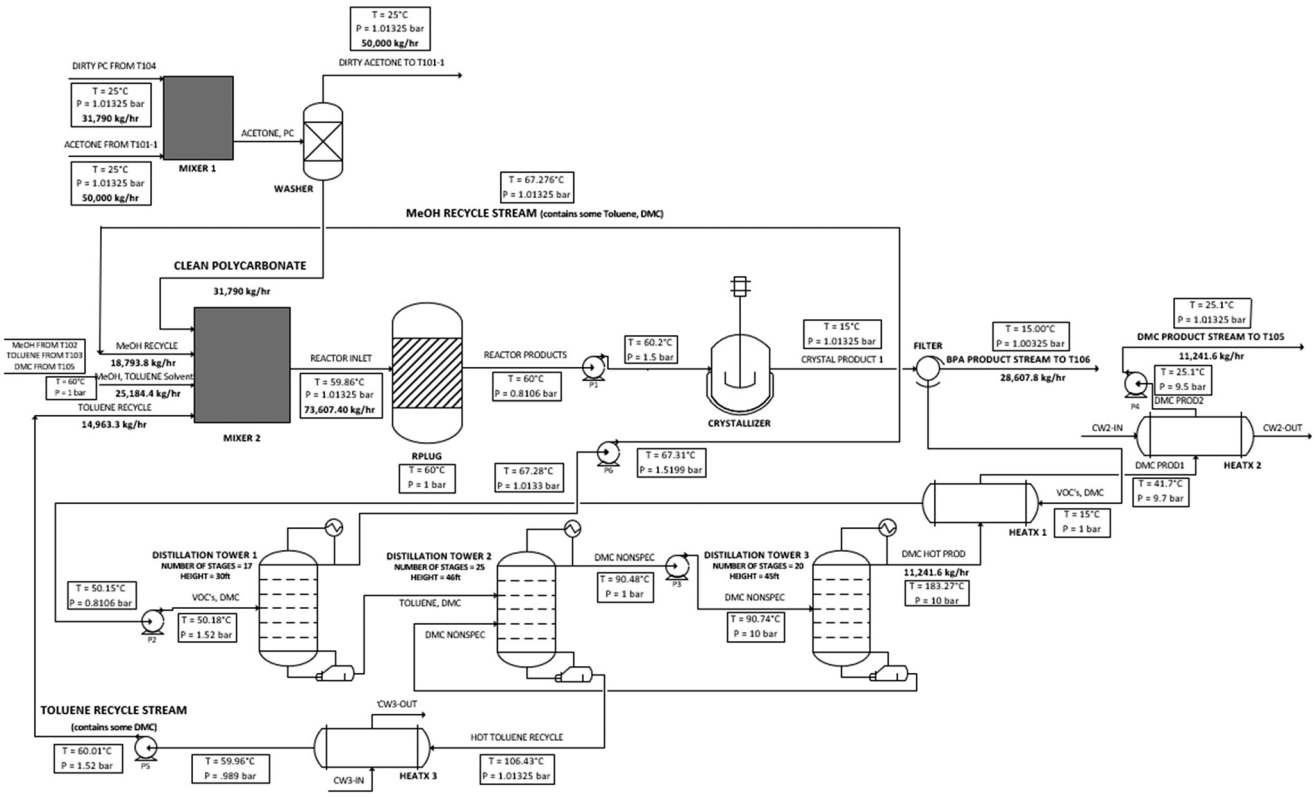

Fig. 1 PFD of the methanolysis of PC to BPA and DMC.

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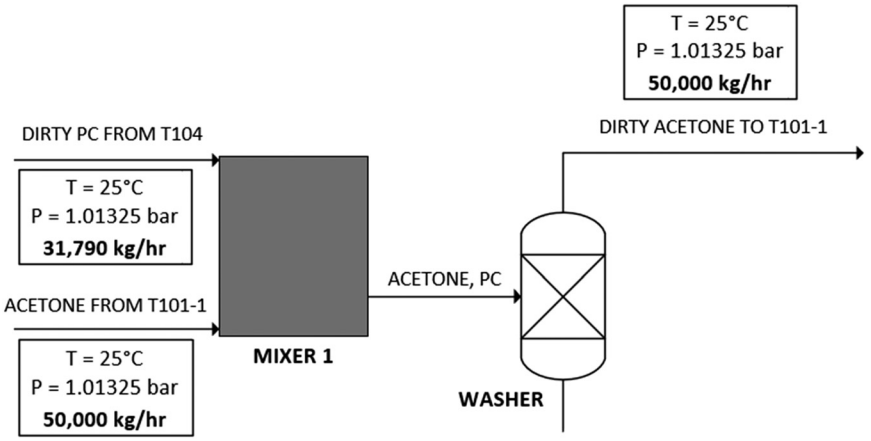

Fig. 2 PFD of the washing step of the process.

contaminant data and the variability in contaminants, this study assumes the PC waste contains flame-retardant impurities and organic material, both of which can be removed *via* an acetone wash. A more rigorous washing step will be required as more data is available on PC waste contaminants and further pilot plant studies are conducted.[7]

Waste PC is mixed with a feed of acetone solvent and is continuously agitated in a washer. The solvent, along with the dissolved contaminants, is separated from the PC and sent to a storage tank. It is then recycled back from the storage tank into the cleaning process to reduce the cost and the volume of acetone required. To produce 200 000 tonnes of BPA per annum, the washing step is estimated to require 50 tonnes of acetone cycling through the system at a given time.[8] As the solvent will eventually become saturated with contaminants, it will have to periodically be replaced with fresh acetone. This washing process is shown below in Fig. 2.

2.3.2 Reactor. The reactor section of the process is shown below in Fig. 3. The reactor conditions are 60 °C and 1 bar. A solvent mixture of methanol and toluene is combined with a fresh feed of PC waste that reacts over a fixed bed of catalytic sodium hydroxide to produce BPA and DMC. The quantity of solvent required at startup was calculated using the PC scale factor. As the system reaches steady state with the recycle streams, a solvent makeup feed is required to replace the amount lost through the product effluent.

To minimize the increased risk of corrosion due to sodium hydroxide, acid neutralization facilities will need to be installed downstream of the reactor to neutralize any sodium hydroxide carry over. A sour water stripper skid will purify the water byproduct and return the water to the process water system. Salt byproducts will be disposed of as process waste. A full detailed description of these systems are outside the scope of this study but allow for further process development.

2.3.3 Crystallization and filtration. The first post-reaction product processing step of the process involves coalescing and purifying the BPA product. This section of the process

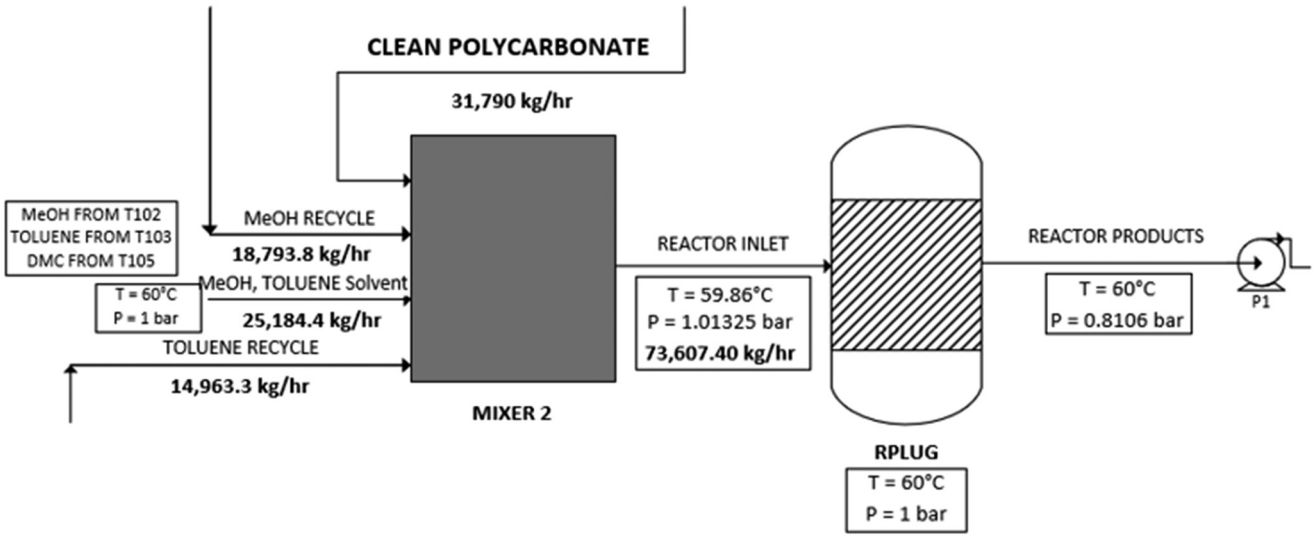

Fig. 3 PFD of the reactor step of the process.

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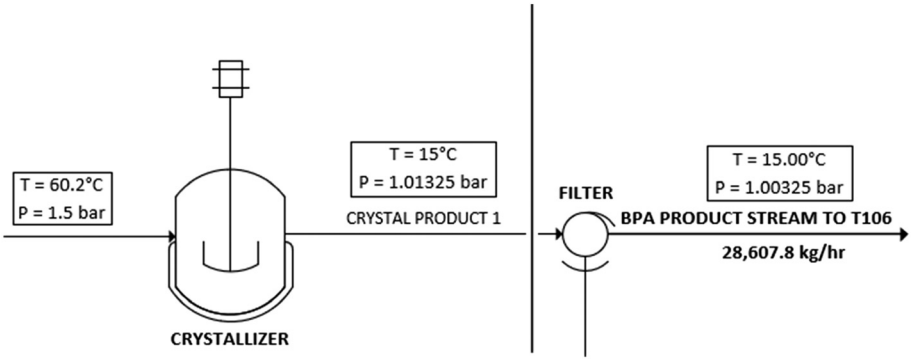

Fig. 4 PFD of the solid separation step of the process.

flow diagram is shown in Fig. 4. Streams of DMC, BPA, methanol, and toluene flow into a surface-cooled crystallizer that is maintained at 15 °C. The temperature change of the process stream from 60 °C to 15 °C drastically reduces the solubility of BPA in toluene, causing most of the BPA to precipitate out from the solution. This slurry of suspended, solid BPA is then fed to a rotary drum vacuum filter (RDVF). This filter will use a mesh with a 5-micron beta 1000 rating to give filtration efficiencies of 99.99%. The filter cake from the RDVF is comprised of the BPA with the desired product concentration of 99.8 wt%. The RDVF effluent stream contains the solvents and the valuable DMC byproduct.

2.3.4 Distillation. The solvent recovery and DMC purification component of the process is comprised of three distillation columns, which are detailed below in Fig. 5. The first distillation column separates nearly all the methanol in the distillate stream, which is then cooled with a heat exchanger and recycled back to the reactor inlet. The second and third distillation columns are used to separate the toluene and DMC streams. However, toluene and DMC form an azeotrope at a 78% molar quantity of DMC. Having the azeotrope at this purity also results in an undesirable amount of DMC to be recycled back to the reactor with the toluene bottoms stream, greatly increasing energy costs and required equipment size. To break this azeotrope, a pressure swing distillation method was used.

Pressure swing distillation uses two columns running at different pressures. The system of toluene and DMC is a minimum-boiling system, in other words, the boiling point of the combined system is lower than the boiling point of the individual components. The lower-pressure column operates at 1.01325 bar while the high-pressure column is operated at 10 bar. The low-pressure column separates the DMC and toluene to the azeotropic limit at atmospheric pressure. The distillate of this lower-pressure column goes through a pump and becomes the feed of the high-pressure column. Because of the much higher pressures, the azeotropic concentration of DMC decreases, bypassing the azeotropic point in column 3. This allows column 3 to achieve higher toluene–DMC separation. The higher purity DMC in the distillate of column 3 is then combined with the initial feed to the low-pressure column. This recycle stream makes it so that the initial feed to the low-pressure column is a higher concentration than the azeotropic concentration at atmospheric pressure, which then bypasses the azeotrope. Fresh toluene will be injected into the reactor to account for solvent loss with DMC. In an

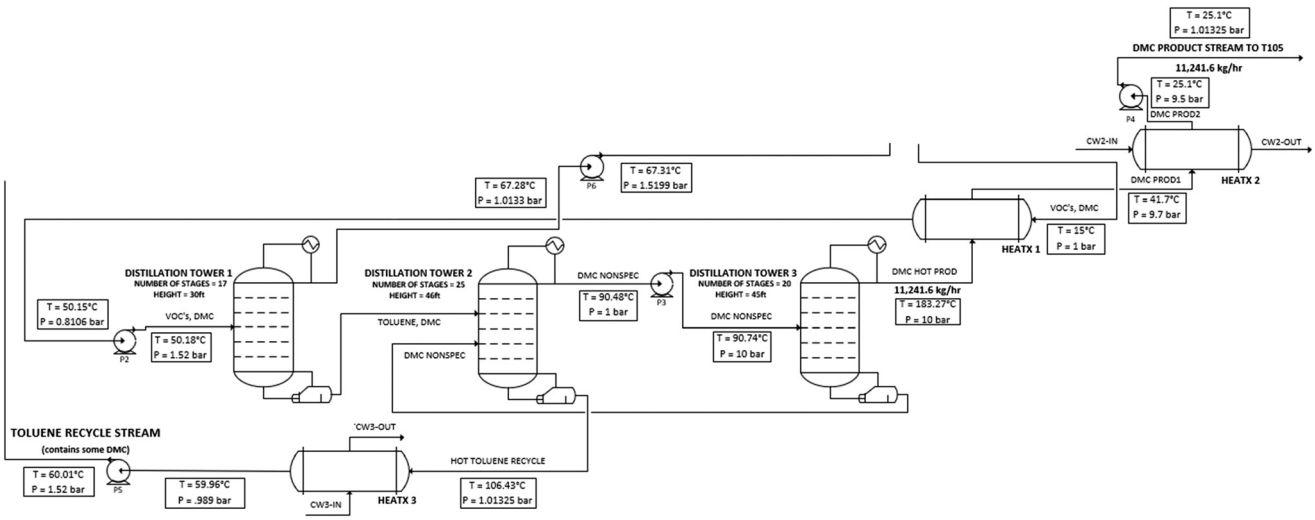

Fig. 5 PFD of the solvent recovery and DMC purification step of the process.

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industrial setting, this will most likely be accomplished by an online product analyzer for DMC (and BPA) with a feedback control loop to open a control valve to allow for more toluene to flow into the reactor. Further pilot plant studies will give valuable experimental data on recycle efficiency and solvent makeup streams.

At steady-state operation, the bottoms of the low-pressure column is at 84.9 wt% toluene and is then cooled and subsequently recycled back to the initial reactor inlet. The distillate of the low-pressure column is pressurized and fed into the high-pressure column. The bottoms of the high-pressure column is recycled back into the feed of the low-pressure column. The distillate of the high-pressure column is 99.9 wt% DMC that is throttled to a lower pressure, passed through a heat exchanger to bring it to room temperature, and sent to storage to be sold as a valuable byproduct. DMC is sold as "high purity" with other studies defining high purity as >99.8 wt% DMC.[9,10]

2.4 Process equipment

2.4.1 Heat exchangers. The heat exchangers were first optimized using heat integration. The process does not require any extreme pressure or temperature differentials, so traditional BEM TEMA-type exchangers were chosen. The hot fluid is in the shell for exchangers 1 and 3. The tube pattern was selected as a 45-rotated square for all exchangers as they ensure optimal heat transfer and ease of maintenance. Single segmental baffle types were selected as they are the standard design. The heat exchanger specifications are shown below in Table 1.

2.4.2 Distillation columns. The distillation columns were optimized with sieve trays spaced 2 ft apart with a diameter of 2.65 m. This data is summarized below in Table 2.

2.4.3 Other process equipment. All other process units are summarized in Table 3 below. The washer was sized using an assumed residence time of 1 hour to determine the necessary volume. The reactor was automatically sized once a rigorous plug-flow reactor was simulated. The crystallizer was sized based on process flow rate and a required tank size was based on the residence time of two hours for surface-cooled crystallizers. The rotary drum vacuum filter was sized using Aspen Process Economic Analyzer®. The distillation columns were sized using rigorous models found in Aspen Plus®. The heat exchangers were sized similarly. The maximum allowable working pressure (MAWP) of the equipment was assumed to be 40 psig if the equipment operates close to atmospheric pressure. Only distillation columns 2 and 3 had the potential to operate at higher pressures. Because column 3 recycles into column 2, column 2 should be built to operate at the same pressure specifications as column 3. The MAWP of column 3 was determined to be 10 barg plus a 10% buffer for safety. The design temperature for all components is based off the operating temperature in the simulation plus a 50 °F safety buffer. The standards used for designing the washer, reactor, crystallizer, RDVF, and columns were ASTM 516. These are the standard specifications for carbon-steel pressure vessels operating for low to moderate temperature service. API 660 was used as the standard for designing the heat exchangers. These standards are relevant for shell-and-tube heat exchangers. The specifications for the other process units are shown below in Table 4. Detailed stream summaries are in Appendix B (Table 10).

Table 2 Distillation tower summaries

Distillation column	Distill 1	Distill 2	Distill 3
No. of stages	17	25	20
Tot. height (m)	9.144	14.021	13.716
Section diameter (m)	3.071	2.650	1.456
Tray type	Sieve	Sieve	Sieve
Tray spacing (m)	0.610	0.610	0.762
Pressure drop (bar)	0.139	0.226	0.177
No. of passes	1	1	1

3. Advantages of the methanolysis of polycarbonate waste process

3.1 Improved process safety

The methanolysis of polycarbonate has several advantages over the traditional Hock process from a process safety perspective. It has become a trend in the past decade to find replacements for hazardous solvents leading to an

Table 1 Heat exchanger specifications

Heat exchanger	HX1 heater	HX2 cooler	HX3 cooler
TEMA type	BEM	BEM	BEM
Hot fluid location	Shell side	Tube side	Shell side
Tube OD/pitch (cm)	1.91/2.38	1.91/2.38	1.91/2.38
Tube length (cm)	540	390	405
Shell ID/OD (cm)	20.5/21.9	20.5/21.9	20.5/21.9
Tube pattern	45-Rotated square	45-Rotated square	45-Rotated square
No. of tubes/passes	40/1	40/1	39/1
Baffle type	Single segmental	Single segmental	Single segmental
No. of baffles	72	18	54
Baffle cut orientation	Vertical	Horizontal	Vertical
Baffle spacing (cm)	7	19	7

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Table 3 Unit summaries

Unit	MAWP (psig)	Design temperature (°F)	Standard	MOC	Characteristic size	Equipment cost (USD)
Washer	40	161	ASTM A516	CS	24 000 gal	544 600
Reactor	40	190	ASTM A516	CS	14 652 gal	218 400
Crystallizer	40	190	ASTM A516	CS	30 000 gal	6 476 500[a]
RDVF	40	161	ASTM A516	CS	1608 sq. ft. filter area	437 800[a]
Distill 1	40	258	ASTM A516	CS	9.1 m tall 3.071 m dia.	668 000
Distill 2	243	273	ASTM A516	CS	14.021 m tall 2.650 m dia.	661 600
Distill 3	243	413	ASTM A516	CS	13.716 m tall 1.456 m dia.	403 300
HeatX1	40	411	API 660	CS	12.74 sq. m	15 800
HeatX2	40	273	API 660	CS	9.15 sq. m	15 200
HeatX3	40	158	API 660	CS	9.28 sq. m	15 200

[a] Starred equipment costs were manually put into Aspen Process Economic Analyzer® because the cost estimates in the main simulations were unreasonably low.

Table 4 Summary of the green metric values for the methanolysis of polycarbonate waste and the Hock process

Green metric	Methanolysis of polycarbonate	Hock process
Atom economy (AE)	1.00	0.93
Reaction mass efficiency (RME)	0.86	0.75
Stoichiometric factor (1/SF)	0.74	0.62
Material recovery parameter (MRP)	1.00	0.63
Yield	0.96	0.90

inherently safer design of a chemical process. The only solvents used in the process are toluene and methanol, both considered green solvents and less hazardous to humans than benzene and sulfuric acid used in the Hock process.[11,12] Furthermore, the methanolysis reaction is run mostly at atmospheric pressure and a mild reaction temperature of 60 °C reducing the risk of overpressure scenarios. There are no explosive intermediates in the methanolysis reaction which is a hazard in the Hock process. Furthermore, by having less hazardous materials at the facility, the risks to the surrounding community are minimized.

3.2 Environmental benefits

The methanolysis of polycarbonate has minimal process wastewater. The Hock process reaction produces water, while the water produced in this process is produced from the neutralization of the catalyst carryover. Furthermore, the emissions from the process are CO_2 from the flare and the production of steam. There is minimal waste from the process as both solvents are recycled and the byproduct, DMC, can be sold. The low process temperatures and pressures result in a lower utility demand than the Hock process. The use of a non-metal catalyst eliminates the need for a non-renewable metal catalyst. Finally, the methanolysis of polycarbonate contributes to a circular economy by reducing the need for virgin materials.

3.2.1 Mass intensity analysis. To compare the mass intensity of the two processes, five standard green metrics were used to analyze the methanolysis of polycarbonate waste and the traditional Hock process. The green metrics evaluated were atom economy (AE), reaction mass efficiency (RME), stoichiometric factor (SF), material recovery parameter (MRP) and yield.[13] The equations to calculate each parameter are shown in Appendix C. The ideal value for each parameter is 1. As shown in Fig. 6 below, the methanolysis of polycarbonate has higher metrics than the Hock process indicating that the process has less waste than the Hock process and therefore less environmental impact. The exact metric values are shown below in Table 4.

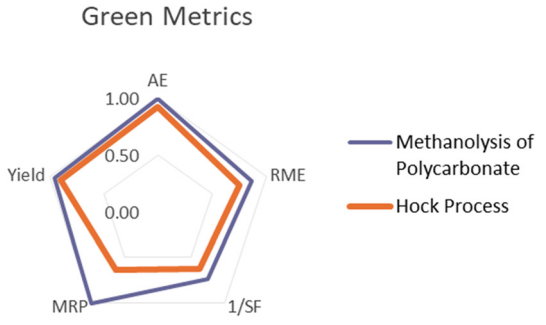

Fig. 6 The comparison of the green metrics for the methanolysis of polycarbonate and the Hock process.

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3.3 Byproduct: dimethyl carbonate (DMC)

DMC is a valuable product since it is a sustainable gasoline additive for meeting oxygenate specifications as outlined in the Clean Air Act. DMC has a high oxygen content of 53.3% and an octane number of 116. Furthermore, DMC in fuels can reduce particle matter, total hydrocarbon and soot emissions.[14] DMC is also used as a nontoxic solvent and an intermediate in pharmaceutical production. In addition, DMC is an electrolyte solvent of lithium-ion batteries mixed with ethylene carbonate. Due to the rising demand for lithium-ion batteries, the demand for DMC is also expected to increase.[10] DMC has a low toxicity and biodegrades quickly. However, the current processes to produce DMC are complex and use toxic reagents such as phosgene. Therefore, there is a global demand for DMC produced from safer and more sustainable sources.[15,16]

4. Challenges with industrial implementation of the methanolysis of polycarbonate waste process

While there are many advantages with the advanced recycling of PC to BPA over the traditional processes, there are still many unknowns about this process that will require a pilot plant and more upfront costs. Further pilot plant scale testing will be required to fully develop this advanced recycling process. One of the biggest challenges with any advanced recycling process is the management of contaminants and balancing the economics of more processing steps to clean the PC feed *versus* higher feed cost by purchasing PC at a higher purity. First, a characterization method will need to be developed with experimental data to efficiently and accurately determine the contaminant composition of PC waste. Some common characterization methods used in advanced recycling today are Fourier transform, infrared spectroscopy (FT-IR), comprehensive 2D gas chromatography (GC × GC), Fourier transform ion cyclotron resonance mass spectrometry (FT-ICR MS), and liquid chromatography.[17] Furthermore, experimental data will be needed to determine how contaminants affect reaction kinetics, catalyst activity, separation efficiency and product purity. In addition to the further development of contaminant impacts, a full detailed mechanism of all byproduct reactions will need to be developed and their impact on the process fully realized. Catalyst life span will need to be determined based on the contaminant load and reaction byproducts. Further economics will need to be developed to determine if an online catalyst change out (with a spare reactor) is feasible. The distillation and product separation will need to be optimized to ensure product quality and energy efficiency. Process utilities will need to be fully developed and integrated into the process including wastewater treatment from the neutralization of sodium hydroxide with acid.

For the sourcing of PC, it may be difficult to find suppliers of the plastic waste for continuous operation of the process for many years. It is recommended that several reliable suppliers are identified to maintain a constant supply of PC waste to the process. A summary of the advantages and disadvantages of the methanolysis of PC process *versus* the traditional Hock process are shown below in Table 5.

5. Economics

5.1 Existing PC advanced recycling facilities

In August of 2023, Covestro® announced successful laboratory trials of polycarbonate advanced recycling *via* chemolysis and are heavily invested into pilot plant trials. While we cannot determine if the Covestro® process is a methanolysis reaction since the company has not released the exact chemical method for the recycling process, Covestro's® investment shows that the advanced recycling of polycarbonate is both possible and profitable.[18]

5.2 Existing PET advanced recycling facilities

Even though the advanced recycling of polycarbonate (PC) has not yet been completed on an industrial scale, advanced recycling of polyethylene terephthalate (PET) plastic is a very similar process that can be used as a proxy for the risk analysis of the advanced recycling of PC. PET is a thermoplastic polymer and is the most used polyester in the plastic industry. There are various PET advanced recycling methods such as methanolysis, glycolysis, and hydrolysis. Since the process we are proposing utilizes the methanolysis reaction, the focus of this section will cover the similarities between the methanolysis of PET and PC. By looking at various successful PET recycling facilities, a rough estimation of process costs, environmental benefits, and capital investments can be extrapolated and compared to the PC recycling process. There are currently several large-scale PET advanced recycling facilities across the globe.

5.2.1 Northwest England PET methanolysis plant. Although the PET and polycarbonate methanolysis processes are not identical, they are similar, and therefore the CO_2 emissions and economics can be extrapolated to give a rough estimates for the methanolysis of polycarbonate process. A table showing estimations of capital investments for the Northwest England PET methanolysis plant can be seen below in Fig. 7.

Although the figure is not directly translatable to the methanolysis of polycarbonate, the process does require similar unit operations and provides a good estimation of capital required.[19]

5.2.2 Eastman Chemical – Kingsport, Tennessee. Currently, Eastman Chemical® is constructing a PET advanced recycling facility with a capacity of 100 000 metric tons. For the first 6 months of 2023, Eastman

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Table 5 Summary of the advantages and disadvantages of the methanolysis of PC *versus* the traditional Hock process

Process parameter	Methanolysis of polycarbonate waste	Hock process
Solvents and reactants used	Methanol and toluene	Benzene and sulfuric acid
Reactor temperature	140 °F	600 °F
Waste streams	All byproducts are profitable	Multiple waste streams
Catalyst	Non-metal catalyst	Precious metal catalysts
Waste reduction	Reducing waste going to landfills	Requires the use of new materials
Financial risk	Process has not been completed on an industrial scale	95% of BPA is produced using this method
Economics	All cases are above a 15% hurdle rate	Some cases fail to meet a 15% hurdle rate

spent $413 MM on capital expenditures. Although not all the capital is being used for the plant, they have stated that "a large majority" of it is. In the future, Eastman is planning to pursue another two PET advanced recycling plants, one in France, and another somewhere in the United States.[20]

5.3 OPEX, CAPEX, ROI, IRR

It was assumed that the polycarbonate waste price is $586 per metric ton.[21–25] This price was determined by the approximate trend associated between the price of various virgin plastics and the price of the respective waste plastic. This assumption is valid since we scaled the waste plastic price of polycarbonate based on data for other plastic and estimated a conservative price. For the government subsidies, there is billions of dollars allocated by the U.S. government through legislation such as the Inflation Reduction Act to advanced recycling projects. While an exact subsidy amount requires an application to the EPA, DOE and other regulatory bodies, a conservative subsidy amount of $160 MM was assumed since other projects have received similar amounts.[26–28]

Due to the uncertainty and variability of a novel process, a sensitivity analysis was conducted to determine which scenarios are profitable and above the hurdle rate of 15%. One estimate for the FCI was the capital investment Eastman Chemical publicly shared for their PET advanced recycling facility. The company states that they spent "$413 MM in capital expenditures" during the first half of 2023 "most of which went to the PET advanced recycling facility". While the exact FCI of Eastman's process is unknown, a conservative estimate of $400 MM for the FCI was used.[20] For a second FCI estimate, the $130 MM FCI of the PET advanced recycling facility in Northwest England was taken as a lower bound.[19] Since the price of waste polycarbonate is widely varied and the cost of cleaning the waste is unknown, two estimates for the price of polycarbonate waste were used. The $586 per metric ton price was determined by the approximate trend associated between the price of various virgin plastics and the price of their respective waste plastics. The $1000 per metric ton price was a conservative estimate to account for unforeseen costs in cleaning and acquiring the waste polycarbonate. For all economic calculations, a $6.4 MM total shipping cost for the polycarbonate was added, the tax rate was assumed to be 25% (estimated total federal and state tax), and a straight-line depreciation method over 10 years with the final salvage value being 10% of the FCI was used. The OPEX was calculated from the sum of the raw material cost, utilities, and the labor/maintenance cost. The results of the sensitivity analysis are shown below in Tables 6 and 7. The yellow color indicates close to the hurdle rate of 15%, while the green color indicates a ROI significantly above the hurdle rate.

As seen in Tables 6 and 7 above, all scenarios show a ROI above the hurdle rate. In addition to the sensitivity analysis, a working estimate of the economics was determined. From the Aspen Process Economic Analyzer®, the purchase price of the equipment for the process was estimated to be $20 MM (scaled to fall 2023 price using CPI, including cost of storage tanks for 30 days of reactants). No extra material multiplier was required since carbon steel will be used for the entire process. The FCI was estimated from a list of multipliers provided to account for installation, process piping, instrumentation,

		M$
Direct Costs	Purchased equipment (22.9% FCI)	**2.80**
	Installation, including insulation and painting (8.3% FCI)	**1.01**
	Instrumentation and controls, installed (6.4% FCI)	**0.78**
	Piping, installed (7.3% FCI)	**0.89**
	Electrical, installed (4.6% FCI)	**0.56**
	Buildings, process and auxiliary (4.6% FCI)	**0.44**
	Service facilities and yard improvements (13.8% FCI)	**1.68**
	Land (1–2% FCI)	**0.24**
Indirect Costs	Engineering and supervision (9.2% FCI)	**1.12**
	Construction expense and contractor's fee (12.8% FCI)	**1.56**
	Contingency (7.3% FCI)	**0.89**
Fixed-capital investment (FCI) = direct costs + indirect costs		**12.21**
Working capital (15% TCI)		**2.15**
Total capital investment (TCI) = fixed-capital investment + working capital		**14.36**

Fig. 7 Estimation of capital investments for PET facility in Northwest England.[19]

Table 6 ROI values for an FCI (CapEx) of $400 MM

PC price ($ per MT)	586	1000
ROI with subsidies	58%	44%
ROI without subsidies	33%	18%

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Table 7 ROI values for an FCI (CapEx) of $130 MM

PC price ($ per MT)	586	1000
ROI with subsidies	176%	131%
ROI without subsidies	98%	53%

Table 8 ROI values for an FCI (CapEx) of $202 MM

PC price ($ per MT)	586	1000
ROI with subsidies	126%	96%
ROI without subsidies	75%	46%

Table 9 Summary of economic parameters

FCI	$201 MM
TCI	$237 MM
Annual operating cost	$284 MM
Cash flow	$585 MM
Annual after-tax profit	$230 MM
ROI	96%
IRR	64%

insulation, electrical, buildings, yard improvement, auxiliary facilities, engineering, construction, contractor's fee, and contingency. For each multiplier, the most conservative value was used. The internal rate of return was calculated by assuming a three-year startup period (with the land cost being $0, FCI for the first and second year being $50 MM and the FCI for the third year being $102 MM) and the after-tax cash flow is a constant $202 MM a year for 10 years. The sensitivity analysis was conducted for the FCI value of $202 MM. The final economic estimations are shown below in Tables 8 and 9 assuming $1000 per metric ton of PC, and government subsidies.

As seen in Table 9 above and in the sensitivity analysis in Table 8, even with conservative estimates, this process is profitable. Further economic analysis will need to be developed for an industrial scale process since there are more upfront costs with a pilot plant design and further research. Economic values for an industrial plant are expected to be lower but still be profitable. Detailed economic calculations are shown in Appendix D.

6. Conclusion

Due to the environmental benefits, minimal process safety risks, and the high profitability of the process, the development of the methanolysis of polycarbonate waste process into BPA should be further researched for commercial applications. For this phase of the process development, further research needs to be done with suppliers of PC waste to determine exact economics for the process and the impurities that will come in the waste. Once the concentration of impurities has been determined, a more detailed washing step can be developed, or it may be determined that it is more economically feasible for a third-party to clean the waste. The full mechanism and chemistry of the reaction (with the known impurities present in the PC waste) will need to be thoroughly analyzed to make sure there are no unwanted side reactions. Eventually, a pilot plant would need to be developed to determine the unknowns associated with the process.

While this process has not been completed on an industrial scale and therefore has an increased economic risk, there has recently been a shift in the chemical industry to implement circular manufacturing into chemical processing. This is seen with companies such as Eastman Chemical® and Covestro® investing in advanced recycling facilities with very similar methods to the methanolysis of polycarbonate waste. More research should be conducted to ensure the development of more sustainable chemical processes to protect our planet for generations to come.

Data availability

The datasets supporting this article have been uploaded as part of the appendices.

Conflicts of interest

The authors have no conflict of interest to declare.

Appendix A: process scale-up calculations

Amount of PC required

$$200\,000 \text{ tonne BPA} = 2 \times 10^{11} \text{ g BPA}$$

*Assuming 80% on-stream efficiency, the facility is operating for 292 days:

$$\frac{2\times10^{11}\text{ g BPA}}{292\text{ days}}\times\frac{1\text{ day}}{24\text{ h}} = 28\,538\,812.79\text{ g BPA h}^{-1}$$

$$28\,538\,812.79\text{ g BPA h}^{-1}\times\frac{1\text{ mol BPA}}{228.29\text{ g BPA}}\times\frac{1\text{ mol PC}}{1\text{ mol BPA}}\times\frac{254.3\text{ g PC}}{1\text{ mol PC}} = 31\,790\,354\text{ g PC h}^{-1}$$

Scale factor

*From the lab-scale experiment done by Hu *et al.*, the amount of PC used was 1.27 g

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$$\frac{31\,790\,354\text{ g PC}}{1.27\text{ g PC}} = 25\,031\,774.8$$

Amount of methanol solvent required

*Assuming an equivalent scale factor for all species.

*The experiment in Hu *et al.* used 1 mL of methanol solvent

$$1\text{ mL MeOH} \times \frac{0.791\text{ g}}{\text{mL}} = 0.791\text{ g MeOH}$$

$$0.791\text{ g MeOH} \times 25\,031\,774.8 = 19\,800\,134\text{ g MeOH} \approx 19\,800.1\text{ kg MeOH}$$

Amount of toluene solvent required

*Assuming an equivalent scale factor for all species.

*The experiment in Hu *et al.* used 1.5 mL of toluene solvent

$$1.5\text{ mL } C_6H_5CH_3 \times \frac{0.866\text{ g}}{\text{mL}} = 1.299\text{ g } C_6H_5CH_3$$

$$0.866\text{ g } C_6H_5CH_3 \times 25\,031\,774.8 = 32\,516\,276\text{ g } C_6H_5CH_3 \approx 32\,516.3\text{ kg } C_6H_5CH_3$$

Appendix B: detailed stream summaries

Appendix C: green metric equations

Atom economy (AE)

$$\text{AE} = \frac{\text{molecular mass of desired product}}{\text{molecular mass of the reactants}}$$

Reaction mass efficiency (RME)

$$\text{RME} = \frac{\text{Mass of desired product}}{\text{mass of reactants}}$$

Stoichiometric factor

$$\text{SF} = 1 + \frac{\text{mass of excess reactants}}{\text{mass of stoichiometric reactants}}$$

Material recovery parameter

$$\text{MRP} = \frac{1}{1 + (\text{yield} \times \text{AE}) \times \frac{(c+s+w)}{\text{SF} \times m_p}}$$

where c is the mass of the catalyst, s is the mass of the solvent, w is the mass of the waste and m_p is the mass of the product.

Table 10 Detailed stream summaries

Stream summaries	Flowrate (kg h^{-1})	Temperature/pressure (C bar^{-1})	PC (mass frac.)	Methanol (mass frac.)	Toluene (mass frac.)	Acetone (mass frac.)	BPA (mass frac.)	DMC (mass frac.)
Dirty PC	31 790	25/1.01325	1	0	0	0	0	0
Acetone	50 000	25/1.01325	0	0	0	1	0	0
Clean PC	31 790	25/1.01325	1	0	0	0	0	0
Acetone recycle	50 000	25/1.10325	0	0	0	1	0	0
Solvent feed	25 184.4	59.9/1.01325	0	0.4771	0.5229	0	0	0
DMC from recycle	8564	60/1.51988	0	0	0	0	0	1
MeOH reactant	8010	60/1.01325	0	1	0	0	0	0
Reactor inlet	73 607.4	59.9/1.01325	0.4319	0.2723	0.1795	0	0	0.1163
Reactor outlet	73 607.4	60/0.8106	0	0.1635	0.1795	0	0.3877	0.2693
High P reactor out.	73 607.4	60/1.51988	0	0.1635	0.1795	0	0.3877	0.2693
Crystallizer outlet	73 607.4	15/1.01325	0	0.1635	0.1795	0	0.3877	0.2693
BPA product	28 607.8	15/1.00325	0	0.0006	0.0006	0	**0.998**	0.0008
VOCs	44 999.6	15/1.00325	0	0.2670	0.2931	0	0	0.4399
Heated VOCs	44 999.6	50.15/0.8183	0	0.2670	0.2931	0	0	0.4399
High P heated VOCs	44 999.6	50.19/1.51988	0	0.2670	0.2931	0	0	0.4399
Distill1 tops	18 794.6	67.28/1.01325	0	0.6392	0.0266	0	0	0.3342
MeOH recycle	18 794.6	67.31/1.51988	0	0.6392	0.0266	0	0	0.3342
Distill 1 bottoms	26 205	98.28/1.01325	0	0	0.4843	0	0	0.5157
Distill 2 bottoms	14 963.3	106.43/1.01325	0	0	0.8475	0	0	0.1525
Cooled tol. recycle	14 963.3	59.97/0.9897	0	0	0.8475	0	0	0.1525
Toluene recycle	14 963.3	60/1.51988	0	0	0.8475	0	0	0.1525
Distill 2 tops	22 492.6	90.5/1.01325	0	0	0.01824	0	0	0.9817
High P distill 3 feed	22 492.6	90.75/10	0	0	0.01824	0	0	0.9817
Distill 3 bottoms (PSD cycle)	11 250.9	184.13/10	0	0	0.0357	0	0	0.9643
Distill 3 tops	11 241.6	183.27/10	0	0	0.0007	0	0	0.9993
Hot DMC prod. 2	11 241.6	41.71/9.74176	0	0	0.0007	0	0	0.9993
High P DMC prod.	11 241.6	25/9.54627	0	0	0.0007	0	0	0.9993
DMC product	11 241.6	25.1/1.01325	0	0	0.0007	0	0	**0.9993**

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Appendix D: detailed economic calculations

CAPEX calculations

The purchase price for the plant from Aspen Process Economic Analyzer® (Aspen®) was $14 MM (2013 dollars). This price was scaled up to the fall 2023 USD price using a ratio of CPI values. The price of the tanks for storage of reactants was calculated to be $5 MM from Aspen®. The price of the crystallizer was underestimated by Aspen® by approximately $3 MM. These three values were summed up to get the total purchase price, which was then scaled to their 2023 USD price resulting in a purchase price of $20 MM. A cost multiplier was used to calculate the FCI. The most conservative values for each multiplier were used as shown in Table 11 below.

From these calculations, the FCI was estimated to be **$201 843 914.95**.

WCI was estimated from 15% of the TCI.

OPEX calculations

*Assume $1000 per MT for PC raw material cost. This is already factored into the total raw material cost

$$\text{OpEx} = \text{Tot. Raw Material Cost} + \text{Utilities} + \text{Labor/Maintenance}$$

$$\text{OpEx} = \$271.4\ \text{MM} + \$11.26\text{MM} + \$2.06\ \text{MM}$$

$$\text{OpEx} = \$284.7\ \text{MM}$$

ROI calculations

Annual Income. *Government subsidies are considered

$$\text{Ann. Income} = \text{BPA Product Cost} + \text{DMC Product Cost} + \text{Subsidies}$$

$$\text{Ann. Income} = \$346.02\ \text{MM} + \$79.64\ \text{MM} + \$160\ \text{MM}$$

$$\text{Ann. Income} = \$585.7\ \text{MM}$$

Table 11 Summary of cost multipliers for FCI

Cost item	Multiplier	Total cost (USD)
Total plant direct cost (TPDC)		
Installation	1.5	31 356 830.04
Process piping	0.6	12 542 732.01
Instrumentation	0.6	12 542 732.01
Insulation	0.05	1 045 227.67
Electrical	0.2	480 910.67
Buildings	3	62 713 660.07
Yard improvement	0.2	4 180 910.67
Auxiliary facilities	1	20 904 553.36
Total plant indirect cost (TPIC)		
Engineering	0.3	6 271 366.00
Construction	0.4	8 361 821.34
Total plant cost		164 100 743.90
Contractor's fee	0.08	13 128 059.51
Contingency	0.15	24 615 111.58
Direct fixed capital		201 843 914.90

Depreciation. *Assuming straight-line depreciation, 10 year period, salvage value = 10% of FCI

$$d = \frac{\$201.8\ \text{MM} - \$20.2\ \text{MM}}{10\ \text{years}} = \$18.2\ \text{MM per year}$$

Annual after-tax profit. *Assuming tax rate is 25% (state and federal taxes)

$$\text{Ann. Net (after tax) Profit} = (\text{Ann. Income} - \text{OPEX} - \text{Depreciation}) \times (1 - \text{tax rate}) + \text{Depreciation}$$

$$(\$585.66\ \text{MM} - \$284.7\ \text{MM} - \$18.2\ \text{MM}) \times (1 - 0.25) + \$18.2\ \text{MM} = \$230.3\ \text{MM}$$

ROI. *Assume WCI is 15% of TCI

$$\text{ROI} = \frac{\text{Ann. Net (after tax) Profit}}{\text{TCI}} \times 100\%$$

$$\text{ROI} = \frac{\$230.3\ \text{MM}}{\$237.5\ \text{MM} + \left(0.15 \times \frac{\$237.5\ \text{MM}}{0.85}\right)} \times 100\%$$

$$\text{ROI} = 96.97\%$$

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