

Environmental Science Nano

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1
2 Plastic pollution is a really concerning topic, inter alia because of the degradation of plastics in
3 nanoparticles that are easily uptaken by living organisms and poorly degradable. Thus, there is a trend
4 toward the development of biobased and biodegradable plastics, and poly hydroxyalkanoates are
5 among the most promising ones. However, they too can fragment into nanoparticles and show effects
6 on living cells. Consequently, we have studied the effect of polylactide nanoparticles on macrophages,
7 i.e. professional phagocytes conserved in evolution, by a combination of proteomic and targeted
8 approaches. No inflammatory response was observed after exposure to polylactide. However, plastics-
9 exposed cells showed a decreased ability to respond to a bacterial stimulus, suggesting slightly
10 impaired immune functions of macrophages after exposure to plastics.
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2 Biobased, Biodegradable but not bio-neutral: about the effects of polylactic acid nanoparticles on
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4 macrophages.
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Abstract

Plastics are persistent pollutants, because of their slow degradation, which suggests that they may lead to cumulative and/or delayed adverse effects due to their progressive accumulation over time. Macroplastics produced by human activity are released in the environment, where they degrade into micro and nanoplastics that are very easily uptaken by a wide variety of organisms, including humans. Microplastics and nanoplastics being particulates, they are handled in the body by specialized cells such as macrophages (or their evolutionary counterparts), where they can elicit a variety of responses. One solution to alleviate the problems due to biopersistence, such as accumulation over life, would be to use biodegradable plastics. One of the emerging biodegradable plastics being polylactide, we decided to test the responses of macrophages to polylactide nanoparticles, using a combination of untargeted proteomics and targeted validation experiments. Proteomics showed important adaptive changes in the proteome in response to exposure to polylactide nanoparticles. These changes affected for example mitochondrial, cytoskeletal and lysosomal proteins, but also proteins implicated in immune functions or redox homeostasis. Validation experiments showed that many of these changes were homeostatic, with no induced oxidative stress and no gross perturbation of the mitochondrial function. However, polylactide particles altered the immune functions such as phagocytosis (-20%) or cytokine production (2-fold increase for TNF production), which may translate into a decreased ability to macrophages to respond to bacterial infections. Furthermore, polylactide particles also induced moderate cross-toxicity with some quinones such as phenanthrene quinone, a combustion by-product that is a suspected carcinogen.

1. Introduction

The wide use of plastics in very diverse areas (e.g. packaging, automotive, textile, electronics, to quote just a few) translates into huge annual production figures, i.e. close to a gigaton/year¹. Unfortunately a huge proportion of these plastics, estimated to 500 Mt/year, is released in the environment², where it has deleterious effects that are more and more documented in detail, on a wide range of marine taxa³, e.g. sea birds⁴. This pollution was first documented in aquatic marine environments⁵⁻⁸, but is now found in marine sediments^{9,10}, freshwater environments¹¹⁻¹³, and also in terrestrial ones^{14,15}.

Although plastics are chemically resistant and degrade very slowly, with half lives in the environment amounting in decades¹⁶, they progressively fragment, first into microplastics (lesser than 5 mm in size and then into nanoplastics (less than 1 μm in size). These nanoplastics are far more difficult to detect in the environment, and their effects are much less known. It is however anticipated that the smaller they are, the more easily they will cross biological barriers and be able to disturb the homeostasis of living organisms. This hypothesis has prompted extensive research on the effects of micro and nanoplastics on a wide variety of living organisms from various phyla, ranging from worms^{17,18} to molluscs¹⁹⁻²¹, crustaceans^{22,23}, insects²⁴⁻²⁶, and then vertebrates such as fish²⁷⁻²⁹ and of course mammals³⁰ including humans models (e.g. in³¹⁻³⁵).

Multicellular organisms defend themselves against particles, including plastic particles, by a series of mechanisms. The first line of defense is represented by biological barriers (e.g. intestinal or epidermal barriers) and research has been devoted to understand how these barriers interact with plastic particles. When translocation across these barriers occurs³¹, then a second line of barriers comes into play, and is represented by professional phagocytes. This cell type is encountered in invertebrates (annelids coelomocytes, insect hemocytes) as well as in vertebrates (macrophages, neutrophils). Indeed, it has been shown that this cell type does respond to plastic particles^{18,26,36-42}.

In view of these deleterious effects and of the fact that fabrication of plastics consumes fossil resources and increases the concentration of greenhouse effect gases, the development of biobased and biodegradable plastics has been investigated and is a very active area of research. Plastics that are both biobased and biodegradable belong to two main families. The first family is represented by modified polysaccharides (e.g. modified starch or cellulose derivatives), and the second family is represented by polyhydroxyalkanoates, i.e. polymers of natural hydroxyacids. Indeed polyhydroxyalkanoates are produced by a variety of bacteria both as a protective medium and as a chemical energy storage⁴³⁻⁴⁷.

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2 One of the simplest poly hydroxyalkanoate, i.e. poly lactic acid (PLA), was first used as a material for
3 surgical sutures ^{48,49}, and it was demonstrated that it was biodegradable ⁵⁰. It also gained popularity by
4 the fact that it could be polymerized as nanospheres for drug encapsulation and controlled release, most
5 often as a copolymer of lactic and glycolic acid ^{51–53}. However, it was also shown that pure PLA is less
6 toxic than the copolymer ^{54,55}, so that PLA is a preferred choice as a model for a safe, biobased and
7 biodegradable plastic available in the nanoplastic format.

8
9 Recent work showed that degradation of PLA produced fragmentation and nanoparticles ⁵⁶, even in
10 domestic use ⁵⁷, which further increases the probability that living organisms may encounter PLA
11 nanoparticles, and especially if the use of this plastic increases. We thus decided to investigate the
12 responses of macrophages to PLA nanoparticles, using a combination of proteomic and targeted
13 approaches, as previously done with polystyrene nanoparticles ⁴¹. In this approach, proteomics is used
14 to gain a wide appraisal of the cell responses, highlighting hypotheses that are then verified by targeted
15 experiments.

16 2. Materials and Methods

17 2.1. Plastic particles

18 Poly lactic acid (PLA) particles (red Poly-lactic acid 150 nm, fluorescent red labelled, catalog number
19 #RFiP-600-150 batch number #230612-ip) were purchased from Adjuvatis (Lyon, France) and
20 provided as sterile 3% (w/v) suspensions. Deep red fluorescent labelled polystyrene (PS) particles
21 (SkyBlue particles, 100-300 nm, catalog number #FP-0270-2, batch number #AL01) were purchased
22 from Spherotech and provided as non-sterile 0.25% (w/v) suspensions. The particles were used for both
23 the proteomic experiments and the validation experiments. The polystyrene suspensions were
24 pasteurized overnight at 80°C before use in cell culture. The PLA suspension was provided sterile by
25 the supplier. The particles were excited at 560 nm and their emission read at 613/18 nm for the PLA
26 particles and at 695/40 nm for the PS particles. It shall be noted that the fluorophore is trapped withing
27 the polymer matrix for both PLA⁵⁸ and PS⁵⁹ nanoparticles.

28 The particles were diluted at 10 µg/mL in PBS 0.001X and characterized by DLS using a Litesizer 500
29 (Anton Paar, USA) and Omega Cuvette (225288, Anton Paar, USA). An average value was obtained
30 from repeated measurements for each sample (n = 3) and analyzed with the instrument-associated
31 Kaliaope software.

2.2. Cell culture

The J774A.1 cell line (mouse macrophages) was purchased from European cell culture collection (Salisbury, UK). Cells were routinely propagated in Dulbecco's Modified Eagle's Medium (DMEM) supplemented with 10% fetal bovine serum (FBS) in non-adherent flasks (Cellstar flasks for suspension culture, Greiner Bio One, Les Ulis, France). For routine culture, the cells were seeded at 200,000 cells/ml and passaged two days later, with a cell density ranging from 800,000 to 1,000,000 cells/ml. For exposure to plastic particles and to limit the effects of cell growth, cells were seeded at 500,000 cells/ml in 6 or 12 wells plates, let settle and recover for 24 hours, and then exposed to the particles at 50 µg/ml for 24 hours before harvesting for the experiments. Proteomic experiments were carried out in 6 well plates, and all the other experiments in 12 well plates. The medium volume was adjusted to keep the same height across all cell culture formats. Cells were used at passage numbers from 5 to 15 post-reception from the repository. Cell viability was measured by the propidium iodide method⁶⁰, or with the SytoxGreen probe (Thermofisher S7020) using the protocol provided by the supplier.

For assessing the persistence of the particles in the cells, we followed the strategy published previously⁴². The cells were seeded in adherent 12-well plates at 400,000 cells/ml in DMEM supplemented with 1% horse serum, in order to limit their proliferation⁶¹. After adaptation to the medium for 72 hours (D0), the medium was changed and the cells were exposed to the plastic particles for 24 hours at a concentration of 50 or 100 µg/ml. The particle-containing medium was removed (D1). The remaining cells were cultivated in DMEM supplemented with 1% horse serum for an additional 3 days without splitting, and with a culture medium change every 2 days. Wells were harvested every day starting at D1 during this process and analyzed by flow cytometry to measure the cell-associated fluorescence. In order to normalize for the cell number, a cell numeration was performed on the harvested cells in each well. This process allowed to determine the remaining fraction of particles over time after the initial loading (D1).

2.3. Confocal microscopy

J774.A1 cells were seeded onto glass coverslips at a density of 200,000 cells/mL in DMEM supplemented with 1% horse serum and 1% streptomycin-penicillin. The cells were then incubated overnight at 37°C with 5% CO₂.

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2 Subsequently, PLA beads were added to the cells at a concentration of 50 μ g/ml, and further incubation
3 was carried out for 24 hours. The medium was then changed for particle-free medium and the cells
4 were let to recover for 24 hours and 48 hours at 37°C.

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7 After each incubation period (without exposure, immediately after exposure and after the 24 and 48
8 hours recovery periods) the cells were fixed with 4% paraformaldehyde for 30 minutes at room
9 temperature and permeabilized with 0.1% Triton X-100 for 20 minutes at room temperature. Following
10 permeabilization, the cells were incubated with Phalloidin Atto 488 (1/500 dilution) and DAPI (1/1000
11 dilution) for 20 minutes and 5 minutes, respectively, at room temperature. Between each step, the cells
12 were washed three times with PBS (1X).

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14
15 Confocal microscopy experiments were performed using a Zeiss LMS880 confocal microscope (Carl
16 Zeiss, Jena, Germany) equipped with a 20 \times objective. Laser tracks were set as following:

- 17 •Dapi: Excitation at 405nm, Emission at [410 – 467]nm
- 18 •Phalloidin Atto 488: Excitation at 488nm, Emission at [494 - 550]nm
- 19 •PLA beads: Excitation at 561nm, Emission at [596 – 632]nm

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22 Laser settings were optimized using the 24-hour-exposed cells labeled with Phalloidin and DAPI to
23 maximize fluorescence intensity before imaging each condition. Confocal laser scanning microscopy
24 was employed to visualize the actin structure, nucleus, and PLA beads.

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27 The acquired images were processed using ImageJ software. Specifically, adjustments were made to
28 enhance the quality and uniformity of the images. Green fluorescence intensity in the B1 merged image
29 was increased to ensure consistency across all panels. Similarly, brightness levels in the red channel
30 were uniformly adjusted for all samples, with or without PLA beads.

31 2.4. Proteomics

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34 Proteomics was carried out essentially as described previously ⁴¹. However, the experimental details
35 are given here for the sake of consistency.

36 2.4.1. Sample preparation

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2 After exposure to the plastic particles, the cells were harvested by flushing the 6 well plates. They were
3 collected by centrifugation (200g, 5 minutes) and rinsed twice in PBS. The cell pellets were lysed in
4 100 μ l of extraction buffer (4M urea, 2.5% cetyltrimethylammonium chloride, 100mM sodium
5 phosphate buffer pH 3, 150 μ M methylene blue). The extraction was let to proceed at room temperature
6 for 1 hour, after which the lysate was centrifuged (15,000g, 30 minutes) to pellet the nucleic acids. The
7 supernatants were then stored at -20°C until use.
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2.4.2. Shotgun proteomics

For the shotgun proteomic analysis, the samples were included in polyacrylamide plugs according to Muller et al.⁶² with some modifications to downscale the process⁶³. To this purpose, the photopolymerization system using methylene blue, toluene sulfinate and diphenyliodonium chloride was used⁶⁴.

As mentioned above, the methylene blue was included in the cell lysis buffer. The other initiator solutions consisted in a 1 M solution of sodium toluene sulfinate in water and in a saturated water solution of diphenyliodonium chloride. The ready-to-use polyacrylamide solution consisted of 1.2 ml of a commercial 40% acrylamide/bis solution (37.5/1) to which 100 μ l of diphenyliodonium chloride solution, 100 μ l of sodium toluene sulfinate solution and 100 μ l of water were added.

To the protein samples (15 μ l), 5 μ l of acrylamide solution were added and mixed by pipetting in a 500 μ l conical polypropylene microtube. 100 μ l of water-saturated butanol were then layered on top of the samples, and polymerization was carried out under a 1500 lumen 2700K LED lamp for 2 hours, during which the initially blue gel solution discolored. At the end of the polymerization period, the butanol was removed, and the gel plugs were fixed for 2x1 hr with 200 μ l of 30% ethanol 2 % phosphoric acid, followed by a 30 minutes wash in 30% ethanol. The fixed gel plugs were then stored at -20°C until use.

Gel plug processing, digestion, peptide extraction and nanoLC-MS/MS was performed as previously described, without the robotic protein handling system and using a Q-Exactive Plus mass spectrometer (Thermo Fisher Scientific, Bremen, Germany). Further details are available in **Methods S1**.

For protein identification, the MS/MS data were interpreted using a local Mascot server with MASCOT

2.6.2 algorithm (Matrix Science, London, UK) against an in-house database containing all *Mus musculus* and *Rattus norvegicus* entries from UniProtKB/SwissProt (version 2019_10, 25,156 sequences) and the corresponding 25,156 reverse entries. Spectra were searched with a mass tolerance of 10 ppm for MS and 0.05 Da for MS/MS data, allowing a maximum of one trypsin missed cleavage. Trypsin was specified as enzyme. Acetylation of protein N-termini, carbamidomethylation of cysteine residues and oxidation of methionine residues were specified as variable modifications. Identification results were imported into Proline software version 2.2 (<http://profiproteomics.fr/proline>) for validation. Peptide Spectrum Matches (PSM) with pretty rank equal to one and a length greater than 7 amino acids were retained. False Discovery Rate was then optimized to be below 1% at PSM level using Mascot Adjusted E-value and below 1% at Protein Level using Mascot Standard score.

Mass spectrometry data are available via ProteomeXchange with the identifier PXD048664.

2.4.3. Label Free Quantification

Peptides abundances were extracted thanks to Proline software version 2.2 (<http://profiproteomics.fr/proline>) using a m/z tolerance of 10 ppm. Alignment of the LC-MS runs was performed using Loess smoothing. Cross Assignment was performed within groups only. Protein Abundances were computed by sum of peptides abundances (normalized using the median).

2.4.4. Data analysis

For the global analysis of the protein abundances data, missing data were imputed with a low, non-null value. The complete abundance dataset was then analyzed by the PAST software ⁶⁵.

Proteins were considered as significantly different if their p value in the Mann-Whitney U-test against control values was inferior to 0.05. No quantitative change threshold value was applied. The selected proteins were then submitted to pathway analysis using the DAVID tool ⁶⁶, with a cutoff value set at a FDR of 0.1.

2.5. Mitochondrial transmembrane potential assay

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2 The mitochondrial transmembrane potential assay was performed essentially as described previously⁴².
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4 Rhodamine 123 (Rh123) was added to the cultures at an 80 nM final concentration (to avoid quenching
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6⁶⁷), and the cultures were further incubated at 37°C for 30 minutes. At the end of this period, the cells
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8 were collected, washed in cold PBS containing 0.1% glucose, resuspended in PBS glucose and
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10 analyzed for the green fluorescence (excitation 488 nm emission 525nm) on a Melody flow cytometer.
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12 As a positive control, butanedione monoxime (BDM) was added at a 30mM final concentration
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14 together with the Rh123⁶⁸. As a negative control, carbonyl cyanide 4-
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16 (trifluoromethoxy)phenylhydrazone (FCCP) was added at 5µM final concentration together with the
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18 Rh123⁶⁸.

2.6. Phagocytosis assay

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20 For this assay, the cells were first exposed to the red fluorescent particles. After 24 hours of exposure,
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22 the cells were the exposed to 0.5 µm latex beads (carboxylated surface, yellow green-labelled, from
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24 Polysciences excitation 488 nm emission 527/32 nm) for 3 hours. After this second exposure, the cells
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26 were collected, rinsed twice with PBS, and analyzed for the two fluorescences (green and red) on a
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28 Melody flow cytometer.
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3 2.7. Lysosomal assay

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32 For the lysosomal function assay, the Lysosensor method was used, as described previously⁴². After
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34 exposure to plastic beads, the medium was removed, the cell layer was rinsed with complete culture
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36 medium and incubated with 1µM Lysosensor Green (Molecular Probes) diluted in warm (37°C)
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38 complete culture medium for 1 hour at 37°C. At the end of this period, the cells were collected, washed
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40 in cold PBS containing 0.1% glucose, resuspended in PBS glucose and analyzed for the green
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42 fluorescence (excitation 488 nm emission 540nm) on a Melody flow cytometer.
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53 2.8. Cytokine release assays

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55 Cells were first exposed to nanoplastics (50µg/ml) for 24 hours. At the end of this exposure period, the
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57 culture medium was removed, the cell layer was rinsed with culture medium and fresh medium was
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2 added to the wells. In half of the wells LPS (1 ng/ml) was added. After another 24 hours, the medium
3 was collected and analyzed for proinflammatory cytokines. Tumor necrosis factor (catalog number
4 558299, BD Biosciences, Le Pont de Claix, France) and interleukin 6 (IL-6) (catalog number 558301,
5 BD Biosciences, Le Pont de Claix) levels were measured using the Cytometric Bead Array Mouse
6 Inflammation Kit (catalog number 558266, BD Biosciences, Le Pont de Claix), and analyzed with
7 FCAP Array software (3.0, BD Biosciences) according to the manufacturer's instructions.
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2.9. Assay for oxidative stress

For the oxidative stress assay, a protocol based on the oxidation of dihydrorhodamine 123 (DHR123) was used, essentially as described previously⁴². After exposure to plastic beads, the cells were treated in PBS containing 500 ng/ml DHR123 for 20 minutes at 37°C. The cells were then harvested, washed in cold PBS containing 0.1% glucose, resuspended in PBS glucose and analyzed for the green fluorescence (same parameters as rhodamine 123) on a Melody flow cytometer. Menadione (applied on the cells for 2 hours prior to treatment with DHR123) was used as a positive control in a concentration range of 25-50µM.

2.10. Cross-toxicity assays

For these assays, the cells were plated at 400,000 cells/ml in a 12-well non adherent plate. After 24 hours, 50µg/ml of beads were added, followed 6 hours later by a variable concentration of the toxicant to be tested, pre-dissolved in ethanol so that the final ethanol concentration in the culture well did not exceed 1% in volume. Toxicants tested included menadione and 9,10-phenanthrene quinone.

3. Results

3.1. Plastics beads characterization

Plastic beads were characterized by DLS and electrophoretic mobility in order to determine their average hydrodynamic size and zeta potential. The results are presented in Table 1

Table 1: characterization of the beads parameters

Beads	Average hydrodynamic diameter	Polydispersity index	Zeta potential
PLA	169±4 nm	18.5±3.7%	-39.6±0.7 mV
PS	477±16 nm	22.2±3.1%	-43.9±5.7 mV

The results are expressed as mean±standard deviation (N= 3)

The PS beads proves larger in size through these measurements than the nominal diameter given by the supplier (0.1-0.3 μm , mean 0.26 μm). This larger size however did not prevent a bead/cell number ratio close to 1000 beads/cell (at 50 $\mu\text{g}/\text{ml}$) for the largest beads and higher than 10,000 beads/cell for the smaller ones.

3.2. Viability of the plastics-treated cells

First, the toxic effects of the PLA beads on J774A.1 cells was determined after a 24 hours exposure. The results, shown in **Fig.1a**, demonstrated a very low toxicity of the PLA beads, with a LD50 that reached 400 $\mu\text{g}/\text{ml}$. As we planned to use polystyrene beads as positive controls for non-degradable plastics in subsequent experiments, we decided to use a dose that was half of the LD20 (100 $\mu\text{g}/\text{ml}$) for the PS beads, i.e. 50 $\mu\text{g}/\text{ml}$. This dose showed very low toxicity for the PLA beads, but was however the first concentration for which the viability was slightly but statistically significantly lower than the one of unexposed, control cells. As the particles were labelled, we could follow their internalization and their degradation. As shown in **Fig.1b**, the amount of internalized PLA particles increased with increasing concentration in the medium. However, the amount of intracellular fluorescence rapidly decreased over time, suggesting a rather fast degradation of the particles. This was further confirmed by confocal microscopy, as shown in Fig. 2. Disappearance of the particles and appearance of a fuzzy intracytoplasmic fluorescence was observed, indicating degradation of the particles. These results are in sharp contrast with those obtained for PS nanoparticles⁴², which showed no degradation in the cells for more than a week.

3.3. Analysis of the proteomic results

The shotgun proteomic analysis was able to detect and quantify 2869 proteins (**Table S1**). A first global analysis of the complete protein list by principle coordinates analysis showed that the two groups (control and particle-treated) appeared separated on the diagram (**Fig. 3**), indicating significant changes in the proteome, even if the chosen concentration was quite remote of the toxicity threshold. The fact that the two proteomes were significantly different, on a statistical point of view,

1
2 was further substantiated by an analysis of similarities ⁶⁹, which lead to a Bonferroni-corrected post-
3 hoc p value of 0.009.

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6 Proteins modulated by the internalization of PLA particles were selected on the basis of a Mann-
7 Whitney U test ≤ 2 in the comparison of plastic-treated cells compared to unexposed controls. This
8 resulted in the selection of 346 modulated proteins (**Table S2**). In order to gain further insight into the
9 significance of the observed changes, this list of modulated proteins was used to perform pathways
10 analyses by the DAVID software, and the results are shown in **Table S3**. Some of the pathways
11 highlighted by this analysis indicated a global stress response (e.g. translation, nucleotide binding,
12 carbon metabolism, endoplasmic reticulum), which is expected for any cellular stress, while other
13 pathways appeared more specific of cellular internalization of particles (e.g. mitochondria, lysosomes).

14
15 As these pathway analysis softwares proceed by aggregation of proteins sharing the same annotations,
16 they require a minimum number of proteins to build a cluster. Thus, the list of modulated proteins was
17 manually combed in addition to this global analysis, in order to retrieve isolated protein changes,
18 which are discarded by pathway analyses but may be of interest in the frame of macrophage
19 physiology. This comprehensive analysis of the proteomic results led us to perform validation
20 experiments on several functions.

21 3.4. Mitochondria and endoplasmic reticulum

22
23 The list of the mitochondrial proteins modulated by cells treatment with PLA nanoparticles is presented
24 in **Table S4**, and included 41 proteins, of which 32 were increased in abundance in response to PLA
25 nanoparticles. Among them, 11 subunits of the respiratory chain were highlighted (8 increases, 3
26 decreases). Although the changes observed in protein abundances were usually of low magnitude for
27 each protein, the changes occurring on several subunits of the same complexes may suggest a
28 functional alteration. To test this hypothesis, we tested the mitochondrial transmembrane potential as a
29 proxy of mitochondrial function. The results, shown in **Fig. 4a**, indicated that PLA particles did not
30 induce any change in the mitochondrial transmembrane potential, while PS particles induced a small
31 but not statistically significant increase in the mitochondrial transmembrane potential, as described
32 previously ^{41,42}.

33
34 Regarding endoplasmic reticulum, 30 proteins were modulated by treatment with PLA nanoparticles, as
35 shown in **Table S5**. These included proteins involved in the endoplasmic reticulum stress response, as
36 the VapB protein (Uniprot Q9QY76) or the tyrosine phosphatase Ptpn1 (Uniprot P35821). This
37 prompted us to test whether endoplasmic reticulum stress response may play a role in PLA
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1 nanoparticles toxicity or not. To this purpose we used salubrinal, an inhibitor of the endoplasmic
2 reticulum stress response⁷⁰, which has been shown to counteract the lethal effects of this stress on cells
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71. The results, shown in **Fig.4b**, indicated no effect of salubrinal on cell survival up to treatment with 200µg/ml PLA particles for 24 hours.

3.5. Lysosomes, phagocytosis

For this important macrophage function, we gathered the proteins selected under the “lysosome” and “actin cytoskeleton” pathways, to which we added some proteins picked manually on the basis of their annotations in the Uniprot database, such as the two “engulfment and cell motility” proteins (accession numbers Q8BPU7 and Q8BHL5, respectively). This led to a set of 36 proteins, which is presented in **Table S6**. Two of the V-type proton ATPase subunits were present in this list (among the 11 V-type proton ATPase subunits that were detected in the whole proteomic analysis) so that we wanted to investigate by the lysosensor probe the lysosome acidification in response to treatment with PLA nanoplastics, PS nanoparticles being used as a control for non-degradable plastics. The results, shown in **Fig. 4c**, indicated a small but statistically significant ($p < 0.001$) increase in the lysosensor signal, indicating an increase in the number of functional lysosomes and/or an increase proton pumping in the existing lysosomes.

When the phagocytic capacity of plastics-treated cells was probed, as described in **Fig. 4d**, a small (-19%) but significant ($p < 0.001$) decrease was observed for PLA-treated cells, while PS-treated cells were not altered compared to control cells.

3.6. Immunity-related proteins, inflammation

As this function was not detected through the classical pathway analysis, opposite to what was observed for polystyrene particles⁴¹, we hand-picked on the basis of Uniprot annotations the proteins that could be associated with this pathway in the shortlist of proteins which expression was significantly modulated in response to exposure to PLA particles. This led to a set of 22 proteins, which is presented in **Table S7**. This relatively low number of proteins may explain why the “innate immunity” annotation was not selected in the classical pathway analysis. Among the selected proteins, several were annotated as regulators of immune response and cytokine secretion (P09581, P10810, P11680, Q60875, Q6P549, Q9D8Y7). This prompted us to investigate the pro-inflammatory response of PLA-exposed macrophages, either in response to PLA alone or to a succession of exposures to PLA and to LPS, by measuring the secretion of TNF-alpha and IL-6. The results, shown in **Fig. 5**, indicated

1
2 that the LPS-induced IL-6 secretion was decreased close to twofold upon prior exposure to PLA (and
3 PS) particles (**Fig. 5a**) No IL-6 secretion could be measured in response to plastic alone. In the case of
4 TNF, the opposite response was observed. A statistically significant increase could be detected for the
5 basal levels after exposure to plastics alone (**Fig. 5b**) and this increase was more pronounced in
6 response to PLA than to PS. After successive exposure to plastics and LPS, the same trend was
7 observed again, of course at much higher levels (**Fig. 5c**).

13 3.7. Redox homeostasis related proteins, oxidative stress

14
15 As an increase in oxidative stress was observed in response to polystyrene particles ⁴², we hand-picked
16 on the basis of Uniprot annotations proteins that could be associated with redox metabolism at large
17 (excluding central metabolism) in the shortlist of proteins which expression was significantly
18 modulated in response to exposure to PLA particles. This led to a set of 13 proteins, which is presented
19 in **Table S8**. This relatively low number of proteins may explain why the “redox homeostasis”
20 annotation was not selected in the classical pathway analysis.

21
22 However, we decided to investigate the level of cellular oxidative stress in response to exposure to
23 PLA (and PS nanoparticles). The results, shown in **Fig. 6**, indicated no increase in oxidative stress in
24 response to PLA nanoparticles, while a small but significant increase was observed in response to PS
25 nanoparticles, as previously described ⁴².

26
27 Furthermore, the proteins highlighted in **Table S8** included a few proteins included in quinone
28 metabolism, such as aflatoxin B1 aldehyde reductase member 2, and quinone oxidoreductase. This
29 prompted us to seek for cross toxicity between quinones and plastic beads. To this purpose, we
30 performed comparative cell viability on cells first treated (or not) with nanoplastics, then with variable
31 concentrations of polycyclic quinones. The results, shown on **Fig. 7**, showed an increased toxicity of
32 phenanthrene quinone on plastic-treated cells, while such increased toxicity was not observed with
33 menadione.

39 4. Discussion

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41 The wide release of plastics in the environment is more than suspected to have major deleterious
42 consequences on life. These can be due to mechanical effects of macroplastics ³, but microplastics
43 arising from the fragmentation of macroplastics are also known to have deleterious effects, e.g. on
44 oysters ¹⁹ on seabirds ⁴. Microplastics fragment even further into nanoplastics, which have been
45 associated with liver fibrosis in rodents ⁷², and more recently with increased cardiovascular risk in
46 humans ⁷³. These deleterious effects are the indirect consequences of the very low biodegradability of
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2 conventional, oil-derived plastics, which makes the particles resulting from their fragmentation long
3 lasting in the environment and within living organisms. Even though this low biodegradability will
4 make this concern a problem for many years to come, one way not to aggravate it would be to switch
5 from poorly-degradable plastics to biodegradable ones, so that the persistent mechanical damage as
6 well as the bioaccumulation issues will be alleviated. Among the few chemical solutions available to
7 produce biodegradable plastics, polymers of organic alkyl hydroxyesters are a promising area of
8 research. Among this class of chemicals, poly-lactic acid (PLA) represents an interesting choice, as it is
9 both biobased and biodegradable ⁵⁰. However, it is also known to fragment easily into nanoparticles
10 ^{56,57}. As the global biodegradability in the environment does not necessarily translate into the absence
11 of accumulation in all cell types, it is necessary to obtain data on the internalization, degradation and
12 effects of PLA nanoparticles on cell types of interest. This is required to figure out if more or less
13 transient effects can be expected on certain cell types. Within the cell types of interest in multicellular
14 organisms, macrophages are important to test, as they are present from invertebrates to vertebrates and
15 play a key role in inflammation, which can be deleterious on a long time frame ⁷⁴.

16
17 First of all, we investigated the acute toxicity of PLA nanoparticles toward macrophages. The LD20
18 proved to be rather high (150µg/ml) i.e. around twice the value observed for polystyrene ⁷⁵. More
19 importantly, PLA particles were actively degraded in macrophages. Half of the fluorescence had
20 disappeared from cells exposed to PLA nanoparticles 48 hours after the end of exposure. As this
21 disappearance of fluorescence, as measured by flow cytometry, implies degradation of the PLA shell of
22 the nanoparticles, then release of the fluorophore within the cell then excretion of the free fluorophore,
23 the real degradation of the PLA shell is probably even faster than this figure. This shows a very active
24 degradation of the PLA particles within macrophages. This shall not come as a surprise, as PLA has
25 been shown to be very sensitive to proteases (which often also have an esterase activity) ⁷⁶, proteases
26 being quite abundant in lysosomes. This easy degradation of PLA by mammalian enzymes may be
27 attributed to the fact that lactic acid is an alpha hydroxy acid, quite close in its structure to the alpha
28 amino acids that make proteins.

29
30 We then investigated the effects of the PLA particles on macrophages immediately after exposure using
31 a proteomic screen. Out of 2869 proteins quantified, the proteomic screen highlighted 346 proteins
32 modulated upon treatment with PLA, covering several pathways.

33
34 Interestingly, we did not see any modulation of metabolic proteins, although lactate has been shown to
35 induce major metabolic reprogramming ⁷⁷. In the same trend, we did not detect a switch to an
36 inflammatory phenotype, as described when macrophages are confronted with a high lactate
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1 concentration ⁷⁸. There can be several explanations to this fact. The first one could just be that
2 immediately after exposure, there is not a sufficient intracellular release of lactate to induce these
3 phenomena. Moreover, as particles are present in the lysosomes and degraded there, a control of the
4 speed of lactate release from the lysosomes to the cytosol may also occur. Finally, regarding the central
5 metabolism, it should be kept in mind that it is heavily regulated by post-translational modifications ^{79–}
6 ⁸¹, which do not appear easily in a global shotgun proteomic screen as in this work.

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13 Nevertheless we detected changes in the abundances of many proteins related to organelles such as
14 mitochondria, lysosomes and endoplasmic reticulum (ER). Regarding ER proteins, on the 30 that
15 showed a significant changes in their abundance, 24 were increased in response to PLA treatment and
16 only 6 decreased. As ER stress may be a factor explaining toxicity, we tested this hypothesis with the
17 drug salubrinal, which is an inhibitor of ER stress ⁷⁰, and has been used to demonstrate the implication
18 of ER stress in toxicity ⁷¹. As we did not detect any effect of salubrinal in the cell viability of cells
19 treated with up to 200µg/ml PLA, we can conclude that this pathway is not a major determinant of PLA
20 toxicity in macrophages. In the same trend, we could not detect any sign of mitochondrial
21 depolarization after treatment with PLA nanoparticles, although the abundance of more than 40
22 mitochondrial proteins is changed, with 32 increases and 9 decreases. It must therefore be concluded
23 that the observed changes are homeostatic, and reflect how the cells adapt to the presence of PLA
24 particles.

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Regarding lysosomes, 36 proteins were modulated in their abundances, with 23 increases and 13
decreases. Many of the observed decreases are indeed lysosome-associated cytoskeletal proteins, such
as tropomyosins, ARP complex subunits, gelsolin and myosin 9. Among the increases were detected
two subunits of the lysosomal proton pump. This prompted us to test some lysosomal functions with
the lysosensor probe. As neutral red, lysosensor probes are pumped in acidic organelles ⁸². The
lysosensor signal thus reflects both the activity of the proton pump and the integrity of the lysosomes,
as damaged lysosomes cannot maintain a proton gradient and thus exhibit a reduced lysosensor signal
⁸³. We detected indeed a small but significant increase in the intensity of the lysosensor signal for PLA-
treated cells, but also for PS-treated cells, suggesting that this effect is not specific for a given
nanoparticle but more a generic effect linked to particle internalization per se.

We then tested more specific functions of the macrophage, starting with phagocytosis. To this purpose,
we first exposed cells to the nanoplastics of interest, and then to a fluorescently-labelled test bead for a
relatively short time, in order to determine whether plastic-treated cells are still active for phagocytosis
or not. In this case we observed a specific decrease in the phagocytic activity of PLA-treated beads,



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2 which did not occur when the cells were pre-treated with PS beads, as described earlier ⁴². A possible
3 explanation for this phenomenon is the induction of the phosphatase SHIP2 (Q6P549) in response to
4 PLA treatment. This phosphatase is known to decrease Fc-gamma-R-mediated phagocytosis in
5 macrophages ⁸⁴, and the beads that we use to probe phagocytosis are incubated with serum, a process
6 known to lead to opsonization ⁸⁵.
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12 Regarding cytokines production, which is another important specialized function of macrophages, the
13 proteomic analysis provided interesting, and somewhat contradictory, hypotheses. On the one hand
14 some proteins increasing the inflammatory response were increased in their abundance in response to
15 treatment with PLA. Examples are the M-CSF receptor (P09581) which mediates the pro-inflammatory
16 effects of M-CSF, the ARHGEF2 protein (Q60875), which plays a role in the signal transduction in
17 response to peptidoglycans ⁸⁶, and Rab-10, which plays a role in the recycling of TLR4 at the
18 membrane and therefore in the response to LPS ⁸⁷.
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40 On the other hand, some proteins acting directly or indirectly as negative regulators of the
41 inflammatory response were also increased in their abundance in response to treatment with PLA.
42 Examples are the SHIP2 protein (Q6P549), which down regulates the signal transduction pathway
43 downstream of the M-CSF receptor, the Tnfaip812 protein (Q9D8Y7), which acts as a negative
44 regulator downstream of some TLRs ⁸⁸, and the TMEM43 protein (Q9DBS1), which acts as a negative
45 regulator of the cGAS/STING interferon response pathway ⁸⁹. Interestingly, the cGAS/STING pathway
46 has been implicated in the PS particles-induced liver inflammation and fibrosis ⁷².
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61 In view of this complex landscape of responses, we decided to perform simple validation experiments
62 in measuring the secretion of the pro-inflammatory cytokines IL-6 and TNF-alpha in response to
63 exposure to PLA or PS nanoparticles, with and without terminal activation with LPS. The results were
64 not consistent between the two cytokines, indicating that their regulations are not strictly similar. For
65 IL-6, we did not detect any secretion without LPS-stimulation, indicating that neither PLA or PS alone
66 induced its secretion. However, when cells were exposed to plastics and then to LPS, an important
67 reduction in IL-6 secretion was observed. This means in turn that plastic-exposed macrophages are less
68 performing in their antibacterial response, compared to plastic-free macrophages. For TNF alpha, the
69 same trend was observed with or without LPS activation, i.e. an increase in TNF secretion upon
70 exposure to plastic, the effect being more pronounced for PLA than for PS. This effect is consistent

1
2 with the one already described on RAW264.7⁴⁰, i.e. another mouse macrophage cell line. These
3 diverging results between IL-6 and TNF-alpha showed how complex the inflammatory response to
4 plastics can be. The important point to be underlined here is that although biodegradable, PLA induces
5 at least a transient pro-inflammatory response at the TNF level.
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9 Finally, we investigated the redox metabolism of plastic-treated macrophages. We started with the ROS
10 production, as overproduction of hydrogen peroxide, which is diffusible, may contribute to damage the
11 adjacent tissues⁹⁰. In macrophages, hydrogen peroxide is produced by the combined action of
12 NADPH oxidase, which produces superoxide, and superoxide dismutase, which transforms superoxide
13 into hydrogen peroxide. Here again, the proteomic screen revealed inconsistent results. The levels of
14 the NADPH oxidase subunits Ncf1 and Ncf4 were constant, while the level of cytochrome b245
15 (Q61093), the protein that transfers single electrons to oxygen atoms to produce superoxide increased,
16 and the level of superoxide dismutase decreased. Thus, no clear trend emerged from the proteomic
17 screen. However, targeted validation experiments showed that PLA nanoparticles did not induce any
18 overproduction of ROS in cells, while PS nanoparticles induced a moderate but significant
19 overproduction, as analyzed in more detail previously⁴². Thus in this regard, PLA particles may induce
20 less oxidative damage to the surrounding tissue than PS particles.
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24 In the proteins grouped under the “redox homeostasis” header, two proteins drew our attention. First
25 the AKR7A2 oxidoreductase (P45376) and then crystallin zeta, also known as the QOR quinone
26 oxidoreductase (P47199). The latter protein is different from the more well-known NQO1 quinone
27 reductase (Q64669). Opposite to NQO1, which catalytic mechanism ensures a two-electron reduction
28 process⁹¹, both QOR (which is known to produce semiquinones⁹²) and AKR7A2 are able to induce
29 redox cycling with their preferential substrate phenanthrene 9-10 quinone^{92,93}. This redox cycling is
30 known to induce in turn adverse effects such as cell death and malignant transformation⁹⁴. We
31 therefore decided to investigate whether treatment of macrophages with PLA could induce a cross-
32 toxicity with phenanthrene 9-10 quinone. As a control, we used menadione, which is not a substrate for
33 at least crystallin zeta⁹². Although the toxicity shown by phenanthrene 9-10 quinone was characterized
34 by high standard deviations, which decreased the statistical significance of the results, we could
35 observe a differential toxicity of phenanthrene 9-10 quinone in cells treated with plastics (either PLA or
36 PS) and control cells. This differential toxicity was not observed with menadione.
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56 **5. Conclusions**

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2 The first important conclusion of this work on PLA particles is that they are biodegradable within
3 macrophages. However, this does not mean that they do not show immediate consequences on the
4 physiology of macrophages, as shown by the proteome alteration that they induce. Comparison with PS
5 nanoparticles shows that there are both common consequences and specific ones. Among the specific
6 consequences are the inhibition of phagocytosis by the PLA nanoparticles and the absence of detectable
7 oxidative stress. Among the shared ones are the impact on cytokine production, the induction of
8 lysosomal acidification and the cross toxicity with phenanthrene quinone. Phenanthrene quinone is
9 found in combustion products and is also one of the compounds produced by the metabolic activation
10 of phenanthrene⁹⁵, a combustion product itself. This suggests that some nanoplastics may induce cross-
11 effects with chemicals present in, e.g. , diesel exhaust particles or cigarette smoke, which may be
12 another mechanism underlying noxious effects of nanoplastics.
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Author Contributions

VCF: methodology, investigation. MV: methodology, investigation, writing (review and editing) . HD:
methodology, investigation, writing (review and editing) . SC: resources, funding acquisition, writing
(review and editing). ED: formal analysis, writing (review and editing). TR: conceptualization, formal
analysis, funding acquisition, writing (original draft).

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Conflict of Interest

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1
2 There are no conflicts of interest to declare
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6 **Legends to Figures.**

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9 **Figure 1: Viability and intracellular fluorescence of cells treated with PLA particles**

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11 In panel 1A, cells were treated with PLA particles for 24 hours, and their viability measured by a flow
12 cytometry fluorophore exclusion assay (sytox green). Results are displayed as mean± standard
13 deviation (N=4)

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15 In panel 1B, cells were treated with 50µg/ml (green curve) or 100µg/ml (blue curve) PLA particles for
16 24 hours, and the cell-associated fluorescence was measured immediately after exposure (D1), or up to
17 3 days post-exposure (D2 to D4). Results are displayed as mean± standard deviation (N=4)

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19 **Figure 2: Confocal Microscopy Imaging of J774.A1 Cells**

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21
22 Confocal microscopy images of J774.A1 cells stained with Phalloidin Atto 488 (green) and DAPI
23 (blue) after exposure to PLA beads for 24, 48, and 72 hours. (A1,C1,E1) Show the overlay of actin
24 cytoskeleton (green) and cell nuclei (blue) captured at each time point. (B1,D1,F1) present the overlay
25 of Phalloidin, DAPI staining and PLA beads. The second column of images (A2,B2,C2,D2,E2,F2)
26 represents the red channel, allowing the observation of the PLA beads (white).

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Figure 3: Global analysis of the proteomic data

The complete proteomic data table (2869 proteins) was analysed by Principal Coordinates Analysis, using the PAST software. The mathematical distance used for the calculations was the Gower distance. The results are represented as the X-Y diagram of the first two axes of the Principal Coordinates Analysis, representing 54% of the total variance. Eigenvalue scale.

Figure 4: Mitochondria, lysosomes and phagocytosis experiments

For panels A, C and D, the cells were treated for 24 hours with 50µg/ml PLA or PS particles, then various physiological parameters were tested.

Panel A: mitochondrial transmembrane potential (rhodamine 123 method). All cells were positive for rhodamine 123 internalization in mitochondria, and the mean fluorescence is the displayed parameter. Results are displayed as mean± standard deviation (N=6)

Panel B: test of the endoplasmic stress response in PLA toxicity. Cells were pre-treated with 4µM salubrinal for 4 hours, and various concentrations of PLA beads were then added for a further 18 hours in culture. At the end of the experiment, the cell viability was measured. Results are displayed as survival curves, with the standard deviations at each tested point (N=4). Blue curve: cells untreated with salubrinal. Green curve: salubrinal-treated cells.

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2 Panel C : lysosomal proton pumping (Lysosensor method). All cells were positive for lysosensor
3 internalization in lysosomes, and the mean fluorescence is the displayed parameter. Results are
4 displayed as mean± standard deviation (N=6). Significance marks: *** p<0.001 (Student t-test method,
5 comparison between control and each treatment)
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7 Panel D: phagocytosis. Cells were first treated for 24 hours with 50µg/ml PLA particles. After removal
8 of the PLA-containing cell culture medium, the cells were treated with green fluorophore labelled
9 carboxylated polystyrene beads for 3 hours. The mean fluorescence, indicating the amount of green
10 beads internalized, is the displayed parameter. Results are displayed as mean± standard deviation
11 (N=6). Significance marks: *** p<0.001 (Student t-test method, comparison between control and each
12 treatment)
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14 Figure 5: Cytokine release

15 The cells were first treated for 24 hours with 50µg/ml PLA or PS particles. The medium was then
16 removed and the cells were treated (or not) with 1ng/ml lipopolysaccharide in complete cell culture
17 medium for 24 hours. The cell medium was then collected for secreted TNF and IL-6 measurements.
18 Results are displayed as mean± standard deviation (N=6). Significance marks: *** p<0.001 (Student t-
19 test method, comparison between control and each treatment)
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21 Panel A: IL-6 release (after stimulation with LPS)

22 Panel B: TNF-alpha release (without stimulation with LPS)

23 Panel C: TNF-alpha release (after stimulation with LPS)

24 Figure 6: Cellular oxidative stress, measured with the dihydrorhodamine 123 (DHR123) indicator

25 The cells were exposed to PLA or PS beads for 24 hours, and finally for 20 minutes to the DHR 123
26 probe. Menadione (25µg/ml for 2 hours) was used as a positive oxidative stress control. Results are
27 displayed as mean± standard deviation (N=6). Significance marks: * p<0.05 ; *** p<0.001 (Student t-
28 test method, comparison between control and each treatment)
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30 Figure 7: Cross toxicities

31 The cells were first exposed to PLA or PS beads (50µg/ml). After 6 hours, the secondary toxicant, i.e.
32 9,10-phenanthrene quinone or menadione, was added at various concentrations and the cells were
33 cultured for an additional 18 hours. After a total of 24 hours, the cell viability was measured.

34 Blue curve: control cells (unexposed to plastics)

35 Green curve: PLA-exposed cells

36 Purple curve: PS-exposed cells

37 Panel A: cross toxicity with 9,10-phenanthrene quinone. The significance of the difference in viability
38 between cells untreated with plastics and plastic-treated cells at 2.25 and 2.5µM 9,10-phenanthrene
39 quinone is p<0.1.

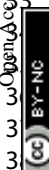
40 Panel B: cross toxicity with menadione. No difference in toxicity between cells untreated with plastics
41 and plastic-treated cells was detected.
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References

- 1 R. Lehner, C. Weder, A. Petri-Fink and B. Rothen-Rutishauser, Emergence of Nanoplastic in the Environment and Possible Impact on Human Health, *Environmental Science & Technology*, 2019, **53**, 1748–1765.
- 2 Y. V. Fan, P. Jiang, R. R. Tan, K. B. Aviso, F. You, X. Zhao, C. T. Lee and J. J. Klemeš, Forecasting plastic waste generation and interventions for environmental hazard mitigation, *Journal of Hazardous Materials*, 2022, **424**, 127330.
- 3 J. G. B. Derraik, The pollution of the marine environment by plastic debris: a review, *Mar Pollut Bull*, 2002, **44**, 842–852.
- 4 H. S. Charlton-Howard, A. L. Bond, J. Rivers-Auty and J. L. Lavers, ‘Plasticosis’: Characterising macro- and microplastic-associated fibrosis in seabird tissues, *J Hazard Mater*, 2023, **450**, 131090.
- 5 K. Enders, R. Lenz, C. A. Stedmon and T. G. Nielsen, Abundance, size and polymer composition of marine microplastics $\geq 10 \mu\text{m}$ in the Atlantic Ocean and their modelled vertical distribution, *Marine Pollution Bulletin*, 2015, **100**, 70–81.
- 6 A. Ter Halle, L. Jeanneau, M. Martignac, E. Jardé, B. Pedrono, L. Brach and J. Gigault, Nanoplastic in the North Atlantic Subtropical Gyre, *Environ. Sci. Technol.*, 2017, **51**, 13689–13697.
- 7 J. A. Brandon, A. Freibott and L. M. Sala, Patterns of suspended and salp-ingested microplastic debris in the North Pacific investigated with epifluorescence microscopy, *Limnology and Oceanography Letters*, 2020, **5**, 46–53.
- 8 M. Kedzierski, M. Palazot, L. Soccalingame, M. Falcou-Préfol, G. Gorsky, F. Galgani, S. Bruzard and M. L. Pedrotti, Chemical composition of microplastics floating on the surface of the Mediterranean Sea, *Marine Pollution Bulletin*, 2022, **174**, 113284.
- 9 E. S. Jones, S. W. Ross, C. M. Robertson and C. M. Young, Distributions of microplastics and larger anthropogenic debris in Norfolk Canyon, Baltimore Canyon, and the adjacent continental slope (Western North Atlantic Margin, U.S.A.), *Mar Pollut Bull*, 2021, **174**, 113047.
- 10 L. Cutroneo, M. Capello, A. Domi, S. Consani, P. Lamare, P. Coyle, V. Bertin, D. Dornic, A. Reboa, I. Geneselli and M. Anghinolfi, Microplastics in the abyss: a first investigation into sediments at 2443-m depth (Toulon, France), *Environ Sci Pollut Res Int*, 2022, **29**, 9375–9385.
- 11 T. Mani, A. Hauk, U. Walter and P. Burkhardt-Holm, Microplastics profile along the Rhine River, *Sci Rep*, 2015, **5**, 17988.
- 12 C. Scherer, A. Weber, F. Stock, S. Vurusic, H. Egerci, C. Kochleus, N. Arendt, C. Foeldi, G. Dierkes, M. Wagner, N. Brennholt and G. Reifferscheid, Comparative assessment of microplastics in water and sediment of a large European river, *Sci Total Environ*, 2020, **738**, 139866.



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- 13 L. Weiss, W. Ludwig, S. Heussner, M. Canals, J.-F. Ghiglione, C. Estournel, M. Constant and P. Kerhervé, The missing ocean plastic sink: Gone with the rivers, *Science*, 2021, **373**, 107–111.
- 14 C. J. Weber, C. Opp, J. A. Prume, M. Koch, T. J. Andersen and P. Chiffard, Deposition and in-situ translocation of microplastics in floodplain soils, *Sci Total Environ*, 2021, 152039.
- 15 H. Golwala, X. Zhang, S. M. Iskander and A. L. Smith, Solid waste: An overlooked source of microplastics to the environment, *Sci Total Environ*, 2021, **769**, 144581.
- 16 A. Chamas, H. Moon, J. Zheng, Y. Qiu, T. Tabassum, J. H. Jang, M. Abu-Omar, S. L. Scott and S. Suh, Degradation Rates of Plastics in the Environment, *ACS Sustainable Chem. Eng.*, 2020, **8**, 3494–3511.
- 17 L. Lei, M. Liu, Y. Song, S. Lu, J. Hu, C. Cao, B. Xie, H. Shi and D. He, Polystyrene (nano)microplastics cause size-dependent neurotoxicity, oxidative damage and other adverse effects in *Caenorhabditis elegans*, *Environmental Science: Nano*, 2018, **5**, 2009–2020.
- 18 X. Jiang, Y. Chang, T. Zhang, Y. Qiao, G. Klobučar and M. Li, Toxicological effects of polystyrene microplastics on earthworm (*Eisenia fetida*), *Environmental Pollution*, 2020, **259**, 113896.
- 19 R. Sussarellu, M. Suquet, Y. Thomas, C. Lambert, C. Fabioux, M. E. J. Pernet, N. Le Goïc, V. Quillien, C. Mingant, Y. Epelboin, C. Corporeau, J. Guyomarch, J. Robbins, I. Paul-Pont, P. Soudant and A. Huvet, Oyster reproduction is affected by exposure to polystyrene microplastics, *Proceedings of the National Academy of Sciences*, 2016, **113**, 2430–2435.
- 20 Z. Liu, P. Yu, M. Cai, D. Wu, M. Zhang, Y. Huang and Y. Zhao, Polystyrene nanoplastic exposure induces immobilization, reproduction, and stress defense in the freshwater cladoceran *Daphnia pulex*, *Chemosphere*, 2019, **215**, 74–81.
- 21 Y. Song, C. Cao, R. Qiu, J. Hu, M. Liu, S. Lu, H. Shi, K. M. Raley-Susman and D. He, Uptake and adverse effects of polyethylene terephthalate microplastics fibers on terrestrial snails (*Achatina fulica*) after soil exposure, *Environmental Pollution*, 2019, **250**, 447–455.
- 22 A. H. D’Costa, Microplastics in decapod crustaceans: Accumulation, toxicity and impacts, a review, *Science of The Total Environment*, 2022, **832**, 154963.
- 23 O. Pikuda, E. Roubeau Dumont, Q. Chen, J.-R. Macairan, S. A. Robinson, D. Berk and N. Tufenkji, Toxicity of microplastics and nanoplastics to *Daphnia magna*: Current status, knowledge gaps and future directions, *TrAC Trends in Analytical Chemistry*, 2023, **167**, 117208.
- 24 A. Muhammad, X. Zhou, J. He, N. Zhang, X. Shen, C. Sun, B. Yan and Y. Shao, Toxic effects of acute exposure to polystyrene microplastics and nanoplastics on the model insect, silkworm *Bombyx mori*, *Environmental Pollution*, 2021, **285**, 117255.
- 25 A. T. B. Guimarães, A. S. De Lima Rodrigues, P. S. Pereira, F. G. Silva and G. Malafaia, Toxicity of polystyrene nanoplastics in dragonfly larvae: An insight on how these pollutants can affect benthic macroinvertebrates, *Science of The Total Environment*, 2021, **752**, 141936.



- 1
2 26 M. Alaraby, D. Abass, J. Domenech, A. Hernández and R. Marcos, Hazard assessment of ingested
3 polystyrene nanoplastics in *Drosophila* larvae, *Environ. Sci.: Nano*, 2022, **9**, 1845–1857.
- 4
5 27 N. R. Brun, P. van Hage, E. R. Hunting, A.-P. G. Haramis, S. C. Vink, M. G. Vijver, M. J. M.
6 Schaaf and C. Tudorache, Polystyrene nanoplastics disrupt glucose metabolism and cortisol levels
7 with a possible link to behavioural changes in larval zebrafish, *Communications Biology*, ,
8 DOI:10.1038/s42003-019-0629-6.
- 9
10 28 I. Brandts, M. Garcia-Ordoñez, L. Tort, M. Teles and N. Roher, Polystyrene nanoplastics
11 accumulate in ZFL cell lysosomes and in zebrafish larvae after acute exposure, inducing a
12 synergistic immune response *in vitro* without affecting larval survival *in vivo*, *Environ. Sci.: Nano*,
13 2020, **7**, 2410–2422.
- 14
15 29 W. Gu, S. Liu, L. Chen, Y. Liu, C. Gu, H. Ren and B. Wu, Single-Cell RNA Sequencing Reveals
16 Size-Dependent Effects of Polystyrene Microplastics on Immune and Secretory Cell Populations
17 from Zebrafish Intestines, *Environ. Sci. Technol.*, 2020, **54**, 3417–3427.
- 18
19 30 X. Li, T. Zhang, W. Lv, H. Wang, H. Chen, Q. Xu, H. Cai and J. Dai, Intratracheal administration of
20 polystyrene microplastics induces pulmonary fibrosis by activating oxidative stress and Wnt/ β -
21 catenin signaling pathway in mice, *Ecotoxicology and Environmental Safety*, 2022, **232**, 113238.
- 22
23 31 V. Stock, L. Böhmert, E. Lisicki, R. Block, J. Cara-Carmona, L. K. Pack, R. Selb, D. Lichtenstein,
24 L. Voss, C. J. Henderson, E. Zabinsky, H. Sieg, A. Braeuning and A. Lampen, Uptake and effects of
25 orally ingested polystyrene microplastic particles *in vitro* and *in vivo*, *Arch Toxicol*, 2019, **93**, 1817–
26 1833.
- 27
28 32 J. Domenech, A. Hernández, L. Rubio, R. Marcos and C. Cortés, Interactions of polystyrene
29 nanoplastics with *in vitro* models of the human intestinal barrier, *Arch Toxicol*, 2020, **94**, 2997–
30 3012.
- 31
32 33 S. Ballesteros, J. Domenech, I. Barguilla, C. Cortés, R. Marcos and A. Hernández, Genotoxic and
33 immunomodulatory effects in human white blood cells after *ex vivo* exposure to polystyrene
34 nanoplastics, *Environ. Sci.: Nano*, 2020, **7**, 3431–3446.
- 35
36 34 C. Meindl, K. Öhlinger, V. Zrim, T. Steinkogler and E. Fröhlich, Screening for Effects of Inhaled
37 Nanoparticles in Cell Culture Models for Prolonged Exposure, *Nanomaterials (Basel)*, 2021, **11**,
38 606.
- 39
40 35 J. Antunes, P. Sobral, M. Martins and V. Branco, Nanoplastics activate a TLR4/p38-mediated pro-
41 inflammatory response in human intestinal and mouse microglia cells, *Environmental Toxicology*
42 *and Pharmacology*, 2023, **104**, 104298.
- 43
44 36 M. Zhang, J. Li, G. Xing, R. He, W. Li, Y. Song and H. Guo, Variation in the internalization of
45 differently sized nanoparticles induces different DNA-damaging effects on a macrophage cell line,
46 *Arch Toxicol*, 2011, **85**, 1575–1588.
- 47
48 37 B. Prietl, C. Meindl, E. Roblegg, T. R. Pieber, G. Lanzer and E. Fröhlich, Nano-sized and micro-
49 sized polystyrene particles affect phagocyte function, *Cell Biol Toxicol*, 2014, **30**, 1–16.
- 50
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- 38 V. Paget, S. Dekali, T. Kortulewski, R. Grall, C. Gamez, K. Blazy, O. Aguerre-Chariol, S. Chevillard, A. Braun, P. Rat and G. Lacroix, Specific Uptake and Genotoxicity Induced by Polystyrene Nanobeads with Distinct Surface Chemistry on Human Lung Epithelial Cells and Macrophages, *PLOS ONE*, 2015, **10**, e0123297.
- 39 I. Florance, S. Ramasubbu, A. Mukherjee and N. Chandrasekaran, Polystyrene nanoplastics dysregulate lipid metabolism in murine macrophages in vitro, *Toxicology*, 2021, **458**, 152850.
- 40 Q. Hu, H. Wang, C. He, Y. Jin and Z. Fu, Polystyrene nanoparticles trigger the activation of p38 MAPK and apoptosis via inducing oxidative stress in zebrafish and macrophage cells, *Environmental Pollution*, 2021, **269**, 116075.
- 41 V. Collin-Faure, B. Dalzon, J. Devcic, H. Diemer, S. Cianfèrani and T. Rabilloud, Does size matter? A proteomics-informed comparison of the effects of polystyrene beads of different sizes on macrophages, *Environ. Sci.: Nano*, 2022, **9**, 2827–2840.
- 42 V. Collin-Faure, M. Vitipon, A. Torres, O. Tanyeres, B. Dalzon and T. Rabilloud, The internal dose makes the poison: higher internalization of polystyrene particles induce increased perturbation of macrophages, *Front Immunol*, 2023, **14**, 1092743.
- 43 D. G. Lundgren, R. Alper, C. Schnaitman and R. H. Marchessault, Characterization of Poly-Beta-Hydroxybutyrate extracted from different bacteria, *J Bacteriol*, 1965, **89**, 245–251.
- 44 R. Bhati, S. Samantaray, L. Sharma and N. Mallick, Poly- β -hydroxybutyrate accumulation in cyanobacteria under photoautotrophy, *Biotechnol J*, 2010, **5**, 1181–1185.
- 45 D. Jendrossek and D. Pfeiffer, New insights in the formation of polyhydroxyalkanoate granules (carbonosomes) and novel functions of poly(3-hydroxybutyrate), *Environ Microbiol*, 2014, **16**, 2357–2373.
- 46 L. H. Gracioso, A. Bellan, B. Karolski, L. O. B. Cardoso, E. A. Perpetuo, C. A. O. do Nascimento, R. Giudici, V. Pizzocchero, M. Basaglia and T. Morosinotto, Light excess stimulates Poly-beta-hydroxybutyrate yield in a mangrove-isolated strain of *Synechocystis* sp, *Bioresour Technol*, 2021, **320**, 124379.
- 47 P. Cabecas Segura, R. Onderwater, A. Deutschbauer, L. Dewasme, R. Wattiez and B. Leroy, Study of the Production of Poly(Hydroxybutyrate-co-Hydroxyhexanoate) and Poly(Hydroxybutyrate-co-Hydroxyvalerate-co-Hydroxyhexanoate) in *Rhodospirillum rubrum*, *Appl Environ Microbiol*, 2022, **88**, e0158621.
- 48 D. E. Cutright and E. E. Hunsuck, Tissue reaction to the biodegradable polylactic acid suture, *Oral Surg Oral Med Oral Pathol*, 1971, **31**, 134–139.
- 49 D. E. Cutright, J. D. Beasley and B. Perez, Histologic comparison of polylactic and polyglycolic acid sutures, *Oral Surg Oral Med Oral Pathol*, 1971, **32**, 165–173.
- 50 R. K. Kulkarni, E. G. Moore, A. F. Hegyeli and F. Leonard, Biodegradable poly(lactic acid) polymers, *J Biomed Mater Res*, 1971, **5**, 169–181.

- 1
2 51 Y. Kawashima, H. Yamamoto, H. Takeuchi, S. Fujioka and T. Hino, Pulmonary delivery of insulin
3 with nebulized DL-lactide/glycolide copolymer (PLGA) nanospheres to prolong hypoglycemic
4 effect, *J Control Release*, 1999, **62**, 279–287.
5
6 52 O. Pillai and R. Panchagnula, Polymers in drug delivery, *Curr Opin Chem Biol*, 2001, **5**, 447–451.
7
8 53 S. Freiberg and X. X. Zhu, Polymer microspheres for controlled drug release, *Int J Pharm*, 2004,
9 **282**, 1–18.
10
11 54 M. S. Shive and J. M. Anderson, Biodegradation and biocompatibility of PLA and PLGA
12 microspheres, *Adv Drug Deliv Rev*, 1997, **28**, 5–24.
13
14 55 S. Jesus, M. Schmutz, C. Som, G. Borchard, P. Wick and O. Borges, Hazard Assessment of
15 Polymeric Nanobiomaterials for Drug Delivery: What Can We Learn From Literature So Far, *Front*
16 *Bioeng Biotechnol*, 2019, **7**, 261.
17
18 56 M. H. Wolf, O. Gil-Castell, J. Cea, J. C. Carrasco and A. Ribes-Greus, Degradation of Plasticised
19 Poly(lactide) Composites with Nanofibrillated Cellulose in Different Hydrothermal Environments, *J*
20 *Polym Environ*, 2023, **31**, 2055–2072.
21
22 57 G. Banaei, A. García-Rodríguez, A. Tavakolpournegari, J. Martín-Pérez, A. Villacorta, R. Marcos
23 and A. Hernández, The release of polylactic acid nanoplastics (PLA-NPLs) from commercial
24 teabags. Obtention, characterization, and hazard effects of true-to-life PLA-NPLs, *Journal of*
25 *Hazardous Materials*, 2023, **458**, 131899.
26
27 58 N. Bhutiani, A. Samykutty, K. M. McMasters, N. K. Egilmez and L. R. McNally, In vivo tracking of
28 orally-administered particles within the gastrointestinal tract of murine models using multispectral
29 optoacoustic tomography, *Photoacoustics*, 2019, **13**, 46–52.
30
31 59 T. Behnke, C. Würth, K. Hoffmann, M. Hübner, U. Panne and U. Resch-Genger, Encapsulation of
32 Hydrophobic Dyes in Polystyrene Micro- and Nanoparticles via Swelling Procedures, *J Fluoresc*,
33 2011, **21**, 937–944.
34
35 60 A. Torres, V. Collin-Faure, D. Fenel, J.-A. Sergent and T. Rabilloud, About the Transient Effects of
36 Synthetic Amorphous Silica: An In Vitro Study on Macrophages, *Int J Mol Sci*, 2022, **24**, 220.
37
38 61 B. Dalzon, A. Torres, J. Devcic, D. Fenel, J.-A. Sergent and T. Rabilloud, A Low-Serum Culture
39 System for Prolonged in Vitro Toxicology Experiments on a Macrophage System, *Front. Toxicol.*,
40 2021, **3**, 780778.
41
42 62 L. Muller, L. Fornecker, M. Chion, A. Van Dorselaer, S. Cianférani, T. Rabilloud and C. Carapito,
43 Extended investigation of tube-gel sample preparation: a versatile and simple choice for high
44 throughput quantitative proteomics, *Sci Rep*, 2018, **8**, 8260.
45
46 63 C. Cavazza, V. Collin-Faure, J. Pérard, H. Diemer, S. Cianférani, T. Rabilloud and E. Darrouzet,
47 Proteomic analysis of *Rhodospirillum rubrum* after carbon monoxide exposure reveals an important
48 effect on metallic cofactor biosynthesis, *J Proteomics*, 2022, **250**, 104389.
49
50 64 T. Lyubimova, S. Caglio, C. Gelfi, P. G. Righetti and T. Rabilloud, Photopolymerization of
51
52
53
54
55
56
57
58
59
60

- polyacrylamide gels with methylene blue, *Electrophoresis*, 1993, **14**, 40–50.
- 65 O. Hammer, D. A. T. Harper, and Ryan P. D, Paleontological statistics software package for education and data analysis, *Palaeontologia Electronica*, 2001, **4**, 9pp.
- 66 D. W. Huang, B. T. Sherman and R. A. Lempicki, Systematic and integrative analysis of large gene lists using DAVID bioinformatics resources, *Nat Protoc*, 2009, **4**, 44–57.
- 67 S. W. Perry, J. P. Norman, J. Barbieri, E. B. Brown and H. A. Gelbard, Mitochondrial membrane potential probes and the proton gradient: a practical usage guide, *Biotechniques*, 2011, **50**, 98–115.
- 68 J. F. Keij, C. Bell-Prince and J. A. Steinkamp, Staining of mitochondrial membranes with 10-nonyl acridine orange, MitoFluor Green, and MitoTracker Green is affected by mitochondrial membrane potential altering drugs, *Cytometry*, 2000, **39**, 203–210.
- 69 K. R. Clarke, Non-parametric multivariate analyses of changes in community structure, *Australian Journal of Ecology*, 1993, **18**, 117–143.
- 70 M. Boyce, K. F. Bryant, C. Jousse, K. Long, H. P. Harding, D. Scheuner, R. J. Kaufman, D. Ma, D. M. Coen, D. Ron and J. Yuan, A selective inhibitor of eIF2alpha dephosphorylation protects cells from ER stress, *Science*, 2005, **307**, 935–939.
- 71 T. L. Pan, P. W. Wang, Y. C. Hung, C. H. Huang and K. M. Rau, Proteomic analysis reveals tanshinone IIA enhances apoptosis of advanced cervix carcinoma CaSki cells through mitochondria intrinsic and endoplasmic reticulum stress pathways, *Proteomics*, 2013, **13**, 3411–3423.
- 72 R. Shen, K. Yang, X. Cheng, C. Guo, X. Xing, H. Sun, D. Liu, X. Liu and D. Wang, Accumulation of polystyrene microplastics induces liver fibrosis by activating cGAS/STING pathway, *Environmental Pollution*, 2022, **300**, 118986.
- 73 R. Marfella, F. Prattichizzo, C. Sardu, G. Fulgenzi, L. Graciotti, T. Spadoni, N. D’Onofrio, L. Scisciola, R. La Grotta, C. Frigé, V. Pellegrini, M. Municinò, M. Siniscalchi, F. Spinetti, G. Vigliotti, C. Vecchione, A. Carrizzo, G. Accarino, A. Squillante, G. Spaziano, D. Mirra, R. Esposito, S. Altieri, G. Falco, A. Fenti, S. Galoppo, S. Canzano, F. C. Sasso, G. Maticchione, F. Olivieri, F. Ferraraccio, I. Panarese, P. Paolisso, E. Barbato, C. Lubritto, M. L. Balestrieri, C. Mauro, A. E. Caballero, S. Rajagopalan, A. Ceriello, B. D’Agostino, P. Iovino and G. Paolisso, Microplastics and Nanoplastics in Atheromas and Cardiovascular Events, *N Engl J Med*, 2024, **390**, 900–910.
- 74 K. Hamidzadeh, S. M. Christensen, E. Dalby, P. Chandrasekaran and D. M. Mosser, Macrophages and the Recovery from Acute and Chronic Inflammation, *Annu Rev Physiol*, 2017, **79**, 567–592.
- 75 V. Olivier, J. L. Duval, M. Hindié, P. Pouletaut and M. D. Nagel, Comparative particle-induced cytotoxicity toward macrophages and fibroblasts, *Cell Biol Toxicol*, 2003, **19**, 145–159.
- 76 D. F. Williams, Enzymic Hydrolysis of Polylactic Acid, *Engineering in Medicine*, 1981, **10**, 5–7.
- 77 G. A. Brooks, Lactate as a fulcrum of metabolism, *Redox Biology*, 2020, **35**, 101454.
- 78 C. Manosalva, J. Quiroga, A. I. Hidalgo, P. Alarcón, N. Anseoleaga, M. A. Hidalgo and R. A.

- 1
2 Burgos, Role of Lactate in Inflammatory Processes: Friend or Foe, *Front Immunol*, 2021, **12**,
3 808799.
4
5
6 79 A. Drazic, L. M. Myklebust, R. Ree and T. Arnesen, The world of protein acetylation, *Biochim.*
7 *Biophys. Acta*, 2016, **1864**, 1372–1401.
8
9 80 E. B. Ketema and G. D. Lopaschuk, Post-translational Acetylation Control of Cardiac Energy
10 Metabolism, *Front Cardiovasc Med*, 2021, **8**, 723996.
11
12 81 Y.-H. Yang, R. Wen, N. Yang, T.-N. Zhang and C.-F. Liu, Roles of protein post-translational
13 modifications in glucose and lipid metabolism: mechanisms and perspectives, *Mol Med*, 2023, **29**,
14 93.
15
16 82 H.-J. Lin, P. Herman, J. S. Kang and J. R. Lakowicz, Fluorescence Lifetime Characterization of
17 Novel Low-pH Probes, *Analytical Biochemistry*, 2001, **294**, 118–125.
18
19 83 Y. Zhao, Z. Ye, D. Song, D. Wich, S. Gao, J. Khirallah and Q. Xu, Nanomechanical action opens
20 endo-lysosomal compartments, *Nat Commun*, 2023, **14**, 6645.
21
22 84 J. Ai, A. Maturu, W. Johnson, Y. Wang, C. B. Marsh and S. Tridandapani, The inositol phosphatase
23 SHIP-2 down-regulates FcγR-mediated phagocytosis in murine macrophages independently of
24 SHIP-1, *Blood*, 2006, **107**, 813–820.
25
26 85 J. M. Rubio, A. M. Astudillo, J. Casas, M. A. Balboa and J. Balsinde, Regulation of Phagocytosis in
27 Macrophages by Membrane Ethanolamine Plasmalogens, *Front. Immunol.*, 2018, **9**, 1723.
28
29 86 Y. Zhao, C. Alonso, I. Ballester, J. H. Song, S. Y. Chang, B. Guleng, S. Arihiro, P. J. Murray, R.
30 Xavier, K. S. Kobayashi and H.-C. Reinecker, Control of NOD2 and Rip2-dependent innate immune
31 activation by GEF-H1, *Inflamm Bowel Dis*, 2012, **18**, 603–612.
32
33 87 D. Wang, J. Lou, C. Ouyang, W. Chen, Y. Liu, X. Liu, X. Cao, J. Wang and L. Lu, Ras-related
34 protein Rab10 facilitates TLR4 signaling by promoting replenishment of TLR4 onto the plasma
35 membrane, *Proc Natl Acad Sci U S A*, 2010, **107**, 13806–13811.
36
37 88 H. Sun, S. Gong, R. J. Carmody, A. Hilliard, L. Li, J. Sun, L. Kong, L. Xu, B. Hilliard, S. Hu, H.
38 Shen, X. Yang and Y. H. Chen, TIPE2, a negative regulator of innate and adaptive immunity that
39 maintains immune homeostasis, *Cell*, 2008, **133**, 415–426.
40
41 89 E. J. Fenech, F. Lari, P. D. Charles, R. Fischer, M. Laëtita-Thézénas, K. Bagola, A. W. Paton, J. C.
42 Paton, M. Gyrd-Hansen, B. M. Kessler and J. C. Christianson, Interaction mapping of endoplasmic
43 reticulum ubiquitin ligases identifies modulators of innate immune signalling, *eLife*, 2020, **9**,
44 e57306.
45
46 90 J. E. Klaunig and L. M. Kamendulis, The role of oxidative stress in carcinogenesis, *Annu Rev*
47 *Pharmacol Toxicol*, 2004, **44**, 239–267.
48
49 91 R. Li, M. A. Bianchet, P. Talalay and L. M. Amzel, The three-dimensional structure of
50 NAD(P)H:quinone reductase, a flavoprotein involved in cancer chemoprotection and chemotherapy:
51 mechanism of the two-electron reduction, *Proc Natl Acad Sci U S A*, 1995, **92**, 8846–8850.
52
53
54
55
56
57
58
59
60

- 1
2
3 92 P. V. Rao, C. M. Krishna and J. S. Zigler, Identification and characterization of the enzymatic
4 activity of zeta-crystallin from guinea pig lens. A novel NADPH:quinone oxidoreductase, *J Biol*
5 *Chem*, 1992, **267**, 96–102.
- 6
7
8 93 C. A. Shultz, A. M. Quinn, J.-H. Park, R. G. Harvey, J. L. Bolton, E. Maser and T. M. Penning,
9 Specificity of human aldo-keto reductases, NAD(P)H:quinone oxidoreductase, and carbonyl
10 reductases to redox-cycle polycyclic aromatic hydrocarbon diones and 4-hydroxyequilenin-o-
11 quinone, *Chem Res Toxicol*, 2011, **24**, 2153–2166.
- 12
13
14 94 J. L. Bolton, M. A. Trush, T. M. Penning, G. Dryhurst and T. J. Monks, Role of quinones in
15 toxicology, *Chem Res Toxicol*, 2000, **13**, 135–160.
- 16
17
18 95 K. Luo, X. Luo, W. Cao, J. B. Hochalter, V. Paiano, C. J. Sipe, S. G. Carmella, S. E. Murphy, J.
19 Jensen, S. Lam, A. P. Golin, L. Bergstrom, D. Midthun, N. Fujioka, D. Hatsukami and S. S. Hecht,
20 Cigarette smoking enhances the metabolic activation of the polycyclic aromatic hydrocarbon
21 phenanthrene in humans, *Carcinogenesis*, 2021, **42**, 570–577.
- 22
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2 The proteomic data have been deposited in the ProteomeXchange Consortium database and are
3 available through the DOI: 10.6019/PXD048664.
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6 The data on targeted experiments are available with the doi 10.6019/S-BSST1470 from the EMBL
7 BioStudies database (<https://www.ebi.ac.uk/biostudies/studies/S-BSST1470>)
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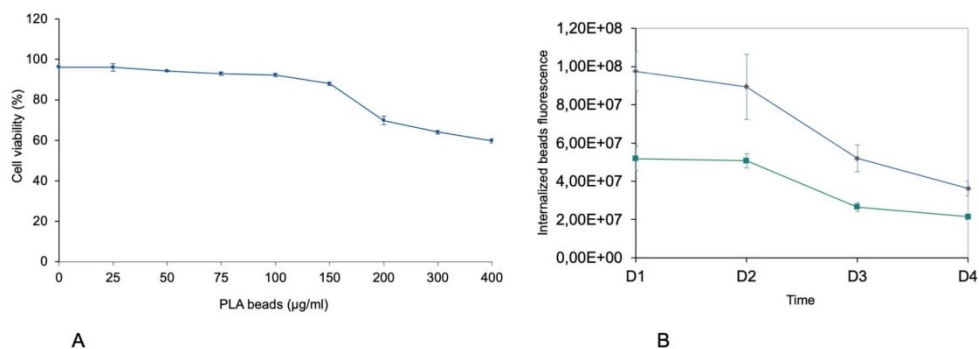


Figure 1: Viability and intracellular fluorescence of cells treated with PLA particles

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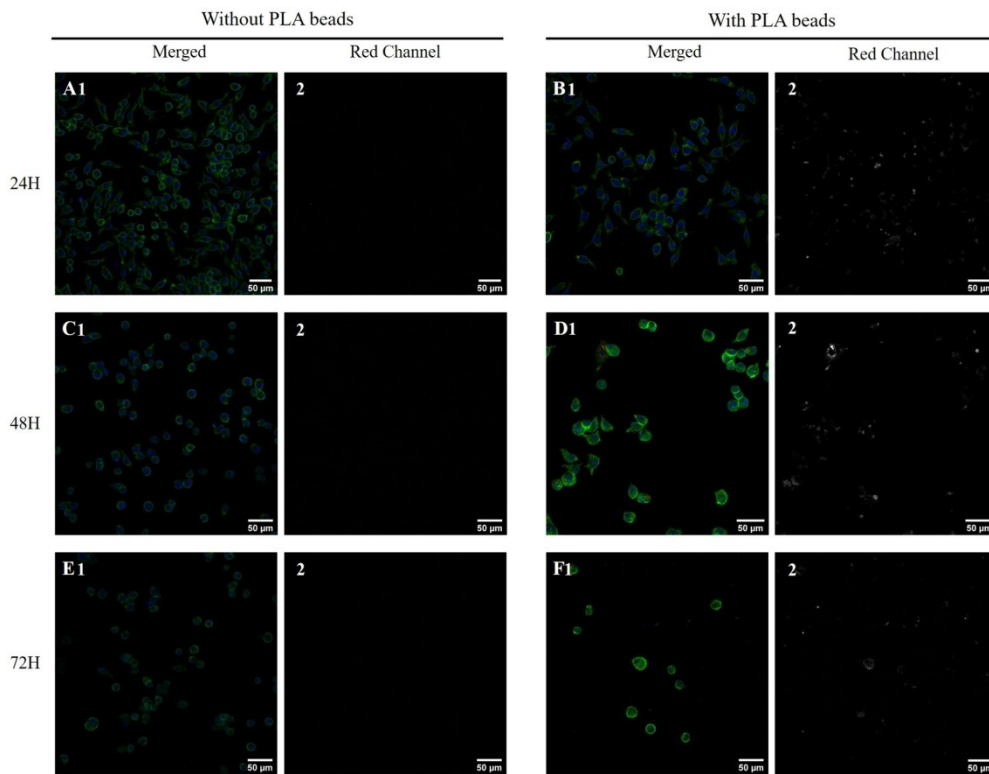
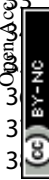


Figure 2: Confocal Microscopy Imaging of J774.A1 Cells

179x139mm (300 x 300 DPI)

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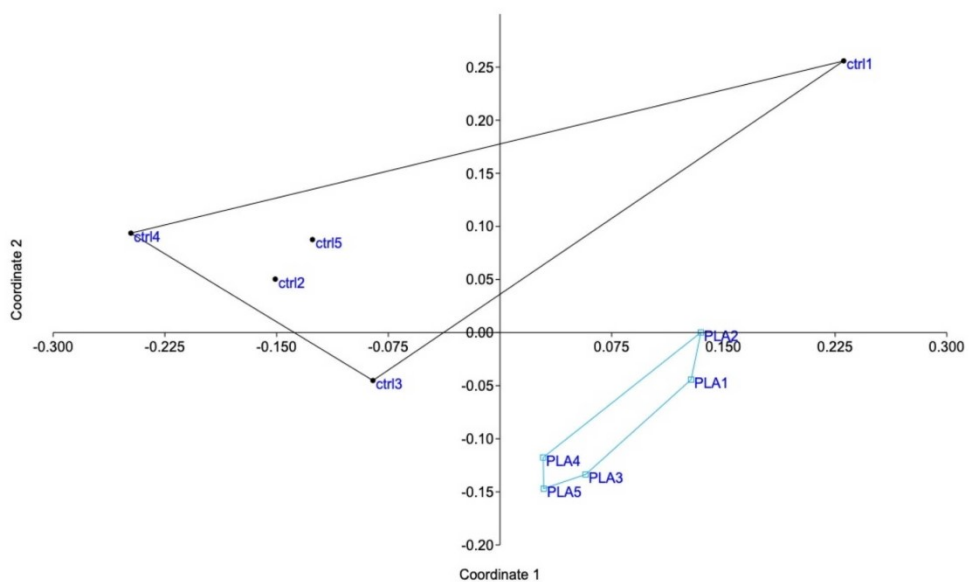


Figure 3: Global analysis of the proteomic data

170x111mm (300 x 300 DPI)

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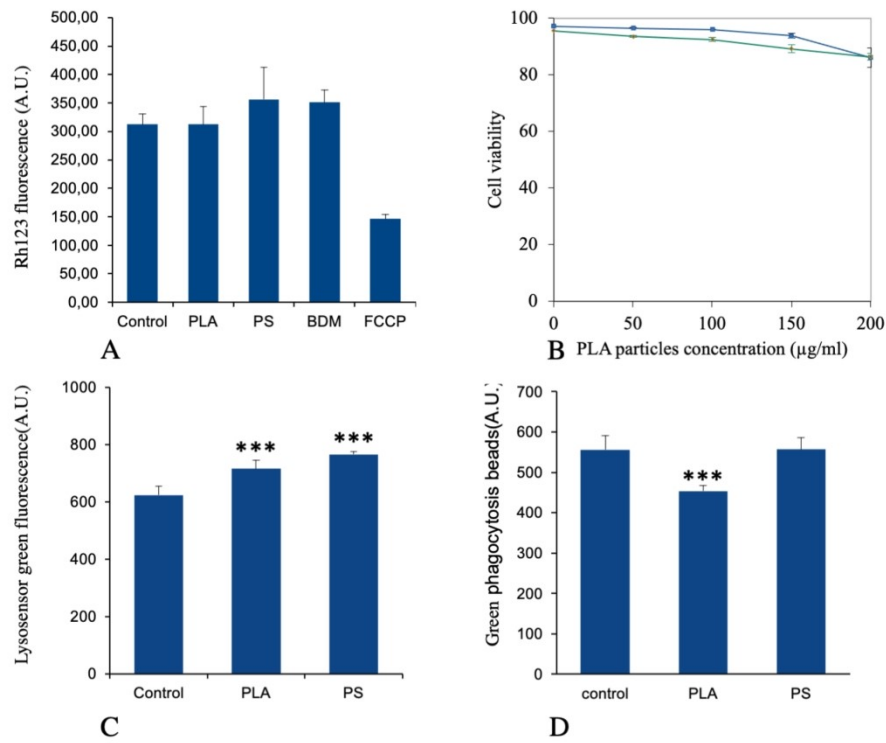


Figure 4: Mitochondria, lysosomes and phagocytosis experiments

170x133mm (300 x 300 DPI)

Figure 5: Cytokine release

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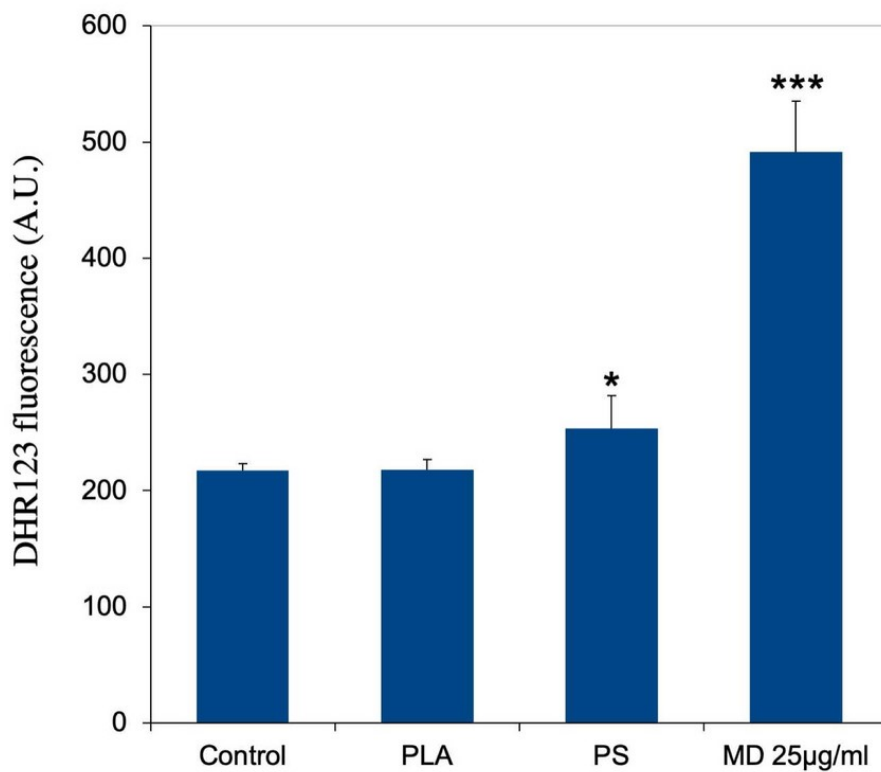


Figure 6: Cellular oxidative stress, measured with the dihydrorhodamine 123 (DHR123) indicator

80x66mm (300 x 300 DPI)

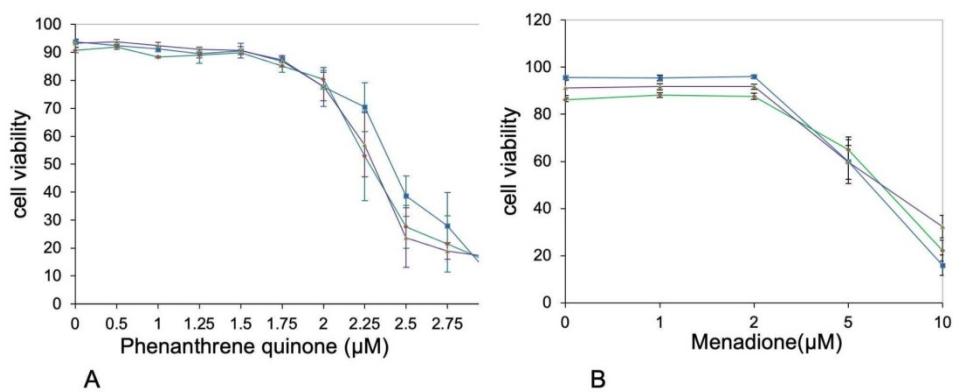


Figure 7: Cross toxicities

169x73mm (300 x 300 DPI)