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1	Removal of basic dye Auramine-O by ZnS: Cu nanoparticles loaded on activated carbon
2	Optimization of parameters using response surface methodology with central composite
3	design
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21 Abstract

This research is focused on the ultrasound-assisted removal of Auramine-O (AO) dve 22 from aqueous solutions using ZnS:Cu nanoparticles loaded on activated carbon (ZnS:Cu-NP-23 AC) as an adsorbent. ZnS:Cu nanoparticles were synthesized and characterized using FESEM 24 (Field-Emission Scanning Electron Microscopy) and XRD (X-Ray Diffraction) analysis. The 25 experiments were designed by response surface methodology. Ouadratic model was used to 26 predict the variables. Analysis of variance was used for investigation of variables and interaction 27 between them. High F-value (48.91), very low P-value (<0.00001), non-significant lack of fit, 28 the determination coefficient ($R^2 = 0.977$) demonstrate good correlation between experimental 29 and predicted values of the response. The highest removal percent (99.76%) was attained, and 30 the optimum parameters are achieved: adsorbent amount (0.02 g), initial concentration dye (20 31 mgL⁻¹), sonication time (3 min) and pH=7. Adsorption processes of AO by ZnS: Cu-NP-AC 32 could be well described with Langmuir isotherm and a pseudo- second-order kinetic model. The 33 maximum adsorption capacity of AO by ZnS: Cu-NP-AC was determined as 183.15 mg g⁻¹, 34 suggesting a highly promising potential for ZnS: Cu-NP-AC to be used as a new adsorbent. 35

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- 38

Keywords: Ultrasound-Assisted; Nanoparticle Loaded Activated Carbon; Response Surface
Methodology; Auramine-O.

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44 **1. Introduction**

The discharge of dyes in to the biological ecosystem is worrying for both toxicological 45 and esthetical aspects [1]. Almost 45% of textile dyes produced worldwide belongs to the 46 reactive class [2]. Reactive dyes are common dyes used for dyeing cellulosic fibres due to their 47 favorable characteristics of bright color, water-fastness, simple application techniques and low 48 energy consumption [3]. Auramine-O (AO) and its hydrochloride salts are used as coloring agent 49 in paper, textiles and leather industries [4]. International agency for research on cancer (IARC) 50 included AO among chemicals for which there is sufficient evidence of carcinogenicity due to its 51 bio-transformation to reactive species in target organs of both rats and humans [5, 6]. Auramine-52 O (AO) is yellow in color and is frequently used in paper mills, textile mills, leather and carpet 53 industry. Presence of these dyes is highly persistent and the manufacturers always go for the 54 most stable dye. For the efficient removal of hazardous impurities, several physical and chemical 55 methods such as coagulation [7], reverse osmosis [8], photo degradation [9], electrochemical 56 oxidation [10], ozonation [11], biosorption [12] and adsorption are used. Among all, adsorption 57 as a popular alternative procedure, especially based on low cost adsorbent benefit from remarks 58 such as, simple design, easy operation and the possibility of using green and non-toxic adsorbent. 59 Many adsorbent were tested on the possibility of the hazardous dye removal such as carbon 60 nanotube [13-21], MWCNTs [22, 23] activated carbon [24, 25], fly ash [26], chitin [27], zeolite 61 62 [28], polymer [29], low cost adsorbents [30-40], lignin [41, 42], barley straw [43], nanocomposites [44-47] and graphene oxide [48]. Design and application of non-toxic adsorbent 63 that was able to remove a huge amount of hazardous dyes molecule in short time are crucial 64 65 requirements for the wastewater treatment. Nanoparticles exhibit intrinsic surface activity, high

surface areas and reactive atom or functional group strongly chemisorbed many hazardous

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chemical compounds. The size, surface structure and interparticle interaction of nanomaterials 67 determine their unique properties which proved nanoparticle an efficient adsorbent as well as 68 several other potential applications in many relevant research areas. The key objective of the 69 present work is the synthesis of ZnS:Cu nanoparticle loaded on AC, which is followed by a 70 characterization via different analytical techniques such as UV-vis, SEM and XRD. The 71 potential feasibility of ZnS:Cu-NP-AC for the adsorption process of AO was investigated and 72 the influence of certain variables was studied and optimized by central composite design (CCD) 73 74 combined with response surface methodology (RSM) using the desirability function (DF) as maximize criterion of the response. The results obtained from the presented models were 75 compared with the experimental values. The adsorption kinetics and isotherms of dye removal on 76 this adsorbent was also investigated. The adsorption rates were evaluated by fitting the 77 experimental data to traditional kinetic models such as pseudo first-order, second-order and 78 intraparticle diffusion models. The proposed sorbent will be useful for quantitative adsorption of 79 the dye with high sorption capacities in short time. 80

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66

82 **2.** Materials and methods

83 2.1. Materials and instruments

Auramine-O (4, 4-dimethylaminobenzophenonimide) (AO) were considered as azo dye. Detailed description of this dye is shown in Table 1. The stock solution (200 mg/ L) of dye was prepared by dissolving 200 mg of solid dye in 1000 mL of double distilled water and the working concentrations were prepared daily by suitable dilution. The ZnS: Cu-NP-AC nanoparticles were prepared and characterized by using BET and SEM. The BET (Brunauer,

89 Emmett, and Teller) surface area of the adsorbent material was measured using TriStar II 3020 (Micrometrics Instrument Corporation) surface area analyzer where N2 gas was used as 90 adsorbate. The nitrogen sorption analysis was accomplished using a Belsorp-BEL. Inc. analyzer 91 92 at 77 K. Prior to measurement, the materials was degassed at 373 K for 12 h. The surface area of the ZnS: Cu-NP-AC was calculated by BET method and the pore size distribution was calculated 93 from the adsorption branch of the isotherm using BJH method. NaOH and HCl with the highest 94 purity were purchased from Merck (Darmstadt, Germany). The pH measurements were carried 95 out using pH/Ion meter model-686 (Metrohm, Switzerland, Swiss) and the AO concentrations 96 were determined using Jusco UV-vis spectrophotometer model V-530 (Jasco, Japan) at 97 wavelength of 434 nm, respectively. An ultrasonic bath with heating system (Tecno-GAZ SPA 98 Ultra Sonic System) at 40 kHz of frequency and 130 W of power was used for the ultrasound-99 100 assisted adsorption procedure.

101

102 2.2. Ultrasound-assisted adsorption method

A batch method was used to appraise the adsorption performance of AO dye from aqueous solutions onto ZnS: Cu-NP-AC in presence of ultrasonic wave. Adsorption experiments were performed in a cylindrical glass vessel by addition of adsorbent (0.02 g) into 50 mL of AO solutions at known concentration (5 and 30 mg L⁻¹) and pH 7. The vessel was immersed in an ultrasonic bath for 3.0 min sonication time at the room temperature. After this time, solutions were analyzed for the final concentration of dyes by using a UV–vis spectrophotometer set at a wavelength of 434 nm for AO, respectively.

110

111 2.3. Measurements of dye uptake

112 The dye concentrations were determined according to calibration curve obtained at 113 maximum wavelength over functioning concentration range. The efficiency of dye removal was 114 determined at different experimental conditions and optimized according to the CCD method 115 discussed in subsequent section. The dye removal percentage was calculated using the following 116 equation:

117 %R JSB =
$$\frac{C_0 - C_t}{C_0} \times 100$$
 (1)

118

Where $C_0 \text{ (mg } L^{-1})$ and $C_t \text{ (mg } L^{-1})$ is the concentration of dye at initial and after time t, respectively [49].

121

122 2.4. Kinetic experiments

Kinetic adsorption is performed to investigate the mechanism of adsorption and to determine the equilibrium time [50]. For this purpose 0.01 and 0.02 g of adsorbent was contacted with 50 mL of AO solution with an initial concentration of AO, 10, 20 and 30 mgL⁻¹, in ultrasonic at temperature for different time intervals at the optimum pH 7. The samples were filtered and determined using Jusco UV-vis spectrophotometer model V-530 (Jasco, Japan) at wavelength of 434 nm, respectively [51].

129

130 2.5. Adsorption isotherms

Adsorption isotherms are used to describe the equilibrium behaviors of adsorbate uptake [52]. For isotherms experiments, various amounts of adsorbents (0.01, 0.015 and 0.02g) were contacted with 50 mL of solution AO with an initial concentration of (5-30 mgL⁻¹), in 6 flasks at the optimum pH 7. The containers were mechanically agitated in a shaker at ultrasonic in room

temperature for 3 min. The samples were filtered and determined using Jusco UV-vis spectrophotometer model V-530 (Jasco, Japan) at wavelength of 434 nm, respectively. The adsorbed dye amount (q_e (mg g⁻¹)) was calculated by the following mass balance relationship: 138

139
$$q_e = \frac{(C_0 - C_e)V}{W}$$
 (2)

140

141 Where q_e is the amount of adsorbed manganese per gram adsorbent at equilibrium (mg/g) and C_0 142 and C_e are the concentrations of the metal ions before and after adsorption (mg/L), V the volume 143 of the aqueous phase (L) and m the mass of the adsorbent (g) [53].

144

145 2.6. Preparation of ZnS:Cu-NP-AC

Analytical reagent grade zinc sulfate (ZnSO₄, 2H₂O), cupper (II) acetate (Cu (CH₃COO) 146 2) and thiourea (SC (NH₂)₂) were purchased from Merck company and used without further 147 purification. The preparation of ZnS: Cu-NPs were carried out at two steps. In the first step, ZnS 148 nanoparticles (ZnS-NPs) were synthesized. The precursor solution for synthesis of ZnS-NPs was 149 prepared as follows: 0.6 mmol of zinc sulfate solution was mixed with 30 ml of 0.5M thiourea 150 solution at pH= 5.5 and deionized water was added to the mixed solutions to make a total volume 151 of 250 ml. Then, 250 ml of the precursor solution in a baker was transferred to an autoclave at 152 pressure of 1.25 bar for 3 h. The temperature of the autoclave was 125 ° C. After 3 hours, the 153 baker containing the reaction solution containing white colored ZnS-NPs in bottom of the baker 154 was removed from autoclave. The obtained ZnS - NPs were filtered and washed several times by 155 deionized water. In the second step, ZnS: Cu-NPs-AC was prepared. 250 ml of deionized water 156 157 was added to the ZnS-NPs prepared from step 1 to form an insoluble suspension. 1 ml 0.2 M (Cu

158 $(CH_3COO)_2$ solution was added to the ZnS-NPs suspension along with vigorous stirring for 5 min. After adding Cu^{2+} solution to ZnS-NPs suspension, its color slightly changes from milky 159 white to light green because of the diffusion of Cu^{2+} ions to ZnS-NPs and formation of ZnS: Cu-160 NPs suspension. Finally, the homogenous deposition of ZnS: Cu-NPs on activated carbon (AC) 161 was carried out by adding 10 g of AC to the obtained ZnS:Cu-NPs suspension and strong stirring 162 for 20 h at room temperature. The prepared ZnS:Cu-NPs-AC were then filtered, washed several 163 times by deionized water, dried at 60 °C for 3h and used as an absorbent for adsorption 164 experiments, the major impurities that may be present during the synthesis of ZnS:Cu 165 nanoparticles are ZnO and Zn (OH)₂ but the XRD pattern confirm high purity of ZnS:Cu-NP and 166 its agreement with reference material without any further impurity. 167

168

169 2.7. Role of ZnS:Cu nanoparticles

ZnS:Cu nanoparticle posses a crucial role on the adsorption phenomenon, it lead to the 170 enhancement of the surface area and number of active sites, hence it is an efficient adsorbent, 171 when loaded on AC, it causes very rapid adsorption of hazardous materials i.e. dyes, even by 172 using a very low amount of adsorbent dose i.e. 0.02, maximum 99.76% of the adsorption takes 173 place. Hence to carry out at an efficient and maximum adsorption it is needed to coat ZnS:Cu 174 nanoparticles on to the activated carbon because several parameters are effected like sonication 175 time, pH, and adsorbent dose, a major change is reported in the sonication time parameter from 176 120 and 300 minutes it reduces to only 3 minutes to achieve the maximum adsorption. 177

178 2.8. Reusable capacity of ZnS:Cu nanoparticles

179 Like the other nanoparticles adsorbent i.e. Fe_2O_3 , MnO etc, they can be reused as 180 adsorbent after magnetic separation for removing the noxious toxic contaminants, it is seen that

it can be used maximum four times without any change and modification in the developedadsorbent [54, 55]

183 2.9. Central composite design (CCD)

A central composite design (CCD) was used to determine the optimal conditions for the critical factors. For the adsorption process, significant variables, such as pH, AO concentration, adsorbent and sonication time, were chosen as the independent variables and designated as X_{1} - X_{4} , respectively. The pH (X_{1}) ranged from 5 to 9, adsorbent (X_{2}) ranged from 0.005 to 0.025 g, sonication time (X_{3}) ranged from 1 to 5 min and the AO concentration (X_{4}) ranged from 10 to 30 mg L⁻¹, as shown in Table 2. The real values of the independent variables (X_{i}) were coded to z_{i} according to Eq. (3) by setting the lowest values as -2 and the highest values as +2:

191

$$192 Z_i = \frac{X_i - X_0}{\Delta X_i} (3)$$

193

194 Where z_i is the dimensionless value of an independent variable, X_i represents the real value of 195 the independent variable, X_0 is the real value of the independent variable at the center point, and 196 ΔX_i is the step change [56, 57]. The number of experiments was equal to Eq. (4), where k and n 197 are the number of factors and center runs, respectively (k=4, n=7), plus two additional points. A 198 total of 31 experiments were performed.

199

200
$$N = 2^{\kappa} + 2k + n_c$$
 (4)

Where k is the number of variables and n_c is the number of central points. A four-factor fivelevel CCD was used to fit the general model of Eq. (5) and to obtain optimal conditions for dependent variables (Y).

205

206
$$y = \beta_0 + \sum_{i=1}^4 \beta_i x_i + \sum_{i=1}^4 \sum_{j=1}^4 \beta_{ij} x_i x_j + \sum_{i=1}^4 \beta_{ii} x_i^2 + \varepsilon$$
 (5)

207

Where y is the response, β_0 , β_i , β_{ii} are the regression coefficients of variables for intercept, linear, 208 quadratic and interaction terms, respectively. X_i and X_i are the independent variables and ε is the 209 210 residual term. The STATISTICA software (Version 10.0) was used for data processing. Experimental data were fitted to a second-order polynomial equation, and regression coefficients 211 were obtained. The analysis of variance (ANOVA) was performed to justify the significance and 212 adequacy of the developed regression model. The adequacy of the response surface models were 213 evaluated by calculation of the determination coefficient (R^2) and also by testing it for the lack of 214 fit. 215

216

217 **2.** Results and discussion

218 **3.1.** Characterization of adsorbent

The optical absorbance spectrum of the prepared ZnS: Cu–NP–AC has a steep absorption edge, indicating good homogeneity in the shape and size of the particles as well as low defect density near the band edge [64], it is well elucidated from Fig. 1. From the absorption data, the band gap energy of ZnS: Cu–NP–AC was estimated using the well-known relation for semiconductors [58, 59]:

$$ahv = k \left(hv - E_g\right)^{\frac{n}{2}}$$
(6)

Where, E_g is the band gap energy, k is a constant, and n is a constant equal to 1 or 4 for direct and indirect band gap materials, respectively. A plot of $(\alpha hv)^2$ versus hv (inset of Fig. 1) is linear at the absorption edge, which means that the mode of transition in these films has a direct nature. The band gap energy, Eg, was calculated about 3.97 eV from an extrapolation of the straight-line portion of the $(\alpha hv)^2$ vs hv plot to zero absorption coefficient value. The obtained band gap energy for the prepared ZnS:Cu –NP is larger than that of the bulk ZnS (3.60 eV) which could be assigned to quantum confinement effects in nanosized material.

234 The morphology and particle size of the prepared ZnS:Cu -NP were studied by FESEM (FE-SEM; Hitachi S-4160, Japan) under an acceleration voltage of 200 kV. The FESEM images of 235 the prepared ZnS:Cu – NP at different magnifications were shown in Figs. 2a - 2c. The surface 236 237 textural and morphology reveals the porous nature and rough surface of the developed adsorbent, it seems that the porous structure was due to the ZnS:Cu nanoparticles loaded on the AC, which 238 should increase the effective surface for adsorption. The XRD patterns of ZnS:Cu nanoparticles 239 prepared at 75 °C has good agreement with standard JCPDS (Joint Committee for Powder 240 241 Diffraction Standards, JCPDS card No. 05-0566) pattern of ZnS (Fig. 3). The three broad peaks observed in the diffractogram at around 28.56°, 47.93° and 57.12° reveal a cubic lattice structure 242 of (β-ZnS phase) planes (111), (220) and (311), respectively. The broad nature of XRD peaks 243 confirms nanosized particles. According to full width at half-maximum (FWHM) of (110) peak 244 and based on the Debeye-Scherrer equation, The average crystalline sizes calculated from the 245 full width at half-maximum (FWHM) of these peaks were about 22 and 21 nm for cubic and 246 hexagonal ZnS:Cu nanoparticles, respectively. 247

248 3.2. Central composite design (CCD)

249 *3.2.1. Model fitting and statistical analysis*

In the CCD step as presented in Table 2, four independent variables (pH (X_1) , adsorbent 250 dosage (X_2) , sonication time (X_3) and AO concentration (X_4)) were prescribed into three levels 251 (low, basal and high) with coded value (-1, 0, +1) and the star points of +2 and -2 for + α and - α 252 respectively, were selected for each set of experiments. 31 experiments and their responses are 253 254 presented in Table 3. In order to optimize AO adsorption, central composite design (CCD) with a 255 total number of 31 experiments was used for the response surface modeling. Step-wise model 256 fitting by STATISTICA 10.0 software was used to get the best fitted model. The software suggested quadratic model by supporting lack of fit and model summary statistics (Table 4). The 257 258 model adequacy was further checked using ANOVA (Table 4).

259 The ANOVA indicates that the model is highly significant through the F value of 48.91. There is only a 0.01% chance that a "model F value" could occur due to noise. Meanwhile, the p value of 260 the model which is at p <0.0001 also implies that the model is highly significant. The lack of fit 261 value of 3.8065 confirms that the lack of fit is not significant relative to the pure error when p=262 0.057729, which is > 0.05. The insignificant lack of fit indicates good predictability. The "R-263 squared" of 0.97717 is in reasonable agreement with the "Adj R-squared" of 0.95719 which also 264 indicates good predictability. Based on data analysis (Table 5), an empirical second order 265 266 polynomial equation was obtained, which in terms of actual factors is as follows:

267

268
$$y_{AO} = -91 + 32x_1 - 3x_1^2 + 7730x_2 - 203042x_2^2 - 2x_3^2 - 0.0549x_4^2 + 0.264x_1x_4 - 32x_2x_4$$
 (7)

Where, y is the percentage removal of AO (%), X_1 , X_2 , X_3 and X_4 are terms for the coded values 270 of pH, adsorbent, sonication time and concentration dye, respectively. (Table 4 according to 271 there should P value of 0.05, it was revealed that pH (X_1) , adsorbent dosage (X_2) , sonication time 272 (X_3) and concentration dye (X_4) , quadratic pH (X_1^2) , adsorbent dosage (X_2^2) , sonication time (273 X_3^2) and concentration dye (X_4^2) and pH× initial AO concentration (X_1X_4) and adsorbent× initial 274 AO concentration (X_2X_4) are significant model terms. The plot of experimental versus calculated 275 values of removal (%) indicate a good fit (as Fig. 4) and presence of linear relationship between 276 them with high correlation coefficient that indicates normal distribution of error around the mean 277 and good applicability of model for explanation of experimental data. These plots are required to 278 check the normality assumption in fitted model. 279

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- 281

3.2.2 Three-dimensional response surface plots

282 The RSM correspond to CCD model was depicted and considered to optimize the critical factors and describe the nature of the response surface in the experiment. The curvature natures 283 of Fig. 5 show the response surface plots of removal (%) confirm strong interaction between the 284 variables. 285

The three-dimensional response surface plots were used to assigning the interaction between the 286 four variables. The relative impresses of two tested variables on the adsorption efficiency, while 287 maintaining all other variables at fixed levels were illustrated in Fig. 5. Based on the quadratic 288 model, the three-dimensional response surface plots were organized. The optimum situations of 289 the relative variables will resemble with the coordinates of the central point in the upmost level 290 in each of these figures. 291

Fig. 5(a) shows the combined effect of adsorbent and pH on adsorption of AO on ZnS: Cu–NP– AC at constant initial concentration of AO (20 mg/L). It is evident from the figure that pH has a profound effect on removal of AO. With increase in pH, removal of AO decreases at lower amount of ZnS: Cu–NP–AC but at higher amount of ZnS: Cu–NP–AC, removal is almost constant with pH. At constant pH, removal increases with increase in amount of ZnS: Cu–NP– AC. It is quite obvious that when amount of adsorbent is high then removal is also high because available surface area for adsorption is much more.

Fig. 5(b) shows response surface plot of the adsorption efficiency as dependent on pH 299 300 and the sonication time. It seems necessary to mention the surface charge of ZnS: Cu-NP-AC in the pH area under pH_{ZPC} is positive, because of the more H^+ in the solution, and it helps the 301 removal of anionic compounds, but in the pH area over pH_{ZPC}, the ZnS: Cu-NP-AC surface 302 charge is negative due to the presence of OH⁻ in the solution, and it helps in the removal of 303 cationic compounds (pH area is considered as pH_{ZPC} in spaces, where the catalyst surface charge 304 is zero.). In a low pH, ZnS: Cu-NP-AC has the positive surface charge and adsorbs the 305 compounds with the negative charges like anionic dyes, but when the solution pH is over pH_{ZPC} , 306 the oxide surface gets the negative charge and can make a complex with cationic compounds. So, 307 according to these considerations, the basic conditions are more ideal for AO dye adsorption 308 because AO dye is a cationic and cationic dye. In basic pH, adsorption happens along with the 309 reduction of dye molecules by ZnS: Cu-NP-AC and in basic pH, adsorption may happen 310 311 through adsorbing the dye molecules on the ZnS: Cu-NP-AC form.

Fig. 5(c) shows the three dimensional response surfaces of the combined effect of initial concentration and pH on percentage removal of dye at constant weight of ZnS: Cu–NP–AC (0.02 g). It is evident from the figure that removal of AO decreases when initial concentration

increases. Removal attains its maximum value when pH is low and initial concentration is alsolow.

Fig. 5(d) clearly states that as the adsorbent amount and sonication time increases, the adsorption efficiency improves. It could be explained by the fact that the more amounts of ZnS: Cu–NP–AC cause an increase in the adsorbent surface and the active surfaces prepare some spaces for capturing AO dye molecules and increasing these spaces makes the dye get out faster.

To study the impact of the dye initial concentration on the adsorption efficiency, some experiments with concentrations (5–25 mg/L) of AO dye and fixed sonication time (2min) and pH (7) were designed and results displayed in Fig. 5(e). The observed decreases in removal percentage at higher initial arrived from lower ratio of vacant sites to candidate dyes molecular that compete for binding to the surface. However at such situation there are not enough spaces for all molecules in high concentration of dye.

The effect of initial AO concentration on its removal percentage and its influence on their 327 factors were shown Fig. 5(f). It was seen that in despite of the increase in the amount of dye 328 uptake, its removal efficiency was decreased. At lower dye concentrations, the ratio of solute 329 concentrations to vacant reactive adsorbent sites is lower and accelerates dye adsorption which 330 causes an increase in dye removal. At higher concentrations, lower adsorption yield is due to the 331 saturation of adsorption sites. On the other hand, the percentage removal of dye was higher at 332 lower initial dye concentrations and smaller at higher initial concentrations, which clearly 333 334 indicate that the adsorption of AO from aqueous solution depend on its initial concentration.

335 3.3. Optimization of CCD by DF for extraction procedure

The profile for desirable option with predicted values in the STATISTICA 10.0 software was used for the optimization of the process (Fig. 6). The profile for desirable responses was

chosen after specifying the DF for each dependent variable (removal percentage) by assigning predicted values. The scale in the range of 0.0 (undesirable) to 1.0 (very desirable) was used to obtain a global function (D) that its maximum (99.764%) and minimum (60.781%) value concern to JSB adsorption was achieved in this research.

Three solutions with different amounts of ideal conditions were used to predict the optimum 342 conditions for AO dye adsorption onto ZnS: Cu-NP-AC (Table 6). The highest removal percent 343 (99, 8) achieved in the experiment number 5, compared to the other two experiments. Optimum 344 parameters of the reaction are achieved at pH (7), (0.02 g) of ZnS: Cu-NP-AC, 3 min sonication 345 time and initial concentration (20 mg L^{-1}). The relative deviation coefficient 0.322% concern to 346 RSM experimental design show good agreement and high correlation between actual and 347 predicted amounts and reveal the suitability of empirical model resulted from the design could be 348 used for well describing the relation between factors and the AO dye removal percentage. 349

350

351 3.4. Adsorption equilibrium study

Adsorption properties and equilibrium parameters of each isotherm model indicate the interaction of adsorbent–adsorbate and give comprehensive information about the nature of interaction [60–65]. The widely used isotherm models such as Langmuir, Freundlich, Dubinin and Radushkevich (D–R) and Temkin were used to analyze the experimental equilibrium data obtained from the sorption process at room temperature over the concentration range of 5–30 mg L^{-1} .

In the Langmuir isotherm [66], the intermolecular forces decrease rapidly with distance and the predicted monolayer coverage of the adsorbate on the outer surface of the adsorbent is represented in linear form as follows:

362
$$\frac{C_{e}}{q_{e}} = \frac{1}{Q_{o}k} + \frac{C_{e}}{Q_{o}}$$
 (8)

363

A plot $1/q_e$ versus 1/Ce should represent a line with slope of $1/Ka \ Q_m$ and $1/Q_m$ intercept of $1/Q_m$ and respective data are presented in Table 7. The high correlation ($R^2>0.999$) coefficient shows that Langmuir isotherms are applicable for the interpretation of AO adsorption onto ZnS: Cu– NP–AC over the whole concentration range studies and maximum adsorption capacity of 92.26 – 183.15 mgg⁻¹.

369

370 The data was analyzed by the linearized form of Freundlich isotherm model:

$$371 \qquad \ln q_e = \ln K_F + \frac{1}{n} \ln C_e \tag{9}$$

Where q_e is the amount of adsorption, k_f is the Freundlich constant related to sorption 372 capacity and 1/n is a constant related to energy or intensity of adsorption. This gives an 373 expression encompassing the surface heterogeneity and the exponential distribution of 374 activated sites and their energies. This isotherm dose not predicts any saturation of the 375 adsorbent surface. The Freundlich exponent's k_f and 1/n can be determined from the linear plot 376 of log q_e vs. log $C_e\!.$ The values of the Freundlich constants K_f and 1/n are 4.982-5.559 and 377 0.4544-1.130 respectively shown in Table 7. The slope 1/n ranging between 0 and 1 is a 378 379 measure of adsorption intensity or surface heterogeneous, becoming more heterogeneous as its value gets closer to zero [67]. 380

Table 7 shows that Langmuir model fits for AO dye (correlation coefficient 0.990-0.998)
adsorption process. Heat of adsorption and the adsorbent-adsorbate interaction on adsorption

isotherms were studied by Temkin [68], its equation is given as:

$$q_e = B LnK_T + B LnC_e$$
(10)

where $B_T = RT/b$, T is the absolute temperature in K, R the universal gas constant, 8.314 J mol⁻¹ K⁻¹, K_T the equilibrium binding constant (L mg⁻¹) and B is related to the heat of

adsorption. The constants obtained for Temkin isotherm are shown in Table 7.

388 The linear form of Dubinin-Radushkevich isotherm equation can be expressed as [69].

$$lnq_e = lnQ_s - B\varepsilon^2$$
(11)

The plot of $\ln q_e vs. \varepsilon^2$ at different temperatures for AO is presented. The constant obtained for D–R isotherms are shown in Table (7). The mean adsorption energy (E) gives information about chemical and physical nature of adsorption [70].

The values of the parameters of the three isotherms and their related correlation coefficients are shown in Table 7, the Langmuir model yields a somewhat better fit ($R^2 = 0.990-0.998$), Temkin isotherm ($R^2=0.925-0.986$) than the Freundlich model ($R^2 = 0.787-0.910$) and Dubinin-Radushkevich model ($R^2 = 0.913-0.984$). Equilibrium data fitted well with the Langmuir model.

398 3.5. Kinetic study

Adsorption of a solute by a solid in aqueous solution usually occurs with a complex kinetics [71]. The adsorption rate is strongly influenced by several parameters related to the state of the solid (generally with very heterogeneous reactive surface) and to physico-chemical conditions under which the adsorption is occurred. To investigate the adsorption processes of AO on the adsorbent, pseudo-first-order and pseudo-second-order adsorption were studied. The

404 Lagergren pseudo-first-order model described the adsorption kinetic data [72]. The Lagergren405 equation is commonly expressed as follows:

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407
$$\frac{dq_t}{dt} = k_1(q_e - q_t)$$
 (11)

408

Where q_e and q_t (mg/g) are the adsorption capacities, at equilibrium and at time t respectively. k_1 is the rate constant of the pseudo-first-order adsorption (L/min). From the following model, plot of log (q_e - q_t) versus t was made and the values of k_1 and q_e were determined by using the slope and intercept of the plot, respectively.

413

414
$$\log(q_e - q_t) = \log q_e - (\frac{k_1}{2.303})t$$
 (12)

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415

The fact that the intercept is not equal to ge, means that the reaction unlikely follows the first-416 order, regardless of the value of correlation coefficient [73]. The variation in rate should be 417 proportional to first power of concentration for strict surface adsorption. However, the 418 relationship between initial solute concentration and rate of adsorption is linear when pore 419 diffusion limits the adsorption process. Furthermore, the correlation coefficient, R² is relatively 420 low for most adsorption data (See Table 8). This indicates that the adsorption of AO onto 421 422 ZnS:Cu -NP-AC was not a first-order reaction. Therefore, it is necessary to fit experimental data 423 to another model. The adsorption kinetic may be described by the pseudo-second order model 424 [74], which is generally given by the following equation:

426
$$\frac{dq_t}{dt} = k_2 (q_e - q_t)^2$$
 (13)

428 Eq. (13) is integrated over the interval 0 to t for t and 0 to q_t for q_t , to give

429

430
$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{t}{q_e}$$
 (14)

As mentioned above, the plot of log (ge- gt) versus t does not show good results for entire 431 sorption period, while the plot of t/qt versus t shows a straight line. The values of k_2 and 432 equilibrium adsorption capacity (ge) were calculated from the intercept and slope of the plot of 433 t/qt versus t (Table 8). For all concentrations and sorbent doses, the calculated q_e values were 434 mainly close to the experimental data and R^2 values for the pseudo-second-order kinetic model 435 were found to be larger than that for the pseudo-first-order kinetic. This indicates that the 436 pseudo-second-order kinetic model applies better for the adsorption of AO dye for the entire 437 sorption period. The intraparticle diffusion equation is given as [75]: 438

439

440
$$q_t = K_{dif} t^{1/2} + C$$
 (15)

441

Where K_F is the intraparticle diffusion rate constant (mg (gmin^{1/2})⁻¹) and C shows the boundary layer thickness. The linear form of Elovich model equation is generally expressed as [76]:

445
$$q_{t} = \frac{1}{\beta} \ln(\alpha\beta) + \frac{1}{\beta} \ln t$$
(16)

The kinetic data from pseudo-first and pseudo-second-order adsorption kinetic models and the intraparticle diffusion and Elovich model are given in Table 8. The linear plots of t/qt versus t indicated a good agreement between the experimental and calculated q_e values for different initial dye concentrations. Furthermore, the correlation coefficients of the pseudo-second-order kinetic model ($R^2 \ge 0.999$) were greater than that of the pseudo-first-order model ($R^2 \le 0.983$). As a result, it can be said that the adsorption fits to the pseudo-second-order better than the pseudofirst-order kinetic model.

454

455 3.6. Comparison with other methods

The ZnS: Cu-NP-AC prepared in this work had a relatively large adsorption capacity on 456 AO compared to some other adsorbents reported in the literature, primarily when ZnS: Cu loaded 457 on AC, it lead to rapid decrease in contact time parameter, it is reported that time taken by 458 activated carbon is about 120 minutes [77], time taken by graphite loaded with titania is about 459 300 minutes [78], but when AC loaded with ZnS:Cu nanoparticle time taken by the adsorbent 460 rapidly decrease and it takes nearly 3 minutes to adsorb the maximum amount of Auramine -O. 461 Secondly the maximum amount of the adsorbate adsorbed on the adsorbent also increases form 462 1.509 mg/g (for AC) it becomes 94.26 mg/g (for ZnS:Cu nanoparticles loaded on AC) 463

Table 9 lists the comparison of maximum monolayer adsorption capacity of AO on various adsorbents. The adsorption capacity and contact time for proposed method in comparison with all of the adsorbents are preferable and superior to the literature which shows satisfactory removal performance for AO as compared to other reported adsorbents [77- 80]. The results indicated that the ultrasound assisted removal method has a remarkable ability to improve the removal efficiency of dyes. The ultrasonic-assisted enhancement of removal could be attributed to the high-pressure shock waves and high-speed microjets during the violent collapse ofcavitation bubbles [81, 82].

472 **4. Conclusion**

The results of the present studies showed that ZnS: Cu-NP-AC prepared by synthesis 473 method was an efficient adsorbent for the Auramine-O removal. The use of response surface 474 methodology involving central composite design for optimization of process parameters was 475 studied. Experiments were performed as a function of initial pH, dve concentration, sonication 476 time and absorbent dosage, these factors are well studied and optimized. The optimized values, 477 at which the highest removal percent (99.76%) was attained, are achieved: pH=7, initial 478 concentration dye (20 mgL⁻¹), sonication time (3 min) and adsorbent amount (0.02 g). In the 479 present study, the analytical utility of experimental design for evaluation of optimum condition 480 for the removal of AO in aqueous solution by ZnS: Cu-NP-AC coupled with ultrasound assisted 481 adsorption method has been investigated. The equilibrium and kinetic studies were investigated 482 for the adsorption process. The isotherm models such as Langmuir, Freundlich, Temkin, and 483 Dubinin-Radushkevich were evaluated and the equilibrium data were best described by the 484 Langmuir model. The high correlation ($R^2 > 0.999$) coefficient shows that Langmuir isotherms are 485 applicable for the interpretation of AO adsorption onto ZnS: Cu-NP-AC over the whole 486 concentration range studies and maximum adsorption capacity of $92.26 - 183.15 \text{ mgg}^{-1}$. The 487 process kinetics can be successfully fitted to the pseudo-second-order kinetic model. 488

489

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621 **Figure captions:**

- **Fig. 1**. UV-vis absorbance spectrum of the prepared ZnS:Cu–NP-AC (inset: Plot of $(\alpha hv)^2$ vs.
- 623 (hv) for the prepared ZnS:Cu-NP-AC)
- **Fig. 2**. FESEM images of the prepared ZnS:Cu–NPs-AC.
- **Fig. 3.** XRD pattern for prepared ZnS: Cu-AC-NP. (H: hexagonal and C: cubic)
- **Fig. 4.** The experimental data vs predicted data for removal of AO.
- **Fig. 5.** Response surfaces for the AO removal: (a) X_1-X_2 ; (b) X_1-X_3 (c) X_1-X_4 (d) X_2-X_3 ; (e) X_3-X_3 ; (e)
- 628 X_4 and (f) X_3 - X_4 .
- **Fig. 6**. Profiles for predicated values and desirability function for removal percentage of AO.
- 630 Dashed line indicated current values after optimization.
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635 Properties of the Auramine-O

Color index number	41000
CAS number	2465-27-2
Chemical Formula	C ₁₇ H ₂₁ N ₃ .HCl
Molecular weight (g mol ⁻¹)	303.83
Maximum wavelength(λ_{max}), nm	434
	H_3C_N H_3C_N CH_3 CH_3 CH_3 CH_3 CH_3
Molecular Structure	
Absorption spectrum	$\left(\begin{array}{c} 1.8\\ 1.6\\ 1.4\\ 1.2\\ 0.8\\ 0.6\\ 0.6\\ 0.4\\ 0.2\\ 0\\ 300\\ 350\\ 400\\ 450\\ 400\\ 450\\ 100\\ 550\\ 600\\ 650\\ 700\\ 550\\ 700\\ 550\\ 700\\ 700\\ 700\\ 7$
Type of dye	Basic Yellow (Cationic)
Use	paper mills, textile mills, leather and carpet industry

643	Design matrix	for the	central	composite	designs.
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Factors	Levels						
	Low (-1)	Central(0)	High(+1)	-α	$+\alpha$		
X ₁ : Ph	6	7	8	5	9		
X ₂ : Adsorbent (g)	0.01	0.015	0.02	0.005	0.025		
X ₃ : time (min)	2	3	4	1	5		
X_3 : AO concentration (mg L ⁻¹)	15	20	25	10	30		
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660 Design matrix for the CCD

					%Removal Auramine-O		
Run	X ₁	X ₂	X ₃	X ₄	Observed ^a	Predictd ^b	
1	7	0.015	3	10	94.98000	94.5652	
2	6	0.020	2	15	96.78000	97.5729	
3	7	0.005	3	20	60.78000	61.5518	
4	7	0.025	3	20	96.78000	93.3768	
5	8	0.010	2	25	73.97000	73.4753	
6	6	0.010	4	25	81.56000	82.5158	
7	9	0.015	3	20	81.89000	81.0778	
8	6	0.020	4	25	95.80000	96.9945	
9(C)	7	0.015	3	20	98.50000	97.7686	
10(C)	7	0.015	3	20	98.67000	97.7686	
11	6	0.020	4	15	99.76000	102.584	
12	8	0.020	2	25	85.36000	87.6090	
13	8	0.010	4	15	75.93000	77.8486	
14	8	0.020	4	25	94.56000	96.3253	
15	5	0.015	3	20	92.42300	90.6038	
16(C)	7	0.015	3	20	98.10000	97.7686	
17	8	0.010	2	15	73.35000	72.4574	
18(C)	7	0.015	3	20	97.60000	97.7686	
19	6	0.010	2	15	80.75000	81.3141	

20	6	0.020	2	25	89.68000	90.0908
21(C)	7	0.015	3	20	97.98000	97.7686
22(C)	7	0.015	1	20	86.67000	85.2468
23	8	0.020	2	15	88.43000	89.8036
24(C)	7	0.015	3	20	95.08000	97.7686
25	7	0.015	3	30	92.21000	89.9935
26	7	0.015	3	20	98.45000	97.7686
27	6	0.010	4	15	86.84000	84.8929
28	8	0.020	4	15	97.80000	96.6274
29	8	0.010	4	25	81.25000	80.7590
30	7	0.015	5	20	98.75000	97.5418
31	6	0.010	2	25	75.57000	77.0445

661 C: Center point

662 ^a Experimental values of response.

^b Predicted values of response by RSM proposed model.

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674 Analysis of variance (ANOVA) for CCD (AO)

Source of variation	Sum of square	Dfª	Mean square	F-value	P-value	
X ₁	136.117	1	136.117	88.5478	0.000082	*
X_1^2	254.271	1	254.271	165.4105	0.000014	*
X ₂	1519.246	1	1519.246	988.3111	0.000000	*
X_2^2	736.809	1	736.809	479.3143	0.000001	*
X_3	226.751	1	226.751	147.5074	0.000019	*
X_{3}^{2}	72.617	1	72.617	47.2393	0.000468	*
X_4	31.350	1	31.350	20.3942	0.004035	*
X_4^2	53.852	1	53.852	35.0325	0.001036	*
$X_1 X_2$	1.183	1	1.183	0.7694	0.414155	**
X_1X_3	3.285	1	3.285	2.1371	0.194089	**
X_1X_4	27.958	1	27.958	18.1872	0.005294	*
X ₂ X ₃	2.052	1	2.052	1.3349	0.291861	**
X_2X_4	10.320	1	10.320	6.7135	0.041154	*
X_3X_4	3.582	1	3.582	2.3299	0.177762	**
Lack-of-Fit	58.514	10	5.851	3.8065	0.057729	**
Pure Error	9.223	6	1.537			
Total	2966.557	30	136.117			

675 ^aDegree of freedom * Significant

** Not significant

677 Regression coefficients (AO)

Factor	Regressn	Std.Err.	Т	Р
Mean/Interc.	-91	19.151	-4.7349	0.003208
\mathbf{X}_{1}	32	3.724	8.5673	0.000139
\mathbf{X}_1^2	-3	0.232	-12.8612	0.000014
X ₂	7730	603.619	12.8053	0.000014
X_2^2	-203042	9274.201	-21.8932	0.000001
X ₃	6	3.018	2.1525	0.074851
X_3^2	-2	0.232	-6.8731	0.000468
X_4	0	0.631	0.4987	0.635739
X_4^2	-0	0.009	-5.9188	0.001036
$X_1 X_2$	54	61.992	0.8771	0.414155
X_1X_3	0	0.310	1.4619	0.194089
X_1X_4	0	0.062	4.2646	0.005294
X ₂ X ₃	72	61.992	1.1554	0.291861
X_2X_4	-32	12.398	-2.5911	0.041154
$X_{3}X_{4}$	0	0.062	1.5264	0.177762

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				Optimal condition	IS	Removal%			
	Exp.	pН	Adsorbent	Sonication time	AO Concentration	Predicted	Predicted	RSE%	
			(g)	(min)	$(mg L^{-1})$	value	value		
	1	7.0	0.02	3.0	20	99.671	100	0.329	
	2	7.0	0.02	3.0	20	99.542	100	0.458	
	3	7.0	0.02	3.0	20	98.877	100	1.123	
	4	7.0	0.02	3.0	20	99.678	100	0.322	
	5	7.0	0.02	3.0	20	98.321	100	1.679	
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682 Optimum conditions derived by RSM design for dye removal.

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Page 37 of 47

696 **Table 7**

697 Various isotherm constants and correlation coefficients calculated for the adsorption of

698 Auramine-O onto ZnS: Cu-NP-AC.

		Value of parameters			
Isotherm	Parameters	0.01 g	0.015 g	0.02 g	
	Qm (mg.g ⁻¹)	183.15	122.54	94.26	
Langmuir	Ka (L mg ⁻¹)	0.402	0.8111	1.102	
	R ²	0.990	0.994	0.998	
	1/n	0.4544	0.4121	0.382	
Freundlich	$K_{\rm F}$ (L mg ⁻¹)	5.559	5.318	4.982	
	R ²	0.864	0.787	0.910	
	B_1	59.697	23.884	16.69	
Temkin	$K_T(L mg^{-1})$	5.0964	10.665	18.56	
	R^2	0.984	0.925	0.986	
	$Qs (mg g^{-1})$	138.29	103.99	71.95	
Dubinin-Radushkevich	B×10 ⁻⁷	2.23	1.4	1.00	
	Е	1497	1889	2236	
	R^2	0.960	0.984	0.913	

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705	Kinetic p	arameters	for the	adsorption	of Aura	amine-O	onto 2	ZnS:	Cu-NP	-AC	adsorben	ts.
										-		

		Value of parameters						
		0.01 g			0.02 g			
Model	Parameters	10	20	30	10	20	30	
		mg L ⁻¹	mg L ⁻¹	mg L ⁻¹	mg L ⁻¹	mg L ⁻¹	mg L ⁻¹	
	$k_1(\min^{-1})$	0.0145	0.0131	0.0175	0.016	0.0135	0.024	
First-order kinetic	q_e (calc) (mg g ⁻¹)	3.907	5.844	8.385	7.23	11.716	25.14	
	R ²	0.969	0.983	0.922	0.972	0.863	0.965	
Pseudo-second-	$k_2(\min^{-1})$	0.001	0.0003	0.00018	0.004	0.0021	0.002	
order kinetic	$q_e(calc) (mg g^{-1})$	49.75	102.04	147.06	25.12	48.309	65.56	
	R ²	0.999	0.998	0.997	0.999	0.999	0.999	
Intraparticle	$K_{diff}(mg g^{-1} min^{-1/2})$	1.363	2.958	4.822	0.372	0.758	0.484	
diffusion	$C (mg g^{-1})$	28.883	44.77	52.579	18.62	34.935	59.12	
	R^2	0.904	0.960	0.936	0.817	0.730	0.832	
	β (g mg ⁻¹)	0.159	0.0804	0.0545	0.475	0.544	0.382	
Elovich	$\alpha (mg g^{-1} min^{-1})$	42.42	84.93	85.77	95.6	2899	28976	
	R^2	0.976	0.976	0.957	0.923	0.940	0.930	
	$q_e(exp) (mg g^{-1})$	46.978	92.64	129.53	24.39	46.597	66.14	

708 Comparison for the removal of Auramine-O by different methods and adsorbents.

Adsorbent	adsorbent	dye	Concentration	Contact time	References
	dosage (g)		$(mg L^{-1})$	(min)	
Bagasse Fly Ash (BFA)	0.001	AO	10	30-240	[77]
Activated Carbon-Commercial	0.020	AO	10	120-240	[77]
grade (ACC)					
Activated Carbon-Laboratory	0.002	AO	10	120-240	[77]
grade (ACL)					
Graphite Loaded with Titania	0.050	AO	50	300	[78]
Ag-NPs- AC	8	MB	2	16	[79]
Ag-NPs-AC	8	CR	2	6	[80]
Au-NPs-AC	10	CR	2	5	[80]
ZnS:Cu-NP-AC	0.020	AO	20	3	Proposed
					method

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Fig. 2. FESEM images of the prepared ZnS:Cu–NPs-AC.

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Fig. 5. Response surfaces for the AO removal: (a) X_1-X_2 ; (b) X_1-X_3 (c) X_1-X_4 (d) X_2-X_3 ; (e) X_2-X_3 ; (e) X_2-X_3 ; (e) X_2-X_3 ; (e) X_3-X_3 ; (e) X_3-X_3 ; (e) X_3-X_3 ; (e) X_3-X_3 ; (f) X_3-X_3 ; (f)











