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Exploring sustainable adsorbents to mitigate micro-/nano-plastic contamination: perspectives on electrospun fibrous constructs, biochar, and aerogels

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In recent years, the issues pertaining to the micro-/nano-plastics (MNP) pollution in urban water have escalated due to their detrimental environmental consequences, which not only disrupt aquatic habitats and harm marine life but also serve as vectors for toxic pollutants, potentially entering the food chain and posing risks to human health. Although conventional techniques such as filtration, sedimentation, and electrocoagulation have been extensively utilized for MNP removal, ongoing concerns persist regarding their effectiveness, sustainability, and cost implications. Hence, it is imperative to critically assess the performance of conventional techniques in addressing MNP-induced pollution and to shed light on the potential of some emerging technologies as promising next-generation solutions. However, the dearth of standardized approaches and the scarcity of comprehensive data contribute to the disturbing extent of uncertainties in utilizing such techniques to address MNP pollution. Therefore, the current review theoretically emphasizes on innovative use of biochar, electrospun fibers, and aerogels as sustainable adsorbents for managing MNP pollution. Further, it offers a systematic overview elucidating the foundational understanding pertaining to the sources, fate, and transport dynamics of microplastics within the environment. Apart from this, the article explores the integration of such adsorbents into existing water treatment systems and examines the associated challenges and future perspectives in real-world applications. Thus, the contextual review provides valuable insights into designing next-generation technologies aimed at controlling MNP pollution in water systems which not only enhances the understanding of the fate and transport mechanisms of emerging MNP pollutants treatment of water to make it portable but also assists material designers in evaluating and refining existing methodologies and thereby promoting a multi-faceted and synergistic approach to combat the complex issue of MNP pollution.

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Environmental significance

The ever-increasing threat pertaining to micro-/nano plastics pollution has sparked alarm across the globe due to the severe repercussions for both the environment and human health. The persistence of these particles in the environment poses a concern due to their ability to absorb and transport toxic contaminants, resulting in bioaccumulation in aquatic creatures and subsequently influencing the food chain. These particles can harm marine life, ranging from plankton to bigger creatures, causing food chain disturbances and ecological imbalances. The assimilation of MPs by marine creatures can lead to physical harm, toxicity, and the transmission of pollutants up the food chain, thereby affecting human populations dependent on seafood. Thus, this comprehensive review article gives insights into the limitations of traditional methods for removing MNP and recommends a paradigm shift towards new sustainable adsorbents, especially biochar, electrospun fibrous constructions, and aerogels. The investigation of these novel materials as next-generation solutions represents a significant divergence from traditional options which not only advances our understanding of MNP pollution but also provides a strategic framework for future research, fostering a multifaceted and synergistic solution to protect the environment from the pervasive threat of MNP.

1. Introduction

The overexploitation of single-use plastics (>300 million tons each year) and their mismanagement led to the accumulation of micro and macro plastics in aquatic, terrestrial, and atmospheric surroundings, resulting in MNP-induced pollution as

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an emerging problem, considering the fact that it has percolated down through the soil streams all through and dispersed up above up to the cloud systems. Reportedly, plastic production globally exceeds ~450 million metric tons (MMT) annually, and without intervention, it is projected to double or triple by 2060.¹ As per the report obtained from the International Union for Conservation of Nature (IUCN), ~3.2 million tons of plastic waste are dumped into the atmosphere in a yearly manner, while ~1.5 million tons of which end up in the oceans. Consequently, the amount of plastic waste generated globally has not only risen up to ~353 MMT in 2019 (OECD 2022) but is also estimated to surge ~6-fold by 2060.² Meanwhile, the prolonged persistence of plastics (which can last up to ~1000 years) in the environment coupled with their potential to release harmful chemicals (such as bisphenol A and related chemicals, phthalates, per and poly-fluoroalkyl substances (PFAS), dioxins, toxic metal ions (lead and cadmium)) and/or other toxic additives during degradation affects the aesthetics of natural habitats but also disrupts thriving ecosystems. For instance, it has shown instances of devastating effects on marine life (such as seabirds, turtles, fish, and marine mammals) as they often mistake plastic debris for food, resulting in the ingestion and entanglement, ultimately causing injury, suffocation, and even death and decimation.³ The most significantly susceptible ecosystem component to microplastics (MPs) is water, which has a vicious nexus and non-escapable interdependence. Due to the invariably lesser density of MPs, generally, they incline to either settle down on the ground or disperse in the air depending on their size and substrate-specific functionality on MPs.⁴ However, most of the MPs eventually get transported to the water system which leaves the water extremely contaminated, to the extent of alarming malignancy.⁵

Further, with passing years, plastic waste can break down into smaller fragments (<5 μm) *i.e.*, MPs (~0.1 μm to 5 mm) and nano-particles (NPs) (0.001–0.1 μm) by the exposure to the ultraviolet (UV) irradiation and weathering processes, thereby causing its infiltration into aquatic and terrestrial environments.⁶ For example, polystyrene (PS)-based MPs (~10 μm) can induce oxidative stress in epithelial cells as well as the human brain and adsorb other toxic organic pollutants (such as polycyclic aromatic hydrocarbons (PAHs), and persistent organic pollutants), which may eventually facilitate their transfer into different organs and tissues.⁷ NPs with smaller size, lesser density, and higher surface area, inherit the negative impacts of MPs on the aqueous environment and may exert more serious impacts, such as ingestion by different aquatic and terrestrial organisms, piling up in various body parts like liver, gills, blood streams of organisms which may cause metabolic disorders, severe infections and disruption of various cycles of the body.⁸ For instance, Wang *et al.* reported that PS-based MNP can not only decrease the cell viability of Caco-2 cells (derived from colon carcinoma) but also tend to induce cytotoxicity linked to oxidative stress, epithelial cell injury, disruption of intestinal barrier function, and altered detoxification processes.⁹ In another *in vitro* simulation experiment conducted by Fournier *et al.* cultured gut microbiota obtained from fecal samples observed a significant shift in the community composition of

the gut microbiota after interaction with polyethylene terephthalate (PET)-based MNP, which had diameters ranging from ~30 to ~140 μm . Such interaction resulted in an increased proportion of clostridium and Bacteroides.¹⁰

Indeed, the concentration of MNP has experienced a significant increase (~10-fold) in the last two decades (*i.e.* from 2000 to 2020), indicating a growing accumulation of these particles in the three functional components of our ecosystem (air, water, and soil). For example, Kawecki *et al.* conducted a study and developed a model with parameterization specific to Switzerland. The findings revealed that ~540 \pm 140 grams per capita per year ($\text{g cap}^{-1} \text{y}^{-1}$) of seven analyzed polymers were released into the soil as macro- and MPs combined. In addition, the distribution of these polymers was further divided compositionally into ~98 \pm 50 $\text{g cap}^{-1} \text{y}^{-1}$ of High-Density Polyethylene (HDPE), ~94 \pm 34 $\text{g cap}^{-1} \text{y}^{-1}$ of Low-Density Polyethylene (LDPE), ~24 \pm 13 $\text{g cap}^{-1} \text{y}^{-1}$ of PS, ~16 \pm 12 $\text{g cap}^{-1} \text{y}^{-1}$ of EPS, ~65 \pm 36 $\text{g cap}^{-1} \text{y}^{-1}$ of polyvinyl chloride (PVC), ~200 \pm 120 $\text{g cap}^{-1} \text{y}^{-1}$ of PET and ~126 \pm 43 $\text{g cap}^{-1} \text{y}^{-1}$ of PP.¹¹ According to data published by the United Nations Educational Scientific and Cultural Organization (UNESCO), it is estimated that as of 2022, the ocean contained a staggering ~50–75 trillion pieces of plastic and MPs.¹² Meanwhile, even in remote areas like the Italian Alps, broken ice was found to contain a maximum of ~74.4 MNP per kg, while on the exterior periphery of the glacier, it was projected to be ~131 million MNP (even in the Tibetan Plateau, which boasts some of the highest altitudes in the world).¹³ Even in high-altitude places (like the Tibetan Plateau), with the highest elevations in the world (with an average height of >4000 meters), the presence of MNP has been identified in water streams like lakes and rivers,¹⁴ which suggests that MNP pollution is not limited by geographic boundaries and can reach even the most remote and seemingly isolated areas. In a recent study, researchers have discovered that the spread of MPs through atmospheric transport depends on their shape. In a model, fibers with a length of up to ~1.5 mm were able to reach the most remote places on earth up to the stratosphere, while spheres of the same mass settled much closer to the respective regional plastic sources.¹⁵

The presence of NPs can disrupt the natural cycles and processes in the bodies of various organisms. For example, NPs can interfere with endocrine functions, affecting hormone production and regulation, which can have cascading effects on reproductive cycles, development, and behavior of organisms.¹⁶ Moreover, the COVID-19 pandemic has further exacerbated this issue by increasing the extensive use of single-use plastics, such as masks, gloves, and personal protective equipment (PPE), combined with the surge in packaging waste from online shopping and food delivery services.¹⁷ Globally, ~8 million tonnes of pandemic-related plastic trash are created, with more than ~25 000 tonnes entering the ocean.¹⁸ In another study, the global amount of face masks entering the oceans experienced a significant increase in 2020, with an estimated additional ~1.56 billion masks contributing to marine pollution during that year alone.¹⁹

Conventional decontamination removal technologies including coagulation, dissolved air flotation, physical



separation technologies, membrane disc-filters, sand filtration, coagulation, enhanced oxidation, magnetic separation, chemical agglomeration, and membrane filtration (such as ultrafiltration, microfiltration, nanofiltration, membrane bioreactors, and reverse osmosis) have demonstrated promising results in addressing the issue of MNP by exhibiting high removal efficiencies of >90%, along with the possibility of retrofitting at minimal cost compared to alternative techniques.²⁰ However, it is critical to comprehend that the performance of the membrane filters is affected by different parameters such as materials used in membrane preparation, orientation, profile, size, wettability feature, and surface area.²¹ The pore sizes of conventional membranes may not be optimal for the removal of certain pollutants such as NPs due to their size scale fluctuations, thereby posing challenges in achieving adequate retention.²² Furthermore, conventional membrane filtration is associated with higher costs and limited process intensification, which can impact the setup and performance of the filtration system.²³ Therefore, considering the challenges associated with conventional membrane-based technologies, there is an urgent need for further research on innovative, sustainable, and renewable technologies for the effective removal of MNP that can contribute to the development of efficient and eco-friendly solutions for the removal of MNP from various environments. In this context, Isaeva *et al.* summarized the report on modern carbon-based adsorbents, including graphene oxide (GO), carbon nanotubes (CNTs), metal-organic framework (MOF)-carbon composites, and MCs (microporous carbons) obtained through controlled pyrolysis of MOF-precursors for isolating MPs.^{24,25} In another study, Yang *et al.* investigated the performance of chitosan (CS)-based hydrogel systems for the removal of various types of pollutants especially MPs in an efficient manner.²⁶ The 3D network structure of hydrogel possesses outstanding water holding and adsorption capacity along with reversible swelling ability enabling the removal of a wide variety of MNP. Recently, Goh *et al.* discussed the significance of multifunctional and flexible nanomaterials by recognizing and eradicating the MPs particles through different mechanistic approaches such as adsorption, advanced oxidation process, magnetic separation, membrane filtration, and many more.²⁷ To resolve the complexity with efficacy and to explore the benefits of nanomaterials in MPs treatment processes in a more sustainable manner, the overview intends to stretch further to multifunctional adsorbents like aerogels, biochar, and electrospun mats. In this regard, Yon *et al.* synthesized durable zinc MOF-based composite material ZIF-8@aerogel by *in situ* grown ZIF-8 on wood aerogel fibers, and it is being utilized for the successful removal of MPs including poly(1,1-difluoroethylene) (60–110 nm) and PS (90–140 nm) from the aquatic medium with a removal efficiency of ~91.4% and 85.8%, respectively.²⁸ Further, Zhu *et al.* synthesized a modified porous cellulose nanofibrous-based aerogel through the chemical modification of bleached eucalyptus pulp by 2,2,6,6-tetra methyl piperidine-1-oxyl and 2,3-epoxy propyl trimethyl ammonium chloride. This aerogel demonstrated highly efficient separation of different PS-MPs, achieving near 100% efficiency for carboxylate-modified PS and ~75% for amine-

modified PS.²⁹ Wang *et al.* have substantiated the effectiveness of corn straw and hardwood-derived biochar in MPs removal through experimental studies. The results revealed that biochar filters exhibit significant capacity for the removal and immobilization of 10 μm diameter MPs spheres (>95%), surpassing the performance of similar grain-sized sand filters studied.³⁰ Rist *et al.* fabricated bio-based polyamide 6.9-based electrospun membrane as filters, prepared from plant oil-based azelaic acid and electrospun from chloroform/formic acid to produce self-standing electrospun nonwovens, with high mechanical properties and resistance to solvents. The electrospun mats demonstrated high efficiencies of up to 99.8% for the filtration of PS microparticles (PS-MPs) from water.³¹

A comprehensive overview of the previously published review articles related to MPs removal is listed in Table 1. To date, there has been no report investigating an overall systematic compilation on the understanding of such emerging frameworks that could be useful for potential MPs removal in water treatment. To address these critical knowledge gaps, the objectives of this work are to provide insights into the sources of nano- and MPs, the implications of pandemics such as COVID-19 on MPs pollution, the journey of such plastics, their environmental fate, followed by standards for the detection of MNP contaminants in water and soil. By critically evaluating the limitations of conventional techniques that are being used for the removal of MNP, the current review aims to shift the attention towards the exploration of the recently developed material-specific sustainable ways *i.e.*, the use of biochar, electrospun micro/nanofibrous assemblies, and aerogels as emerging solutions in the field of MNP filtration from soil and water ecosystems. Additionally, it further delves into the challenges in implementing these materials for the removal of MNP and highlights the research gaps in this critical aspect of concern. Summatively, the paper offers strategies and recommendations for future research to improve the application of these adsorbents in real water treatment systems, thereby paving the way for a more comprehensive and synergistic approach to combatting the complex issue of MNP pollution.

2. Methodology

A systematic literature review was meticulously conducted over a span from December 2022 to November 2023, employing a three-step methodology that encompassed (i) identification and compilation of scholarly materials, (ii) screening of articles and meticulous data acquisition, and (iii) comprehensive reporting and widespread dissemination of findings. Various esteemed electronic and scientific databases such as Google Scholar, ACS Publication, ScienceDirect, Springer Link, and Web of Science were utilized to explore the originally published research and review articles from 2016 to 2023. The keywords/search terms such as the emerging level of MP, their harmful effect, toxicity assessment, and emerging adsorbents for MNP removal were used. Based on the title, keyword, and abstract of the extensive pool of scholarly works, similar types of articles (~180) were selected for comparative analysis. The conventional techniques utilized for the removal of MNP are thoroughly analyzed, taking into account their limitations including sustainability, and



Table 1 A comprehensive overview of the previously published review articles related to MPs removal

S. no.	Paper title	Highlights	Aspects not discussed/ limitations	Ref.
1	Review of MPs distribution, toxicity, analysis methods, and removal technologies	<p>Major MPs were reported from PET, polyurethane (PU), PS, polyvinyl alcohol (PVC), polypropylene (PP), polyethylene (PE), and polyamide (PA)</p> <p>Different sampling, pretreatment, and separation techniques (density-based), were discussed</p> <p>Analytical analysis methods for different MPs applications have been discussed in detail</p> <p>Conventional removal techniques such as wastewater treatment plants (WWTPs), physical (flotation, sedimentation, and filtration), chemical, and biological technologies (activated sludge treatment, aerobic and anaerobic digestion, lagoons, and septic tanks) have been discussed in full detail</p>	<p>Risk factors and toxicity of the MPs in the ecosystem and human body have not been mentioned fully</p> <p>All the mentioned MPs removal techniques are not that efficient and the efficiency is only 70%</p>	32
2	A critical review of MPs removal from the environment	<p>The study focussed on the source and occurrence of MPs studies on the physical, chemical, and biological techniques for MPs removal from the environment have been critically carried out along with the advantages and limitations of each removal technique are identified</p>	<p>Major research has been done on the dissemination pathway, environmental occurrence, and distribution in the air, soils, sediments, rivers, lakes, and oceans</p> <p>The lack of investigation on MPs ecotoxicity and remediation technologies has been profound and needs attention to address the research gaps</p>	33
3	MPs pollution focused on sources, distribution, contaminant interactions, analytical methods, and wastewater removal strategies: a review	<p>The evolution of MPs sources, the toxicological effect on the environment and human life, its extraction technique, and analysis have been briefly mentioned</p> <p>A comprehensive study on the existing and conventional removal techniques such as WWTPs, biological wastewater treatment with microfiltration or ultrafiltration in a membrane bioreactor, electrocoagulation, and sol-gel method were discussed</p> <p>A new approach to the utilization of coated Fe nanoparticles to magnetize plastics, allowing magnetic</p>	<p>Standardization is not available, leading to the misidentification of MPs, future work on the precise sampling, treatment, solution preparations, and equipment was required for more precise evaluation</p> <p>Several new hybrid and advanced treatment processes have not been explored properly, whereas the existing mentioned technology has been tested on a laboratory scale only with a few MPs</p> <p>New treatment alternatives must continue to be explored and tested in real samples, where a mixture of</p>	34



Table 1 (Contd.)

S. no.	Paper title	Highlights	Aspects not discussed/ limitations	Ref.
4	A review of methods for extraction, removal, and stimulated degradation of MPs	<p>extraction and the isolation of MPs, and dynamic membrane technology was explored</p> <p>The extraction of MPs from water probes, techniques for identification of polymer types, and strategies for the stimulated natural and artificial degradation of the material are comprehensively discussed</p> <p>Recent trends in the recycling of MPs and pollution control were explored in this study</p> <p>The review covers about two hundred recent sources and briefly explains the most appropriate techniques for a given task</p>	<p>MPs and different contaminants is present. It is also important to approach both lab tests and the industrial scale to ensure the viability of strategies to be implemented in future wastewater treatments</p> <p>Sophisticated laboratory devices combining optical microscopy with FTIR or Raman as well as FTIR-TGA couples for the analysis of MPs are not explored</p>	35
5	A review on MPs separation techniques from environmental media	<p>Separation techniques for MPs removal from environment matrices are explored</p> <p>Different environmental media like soil, water, and sediments have been considered</p> <p>Removal efficiency of the proposed system with advantages and limitations has been mentioned</p> <p>MPs separation helps to understand system performance and pollution estimation</p>	<p>Recovery rate, lack of comparative data, low fiber separation, contamination, quality assurance, and separation measures are not thoroughly discussed</p> <p>Lack of information about MPs separation from the soil</p>	36

Thematic focus of the present review

Provides insights into the sources of nano- and micro-plastics, and implications of the pandemics such as COVID-19 on MPs pollution

Describes environmental consequences followed by the conventional methods used for the removal of MNP contaminants in water and soil

Explores some of the recently developed material-specific sustainable ways *i.e.*, the use of biochar, electrospun micro/nanofibrous assemblies, and aerogels as emerging solutions in the field of MNP filtration

Delves into the challenges in implementing these materials for the removal of MNP and highlights the research gaps in this critical aspect of concern

Offers strategies and recommendations for future research to improve the application of these adsorbents in real water treatment systems

efficacy parameters. Additionally, recently available relevant references pertaining to the utilization of electrospun fibers, biochar, and aerogels as potential adsorbent systems for MNP were meticulously curated and integrated into the review.

The book chapters, letters, opinions/perspectives, abstracts from conferences/seminars, encyclopedias, and viewpoints were not considered for the future purpose. Throughout the

research endeavor, any kind of queries, disputes, and confusions were effectively resolved through collaborative group discussions, paper revisions, and laboratory meetings. Furthermore, the review, as well as the research article, were systematically categorized according to various aspects such as the escalating MNP from diverse origins, their direct and indirect consequences on the environment and living species,



conventionally available methodology, and emerging techniques/systems for their treatment in eco-friendly/sustainable manner. The obtained articles were incorporated into Mendeley software for reference citation purposes, with duplicate results systematically removed. The schematic illustration of the performed screening process is shown in Fig. 1.

3. A quantitative outlook on the sources of MPs

MNP originate from a diverse range of sources, including adhesives, paints, coatings, pharmaceuticals, biomedical materials, electronics, magneto/optoelectronic devices, medical equipment, plastic sheets in the food department, marine fisheries, disposable materials for carriers, precursors used in industries and

agricultural films.³⁷ The schematic representation concerning the diverse sources of MNP generation is illustrated in Fig. 2a.

3.1 MPs from packaging, cosmetics, and textile industries as fast-moving consumer goods (FMCG) sources

The cosmetics industry has posed a significant challenge concerning MPs pollution primarily attributed to the use of microbeads (composed of materials like polyethylene (PE) or polypropylene (PP)) in a wide range of personal care products such as scrubs, toothpaste, and shower gels. For example, a study conducted by Guerranti *et al.* revealed that cosmetics usage in the countries from the European Union is found to be ~4000 tons of microbeads annually.⁴¹ Furthermore, research by Cheung and Fok in 2017 highlighted that the aquatic environment in mainland China is facing contamination with an estimated average number of microbeads of ~210 trillion, which

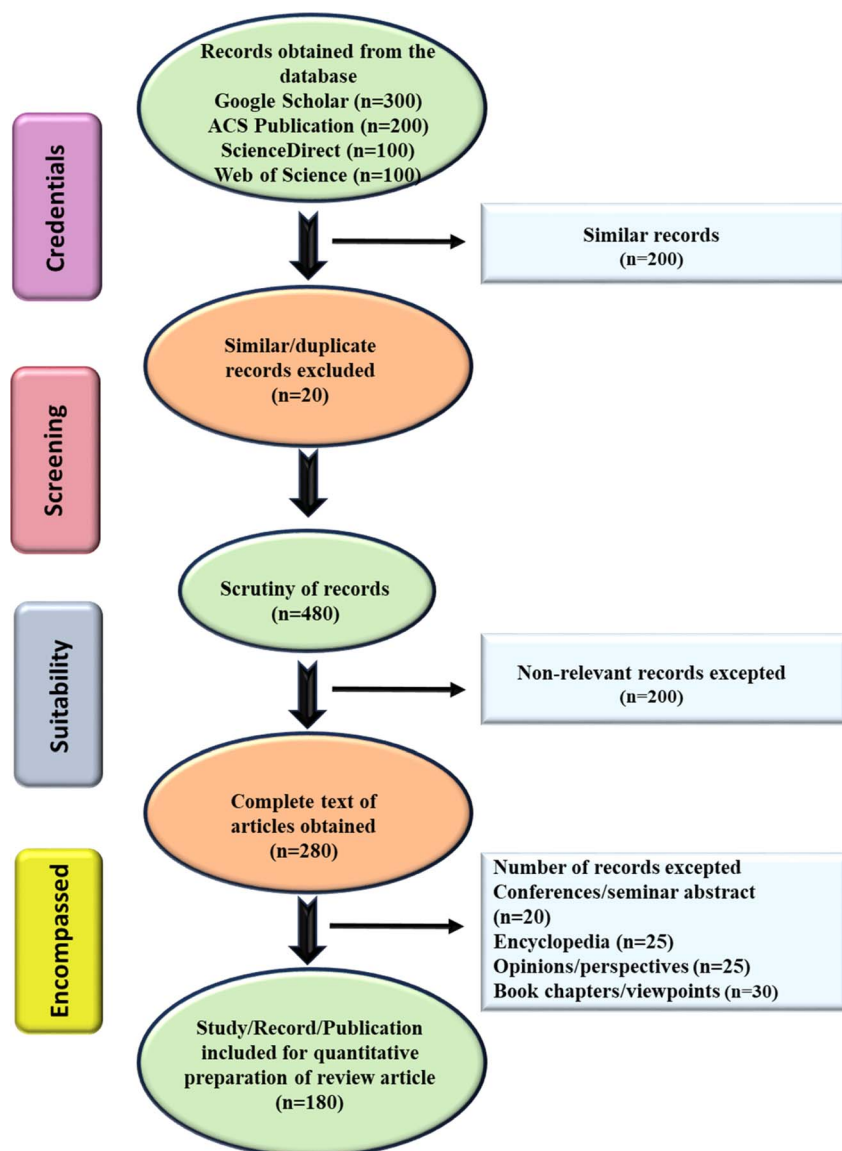


Fig. 1 A schematic representation of the process involved in screening the databases and records which helped in the effective outcome of this study.



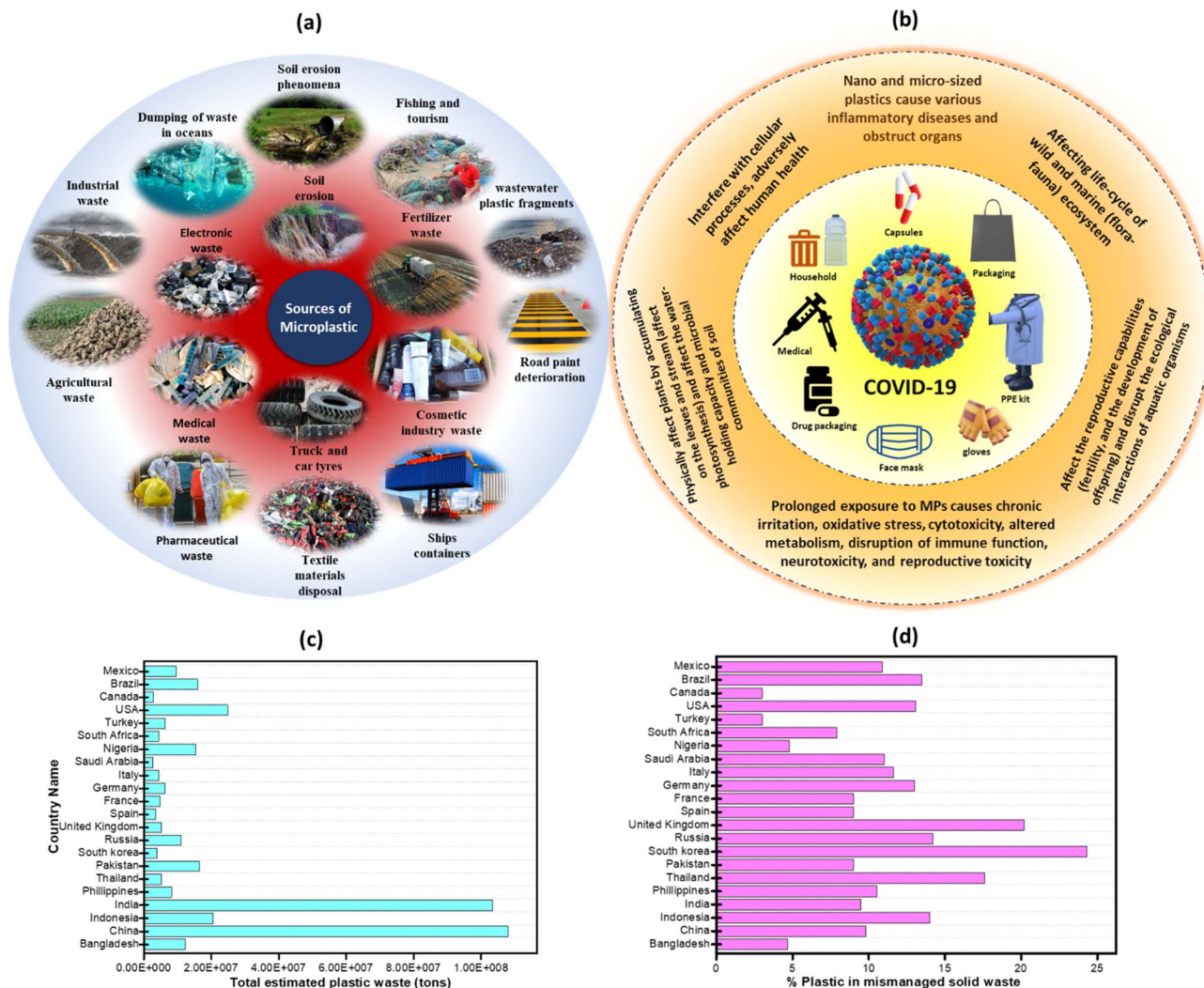


Fig. 2 (a) Pictorial representation of various sources of MPs, (b) effect of the Covid-19 pandemic on MNP generation, (c) graphical representation of total estimated plastic waste (in tons)^{38,39} and (d) percentage of plastics in mismanaged solid waste.⁴⁰

corresponds to ~ 307 tons in weight.⁴² Additionally, synthetic textiles including polyester and nylon are major contributors to MPs pollution. For instance, Hernandez *et al.* presented the amount of MPs fiber generation from synthetic textiles (polyester) in terms of their mass and size while carrying out the imitated home washings taking into account washing parameters such as the utilization of detergent, washing time, consecutive washing and temperature condition along with fiber structure orientation.⁴³ The study demonstrated that, of all the factors assessed the use of detergent exhibited the greatest effect on the overall amount of fibres released. However, detergent composition (liquid or powder) and overdose failed to impact MPs release significantly. Despite various release parameters while carrying out the insertion of surfactant (~ 0.025 and ~ 0.1 mg fibers per g textile washed out deprived of and with detergent, respectively), average MPs-based fabric orientation was found consistent regardless of wash conditions, with maximum of fibres found between ~ 100 μm and ~ 800 μm in length, regardless of the number of washing cycles.⁴³

MPs can be generated from packaging films during everyday activities such as opening packages, consuming food, drinking from plastic bottles, or disposing insensitively into sewerage or garbage heaps. For instance, Sobhani *et al.* demonstrated that plastic packages from consumables in everyday life may result in a significant mass of MPs generation (*i.e.*, within the range of ~ 10 to ~ 30 nanograms, which is equivalent to $\sim 14\,000$ to $\sim 75\,000$ isolated number of MPs).⁴⁴ The quantity of MPs generated (ranging from ~ 46 to ~ 250 MPs per cubic centimeter (size of $10\ \mu\text{m} \times 10\ \mu\text{m} \times 10\ \mu\text{m}$), and/or ~ 0.46 to ~ 2.5 MPs per cubic centimeter (size of $1\ \text{mm} \times 10\ \mu\text{m} \times 10\ \mu\text{m}$)) depends on various factors, including the stiffness, thickness, anisotropy, density of the plastic materials, and the size of the MPs.⁴⁴

3.2 MPs from agriculture, construction, and infrastructure demolition waste

MNP can also be deliberately introduced into the soil through the use of industrially produced chemical fertilizers such as diammonium phosphate and single superphosphate (which



predominantly possess nitrogen, phosphorus, calcium, magnesium, and potassium elements) and pesticides such as herbicides, fungicides, nematicides, rodenticides, and insecticides for agriculture.⁴⁵ These products often come with a polymer shell or coating, contributing to MNP contamination.⁴⁶ For instance, China, being one of the largest producers and consumers of plastic globally, utilizes over ~70 million tons of plastic products annually.⁴⁷ Among these, mulch films account for ~57.27% of agricultural film consumption, which has reached over 2.408 million tonnes.⁴⁸ Research conducted in China has revealed significant stages of MPs contamination, with soil samples in localities where plastic mulching has been employed for consecutive ~30 years showing MPs concentrations of ~40.35 mg kg⁻¹.⁴⁹ A recent study in Japan found that plastic-based microcapsules from coated fertilizers collected in rice fields resulted in ~144 mg of MPs per kg of soil.⁵⁰ Additionally, the utilization of sewage sludge from WWTPs has been identified as a major source of MPs in soil used for agriculture.⁵¹ For instance, average MPs concentrations of ~22.7 ± 12.1 × 10³ particles per kg dry sludge were detected in sewage sludge from 28 WWTPs in China,⁵² which indicates a relatively high abundance of MPs in the sludge samples. In Europe, sludge reuse releases ~63 000 to ~430 000 tonnes of MNP into farmland each year, out of which only Finland uses up to 72% of sludge for agricultural use.^{53,54} Furthermore, the wear and tear of agricultural equipment (such as agrochemicals loaded plastic films coated crop seeds) including tire abrasion (from tractors and mechanical harvests), contribute to the presence of MPs in soil samples from agro-fields, with the accumulation of tire wear debris being a significant source of soil MNP.⁵⁵ Afterward, Kole *et al.* estimated that ~67% of tire constituent parts unconstrained in the surroundings in the Netherlands end up in the soil, emphasizing the importance of tire wear as a substantial contributor to soil MNP. Moreover, MNP can enter into the soil and aquatic ecosystem through road paint deterioration, truck and car tires, ship containers, and by disposal of textile materials.⁵⁶ Further, massive numbers of MNP as fibers (synthetic and natural based), such as rayon, polyester, and carpets are released from the common textile garments industry during the domestic and industrial laundering processes.

3.3 MPs from aircraft, aviation, and other transport modes such as vehicular sources

The aviation sector has emerged as one of the newly realized and aggressive potential sources for the generation of MNP. The major envisaged constituents comprised of tire wear, cabin materials, and engine emissions. As documented, a single take-off and landing of an aircraft (not only commercial flights but also private jets and military aircraft) generates an average of up to ~2.5 kg of MNP getting into the atmosphere dispersed and the majority of it primarily consists of rubber particles from tire abrasion.⁵⁶ In the context of automotive sources, MPs most significantly arise from tire wear and road abrasion. Tires shed tiny plastic particles as they wear down due to friction with the road surface. As an example, Baensch-Baltruschat *et al.* used the specific emission factors per vehicle-kilometer approach to assess wear and tear on urban, rural, and highway roads. The

cumulative estimated wear and tear for these three types of roads amounted to ~17 300 tonnes annually.⁵⁷ Worek *et al.* tested samples of road dust and soil from high-traffic areas, finding an average of ~372 ± 50 MPs fragments/kg of dry weight in soil samples and ~515 ± 20 fragments per kilogram of dry matter in road dust.⁵⁸ Notably, the road dust samples included a larger concentration of MPs than the soil samples, possibly due to the direct action of braking, which causes a greater buildup of MPs at the emission source than in the surrounding soil. Furthermore, there is a clear variation in the size distribution of MPs fractions, as most instances in soil samples lie below ~50 µm, whereas road dust samples largely consist of fractions between ~50 and ~200 µm.⁵⁸ It must be noted that the extent of MPs-induced contamination in soil that can directly be attributed to tire dust or tire- or origin remains mostly unreported or scattered to the extent that it is hard to reproduce.

3.4 MPs from the pandemic, epidemic, healthcare and medical-safety products

The COVID-19 pandemic has led to a significant increase in single-use plastic pollution due to the extensive use of single-use-plastic-based biowaste and medical waste such as gloves, face masks, and PPE kits, hand sanitizer bottles, gloves, and water-logged masks. However, the persistence of plastics for prolonged periods, and improper disposal, including these single-use medical devices and PPE items, can contribute to their prevalence in the environment for many thousand years, thereby leading to detrimental effects on wildlife, including entanglement, ingestion, and habitat disruption.^{59,60} Fig. 2b provides an overview of the impact of plastic waste generation at the time of the pandemic. Reportedly, global demand for PPE has risen to ~129 billion face masks and 65 billion gloves every month, ensuring the protection of the public and frontline workers globally.⁶¹ The total quantity of plastic garbage generated increased as a result of individual choices about utilizing online delivery services in order to address safety concerns about dining out or shopping during the pandemic or any other extended period of illness. As an example, the waste produced due to online shopping during the pandemic varied across countries, with the highest amounts recorded in China (~402 000 tons), followed by Germany (~36 500 tons), the USA (~2700 tons), India (~520 tons), and Italy (~450 tons).⁶² Conversely, hospital waste generation, ranked from highest to lowest, was observed in India (~100 865 000 tons), the USA (~685 200 tons), Germany (~500 300 tons), Italy (~404 000 tons), and China (~146 000 tons).^{38,39} A study by Peng and Kathuria estimated that ~22–30 thousand tonnes of plastics related to the epidemic, including ~12 000 tonnes of MPs, were released into worldwide seas. This highlights the significant contribution of the COVID-19 pandemic to the accumulation of MPs in marine and inland waters.⁴⁰ Further, Saliu *et al.* revealed that in 2020, the worldwide manufacturing of masks was found to be range from ~2.4 to ~52 billion pieces, resulting in the generation of ~7200 to ~312 000 tons of waste containing MPs. Among these wastes, a significant portion, *i.e.*, ~1% to ~10% was released into the ocean, which accounts for ~72 to ~31 200 tonnes of



MNP waste.⁶³ Further, Mintenig *et al.* demonstrated that facial cleanser beads exhibit a higher mass of MPs compared to other personal care products (PCPs). For instance, they found that facial cleanser beads contained ~ 0.05 grams of MNP per gram or ~ 2450 particles per gram, whereas shower gels exhibited only ~ 0.02 grams per gram or ~ 2.15 particles per gram.⁶⁴ Plastic particles in shower gel exhibited an average diameter of ~ 419 μm , whereas those in face cleanser beads showed an average diameter of ~ 197 μm . A previous study conducted in China on characterizing MPs in facial cleanser beads indicated that WWTP was responsible for $>80\%$ of MPs entering the aquatic environment. The residual portion ($\sim 20\%$) may be ascribed to direct emissions, of which $\sim 18.2\%$ come from cities and $\sim 0.5\%$ from rural regions.^{42,65} The detailed outline of the total estimated plastic waste in various countries is shown in Fig. 2c and d.

4. Gross overview of physical consequences of MPs on the “soil–water–atmosphere” nexus

4.1 Physical consequences of MPs in water

The occurrence of MNP in earthly and marine environments raises concerns due to their potential impacts on organisms, ecosystems, and human health assisted by various mechanistic routes such as ingestion, contamination, infection, inflammation, assimilation, blockage of various tracts in active biological systems such as plants and animals (both aquatic and terrestrial).⁶⁶ For example, MPs can be consumed by a variety of organisms, ranging from small invertebrates to larger marine animals, and have the ability to accumulate and magnify in the food chain. Such undesirable effects may cause nutritional imbalances, and changes in feeding behavior, leading to related growth facts, reproduction malfunction, and eventually threatening the overall growth of such species.⁶⁷ Further, nanoparticles (NPs) can be ingested by a diverse range of aquatic organisms, including invertebrates, fish, and plankton through various pathways, such as filtration feeding or direct absorption across cellular membranes. Once ingested, NPs can accumulate in different body parts and tissues, including the gills, liver, intestines, and blood.⁶⁸ While keeping this fact in mind, Goh *et al.* explored detailed profiling of nanoparticles for MPs remediation from aquatic environments which may have long-persisting issues like depth-specific susceptibility of marine creatures/microorganisms.²⁷ Additionally, it has been observed that copepods, when exposed to high concentrations of MPs, exhibited reduced algal feeding, as highlighted in a study conducted by Piyal Bhattacharya.⁶⁹ Furthermore, research by Moos *et al.* demonstrated that mussels ingesting MPs experienced a pronounced inflammatory response.⁷⁰ Further, the regions with elevated MPs concentrations such as certain coastal areas and bodies of water often experience diminished water clarity due to the scattering of light by these particles, resulting in a cloud/hazy or murky appearance.⁷¹ Consequently, this can disrupt the process of photosynthesis in underwater plants due to a significant drop in photon flux density with increased

concentration of MPs and alter the visual hunting tactics of marine underwater species employed by certain species.

The higher exposure of humans to freshwater environments increases the significance of MNP as a threat in these ecosystems compared to marine environments. For example, in the urban waters of Changsha, Hunan Province, China, surface sediments were found to contain MNP concentrations ranging from ~ 270 to 866 items per kg.⁷² In Qinghai Lake, the largest inland lake in China, MNP concentrations in the lake water ranged from $\sim 0.05 \times 10^5$ to 7.58×10^5 particles per km^2 , while sediment samples contained ~ 50 to ~ 1292 particles per m^2 .⁷³ Similarly, in the Wei River in northwestern China, MPs concentrations in surface waters ranged from ~ 4 to ~ 11 items per L, while sediment samples had concentrations of ~ 360 to ~ 1320 items per kg.⁷⁴ In humans, the MPs incorporation takes place *via* gastrointestinal tracts followed by their permeation into both the lymph and circulatory systems of unborn fetuses through the placentas, as well as being absorbed in the lungs when inhaled, which can trigger various biological responses from the immune system and have detrimental effects on the health of body cells. Reportedly, MPs smaller than <150 nm may translocate across the gut epithelium.⁷⁵

4.2 Consequences of MPs across soil–water interface

Contamination of soil by MNP can occur through diverse pathways, such as the incorporation of MPs in sewage sludge or compost direct interference, the fragmentation of larger plastic objects into smaller particles (fracture and collapse of structural integrity), or direct deposition from the atmosphere (settlement of contaminated MNP aerosols).⁷⁶ They have several effects on soil environments as they can enhance water evaporation rates by influencing soil porosity, potentially leading to increased water loss, increasing amount of dissolved nitrogen, phosphorus, and other organic carbonous compounds in soil through enzymatic stimulation activity, altering the structure and diversity of the bacterial community, and eventually leading to significant extent of soil–moisture interaction imbalance over time.⁷⁷ Additionally, MNP in the soil can adsorb various pollutants like antibiotics, heavy metals, and persistent organic substances (also known as “forever chemicals”) and release internal and non-bonded additives through dynamic leaching activity, which in turn may cause pollution and alter the movement of pollutants in the soil, thereby endangering the soil quality and its associated microorganisms.⁷⁸ Further, the presence of MNP in the soil environment can alter the overall bulk density (making it lighter) and pH of the soil. For instance, Tan *et al.* noticed that the pH decreased with increasing levels of MPs in a pot experiment, which may account for the increase in soil aggregate size, soil porosity, and aeration when the MPs were put into the soil.⁷⁹

Similar to the marine environment, MNP pollution in soil has reportedly increased the uptake of MNP by various soil organisms such as earthworms, and rhizobium. For example, PS particles can induce a malfunctioning of intestinal microflora, impair intestinal boundary function, and thereby causing metabolic disorders in mice as a secondary consequence.⁸⁰ Earthworms, which are important contributors to soil health,



also experience inhibited growth when exposed to PE-based MPs, and higher concentrations of MPs that can even lead to increased mortality rates in these organisms. Similarly, nematodes, another important group of soil organisms, efficiently ingest MNP, which can be detected in their offspring as well. MNP is extremely dangerous to plant species, causing growth suppression, intestinal damage, immune system disturbance, reproductive consequences, and death. MNP can also inhibit the ability of roots to absorb water and nutrients, affecting plant development and germination.⁸¹ Similarly, MPs in soil may induce a new microbial habitat with its consequences on bulk soil microbiomes.⁸² MPs tend to accumulate in pores indeed capsules and also impact seed water adsorption and may thus delay germination and root growth. Thus, MPs in soil may disturb the soil nutrient cycle alter the microbial structure, and may even ultimately affect crop growth.⁸³ Through the complex food chain and species interactions, MNP can accumulate and transfer through trophic levels, ultimately posing a threat to human health *via* bioaccumulation.⁸⁴

4.3 Consequences of MPs across atmosphere–water interface

MNP possess the potential to remain airborne due to their corresponding tiny size and lower density values, which can function as carriers for different other contaminants present in the air, and their transportation with more time is helped by wind.⁸⁵ While rivers and coastal areas are recognized as major sources of marine MNP, present research indicates that atmospheric deposition may also contribute to the presence of MNP in the ocean and their concentrations in the atmosphere can vary across different regions depending on climate conditions. For example, MNP concentrations ranged from ~ 4.18 items per m^3 in Shanghai, China, to ~ 14.1 items per m^3 in Beijing, and up to ~ 275 items per m^3 in Germany.^{86–88} Indoor air tests from 12 countries globally revealed a range of PET MPs concentrations from ~ 38 to $\sim 120\,000$ $\mu\text{g g}^{-1}$, with the average value in Colombia.⁸⁹ MPs can indirectly influence climate by modifying the reflectivity (albedo) of the surfaces they settle on, thereby not only impacting local temperature and climate patterns but also acting as radiative surfaces. Further, Zhang *et al.* addressed the pressing issue of anthropogenic pollutants, specifically focusing on the MPs in cryosphere regions where they conducted experiments in snow, river, glaciers, lake water, and sea ice with a primary focus on the physical and chemical characteristics of MPs and their transportation dynamics in these fragile ecosystems.⁹⁰ However, there has been relatively minimal attention directed toward the significant consequences of cryosphere changes so far. Revell *et al.* observed that atmospheric MPs had an effective radiative forcing of $\sim 0.044 \pm 0.399$ $\mu\text{W m}^{-2}$ in the ecosystem, presuming a uniform surface concentration of one MPs particle per cubic meter and a vertical distribution up to ~ 10 km height.⁹¹ Despite the significant uncertainty involved with these findings, the research provides an innovative perspective on the potential impact of MPs on the climate and environment, especially in the context of global warming and increasing plastic production and consumption. It further highlighted that deposited MPs on glacier surfaces are

transported over long distances through wet and dry deposition, which can have significant effects, particularly when they possess dark colors like black, grey, or brown.⁹² These MPs can absorb solar radiation, leading to the heating of surface snow, a phenomenon akin to the impact of carbon black in snow. This suggests that MPs in snow may lower glacier surface albedo and, consequently disrupt the energy balance in cryosphere regions. Furthermore, the accelerated melting of glaciers could introduce MPs into downstream rivers and lakes, creating threats to ecosystems.⁶⁴ Unfortunately, existing studies have failed to quantify the impacts of airborne MPs deposition on surface snow.⁹³

Human exposure to MNP through inhalation poses a significant risk to human health. Inhalation is recognized as one of the primary routes of exposure to MNP, and it is estimated that the intake of MNP through the air exceeds other exposure routes.⁹⁴ When inhaled, MNP can deposit in the lungs, and natural mechanisms such as cilia movement, sneezing, and phagocytosis aid in the removal of inhaled particles. Consequently, MNP has the potential to accumulate in the body over time, leading to chronic inflammation, which may increase the risk of cancer.⁹⁵ Fig. 3a highlights the numerous phases of the plastic life cycle, including manufacture, use, and final disintegration, with an emphasis on the formation of MPs.

Understanding the various mechanisms contributing to MPs generation is crucial for comprehending their pervasive presence in various environments. Thus, the paper further gives insights into various factors influencing MNP generation, namely the influence of photo-oxidation, UV-irradiation, and different solvent systems on polymer structures.⁹⁶ The entire lifespan of plastic items starting from critical phases of production to use to final environmental degradation exhaustion of utility shelf life is shown in Fig. 3, which highlights the critical stages at which structural changes take place and, as a result of various environmental stresses (including photooxidation, exposure to ultraviolet light, and contact with solvents and aquatic environments), macroscopic plastic entities break down fragment with their MPs counterparts. These stressors include both human activity and natural factors, which create an environment that is favourable for the disintegration of plastics into components of MPs. When exposed to sunlight, high-energy UV photons are absorbed by polymer molecules and it generates free radicals (which are reactive entities with unpaired electrons), which further combine with ambient oxygen, thereby triggering oxidative chain reaction.⁹⁷ Oxygen radicals attack the polymer backbone, causing molecular bond breaking and the creation of smaller fragments like MNP. In marine environments, the combination of UV exposure and ocean dynamics, such as temperature variations and wave activity, accelerates the breakdown of plastics. Temperature variations cause polymers to expand and compress, increasing their susceptibility to mechanical stress.⁹⁸ Furthermore, wave motion exposes plastic waste to ongoing physical abrasion from other materials and surfaces, hastening the fragmentation process.⁹⁹ Solvents, which are commonly employed in industrial processes and products, can weaken the polymer matrix, allowing larger plastic objects to break down into micro-sized particles. For example, Ghatge *et al.* evaluated the influence of organic solvents on the breakdown of





Fig. 3 (a) Schematic representation showing a comprehensive overview of the stages/life cycle of plastic products and its disorientation, followed by disintegration into different stages of MPs through various environmental, mechanical, and human factors along with their adverse impact on human life. (b) Effect of photo-oxidation on the structure of the polymer for MNP generation, (c) effect of UV-irradiation and marine environment on the structure of the polymer for MNP generation, and (d) effect of solvent or water on the structure of the polymer for MNP generation.

PE, a widely used plastic polymer.¹⁰⁰ The results showed that exposure to solvents enhanced PE fragmentation, resulting in micro- and NPs. Water has an important role in plastic transport and transformation, altering structural integrity and contributing to MPs dispersion across ecosystems. Arhant *et al.* investigated the hydrolytic breakdown of PET and discovered that water contact induced considerable chemical changes and structural damage to the polymer.¹⁰¹

5. Conventional methods for the removal of MNP from contaminated water

As the existence of MNP in the ecosystem causes great concern in the current circumstances, numerous traditional strategies for

eliminating these new plastic pollutants have previously been thoroughly researched. As previously documented, the bulk of MPs are $>20 \mu\text{m}$ and easily removed using filtering and Membrane Bioreactor (MBR).¹⁰² These removal methods are classified based on the nature of the removal process, such as (i) capture and surface attachment methods: coagulation, flocculation, and sedimentation (CFS), electrocoagulation, adsorption, magnetization, and microorganism aggregation, (ii) filtration and separation methods, which include filtration and separation *via* membrane and constructed wetlands, and (iii) degradation methods, which include photocatalytic degradation, microorganism degradation, and thermal degradation.^{103,104}

Indeed, filtration and separation methods are commonly employed for the removal of MNP from the aqueous environment due to their practicality and effectiveness.¹⁰⁵ However, it is important to note that the efficiency of these methods can vary



depending on the size and nature of the particles. When it comes to small-sized and fibrous MNP especially those that are in the nanometer range, the efficiency of conventional filtration techniques may be relatively low. This may be attributed to the fact that smaller particles have a higher tendency to pass through filter media or pores, resulting in reduced removal efficiency.¹⁰⁶ The small size and mobility of these particles make them more challenging to capture and separate effectively. To address this limitation, alternative techniques, and advanced filtration systems are being explored. For example, hydrogels, MOFs, and membranes have gained widespread acceptance in usage for the removal of MPs from various components of the environment.^{77,107} Despite their numerous advantages, these materials come with certain limitations, such as selectivity issues, fouling problems, challenges in scaling up, and complex treatment processes, which restrict their potential application for MPs removal.¹⁰² To impress upon the scenario, various systems (hydrogel, membrane, and MOF), their respective synthesis methods, the solvents utilized, and their flocculation activity concerning MPs are listed in Table 2.

6. Emerging techniques for the removal of MNP: new-directions and alternative material frameworks as adsorbents

The most emerging technologies cover the functional usage of biochar, electrospun fibrous constructs, and aerogels as they demonstrate significant potential in tackling the challenges posed by MNP pollution. These newer framework approaches offer distinct characteristics (specific to the chemical structure of MPs, specific to functional groups, specific to absorbability, specific to surface capability) and novel strategies (trapping, sequestration, coagulation, sedimentation, filtration, degradation) for effective MNP removal from water and wastewater.^{29,124} Biochar, renowned for its extensive surface area ($\sim 250\text{--}500\text{ m}^2\text{ g}^{-1}$) and exceptional adsorption capacity, exhibits promising ability in sequestering MNP.³⁰ Electrospun fibrous constructs, on the other hand, feature high porosity (controlled pore size, allows size-specific MNP trapping) and customizable surface functionalities, enabling efficient capture of MNP.³¹ Aerogels, characterized by their remarkably low density ($\sim 0.01\text{--}0.4\text{ g cm}^{-3}$) and large surface area ($\sim 30\text{--}600\text{ m}^2\text{ g}^{-1}$), present opportunities for advanced MNP adsorption and removal.¹²⁵ Therefore, by harnessing the potential of biochar, electrospun fibrous constructs, and aerogels, the challenges associated with MNP pollution including mitigating their detrimental impacts on human health and the environment may be effectively addressed.

6.1 Biochar-based adsorbents

Biochar, a pyrogenic carbonaceous substance with a high surface-to-volume ratio, and the ability to functionalize its surface, porosity, and chemical composition, has sparked widespread attention as a bio-based sustainable adsorbent for the removal of MNP. The conventionally adopted procedure to

make biochar is pyrolysis, which involves heating the feedstock (e.g., biomass waste, plant waste, animal waste) under minimally aerated (low oxygen) conditions.¹²⁶ The utilization of precursor materials, temperature, pressure conditions, and pyrolysis technology, have a significant impact on the various properties of biochar, including adsorption capabilities and the elemental composition of the adsorbent of the resulting biochar.¹²⁷ In this context, Chen *et al.* demonstrated that increasing the pyrolytic temperatures showed a direct impact on the degree of carbonization, which in turn led to improvements in the structural, compositional, and morphological aspects of the resulting biochar, accompanied by improved sorption capabilities.¹²⁸ Furthermore, the abundance of feedstocks for the production of biochar, such as biomass of agricultural waste and solid waste, makes it a renewable resource with immense potential to remove MNP.¹²⁹ For instance, Singh *et al.* developed eco-friendly iron-nanoparticles surface-modified biochar adsorbents with improved magnetic and surface properties to remove NPs under various pH conditions. The results showed that iron-modified biochar achieved $\sim 100\%$ removal efficiency compared to $\sim 75\%$ removal efficiency with untreated biochar.¹³⁰ Further, Siipola *et al.* utilized pine and fir bark for the development of activated biochar by slow pyrolysis at $\sim 475\text{ }^\circ\text{C}$ and steam activation at $800\text{ }^\circ\text{C}$. Such modification tends to alter the shape of the biochar and improve its adsorption capabilities and was found to retain $\sim 100\%$ of cylindrical pieces of PE, including spherical microbeads ($\sim 10\text{ }\mu\text{m}$), cylindrical pieces of PE ($\sim 2\text{--}3\text{ mm}$), and fleece woven fibers.¹³¹

Employing magnetic adsorbents made from zinc/magnesium-modified biochar (Zn/Mg-MBC) has been explored for the removal/separation of Wang *et al.*¹³² The study demonstrated that even after undergoing five consecutive cycles of adsorption, the elimination efficiency using Mg-MBC, Zn-MBC, and MBC remained consistently high at $\sim 94.6\%$, $\sim 95.8\%$, and $\sim 95.0\%$, respectively. The primary mechanisms responsible for pollutant retention are chemical binding and electrostatic interaction between biochar and MP. Table 3 summarizes various reported methods of adsorptive removal of MPs using biochar. For example, Shi *et al.* developed biochar by applying the pyrolysis method of rape straw for MPs separation applications.¹³⁶ Further, the magnetic biochar (MB) was prepared by using ferric chloride salt as an agent through the ultrasonication method. The performance of magnetic biochar was further improved by modifying biochar with CTAB surfactant (CMB) (Fig. 4a). The developed magnetic biochar system was utilized to isolate and segregate NPs from the environment surrounding in the form of PS (PS) and carboxylate-modified PS (CPS) nanoparticles (Fig. 4b). The maximum removal efficacy of $\sim 90\%$ was attained by using the modified magnetic biochar system with quantitative values of ~ 163.4 and $\sim 159.6\text{ mg g}^{-1}$ for PS and CPS respectively through an aggregation-based mechanistic approach. The prepared system was found recyclable in nature for up to five cycles suggesting their excellent selectivity and sensitivity (Fig. 4c). The system was highly effective at neutral pH and highly selective in the presence of a wide variety of metal ions (Fig. 4d and e). Further, Shi *et al.* utilized pinewood sawdust as a precursor for the fabrication of



Table 2 An overview of the compilation of different systems (hydrogel, membrane, and MOF) along with their respective synthesis method, solvent media, and flocculation activity towards MPs

Hydrogel				
System	Synthesis method	Solvent used	Flocculation activity	Ref.
Bacterial cellulose hydrogel	Biofilm formation method	Distilled water	~88.6%: ~1.77 g L ⁻¹ removal of MPs	108
Polyethyleneimine and polydopamine copolymer, graphene oxide nanosheet, and poly(<i>N</i> -isopropyl acrylamide) hydrogel	Self-polymerization reaction	Deionized water	~97.09% efficacy for ferric ion adsorbed MPs compound polymer	109
Ferrofluid/aminated alginate/psyllium hydrogel	Chemical method	Acetone	~94% removal of PMMA (coagulation adsorption)	110
Sericin-CS-based hydrogel	Crosslinking and chemical method	Organic solvents	Removal of PET by degradation method	111
Membrane system				
System	Synthesis method	Removal efficiency	Mechanism	Ref.
Membrane bioreactors	Membrane unit and ultrafiltration process	13 different polymers found in MPs (mainly polyester and PE) ~99.9% removal efficiency	Entrapment of MPs fiber due to their size and morphology	112
Pilot scale membrane bioreactors	Membrane formation method	Removal of ~99.4% MPs in activated sludge	Opting for different processing steps	113
Membrane microfiltration (polyvinylidene fluoride used as a substrate)	Membrane formation method	—	~100% removal efficiency	114
Polycarbonate, cellulose acetate, and polytetrafluoroethylene membrane	Filtration method	Distilled water	Removal of polyamide and PS MPs (~94%)	21
Silica-based ceramic hollow fiber microporous membrane by using Guinea cornhusk ash	Membrane formation method and phase inversion	Absolute ethanol and ammonia	Polyacrylonitrile (PAN), PVC, polyvinylpyrrolidone (PVP), and polymethyl methacrylate (PMMA) MPs (~99%)	115
Holey Ti ₃ C ₂ nanosheet-based membrane	Membrane method	Deionized water and hydrochloric acid	~99.3% removal (water flux 196.7 L h ⁻¹ m ⁻² kPa ⁻¹)	116
Wastewater treatment plant samples were utilized	Microfiltration membrane technique	—	~98% MPs recovery in the form of fibers, pellets	117
MOF				
Zr-based MOF (UiO-66-OH@MF-3)	Chemical method	Acetone, ethanol, and dimethylformamide (DMF)	~95.5% removal efficiency	77
ZIF-67 MOF	Chemical method	Deionized water, concentrated HCl, and NaOH	~92.1% adsorption ratio (PS MPs)	118
Zinc-based MOF composite (ZIF-8@aerogel by <i>in situ</i> growing ZIF-8 on wood aerogel fibers)	Chemical method	Deionized water and nitric acid solution	Micro/nano plastic (poly(1,1-difluoroethylene and PS)) = ~91.4 and ~85.8% removal efficiency	28
MOF (MIL-100) and Fe nanoparticles were incorporated into polysulfone matrix (PSF)-PSF/MIL-100 (Fe) nanoparticle composite membrane	Chemical method and membrane formation method	Deionized water	~99% rejection efficiency for methylene blue dye adsorbed in MPs	119
<i>In situ</i> formation of Ag ₂ O/Fe-MOF	Chemical precipitation method	Deionized water	Conversion of MPs into useful chemical (acetic acid) and hydrogen production	120
	Hydrothermal method			121



Table 2 (Contd.)

Membrane system				
System	Synthesis method	Removal efficiency	Mechanism	Ref.
Cr-based MOF (Cr-MOF/MIL-101)		HF acid + water (pseudo-second-order kinetic)	~96% removal efficiency of PS nano-plastic (~800 mg g ⁻¹) follows	
Nano-pillared structure composed of 2D MOF separated by carbon-encapsulated FeO nanoparticle	Hydrothermal method	Distilled water	~100% removal of MPs	122
Zirconium MOF	Hydrothermal method	DMF used as a solvent	~95.5% removal efficiency of MPs	77
Composite membrane made up of PAN with the addition of reduced graphene oxide	Chemical and membrane formation method	DMF, H ₂ SO ₄ and H ₂ O ₂	>82% removal efficiency of MPs	123

Table 3 Representation of biochar system in MPs removal along with their solvent, mechanism, and removal efficiency

Biochar				
System	Experimental medium of MPs dispersion	Mechanism	Removal efficiency	Ref.
Straw magnetic biochar with porosity and graphitization	Ultrapure water	Magnetic force	Isolation of ~31.29 mg g ⁻¹ MPs	126
Corn straw and hardwood biochar (pyrolysis method)	—	Stuck, trapped, and entangled mechanism	~10 μM MPs spheres (~90% adsorptive removal of MPs)	30
Waste rice husk-derived magnetic biochar (impregnation pyrolysis)	Deionized water	Electrostatic interaction and complexation	~99.96% removal of MPs	133
Biochar of bark of Scots pine and spruce (slow pyrolysis)	Deionized water	Adsorption phenomena	~90% removal of MPs	131
Woodchip-derived biochar (pyrolysis)	Ultrapure water	Adsorption mechanism	~90% removal of MPs from sand columns	134
Magnetic corncob biochar (hydrothermal)	Ultrapure water	Hydrophobic and electrostatic interaction	~90% removal of polyamide MPs	127
Mg/Zn modified magnetic biochar (impregnation method)	Deionized water and ethanol	Adsorption and thermal degradation	~99.46% removal of PS MPs	132
(Cow manure and sawdust) livestock manure biochar (biodegradation method)	Deionized water	Biological degradation through different species	~90% removal of polyhydroxyalkanoates MPs	135

biochar and modified the biochar to develop a magnetic biochar system.¹³⁷ The magnetically active biochar system was utilized to remove different types of PS such as unmodified, amine-modified, aged under UV radiation, and carboxylate-modified systems (as dummy MPs representation) with a maximum removal efficiency of ~95%. The Langmuir adsorption isotherm for the system elicited an adsorption capacity of ~107.71 mg g⁻¹ and it followed the H-bonding and electrostatic interaction-based mechanisms.

6.2 Electrospun fiber-based adsorbents

Electrospun mats/membranes have attracted substantial consideration in the field of MNP separation and segregation due to their extraordinary properties such as excellent surface

area (~8.93 m² g⁻¹) depending upon the porosity, wide range of diameters (ranging from ~40 to ~2000 nm) permeability, easy surface functionalization process, controllable surface charge along with exceptional pliability/flexibility as a macrostructural framework.¹³⁸ Additionally, the highly interrelated and porous network structure of such electrospun mat-based systems results in a very low resistance toward water transport (permeability/diffusibility) which ultimately enhances the membrane flux to a favorable desired value (Table 4). Further, electrospun nanofibrous membranes (when suitably designed as mats) found domination as compared to conventionally reported membranes in terms of hydrophilicity, mechanical stability, selectivity, flexibility, resistance to water flow, and surface area.^{31,147} An interesting aspect of this method is the



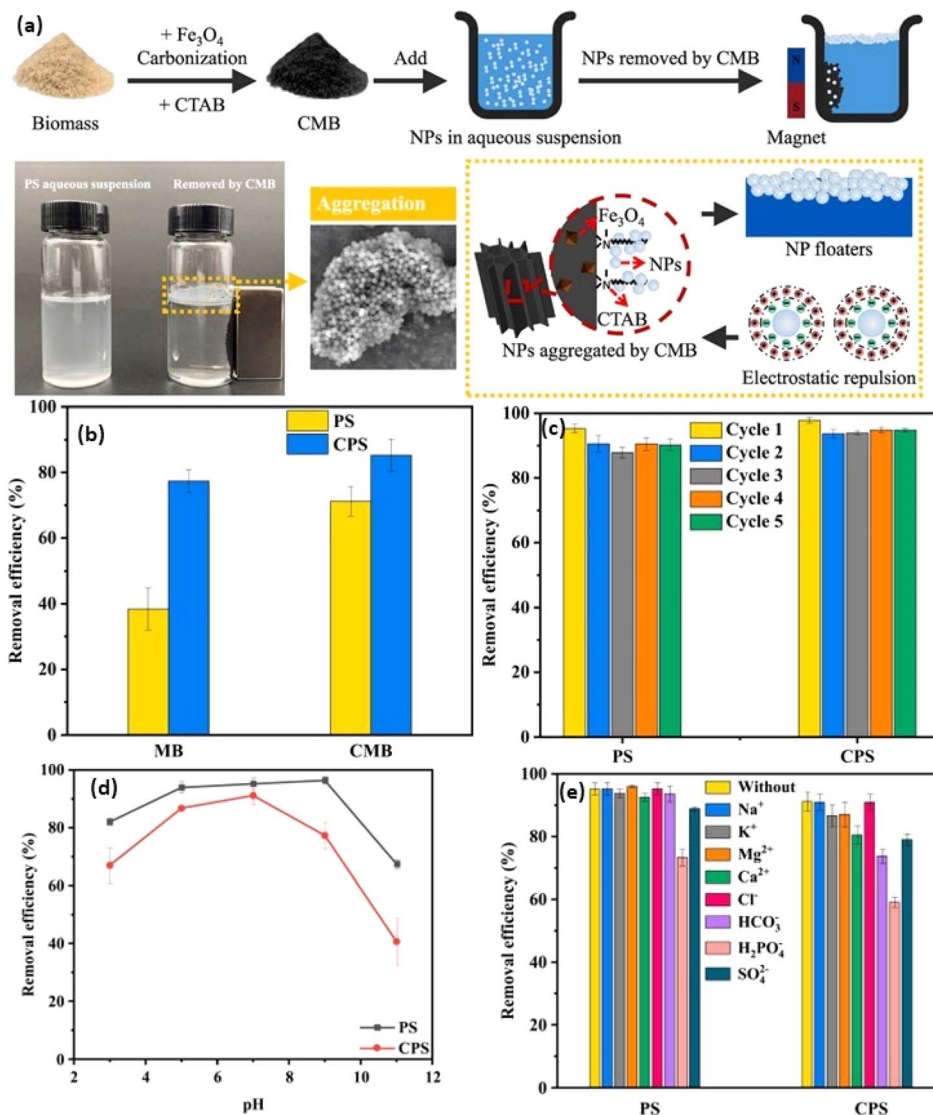


Fig. 4 (a) Synthesis of magnetic biochar (MB) and CTAB modified biochar (CMB), (b) removal efficiency of PS (PS) and carboxylate modified PS (CPS), (c) reusability, (d) pH stability, and (e) highly selective in the presence of a wide variety of metal ions.¹³⁶

production of electrospun nanofiber mats of various sizes and shapes with porous properties, which allows them to eradicate smaller contaminant particles isolation *via* trapping, such as PM 2.5 (<2.5 μm), PM10 (<10 μm), and PM1.0 (<1 μm), over the last few decades owing to their size exclusion mechanism.¹⁴⁸ The performance and adsorption sieving characteristics of mats could be improved through functionalization/modification mode (chemical grafting, layer by layer, self-assembly, co-axial spinning, side-by-side deposition, and nanomaterial implementation *via* electrospaying) and incorporation of polymer matrix and chemical agents.¹⁴⁹ Few reports have dealt with the electrospun mats, concerning their potential for MNP removal. For instance, Wang *et al.* fabricated low-pressure-driven and highly porous electrospun PAN membrane by adopting a layer-by-layer assembly protocol (where three layers *i.e.* polyethyleneimine (PEI), poly(acrylic acid) (PAA) and PAN were deposited over one another¹³⁹ (as shown in Fig. 5a). The

electrospun mat was hydrolyzed to add carboxylic groups that were negatively charged, so as to initiate the easy formation of LbL assembly, as seen in Fig. 5b. The prepared system reflected outstanding mechanical strength ($\sim 58.5 \text{ MPa} \pm 20.1$) and hydrophilicity (water contact angle = $38.3 \pm 2.3^\circ$) at a neutral pH (pH = 7) condition and subsequently acted as an efficient model for the removal of $\sim 89.9\%$ PS (PS) based micro-/nanosphere with its size in the range of $\sim 50\text{--}500 \text{ nm}$. As seen in Fig. 5c, all of the modified membranes exhibited retention rates >99%. The presence of electrostatic attraction mainly contributes to the isolation/removal of MNP in an effective manner from the micro-plastic dispersed polluted water. Further, Risch and Adlhart developed CS-glutaraldehyde/PE electrospun nanofiber by applying a high-throughput nozzle free-surface electrospinning methodology (Fig. 5d). The CS-based nano-sponge is highly biocompatible in nature and possesses outstanding bulk density and porosity. The nano-



Table 4 Representation of electrospun material systems in MPs removal along with different parameters (solvent, method, and pH)

Electrospun system	Solvent	Method	pH	Removal efficiency	Ref.
Highly porous electrospun PAN membrane	Ultrapure water and organic solvent	Layer-by-layer assembly protocol	Physiological pH = 7.2	PS (PS) microspheres (89.9% removal through electrostatic attraction)	139
Graphene oxide incorporated PU-based electrospun composite membrane	Tetrahydrofuran (THF), dimethylformamide (DMF)	Chemical and electrospun method	Neutral pH = 7	Removal of micro and nano plastic of butadiene styrene (BS), PS, and poly(methyl methacrylate) (PMMA) (cake formation and blocking mechanism)	124
CS–polyethylene oxide (PEO) nanofiber electrospun mats	Methanol, ethanol, and acetic acid	Cross-linking method	pH = 8	91% removal of PET MP (hydrostatic filtration)	140
		Bottom-up electrospinning method	—		141
Tubular nanofiber electrospun membrane of PAN polymer	DMF	Hot pressing method	—	Removal of 0.2 μM MPs through cake formation mechanism	142
PAN electrospun nanofiber membrane	DMF	Hot pressing method	—	Removal of 0.2 μM MPs through cake formation mechanism	142
Polyvinyl alcohol (PVA) electrospun membrane followed by cross-linking with glutaraldehyde	HCl, acetone, and Milli-Q water	Conventional non-woven micro fibrous approach	Wide pH range	98% removal of polycarboxylate microsphere particles through particle retention scheme	143
Electrospun polyimide (PI) nanofiber membranes	Deionized water	Dead-end filtration system	—	0.22 μM microfiltration membrane	144
Polyvinylidene fluoride (PVDF) ultrafiltration membranes	Millipore, Milli-Q water, and other systems	Top-down approach by considering the electrospinning method	—	Removal of PET MPs	145
Silica-based ceramic-supported thin membrane and Guinea corn husk were utilized as precursor	<i>N</i> -Methyl 2-pyrilidone	Phase inversion, interfacial polymerization, and sintering approach	Neutral pH	Removal of PVC, PMMA, PAN, and PVP	115
Ultrafiltration membranes of PET nanofibers and organic contaminants	Ethanol, trifluoroacetic acid, hydrochloric acid and acetonitrile	Electrospinning method for nanofiber preparation and cryo-sectioning through freezing agent	pH = 7	Removes 75% of nano/microfibers	146

sponge was found to be highly selective towards the adsorption of PET-based MPs from the micro-plastic-induced pollution in water and wastewater mediums.¹⁴⁰ The turbidity test of MPs was performed on nanofiber which decreases with the time of measurement and concentration of MPs (Fig. 5e). The non-linear fit of data for the adsorbed amount indicated Lagergren's pseudo-first-order model (Fig. 5f). A maximum of $\sim 99.49\%$ removal was observed for MPs which suggests their application in tackling environmental challenges (Fig. 5g).

A highly economical silica-based ceramic-supported thin hollow fibrous membrane *via* phase inversion, interfacial polymerization, and sintering approaches has been developed by Yogarathinam *et al.* (as shown in Fig. 6a). The prepared silica membrane was utilized for the eradication of various types of MPs from the aquatic environment such as PVC, polymethyl methacrylate (PMMA), PAN and PVP to a greater extent approximated to $>90\%$.¹¹⁵ The order for the removal of MPs was removed as PVP ($\sim 97.2\%$) $>$ ($\sim 93.7\%$) $>$ PMMA ($\sim 92.1\%$) $>$ PVC ($\sim 88.8\%$). In a recent study, Juraji *et al.* developed a polyurethane (PU)-based electrospun composite membrane by the incorporation of graphene oxide montmorillonite hybrid filters (Fig. 6d). The developed membrane exhibited super hydrophilicity (due to no/zero water contact angle and presence of

various hydrophilic groups (carboxylic and hydroxyl)), maximum extent of pressure-driven as well as gravity-driven water flux, and outstanding capacity in the removal of PS (PS), acrylonitrile butadiene styrene (ABS), and PMMA type MNP.¹²⁴ The mechanistic behaviour is characterized by individual and combined impacts of the size exclusion effect and adsorption phenomena (Fig. 6e). The maximum adsorption capacity was $\sim 417 \text{ mg g}^{-1}$, following the pseudo-second-order kinetic model and Langmuir adsorption isotherm. Removal effectiveness was $\sim 95\%$ for various MPs (PMMA, ABS, and PS) (Fig. 6f). The present electrospun composite was found satisfactory for up to ten cycles for the removal of MPs (Fig. 6g). The time-dependent gravity flux of different MPs was found to be reduced as compared to the water system due to the fouling of MPs (Fig. 6h). The fabricated PU-composite membrane acts as an efficient approach for the removal of MNP from polluted water.

6.3 Aerogels as adsorbents

Aerogel-based adsorbents have gained significant attention due to their remarkably high specific surface area ($30 \text{ m}^2 \text{ g}^{-1}$ to as high as $600 \text{ m}^2 \text{ g}^{-1}$) and adjustable surface chemistry, which is approximate for the removal and segregation of MNP.¹⁰⁶ Aerogels





Fig. 5 (a) Surface modification of electrospun mats using a Layer-by-Layer (LbL) approach. (b) Membrane surface modification process (c) retention rates of the prepared membranes for PS microspheres and two filtration cycles of MC and M+ to assess the removal efficiency of PS-500 MPs.¹³⁹ (d) Schematic representation showing the synthesis of CS-based electrospun nanofiber, (e) turbidity reduction test of MPs, (f) adsorbed amount of MPs, and (g) highest concentration with surface area of MPs on nanofibers.¹⁴⁰

are essentially highly porous structures with a high percentage of air, that can be derived from different types of precursors, including organic, inorganic, or hybrid molecules.¹⁵⁰ Aerogels are available in various forms such as powders, monoliths, and films, and exhibit different types of porosity, including mesoporous, microporous, or mixed porous structures. The porous structure

of aerogels enables the physical entrapment of MPs particles, while the surface chemistry can be suitably tailored to enhance the adsorption affinity for a specific plastic.^{151,152} Additionally, the lightweight nature of aerogels facilitates their deployment and handling in water treatment systems. Furthermore, the regeneration capabilities of aerogels make them appealing for the





Fig. 6 (a) Fouling analysis MPs fouling in (Guinea cornhusk ash) GCHA and TFC-GCHA (thin-film composite Guinea cornhusk ash) membranes with their FTIR spectra of before and after MPs analysis, (b) schematic of the interaction of MPs-HA, (c) brine solution filtration analysis, (d) schematic representation showing the synthesis of a PU-based electrospun composite, (e) mechanistic representation showing the adsorption of ABS, PMMA, and PS nano plastics, (f) removal efficacy of MPs, (g) reusability of the electrospun composite after ten cycles, and (h) time-dependent flux of different MNP.¹²⁴

treatment of MP-contaminated water. After adsorption, the aerogel-based adsorbents can undergo regeneration processes to recover the trapped plastics and restore their adsorption capacity.

Several studies have explored the potential of aerogel-based adsorbents for MNP removal, as listed in Table 5. For instance, Zheng *et al.* fabricated coral-inspired environmentally durable aerogels by utilizing polydopamine-enhanced magnetic chitosan (PDA-MCS) (Fig. 7a).¹⁰⁶ The adsorption and removal/separation capacity of different CS, magnetic chitosan (MCS), and PDA-MCS-based systems were investigated (Fig. 7b and c). The advanced aerogel was found to be highly efficient in the adsorption of PET, PE, and PS type MPs from the contaminated water under different pH conditions (pH = 6–9) with ~94.6%, ~92.3%, and ~97.3% removal efficiency, respectively (Fig. 7d and e). The developed aerogel-based system was found acceptable for up to four recycles in the removal of MPs (Fig. 7f). The

mechanism behind the formation of aerogel-assisted MPs disposed of water involves electrostatic interactions and physical adhesion. Different reaction conditions such as no light, daylight, and UV light were explored to assess the adsorption of MPs through such aerogel systems (Fig. 7g). Further, Zhuang *et al.* selected cellulose nanofiber as a standard matrix material for the preparation of modified cellulose nanofibers-based aerogel.¹²⁵ Here, 2,3-epoxypropyl trimethyl ammonium chloride acts as a modifier whereas polyvinyl alcohol acts as a crosslinker and liquid nitrogen freezing method was used during the preparation of these aerogels. The prepared aerogel system possesses a very high tendency to adsorb small-size MPs (~2 to 5.5 μm) with an adsorption capacity of ~146.38 mg g^{-1} .¹²⁵ The adsorption phenomena follow the pseudo-second-order kinetic model and Langmuir adsorption isotherm acts as an efficient model to be operational mechanistically in removing



Table 5 Representation of various types of aerogel systems in MPs removal along with different parameters

Aerogel				
System	Model system employed for efficacy assessment of the system	Underlying mechanistic principle	Removal efficiency	Ref.
Cellulose nanofiber aerogel modified by quaternary ammonium salt (chemical synthesis)	Deionized water	Adsorption mechanism/ electrostatic action	~146.38 mg g ⁻¹ adsorption capacity of MPs by fluorescence characteristics	125
Zinc MOF-based wood aerogel (<i>in situ</i> method)	Deionized water	Electrostatic interaction	~91.4 and 85.8% removal of polydifluoroethylene and PS MNP	28
Coral bio-inspired aerogel (polydopamine and CS) (self-polymerization and freeze-drying)	Ethanol	Surface and intra-pore adsorption	~91.6% removal of PET MPs	106
Egg protein-derived ultralightweight hybrid monolithic aerogel (pyrolysis)	Water	Chemical interconnection	~99.9% MNP contaminants	153
Microporous carbon aerogels (fast polycondensation of dialdehyde and silsesquioxane)	Methanol, ethanol, HF	Electrocatalyst support/ adsorption	Rapid removal of ~90% MPs in <10 s	154
Cyclodextrin-graphene oxide aerogel microspheres (electrospraying, freeze casting, and self-assembly)	Organic solvents	π - π and adsorptive chemical interaction	~17 mg g ⁻¹ removal of organic micropollutants (2,4-dichlorophenol)	155
Modified aerogel system where cellulose nanofiber acts as a matrix and 3-glycidoxypropyltrimethoxysilane/polyethyleneimine acts as a modifier	Acetic acid and KOH	Ultrasonic and freeze-drying methods	~117.04 mg g ⁻¹ adsorption capacity for MPs (pseudo-second-order kinetic model)	156
Lanthanum-modified k-carrageenan/sodium alginate aerogel system	Deionized water	Sol-gel technique	Removal of ciprofloxacin hydrochloride antibiotic along with existing MPs	150
Facile crosslinking brittle oxide nanofibers into elastic 3D fibrous aerogels	Organic solvents	Smart screening	100% removal efficiency	157

MPs from contaminated water. In another study, Meng *et al.* developed facile crosslinking of brittle oxide-based nanofibers into 3D-based aerogels on the basis of smart screening.¹⁵⁷ The methodology exhibited high efficiency in evaporating water from MP-polluted seawater sources powered through solar energy which basically enables the vapor to condense naturally and harvest the same in an electricity-free system. A 100% removal efficiency was observed for the MPs from contaminated water by using the aerogel system, further, it can be reused in pressure sensing post-solvent evaporation. Very recently, Zhuang *et al.* developed a highly modified aerogel system where cellulose nanofiber acts as a matrix and 3-glycidoxypropyltrimethoxysilane/polyethyleneimine as a modifier.¹⁵⁶ The fabricated aerogel system was successful in the removal/isolation/separation of MPs from water with an adsorption capacity of ~117.04 mg g⁻¹ at room temperature conditions. The adsorption phenomenon followed the pseudo-second-order kinetic model and the Langmuir isotherm, suggesting that MPs particles are adsorbed as a single molecular layer on an aerogel's non-homogeneous surface.

However, research on the application of aerogel-based adsorbents for MPs and NPs removal/isolation/separation is

still in its early stages, and further investigations are needed to optimize the adsorption performance *via* a complete understanding of the underlying mechanisms. The properties of the aerogels, including their surface chemistry, pore structure, and composition, should be tailored specifically to target the unique selective characteristics of MNP. A comparative analysis of the conventional as well as our proposed emerging frameworks in terms of their efficacy, intrinsic properties, and operational efficiencies is shown in Fig. 8a and b. Moreover, it addresses the increased effectiveness and specificity demonstrated by emerging adsorption methods in comparison to traditional ones, supporting their incorporation into contemporary water treatment systems. The comparative analysis presented here provides insightful information about how pollutant removal technologies are developing and steers the field of water purification efforts in the direction of greater sustainability and environmental stewardship.

Thus, the present review offers an in-depth investigation of the theoretical and methodological issues concerning the removal of MNP from diverse environmental matrices using aerogels, electrospun frameworks, and biochar-based constructs. Further theoretical exploration and empirical



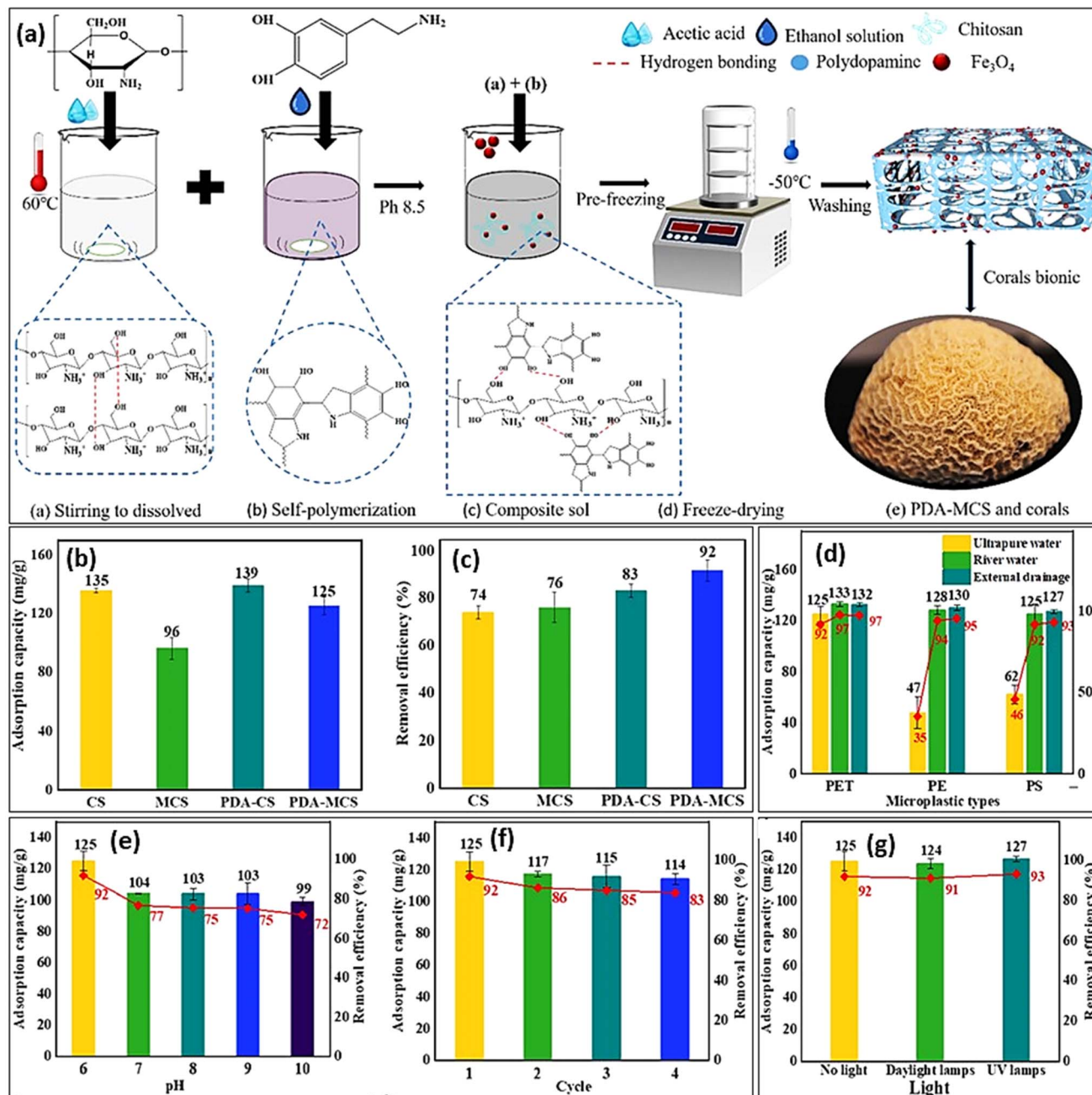


Fig. 7 (a) Synthesis of polydopamine-enhanced magnetic chitosan (PDA-MCS) aerogel and corals, (b) adsorption and (c) removal efficiency of different synthesized systems, (d) adsorption capacity in different water media for various MPs, (e) adsorption capacity and removal efficiency of developed aerogel at different pH values, (f) after various cycles of reuse, and (g) in the presence of different light conditions.¹⁰⁶

investigations are needed to validate performance under varying environmental conditions, assess long-term effectiveness, and explore synergistic effects when combining these materials. To address the underlying causes of plastic pollution, policy interventions may also focus on consumer behaviour, waste management techniques, and the manufacturing process of plastics. Collaborative approaches should be given priority in future research and policy endeavors in order to expedite the conversion of scientific information into practical solutions for ensuring sustainability.

7. Challenges and future perspectives

Aerogels, biochar, and electrospun micro-/nano fibrous constructs though offer potential solutions for the removal of MNP from various environments. Studies on this perspective are still in the preliminary stages with a broad set of open questions. In the coming decades, the demand for advanced smart, and efficient systems dealing with MNP contamination in water will likely increase as water scarcity and its associated environmental sustainability issues become more pressing.^{30,151}



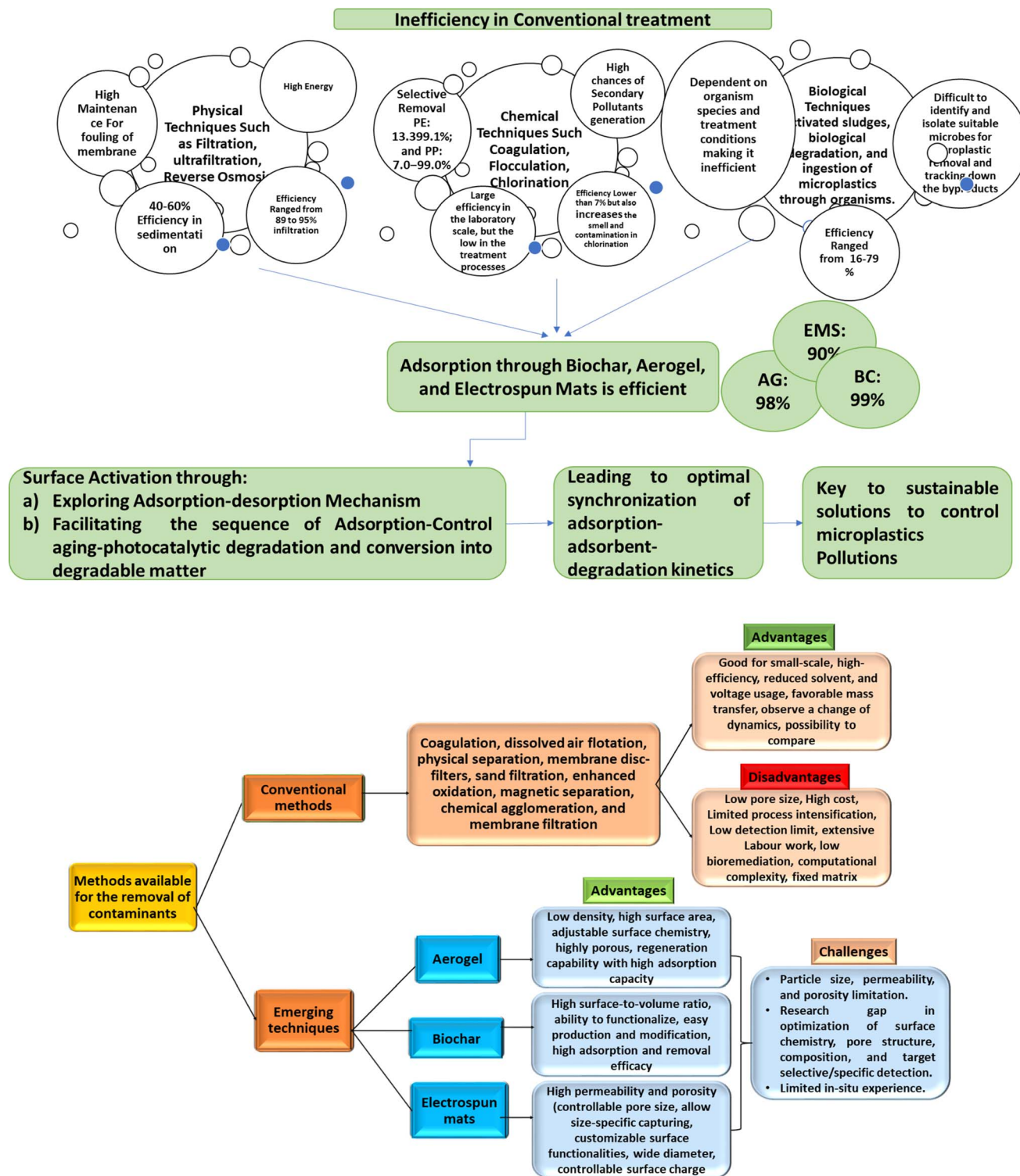


Fig. 8 Schematic representation highlighting the inefficiency of conventional treatment and efficiency (in terms of effectiveness, properties, and removal efficacy) of biochar, aerogel, and electrospun mats for the removal of MPs.

Such efforts may lead to the creation of new and improved classes of hybrid adsorbents to be employed in a range of domains, from large-scale WWTPs to small-scale household water management systems. Meanwhile, at this nascent stage, the utilization of biochar, aerogel, and electrospun-based adsorbent systems for MNP removal/separation/isolation faces

several challenges that hinder the development of more efficient, sustainable, and viable technologies of which scalability, cost-effectiveness, and long-term environmental impacts prior to their integrated implementation as full-scale technology are the significant ones. Therefore, it is crucial to conceptually address the following aspects in their further exploration and



development while dealing with technical challenges pertaining to (a) their practical implications, (b) their long-term interactions with the water ecosystem with regard to their influence on aquatic micro-organisms, and (c) the post-removal/segregation strategy to handle/neutralize such MNP. Some of the challenges associated with these techniques are as below.

7.1 Particle size, porosity, and permeability limitations of adsorbent systems

The effectiveness of aerogels, beads, and electrospun constructs in removing MNP depends on their pore dimensions, pore structure, and surface area. Aerogels are typically expected to capture a wide range of MPs particles and rather non-selectivity where designed with suitable shapes along with pore size, and their application mostly may remain regiospecific and in a certain locality of concern. In contrast, micro/nano-sized and micro/nanoporous biochar particles with a significant surface area are envisaged to be able to effectively capture and retain smaller anisotropic particles, considering the fact that the MNP capturing mechanism involves site-specific locking. For example, Wang *et al.* developed corn straw biochar with $\sim 808.3 \pm 81.9 \text{ m}^2 \text{ g}^{-1}$ total specific surface area, $\sim 1.318 \text{ mg m}^{-3}$ particle density value, and $\sim 47.28 \text{ mL}$ pore volume. The formed biochar was found effective in the removal of $\sim 95\%$ of MPs spheres.¹⁵⁸ Reportedly, biochar is effective in the removal of MPs spheres, provided biochar particle size remains appropriate and consistent since size-scale incompatibility between pores and MPs size may impede such a process. In contrast, electrospun fibrous constructs are used to favourably alter the passage of water and unwanted smaller particles while selectivity capturing MNP. For instance, Wang *et al.* developed a low-pressure driven electrospun membrane for the removal of PS. The electrospun membrane possesses outstanding mechanical stability as it was highly dependent upon the tensile strength ($\sim 8 \text{ MPa}$) and tensile strain ($\sim 58.5 \pm 20.1\%$). Their high values suggest mechanical strength augmentation along with durability.¹³⁹ Hence, apart from the porosity of such frameworks striking the right balance between capturing efficiency and water permeability may be a challenge. The typical mechanism behind the functional role of the electrospun membrane remains (a) permeation-assisted trapping and (b) ligand-induced selective binding of the membrane.¹⁵⁹

7.2 Heterogeneity and polydispersity (size and nature) of MNP

MPs and NPs come in various shapes, sizes, and chemical compositions making their chemical functionality specific and size-selective removal/separation a complex task. The differential affinity of a diverse range of MNP to biochar functionality makes the process of separation adsorption/absorption extent dependent. In an effort to demonstrate the same, for example, Wu *et al.* developed rice straw-derived biochar with an excellent specific surface area of $\sim 327.65 \text{ m}^2 \text{ g}^{-1}$.¹⁶⁰ The result showed that the developed biochar with remarkable surface area and porosity offers excellent removal of MPs with a diameter of $\sim 300 \text{ nm}$ from contaminated water. Similarly, for instance,

Juraj *et al.* electrospun PU nanofiber membranes for the complete removal and separation of MPs and NPs (PS, ABS, and PMMA) from aqueous mediums.¹²⁴ The prepared membrane possesses a fouling nature that follows a standard blocking model where foulants adhere to the inner wall of fiber and help in the removal of MNP.

7.3 Adsorption capacity selectivity and specificity

High adsorption capacity to effectively capture and retain MNP without compromising much in their chemical functionality and other desirable attributes such as mechanical stability or liability or hydrolytic stability may be extremely essential. For example, Ozden *et al.* developed egg protein-derived ultra-lightweight hybrid monolithic aerogel for the removal of MP. The adsorption capacity of developed aerogel was found $\sim 32.6 \text{ mg g}^{-1}$ for the effective removal of MPs.¹⁵³ However, the adsorption capacity of biochar for MNP may be limited by other substances present in water. Natural organic matter, dissolved ions, and other contaminants can compete with plastic particles for adsorption sites on biochar surfaces, reducing the overall removal efficiency. For instance, when biochar is introduced to a substrate over a specified concentration, it may have the effect of agglomeration, which lowers the specific surface area overall, preventing binding sites from being exposed and subsequently weakening the adsorption capacity.¹⁶¹ To demonstrate this effect, it has been observed that the adsorption of diclofenac by biochar fell from around 107 to approximately 25% when the quantity of biochar was raised from ~ 2 to $\sim 20 \text{ g L}^{-1}$.¹⁶¹ Similarly, in electrospun construct, the selective capture of MNP generally remains controlled by their network structure and is challenging in complex water matrices with diverse particle sizes and compositions.

7.4 Scale-up and cost-effectiveness

The scalability and cost-effectiveness of aerogels and biochar-based adsorbent systems for large-scale water treatment/decontamination with regard to MNP involve severe manufacturing engineering challenges. The design, operation, and maintenance of such decontamination-aiming filtration systems remain complex due to factors such as energy-intensive supercritical drying process as necessary for aerogels or the necessary optimization to ensure maximized efficiency while keeping the flow rate, contact time, and system stability in view for biochar-based set-ups. In the case of electrospun constructs as nanofibrous membranes, the large-scale manufacturing costs need to be connected by enhanced efficiency. For example, Risch and Adhart developed a highly cost-effective CS/polyethylene oxide nanofiber sponge through the electrospinning method. The fabricated sponge was highly effective in the removal of poly(ethylene terephthalate) MPs up to $\sim 99.6\%$.¹⁴⁰ The nanofiber sponge was found to be easy to synthesize in bulk amounts which suggests their outstanding scalability.

7.5 Stability, durability, regeneration and reusability

The adsorption capacity of aerogels, biochar, or even electrospun materials tends to suffer from fouling over a period of



time, attributed to the saturation of the adsorption sites by the MPs and other contaminants. Therefore, suitable methods need to be adopted to ensure the regeneration and reusability of such frameworks. For instance, Wu *et al.* developed a rice straw-modified biochar system for the effective removal of MPs from the water. The formed biochar can be easily recovered through thermal treatment after the adsorption of MPs and can be reused again in the same manner. The system works efficiently and acts as regenerative in nature for up to >3 cycles.¹⁶⁰ Similarly, in electrospun frameworks surface characteristics modulation *via* premodification during membrane fabrication by incorporating functional agents into spinning solution and post modification as the deposited membrane surface may enhance anti-fouling performance. Interestingly, aerogels may get fouled by water organic matter and other minerals hindering their ability to capture MNP. Hence, strategies to mitigate such surface fouling by smarter approaches may be necessary.

7.6 Impact on the environment

The environmental consequences of using such material frameworks need to be carefully assessed in such a manner that they should not enhance secondary MNP concentration post-use or production. Aerogels remain highly porous lightweight materials whose manufacturing, use, and disposal may leave their footprints. For example, Zheng *et al.* coral inspired environmental durability aerogel for the eradication of MPs from the aquatic environment.¹⁰⁶ The formed aerogel system was found highly biocompatible, reusable, and environmentally friendly/durable in nature. Similarly, biochar may require a significant amount of biomass and hence sustainable approach is necessary. Disposal of spent biochar post MNP removal/separation must be assessed while aligning to maximized benefits of life-cycle analysis. Electrospun nanoporous membrane systems should also be looking from the angle that the membranes themselves do not become the source of MNP. Therefore, the utilization of biopolymers and non-toxic solvents in electrospinning referred to as bio-electrospinning, offers a promising avenue for developing efficient nanofibrous membranes for removal of MNP, aligning with sustainability considerations.

Furthermore, one of the primary sources of error lies in the accuracy of MNP quantification (in size, volume, and chemical constitutions) methods. Another source of error stems from assumptions about MNP behavior in aqueous environments, exhibiting complex transport dynamics influenced by factors such as particle size, surface properties, and environmental conditions. For example, Zhang *et al.* developed coagulant-coated bubbles for the removal of MPs and organic matter. Here the saline condition of media plays a significant role in the separation of MPs whereas the interaction behavior of MPs and organic matter influences the synergistic removal in different water mediums.¹⁶² Additionally, the effectiveness of the proposed emerging adsorbents, such as biochar, aerogel, and electrospun fiber, for MNP removal may be influenced by various factors such as adsorbent morphology, surface

chemistry, and the presence of other dyes and heavy metal ions. For instance, the work reported by Li *et al.* involves the removal of MPs by considering different aging times through magnetic corncob biochar under the specified environmental condition.¹²⁷ Recently, Zhuang *et al.* carried out the preparation of cellulose-based aerogels with structural direction through the freezing mechanism of liquid nitrogen. The developed aerogel can significantly adsorb MPs by reaching the stage of equilibrium in 20 min.¹⁵⁶ Further, Juraji *et al.* (2023) developed electrospun PU nanofiber for the removal of MPs and nanoplastic. The mechanistic behavior depends upon the adsorptive separation and size-exclusive interception analysis respectively. Here, Langmuir adsorption isotherm defines the adsorption behavior of MNP.¹²⁴ Assumptions regarding the adsorption mechanisms and kinetics of MNP onto adsorbents should be validated through experimental studies conducted under specified controlled conditions. For example, Heo *et al.* designed magnetic iron oxide nanoparticles for the removal of PS-based MPs through various adsorption isotherms and kinetic modeling. However, the maximum adsorption capacity was observed through the Langmuir adsorption isotherm model and pseudo-second kinetic model.¹⁶³ Characterizing the adsorption capacity and affinity of adsorbents for different MNP types can improve the understanding of adsorption processes and optimize the adsorbent design for enhanced MNP removal efficiency. For example, Wang *et al.* performed the thermal degradation and adsorption of MNP through Mg/Zn modified magnetic biochar adsorbents. The modified ability of biochar highly augments the removal efficacy in a significant manner and through adsorbent regeneration/recyclability.¹³² In this regard, integrating advanced analytical techniques, computational modeling approaches, and experimental validation studies can provide comprehensive insights into MNP behavior and facilitate the development of effective strategies for MNP pollution mitigation. For instance, Mendonca *et al.* fabricated bacterial cellulose based hydrogel as a significant bio-flocculants for the eradication of MPs through computational analysis which implement the ratio and grinding times acts as potential variables during reliability and flocculation rate.¹⁰⁸

8. Conclusion

The review paper delves into the escalating concerns on MNP pollution, along with its detrimental impacts on both the environment and human health. A systematic overview of the limitations of conventional techniques like filtration, sedimentation, and electrocoagulation, which have been widely employed for MNP removal, exposed challenges related to their efficacy, sustainability, and cost-effectiveness. Considering these challenges, further development has been stressed regarding the effective removal of MNP, contributing to the development of eco-friendly solutions. In this context, the current attempt aims at a comprehensive analysis of the three promising materials including biochar, electrospun fibrous constructs, and aerogels, that not only offer promising avenues for the effective removal of MNP from diverse environments but also align with the principles of circularity by utilizing waste



materials, reducing the demand for single-use plastics, and promoting circular economy practices. However, it is important to note that further research, innovation, and collaboration are essential to optimize their performance and scale up production to fully exploit their long-term environmental advantages. Challenges such as scalability, cost-effectiveness, and long-term environmental impacts need to be addressed to ensure the practical implementation and commercial viability of these solutions, thereby realizing effective and implementable solutions for MNP removal. Thus, the current review offers a visionary glimpse into the transformative potential of these innovative adsorbent materials in combating the escalating crisis of plastic pollution, thereby beckoning material designers toward a future where environmental sustainability and cutting-edge technology join forces to safeguard the planet from such threats. Conclusively, future directions for developing and implementing such sustainable framework-based adsorbent systems to mitigate micro- and nano-plastic pollution encompass advancing innovative material development, comprehensively understanding adsorption mechanisms, rigorously evaluating real-world performance while ensuring environmental safety, optimizing procedures for scalability, and integrating further interdisciplinary research.

Data availability

No primary research results, software or code have been included and no new data were generated or analysed as part of this review.

Author contributions

Krishna Priyadarshini Das: conceptualization; writing – original draft & editing; Pooja Chauhan: writing – review & editing; Ulrike Staudinger: writing – review & editing; Bhabani Kumar Satapathy: conceptualization, resources; writing – review & editing.

Conflicts of interest

There are no conflicts to declare.

References

- M. Bergmann, B. C. Almroth, S. M. Brander, T. Dey, D. S. Green, S. Gundogdu, A. Krieger, M. Wagner and T. R. Walker, *Science*, 2022, **376**, 469–470.
- S. Afzal, A. Singh, S. R. Nicholson, T. Uekert, J. S. DesVeaux, E. C. D. Tan, A. Dutta, A. C. Carpenter, R. M. Baldwin and G. T. Beckham, *Green Chem.*, 2023, **25**, 5068–5085.
- S. Hohn, E. Acevedo-Trejos, J. F. Abrams, J. Fulgencio de Moura, R. Spranz and A. Merico, *Sci. Total Environ.*, 2020, **746**, 141115.
- S. Vijayaraman, P. Mondal, A. Nandan and N. A. Siddiqui, *Presence of Microplastic in Water Bodies and Its Impact on Human Health*, 2020, pp. 57–65, DOI: [10.1007/978-981-15-0954-4_4](https://doi.org/10.1007/978-981-15-0954-4_4).
- M. N. Issac and B. Kandasubramanian, *Environ. Sci. Pollut. Res.*, 2021, **28**, 19544–19562.
- S. N. Dimassi, J. N. Hahladakis, M. N. D. Yahia, M. I. Ahmad, S. Sayadi and M. A. Al-Ghouti, *Arabian J. Chem.*, 2022, **15**, 104262.
- M. S.-L. Yee, L.-W. Hii, C. K. Looi, W.-M. Lim, S.-F. Wong, Y.-Y. Kok, B.-K. Tan, C.-Y. Wong and C.-O. Leong, *Nanomaterials*, 2021, **11**, 496.
- M. Zhang, J. Yang, Z. Cai, Y. Feng, Y. Wang, D. Zhang and X. Pan, *Environ. Sci.: Nano*, 2019, **6**, 709–735.
- Q. Wang, J. Bai, B. Ning, L. Fan, T. Sun, Y. Fang, J. Wu, S. Li, C. Duan, Y. Zhang, J. Liang and Z. Gao, *Chemosphere*, 2020, **254**, 126788.
- E. Fournier, J. Ratel, S. Denis, M. Leveque, P. Ruiz, C. Mazal, F. Amiard, M. Edely, V. Bezirard, E. Gaultier, B. Lamas, E. Houdeau, E. Engel, F. Lagarde, L. Etienne-Mesmin, M. Mercier-Bonin and S. Blanquet-Diot, *J. Hazard. Mater.*, 2023, **443**, 130383.
- D. Kawecki, P. R. W. Scheeder and B. Nowack, *Environ. Sci. Technol.*, 2018, **52**, 9874–9888.
- S. S. Paul, R. Anirud, B. Bahl, K. Maheshwari and A. Banerjee, in *Visualization Techniques for Climate Change with Machine Learning and Artificial Intelligence*, Elsevier, 2023, pp. 77–100.
- R. Ambrosini, R. S. Azzoni, F. Pittino, G. Diolaiuti, A. Franzetti and M. Parolini, *Environ. Pollut.*, 2019, **253**, 297–301.
- C. Jiang, L. Yin, Z. Li, X. Wen, X. Luo, S. Hu, H. Yang, Y. Long, B. Deng, L. Huang and Y. Liu, *Environ. Pollut.*, 2019, **249**, 91–98.
- D. Tatsii, S. Bucci, T. Bhowmick, J. Guettler, L. Bakels, G. Bagheri and A. Stohl, *Environ. Sci. Technol.*, 2024, **58**, 671–682.
- I. Iavicoli, L. Fontana, V. Leso and A. Bergamaschi, *Int. J. Mol. Sci.*, 2013, **14**, 16732–16801.
- A. L. Patrício Silva, J. C. Prata, T. R. Walker, A. C. Duarte, W. Ouyang, D. Barcelò and T. Rocha-Santos, *Chem. Eng. J.*, 2021, **405**, 126683.
- Y. Peng, P. Wu, A. T. Schartup and Y. Zhang, *Proc. Natl. Acad. Sci. U. S. A.*, 2021, **118**(47), DOI: [10.1073/pnas.2111530118](https://doi.org/10.1073/pnas.2111530118).
- S. Dey, P. Samanta, D. Dutta, D. Kundu, A. R. Ghosh and S. Kumar, *Environ. Sci. Pollut. Res.*, 2023, **30**, 93363–93387.
- R. Y. Krishnan, S. Manikandan, R. Subbaiya, N. Karmegam, W. Kim and M. Govarthan, *Sci. Total Environ.*, 2023, **858**, 159681.
- A. R. P. Pizzichetti, C. Pablos, C. Álvarez-Fernández, K. Reynolds, S. Stanley and J. Marugán, *Case Stud. Chem. Environ. Eng.*, 2021, **3**, 100075.
- S. Bolan, L. P. Padhye, T. Jasemizad, M. Govarthan, N. Karmegam, H. Wijesekara, D. Amarasiri, D. Hou, P. Zhou, B. K. Biswal, R. Balasubramanian, H. Wang, K. H. M. Siddique, J. Rinklebe, M. B. Kirkham and N. Bolan, *Sci. Total Environ.*, 2024, **909**, 168388.
- S. H. Joo, Y. Liang, M. Kim, J. Byun and H. Choi, *Environ. Challenges*, 2021, **3**, 100042.



- 24 V. I. Isaeva, M. D. Vedenyapina, A. Y. Kurmysheva, D. Weichgrebe, R. R. Nair, N. P. T. Nguyen and L. M. Kustov, *Molecules*, 2021, **26**, 6628.
- 25 D. Pedrero, C. Edo, F. Fernández-Piñas, R. Rosal and S. Aguado, *Sep. Purif. Technol.*, 2024, **333**, 125816.
- 26 J. Yang, X. Chen, J. Zhang, Y. Wang, H. Wen and J. Xie, *Int. J. Biol. Macromol.*, 2021, **189**, 53–64.
- 27 P. S. Goh, H. S. Kang, A. F. Ismail, W. H. Khor, L. K. Quen and D. Higgins, *Chemosphere*, 2022, **299**, 134418.
- 28 D. You, Y. Zhao, W. Yang, Q. Pan and J. Li, *Chem. Res. Chin. Univ.*, 2022, **38**, 186–191.
- 29 G. Zhu, J. Wang, X. Wang, Y. Liu, W. Wu, H. Xiao, L. Huang, H. Dai, X. Zhou and H. Bian, *ACS Sustain. Chem. Eng.*, 2023, **11**, 13928–13938.
- 30 Z. Wang, M. Sedighi and A. Lea-Langton, *Water Res.*, 2020, **184**, 116165.
- 31 M. Rist and A. Greiner, *RSC Applied Polymers*, 2024, DOI: [10.1039/D3LP00201B](https://doi.org/10.1039/D3LP00201B).
- 32 H. Park and B. Park, *Water*, 2021, **13**, 2736.
- 33 R. Ahmed, A. K. Hamid, S. A. Krebsbach, J. He and D. Wang, *Chemosphere*, 2022, **293**, 133557.
- 34 S. D. Martinho, V. C. Fernandes, S. A. Figueiredo and C. Delerue-Matos, *Int. J. Environ. Res. Public Health*, 2022, **19**, 5610.
- 35 T. A. Lastovina and A. P. Budnyk, *Journal of Water Process Engineering*, 2021, **43**, 102209.
- 36 I. Nabi, A.-U.-R. Bacha and L. Zhang, *J. Cleaner Prod.*, 2022, **337**, 130458.
- 37 V. Kumar, M. Umesh, P. Chakraborty, P. Sharma, S. Sarojini, T. Basheer, K. Kaur, R. Pasrija and D. Barcelo, *TrAC, Trends Anal. Chem.*, 2024, **170**, 117392.
- 38 N. Parashar and S. Hait, *Sci. Total Environ.*, 2021, **759**, 144274.
- 39 N. U. Benson, O. H. Fred-Ahmadu, D. E. Bassey and A. A. Atayero, *J. Environ. Chem. Eng.*, 2021, **9**, 105222.
- 40 M. W. Peng and N. Kathuria, *Journal of Management Studies*, 2021, **58**, 1431–1435.
- 41 C. Guerranti, T. Martellini, G. Perra, C. Scopetani and A. Cincinelli, *Environ. Toxicol. Pharmacol.*, 2019, **68**, 75–79.
- 42 P. K. Cheung and L. Fok, *Water Res.*, 2017, **122**, 53–61.
- 43 E. Hernandez, B. Nowack and D. M. Mitrano, *Environ. Sci. Technol.*, 2017, **51**, 7036–7046.
- 44 Z. Sobhani, Y. Lei, Y. Tang, L. Wu, X. Zhang, R. Naidu, M. Megharaj and C. Fang, *Sci. Rep.*, 2020, **10**, 4841.
- 45 M. G. A. Barros, R. dos Santos Grignet, S. P. B. Bernal, C. D. C. S. Gonçalves, M. R. Z. Passarini and J. R. Ottoni, in *Microbial Inoculants*, Elsevier, 2023, pp. 207–228.
- 46 J. Wang, X. Liu, Y. Li, T. Powell, X. Wang, G. Wang and P. Zhang, *Sci. Total Environ.*, 2019, **691**, 848–857.
- 47 Y. Chen, A. K. Awasthi, F. Wei, Q. Tan and J. Li, *Sci. Total Environ.*, 2021, **752**, 141772.
- 48 Y. Zhang, Y. Li, F. Su, L. Peng and D. Liu, *Sci. Total Environ.*, 2022, **802**, 149658.
- 49 W. Li, R. Wufuer, J. Duo, S. Wang, Y. Luo, D. Zhang and X. Pan, *Sci. Total Environ.*, 2020, **749**, 141420.
- 50 N. Katsumi, T. Kusube, S. Nagao and H. Okochi, *Chemosphere*, 2021, **267**, 129185.
- 51 R. R. Hurley and L. Nizzetto, *Current Opinion in Environmental Science & Health*, 2018, **1**, 6–11.
- 52 J. Li, H. Liu and J. Paul Chen, *Water Res.*, 2018, **137**, 362–374.
- 53 M. Bläsing and W. Amelung, *Sci. Total Environ.*, 2018, **612**, 422–435.
- 54 D. R. da Silva Souza, L. D. Caminhas, J. P. de Mesquita and F. V. Pereira, *Mater. Chem. Phys.*, 2018, **203**, 148–155.
- 55 N. Khalid, M. Aqeel, A. Noman and Z. Fatima Rizvi, *J. Hazard. Mater.*, 2023, **445**, 130455.
- 56 P. J. Kole, A. J. Löhr, F. Van Belleghem and A. Ragas, *Int. J. Environ. Res. Public Health*, 2017, **14**, 1265.
- 57 B. Baensch-Baltruschat, B. Kocher, F. Stock and G. Reifferscheid, *Sci. Total Environ.*, 2020, **733**, 137823.
- 58 J. Worek, X. Badura, A. Białas, J. Chwiej, K. Kawoń and K. Styszko, *Energies*, 2022, **15**, 2816.
- 59 M. Lee and H. Kim, *Nanomaterials*, 2022, **12**, 851.
- 60 F. Haque and C. Fan, *J. Cleaner Prod.*, 2022, **367**, 133027.
- 61 J. C. Prata, J. P. da Costa, I. Lopes, A. C. Duarte and T. Rocha-Santos, *Sci. Total Environ.*, 2020, **702**, 134455.
- 62 N. U. Benson, D. E. Bassey and T. Palanisami, *Heliyon*, 2021, **7**, e06343.
- 63 F. Saliu, M. Veronelli, C. Raguso, D. Barana, P. Galli and M. Lasagni, *Environ. Adv.*, 2021, **4**, 100042.
- 64 S. M. Mintenig, I. Int-Veen, M. G. J. Löder, S. Primpke and G. Gerdt, *Water Res.*, 2017, **108**, 365–372.
- 65 J. Sun, Z.-R. Zhu, W.-H. Li, X. Yan, L.-K. Wang, L. Zhang, J. Jin, X. Dai and B.-J. Ni, *Water Res.*, 2021, **190**, 116784.
- 66 S. Agarwal, *Macromol. Chem. Phys.*, 2020, **221**(6), DOI: [10.1002/macp.202000017](https://doi.org/10.1002/macp.202000017).
- 67 Y. Xiang, L. Jiang, Y. Zhou, Z. Luo, D. Zhi, J. Yang and S. S. Lam, *J. Hazard. Mater.*, 2022, **422**, 126843.
- 68 T. J. Baker, C. R. Tyler and T. S. Galloway, *Environ. Pollut.*, 2014, **186**, 257–271.
- 69 P. Bhattacharya, *Acta Biomedica Scientia*, 2016, **3**, 47–52.
- 70 N. von Moos, P. Burkhardt-Holm and A. Köhler, *Environ. Sci. Technol.*, 2012, **46**, 11327–11335.
- 71 W. Luo, L. Su, N. J. Craig, F. Du, C. Wu and H. Shi, *Environ. Pollut.*, 2019, **246**, 174–182.
- 72 L. Yin, C. Jiang, X. Wen, C. Du, W. Zhong, Z. Feng, Y. Long and Y. Ma, *Int. J. Environ. Res. Public Health*, 2019, **16**, 1650.
- 73 X. Xiong, K. Zhang, X. Chen, H. Shi, Z. Luo and C. Wu, *Environ. Pollut.*, 2018, **235**, 899–906.
- 74 L. Ding, R. Fan Mao, X. Guo, X. Yang, Q. Zhang and C. Yang, *Sci. Total Environ.*, 2019, **667**, 427–434.
- 75 D. Barceló, Y. Picó and A. H. Alfarhan, *Environ. Toxicol. Pharmacol.*, 2023, **101**, 104204.
- 76 D. Gu, P. Zhang, L. Zhang, H. Liu, Z. Pu and S. Shang, *Opt. Mater.*, 2018, **83**, 272–278.
- 77 Y.-J. Chen, Y. Chen, C. Miao, Y.-R. Wang, G.-K. Gao, R.-X. Yang, H.-J. Zhu, J.-H. Wang, S.-L. Li and Y.-Q. Lan, *J. Mater. Chem. A*, 2020, **8**, 14644–14652.
- 78 S. Dong, Q. Ji, Y. Wang, H. Liu and J. Qu, *J. Environ. Sci.*, 2020, **89**, 102–112.
- 79 J. Tan, Y. Chen, Z. Mo, C. Tan, R. Wen, Z. Chen and H. Tian, *Environ. Sci. Pollut. Res.*, 2022, **29**, 61534–61546.



- 80 G. Zhou, Q. Wang, J. Li, Q. Li, H. Xu, Q. Ye, Y. Wang, S. Shu and J. Zhang, *Sci. Total Environ.*, 2021, **752**, 141837.
- 81 N. Khalid, M. Aqeel and A. Noman, *Environ. Pollut.*, 2020, **267**, 115653.
- 82 S. Kublik, S. Gschwendtner, T. Magritsch, V. Radl, M. C. Rillig and M. Schloter, *Front. Environ. Sci.*, 2022, **10**, DOI: [10.3389/fenvs.2022.989267](https://doi.org/10.3389/fenvs.2022.989267).
- 83 M. Sajjad, Q. Huang, S. Khan, M. A. Khan, Y. Liu, J. Wang, F. Lian, Q. Wang and G. Guo, *Environ. Technol. Innovation*, 2022, **27**, 102408.
- 84 C. Wang, J. Zhao and B. Xing, *J. Hazard. Mater.*, 2021, **407**, 124357.
- 85 Y. Zhang, T. Gao, S. Kang and M. Sillanpää, *Environ. Pollut.*, 2019, **254**, 112953.
- 86 Z. Wang, Y. Qin, W. Li, W. Yang, Q. Meng and J. Yang, *Environ. Chem. Lett.*, 2019, **17**, 1821–1830.
- 87 Z. Li, Q. Li, R. Li, Y. Zhao, J. Geng and G. Wang, *Environ. Sci. Pollut. Res.*, 2020, **27**, 30306–30314.
- 88 M. Klein and E. K. Fischer, *Sci. Total Environ.*, 2019, **685**, 96–103.
- 89 J. Zhang, L. Wang, L. Trasande and K. Kannan, *Environ. Sci. Technol. Lett.*, 2021, **8**, 989–994.
- 90 Y. Zhang, T. Gao, S. Kang, H. Shi, L. Mai, D. Allen and S. Allen, *Earth-Sci. Rev.*, 2022, **226**, 103924.
- 91 L. E. Revell, P. Kuma, E. C. Le Ru, W. R. C. Somerville and S. Gaw, *Nature*, 2021, **598**, 462–467.
- 92 X. Li, S. Yao, Z. Wang, X. Jiang, Y. Song and S. X. Chang, *Environ. Pollut.*, 2022, **315**, 120433.
- 93 Y.-L. Zhang, S.-C. Kang and T.-G. Gao, *Advances in Climate Change Research*, 2022, **13**, 455–458.
- 94 J. C. Prata, A. L. P. Silva, J. P. da Costa, C. Mouneyrac, T. R. Walker, A. C. Duarte and T. Rocha-Santos, *Int. J. Environ. Res. Public Health*, 2019, **16**, 2411.
- 95 X. Chen, X. Chen, Y. Zhao, H. Zhou, X. Xiong and C. Wu, *Sci. Total Environ.*, 2020, **719**, 137276.
- 96 M. Zhang, K. Lin, X. Li, L. Wu, J. Yu, S. Cao, D. Zhang, L. Xu, S. J. Parikh and Y. S. Ok, *Environ. Pollut.*, 2022, **293**, 118521.
- 97 V. Pillay, A. Seedat, Y. E. Choonara, L. C. du Toit, P. Kumar and V. M. K. Ndesendo, *AAPS PharmSciTech*, 2013, **14**, 692–711.
- 98 I. E. Gilani, S. Sayadi, N. Zouari and M. A. Al-Ghouti, *Bioresour. Technol. Rep.*, 2023, **24**, 101606.
- 99 R. C. Hale, A. E. King and J. M. Ramirez, in *Microplastic Contamination in Aquatic Environments*, Elsevier, 2024, pp. 1–31.
- 100 S. Ghatge, Y. Yang, J.-H. Ahn and H.-G. Hur, *Appl. Biol. Chem.*, 2020, **63**, 27.
- 101 M. Arhant, M. Le Gall, P.-Y. Le Gac and P. Davies, *Polym. Degrad. Stab.*, 2019, **161**, 175–182.
- 102 T. Poerio, E. Piacentini and R. Mazzei, *Molecules*, 2019, **24**(22), 4148.
- 103 L. Feng, Q. Zhang, F. Ji, L. Jiang, C. Liu, Q. Shen and Q. Liu, *Chem. Eng. J.*, 2022, **430**, 132754.
- 104 L. Liu, M. Xu, Y. Ye and B. Zhang, *Sci. Total Environ.*, 2022, **806**, 151312.
- 105 M. Rashid, N. T. Price, M. Á. Gracia Pinilla and K. E. O'Shea, *Water Res.*, 2017, **123**, 353–360.
- 106 B. Zheng, B. Li, H. Wan, X. Lin and Y. Cai, *J. Hazard. Mater.*, 2022, **431**, 128611.
- 107 M. Malankowska, C. Echaide-Gorriz and J. Coronas, *Environ. Sci.: Water Res. Technol.*, 2021, **7**, 243–258.
- 108 I. Mendonça, J. Sousa, C. Cunha, M. Faria, A. Ferreira and N. Cordeiro, *Chemosphere*, 2023, **314**, 137719.
- 109 Q. Guo, Y. Liu, J. Liu, Y. Wang, Q. Cui, P. Song, X. Zhang and C. Zhang, *Chem. Mater.*, 2022, **34**, 5165–5175.
- 110 M. E. Mahmoud, M. F. Amira, S. Daniele, M. E. Abouelanwar and B. M. Morcos, *J. Taiwan Inst. Chem. Eng.*, 2023, **146**, 104793.
- 111 A. Ullah, G.-J. Lee, H. Kim, H. T. Kwon and S. I. Lim, *Int. J. Biol. Macromol.*, 2023, **242**, 124698.
- 112 J. Talvitie, A. Mikola, A. Koistinen and O. Setälä, *Water Res.*, 2017, **123**, 401–407.
- 113 K. Gurung, M. C. Ncibi and J. M. Fontmorin, *J. Membr. Sci. Technol.*, 2016, **6**(3), DOI: [10.4172/2155-9589.1000158](https://doi.org/10.4172/2155-9589.1000158).
- 114 C. Akarsu, H. Kumbur and A. E. Kideys, *Water Sci. Technol.*, 2021, **84**, 1648–1662.
- 115 L. T. Yogarathinam, J. Usman, M. H. D. Othman, A. F. Ismail, P. S. Goh, A. Gangasalam and M. R. Adam, *J. Hazard. Mater.*, 2022, **424**, 127298.
- 116 L. Yang, X. Cao, J. Cui, Y. Wang, Z. Zhu, H. Sun, W. Liang, J. Li and A. Li, *J. Colloid Interface Sci.*, 2022, **617**, 673–682.
- 117 N. Yahyanezhad, M. J. Bardi and H. Aminirad, *Water Practice & Technology*, 2021, **16**, 782–792.
- 118 H. Wan, J. Wang, X. Sheng, J. Yan, W. Zhang and Y. Xu, *Toxics*, 2022, **10**, 70.
- 119 G. Gnanasekaran, A. G and Y. S. Mok, *Sep. Purif. Technol.*, 2021, **277**, 119655.
- 120 J. Qin, Y. Dou, F. Wu, Y. Yao, H. R. Andersen, C. Hélix-Nielsen, S. Y. Lim and W. Zhang, *Appl. Catal., B*, 2022, **319**, 121940.
- 121 S. Modak, M. Kasula and M. R. Esfahani, *ACS Appl. Eng. Mater.*, 2023, **1**, 744–755.
- 122 M. Haris, M. W. Khan, A. Zavabeti, N. Mahmood and N. Eshtiaghi, *Chem. Eng. J.*, 2023, **455**, 140390.
- 123 B. Fryczkowska and L. Przywara, *Desalin. Water Treat.*, 2021, **214**, 252–262.
- 124 K. Juraj, S. P. Ammed, C. Chingakham, B. Ramasubramanian, S. Ramakrishna, S. Vasudevan and A. Sujith, *ACS Appl. Nano Mater.*, 2023, **6**, 4636–4650.
- 125 J. Zhuang, N. Rong, X. Wang, C. Chen and Z. Xu, *Sep. Purif. Technol.*, 2022, **293**, 121133.
- 126 S. Ye, M. Cheng, G. Zeng, X. Tan, H. Wu, J. Liang, M. Shen, B. Song, J. Liu, H. Yang and Y. Zhang, *Water Res.*, 2020, **179**, 115876.
- 127 J. Li, X. Chen, S. Yu and M. Cui, *Sci. Total Environ.*, 2023, **875**, 162647.
- 128 T. Chen, Y. Zhang, H. Wang, W. Lu, Z. Zhou, Y. Zhang and L. Ren, *Bioresour. Technol.*, 2014, **164**, 47–54.
- 129 M. Verma, P. Singh and M. Dhanorkar, *J. Chem. Technol. Biotechnol.*, 2024, **99**, 330–342.
- 130 N. Singh, N. Khandelwal, Z. A. Ganie, E. Tiwari and G. K. Darbha, *Chem. Eng. J.*, 2021, **418**, 129405.
- 131 V. Siipola, S. Pflugmacher, H. Romar, L. Wendling and P. Koukkari, *Appl. Sci.*, 2020, **10**, 788.



- 132 J. Wang, C. Sun, Q.-X. Huang, Y. Chi and J.-H. Yan, *J. Hazard. Mater.*, 2021, **419**, 126486.
- 133 J. Wu, C. Yang, H. Zhao, J. Shi, Z. Liu, C. Li and F. Song, *Environ. Sci. Pollut. Res.*, 2022, **30**, 26914–26928.
- 134 L. Hsieh, L. He, M. Zhang, W. Lv, K. Yang and M. Tong, *Water Res.*, 2022, **221**, 118783.
- 135 Y. Sun, S. M. Shaheen, E. F. Ali, H. Abdelrahman, B. Sarkar, H. Song, J. Rinklebe, X. Ren, Z. Zhang and Q. Wang, *Environ. Pollut.*, 2022, **306**, 119339.
- 136 Y. Shi, J. Du, T. Zhao, B. Feng, H. Bian, S. Shan, J. Meng, P. Christie, M. H. Wong and J. Zhang, *Environ. Pollut.*, 2023, **318**, 120897.
- 137 Q. Shi, S. Guo, J. Tang, H. Lyu, C. Ri and H. Sun, *Environ. Pollut.*, 2023, **316**, 120696.
- 138 M. N. Pervez, M. R. Mishu, M. E. Talukder, G. K. Stylios, A. Buonerba, S. W. Hasan, Y. Cai, Y. Zhao, A. Figoli, T. Zarra, V. Belgiorno and V. Naddeo, *Water Emerging Contam. Nanoplast.*, 2022, **1**, 10.
- 139 R. Wang, L. Zhang, B. Chen and X. Zhu, *J. Membr. Sci.*, 2020, **614**, 118470.
- 140 P. Risch and C. Adlhart, *ACS Appl. Polym. Mater.*, 2021, **3**, 4685–4694.
- 141 T. Aslan, S. Arslan, M. Eyvaz, S. Güçlü, E. Yüksel and İ. Koyuncu, *J. Membr. Sci.*, 2016, **520**, 616–629.
- 142 Z. Wang, C. Crandall, R. Sahadevan, T. J. Menkhaus and H. Fong, *Polymer*, 2017, **114**, 64–72.
- 143 Y. Liu, R. Wang, H. Ma, B. S. Hsiao and B. Chu, *Polymer*, 2013, **54**, 548–556.
- 144 A. K. Gautam, C. Lai, H. Fong and T. J. Menkhaus, *J. Membr. Sci.*, 2014, **466**, 142–150.
- 145 C. Hachemi, M. Enfrin, A. O. Rashed, V. Jegatheesan, P. D. Hodgson, D. L. Callahan, J. Lee and L. F. Dumée, *Chemosphere*, 2023, **310**, 136891.
- 146 M. Enfrin, C. Hachemi, D. L. Callahan, J. Lee and L. F. Dumée, *Sep. Purif. Technol.*, 2021, **278**, 119592.
- 147 F. Zhu, Y.-M. Zheng, B.-G. Zhang and Y.-R. Dai, *J. Hazard. Mater.*, 2021, **401**, 123608.
- 148 R. Senthil, V. Sumathi, A. Tamilselvi, S. B. Kavukcu and A. W. Aruni, *Sci. Rep.*, 2022, **12**, 8411.
- 149 H. Chen, M. Huang, Y. Liu, L. Meng and M. Ma, *Sci. Total Environ.*, 2020, **739**, 139944.
- 150 F. Yu, Z. Yang, X. Zhang, P. Yang and J. Ma, *Environ. Technol. Innovation*, 2021, **24**, 102052.
- 151 F. Liu, Y. Zhang, L. Zhong, Q. Feng, Z. Dong and Z. Xu, *Int. J. Biol. Macromol.*, 2024, **256**, 128326.
- 152 Q. Liu and S. M. Khor, *TrAC, Trends Anal. Chem.*, 2024, **170**, 117465.
- 153 S. Ozden, S. Monti, V. Tozzini, N. S. Dutta, S. Gili, N. Caggiano, A. J. Link, N. M. Pugno, J. Higgins, R. D. Priestley and C. B. Arnold, *Mater. Today*, 2022, **59**, 46–55.
- 154 Z. Chen, S. Liu, L. Chen, J. Huang, B. Zheng, W. Huang, S. Li, Y. Lu and R. Fu, *Chem. Eng. J.*, 2021, **418**, 129315.
- 155 Z.-J. Nie, Q.-F. Guo, H. Xia, M.-M. Song, Z.-J. Qiu, S.-T. Fan, Z.-H. Chen, S.-X. Zhang, S. Zhang and B.-J. Li, *J. Environ. Chem. Eng.*, 2021, **9**, 104749.
- 156 J. Zhuang, M. Pan, Y. Zhang, F. Liu and Z. Xu, *Int. J. Biol. Macromol.*, 2023, **235**, 123884.
- 157 X. Meng, X. Peng, Y. Wei, S. Ramakrishna, Y. Sun and Y. Dai, *Chem. Eng. J.*, 2022, **437**, 135444.
- 158 H. Wang, H. Teng, X. Wang, J. Xu and L. Sheng, *J. Environ. Manage.*, 2022, **310**, 114758.
- 159 Z. Meng, L. Zhu, X. Wang and M. Zhu, *Acc. Mater. Res.*, 2023, **4**, 180–192.
- 160 W. Wu, M. Yang, Q. Feng, K. McGrouther, H. Wang, H. Lu and Y. Chen, *Biomass Bioenergy*, 2012, **47**, 268–276.
- 161 M. Dong, L. He, M. Jiang, Y. Zhu, J. Wang, W. Gustave, S. Wang, Y. Deng, X. Zhang and Z. Wang, *Int. J. Environ. Res. Public Health*, 2023, **20**, 1679.
- 162 M. Zhang, J. Yang, Z. Kang, X. Wu, L. Tang, Z. Qiang, D. Zhang and X. Pan, *J. Hazard. Mater.*, 2021, **404**, 124095.
- 163 Y. Heo, E.-H. Lee and S.-W. Lee, *Chemosphere*, 2022, **307**, 135672.

