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# Amorphous $Fe^{2+}$ -rich $FeO_x$ loaded in mesoporous silica as a highly efficient heterogeneous Fenton catalyst

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A simple physical-vapor-infiltration (PVI) method using ferrocene as iron source, has been developed to load  $FeO_x$  into pore channels of mesoporous silica SBA-15. The obtained  $FeO_x/SBA-15$  composite has a high loading amount of  $FeO_x$  (e.g. 26.64 wt% Fe content obtained at PVI duration 17 h and calcination

- <sup>10</sup> temperature 450 °C) but unblocked pore channels thanks to the unique preparation strategy. The FeO<sub>x</sub> species are amorphous, rich of Fe<sup>2+</sup> and have been highly dispersed as a nanocoating onto the pore channel surface. The FeO<sub>x</sub>/SBA-15 were used as heterogeneous Fenton catalysts to degrade Acid orange 7 (AO7). They showed high catalytic activity and degradation efficiency, which was attributed to the high proportion of Fe<sup>2+</sup> in amorphous FeO<sub>x</sub> and their favorable adsorption capability for the dye. The <sup>15</sup> influences of PVI duration, calcination temperature and Fenton reaction conditions (FeO<sub>x</sub>/SBA-15
- dosages,  $H_2O_2$  dosages and initial pH value) on their catalytic activity were investigated in detail.

#### Introduction

Numerous dyes are being produced in various chemical industries, especially in textile industry, in large scale every day. The heavy

- <sup>20</sup> emissions of dyeing wastewater cause serious water pollution and ecological deterioration. Even a very small amount of dyes of as low as 1 ppm in concentrations can result in distinct color change of water <sup>1</sup>. More terribly, these dyeing wastewaters are considered to be carcinogenic, mutagenic and toxic threats to human being
- <sup>25</sup> and aquatic lives <sup>2</sup>. However, many biologically toxic molecules are rather environmentally stable and cannot be well-degraded by commonly used physical, chemical and biological methods for water treatments <sup>3</sup>. Advanced oxidation process (AOP) which relies on highly oxidizing HO• (oxidation potential  $E_0 = 2.80 V^4$ )
- <sup>30</sup> to non-selectively degrade these biologically toxic or nondegradable organic contaminants into CO<sub>2</sub> or biodegradable small organic molecules <sup>5</sup> is the most promising strategy for wastewater treatment.

Fenton reaction, based on  $H_2O_2$  and  $Fe^{2+}$ , is the most common <sup>35</sup> AOP for its mild reaction conditions, convenient generation of HO• and satisfying degradation of contaminants without toxic byproducts <sup>6</sup>. However, in homogeneous Fenton reactions, the difficulty in separating or recovering/recycling Fe catalysts from water and the narrow usable pH range are main obstacles for

- <sup>40</sup> practical use <sup>7</sup>. Therefore, heterogeneous Fenton reaction is a promising process because of its long-term activity, easy recovery/recycling of catalysts from water, and no iron hydroxide precipitation <sup>8</sup>, <sup>9</sup>. In general, there are three types of heterogeneous Fenton systems, that is, natural iron minerals <sup>10</sup>, <sup>45</sup> iron-oxides (Fe@Fe<sub>2</sub>O<sub>3</sub> <sup>11</sup>, Fe<sub>3</sub>O<sub>4</sub> <sup>12</sup> etc.) and Fe-immobilized
- porous materials (such as clay <sup>13, 14</sup>, mesoporous silica <sup>15, 16</sup>,

zeolite <sup>17, 18</sup> and carbon <sup>19, 20</sup> etc.).

In Fenton reaction, the reaction rate constant between  $Fe^{2+}$ and  $H_2O_2$  (K = 70 M<sup>-1</sup>·s<sup>-1</sup>)<sup>21</sup> is remarkably larger than that <sup>50</sup> between  $Fe^{3+}$  and  $H_2O_2$  (K = 0.0001 ~ 0.01 M<sup>-1</sup>·s<sup>-1</sup>)<sup>22</sup>. Therefore, the proportion of  $Fe^{2+}$  in Fenton catalysts is a key point in determining their catalytic activity <sup>23, 24</sup>. Fortunately, amorphous  $FeO_x$  usually contains large proportions of  $Fe^{2+}$  due to its low crystallization and large amount of defects. Furthermore, <sup>55</sup> amorphous  $FeO_x$  generally has higher surface area, more active sites than iron oxide crystals <sup>25</sup> caused by its higher surface-bulk ratio and more "dangling bonds" <sup>26</sup>. All these are believed to endow amorphous  $FeO_x$  with high catalytic activity in Fenton reaction.

<sup>60</sup> Ferric oxide-immobilized porous catalysts have been extensively used in Fenton reaction because of their large pore volume and high surface area which are believed to improve their catalytic activity. Nevertheless, the wet impregnation method commonly used to incorporate Fe-based catalysts into porous <sup>65</sup> matrix usually leads to aggregation of iron oxide particles and blockage of pore channels <sup>27</sup>, thus resulting in generation of large amount of inaccessible active sites and consequently low catalytic activity in Fenton reaction.

Herein, a simple physical-vapor-infiltration (PVI) method has <sup>70</sup> been developed to incorporate extraordinarily large amount of amorphous  $FeO_x$  without blocking the pore channels of mesoporous silica SBA-15. As heterogeneous Fenton catalysts, the obtained  $FeO_x/SBA-15$  showed high catalytic activity in degrading Acid orange 7 (AO7).

#### 75 **Experimental**

PVI incorporation of amorphous FeO<sub>x</sub> into mesoporous silica

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**SBA-15.** Mesoporous silica SBA-15 was synthesized according to a procedure published elsewhere <sup>28</sup>. 1 g ferrocene and 0.2 g calcined SBA-15 were together put into a tightly sealed jar. Then the PVI process was carried out at 110 °C for different duration

- s time. The obtained yellow brown powders were finally calcined at 400, 450, 500, 550 and 600 °C in air, respectively. The obtained composites were denoted as FeO<sub>x</sub>/SBA-nh-T, where nh and T represent the PVI duration time and calcination temperature, respectively.
- <sup>10</sup> **Characterizations.** Powder X-ray diffraction (XRD) patterns were recorded on a Rigaku D/MAX-2250 V D/Max 2200PC diffractometer using Cu-Kalpha radiation (40 kV and 40 mA). The scanning rate was 4 °·min<sup>-1</sup> and 0.6 °·min<sup>-1</sup> for the wide angle and small angle XRD measurements, respectively. Field Emission
- <sup>15</sup> Transmission Electron Microscopy (FETEM) analysis was conducted with a JEOL 200CX electron microscope operated at 200 kV. Energy dispersive X-ray (EDX) spectrum was obtained from an attached Oxford Link ISIS energy-dispersive spectrometer fixed on a JEM-2010 electron microscope operated
- <sup>20</sup> at 200 kV. Scanning transmission electron microscopic (STEM) were obtained on a JEOL-2010F electron microscope operated at 200 kV. Nitrogen adsorption/desorption isotherms at 77 K were measured on a Micromeritics TriStar 3000 instrument. All samples were outgassed at 150 °C for 12 h under flowing N<sub>2</sub>
- <sup>25</sup> before the measurement. The specific surface areas were calculated by the Brunauer–Emmet–Teller (BET) methods. X-ray photoelectron spectroscopy (XPS) signals were collected on VG Micro MKII instrument using monochromatic Mg Kalpha X-ray at 1253.6 eV operated at 120 W. All the element binding energies
- <sup>30</sup> were referenced to the C (1s) line situated at 284.6 eV. Fourier Transform Infrared (FTIR) spectra were obtained in the range of 1000-2000 cm<sup>-1</sup> with a Nicolet iS10 FTIR spectrometer of a 4 cm<sup>-1</sup> resolution with powders dispersed in KBr. UV-vis absorption spectra were conducted using a UV-3101PC instrument
- <sup>35</sup> manufactured by Shimadzu Corporation. The Fe content and the leaching of Fe concentration of catalysts was evaluated by measuring the Fe concentration in the supernatant after Fenton reaction with ICP-OES (Vista AX). Total organic carbon (TOC) was measured using a Shimadzu TOC-Vcpn analyzer.
- <sup>40</sup> Adsorption performance and catalytic activity of FeO<sub>x</sub>/SBA-15 in Fenton reaction. Acid orange 7 (AO7) was used as a probe dye to evaluate the catalytic activity of FeO<sub>x</sub>/SBA-15. The AO7 concentrations were identified by the UV-vis spectrophotometer at 484 nm. All adsorption tests and
- <sup>45</sup> degradation tests of AO7 by FeO<sub>x</sub>/SBA-15 suspension were carried out in dark. Typically, 25 mL 100 mg·L<sup>-1</sup> AO7 solution was magnetically stirred at ambient temperature and the desired pH values were adjusted using HCl solution or NaOH solution. Then 15 mg FeO<sub>x</sub>/SBA-17-450 was added. For the adsorption
- <sup>50</sup> kinetics studies, concentrations of AO7 at different time were measured on UV-vis spectrophotometer. The adsorption isotherms were carried out at the same conditions with different initial concentrations of AO7 between 50 and 500 mg·L<sup>-1</sup>. For Fenton reaction, 15 mg FeO<sub>x</sub>/SBA-15 was added into the AO7
- ss solution of 100 mg·L<sup>-1</sup> and the mixture was stirred for 2 h in dark to reach the adsorption equilibrium. The heterogeneous Fenton reaction was initiated by the addition of an appropriate amount of H<sub>2</sub>O<sub>2</sub> (30 wt%). 3.0 mL reaction suspension was taken out by 115

injection syringe and filtered through a 0.22 μm nylon membrane.
 The AO7 concentrations in filtrates were analyzed by a UV-vis spectrophotometer. The degradation efficiency (%) of dye was calculated by the following equation:

Degradation efficiency =  $(C_0 - C) \times 100\% / C_0$  (1) where  $C_0$  and C stand for the initial and instant concentrations at

 $_{65}$  different reaction time of AO7, respectively.

#### **Results and discussion**

Characterizations of FeO<sub>x</sub>/SBA-15



**Fig. 1** TEM images (A, B) of the sample FeO<sub>x</sub>/SBA-17h-450 and the EDX spectrum (C) collected from the circled area shown in (A).

Fig. 1A and B show the TEM images of the sample FeO<sub>x</sub>/SBA-<sup>90</sup> 17h-450 taken with electron beams being perpendicular and parallel to the mesopore channels of SBA-15, respectively. It can be seen that the FeO<sub>x</sub>/SBA-17h-450 has highly ordered hexahedral mesoporous structure as same as SBA-15 <sup>28</sup>. The Fe content in FeO<sub>x</sub>/SBA-17h-450 obtained by EDX spectrum (Fig. <sup>95</sup> 1C) is as high as 26.64 wt%. Surprisingly, there are no observable particle aggregations both inside and outside of its pore channels, demonstrating the thoroughly uniform dispersion of FeO<sub>x</sub> species in the pore channels of the substrate, which ensures their high accessibility for reactants during catalytic reaction. This can be <sup>100</sup> ascribed to the fact that it was sublimated metal-organic precursor, which can easily penetrate deep into the pore channels of SBA-15, that the PVI strategy used, to get rid of particle aggregates resulting from solvent surface tension <sup>29</sup>.



**Fig. 2** STEM images of the FeO<sub>x</sub>/SBA-17h-450 and the corresponding element mappings of the sample: D (Si, Fe, O), (E) Si, (F) Fe, (G) O.

Fig. 2 (A, B, C) show secondary electron (SE) image, highangle annular dark-field (HAADF) image, bright filed (BF) image of  $FeO_x/SBA-17h-450$ , respectively. No  $FeO_x$  particles or aggregates could be observed on the outside surface of catalyst, even when the  $FeO_x$  content is as high as 26.64wt%, which is

<sup>35</sup> further identified by the HAADF image since no bright spots can be observed. BF image further shows the unblocked channel of the obtained composites. Element mapping by STEM confirmed that ferric species were highly dispersed in SBA-15 matrix (Fig. 2F).



Fig. 3 Wide angle (A, B) and Small angle (C, D) XRD patterns of 55 FeO<sub>x</sub>/SBA-15 samples.

Fig. 3 shows the wide (A, B) and small angle XRD (C, D) patterns of  $FeO_x/SBA-15$  samples. As can be seen in Fig. 3A and B, there is only one highly broadened peak corresponding to

<sup>60</sup> amorphous SiO<sub>2</sub> of SBA-15. No peaks of ferric oxide can be identified, indicating that FeO<sub>x</sub> is amorphous and well-confined in the mesopore channels, as also verified by TEM (Fig. 1A, B). Three peaks indexed as (100), (110) and (200) reflections of the P6mm hexagonal mesoporous structure of SBA-15, respectively,
<sup>65</sup> can be found in their small angle XRD patterns as shown in Fig. 3(C, D), demonstrating their highly ordered mesoporous structure have well-maintained after FeO<sub>x</sub> loading.



Fig. 4 The  $N_2$  adsorption–desorption isotherms (A, B) and corresponding pore size distributions (C, D) of FeO<sub>x</sub>/SBA-15 samples.

Their pore structures were further characterized by  $N_2$ adsorption-desorption isotherms. As shown in Fig. 4, all the samples show a type IV isotherm with a H1 hysteresis loop at  $P/P_0 > 0.5$  associated with capillary condensation, and steep jumps in the  $N_2$  adsorption volume, indicating their mesoporous structures and narrow pore size distributions. Their specific <sup>90</sup> surface area, pore size and pore volume data are all listed in Table. S1. For longer PVI durations, larger amount of FeO<sub>x</sub> can be loaded, which results in decreased surface areas, pore volumes/diameters. Even so, their pore channels still remain open and accessible, as also confirmed by TEM, indicating the ferric <sup>95</sup> oxide has been most probably deposited onto the inner mesopore channel surface of SBA-15 as a thin coating.



Fig. 5 XPS spectra of Fe 2p of FeO<sub>x</sub>/SBA-17h-450.

The surface chemical composition and chemical state of  $FeO_x/SBA-15$  were analyzed by XPS. Fig. 5 shows the Fe 2p <sup>115</sup> spectrum of  $FeO_x/SBA-17h-450$ , in which two characteristic peaks at about 711 eV and 713 eV corresponding to  $Fe^{2+}$  and  $Fe^{3+}$  <sup>30</sup>, respectively, can be found. The proportion of  $Fe^{2+}$  was

calculated to be 58% relative to 42% of Fe<sup>3+</sup>, which could be beneficial for the high degradation efficiency of AO7 as mentioned above. The high proportion of Fe<sup>2+</sup> can be due to the low crystallization and large amount of defects of amorphous 5 FeO<sub>x</sub>. Fig. S1 shows the XPS spectra of Fe 2p of FeO<sub>x</sub>/SBA-17h-T. And Table S2 shows the proportion of Fe<sup>2+</sup> and Fe<sup>3+</sup> of FeO<sub>x</sub>/SBA-17h-T, from which we can see all the materials

contain high proportion of Fe<sup>2+</sup>. Adsorption performance and Fenton catalysis activity of <sup>10</sup> FeO<sub>v</sub>/SBA-15.



Fig. 6 (A) The adsorption kinetic curves of AO7 on  $FeO_x/SBA-17h-450$ , (B) The adsorption isotherm of AO7 on  $FeO_x/SBA-17h-450$ , (C) Langmuir and (D) Freundlich adsorption isotherms of AO7 on  $FeO_x/SBA-17h-450$ .

The adsorption of AO7 on FeO<sub>x</sub>/SBA-15 was investigated prior to the Fenton reaction. Fig. 6A shows the adsorption kinetics curve of AO7 on FeO<sub>x</sub>/SBA-17-450. The AO7 concentration decreased rather quickly at the initial stage and <sup>35</sup> afterwards decreased further but slowly. Adsorption equilibrium was achieved in 2 h and about 73% of the dye AO7 was absorbed. The isoelectric point of FeO<sub>x</sub>/SBA-17h-450 was indentified to be at pH 2.5-2.7 and AO7 is an anionic dye with -SO<sub>3</sub>. Therefore, it was electrostatic favorable adsorption for AO7 on FeO<sub>x</sub>/SBA-15,

<sup>40</sup> which has been further verified by the fact that the equilibrium concentration of AO7 decreased with the decrease of pH values, as shown in Fig. S6C. The adsorption isotherm of AO7 on FeO<sub>x</sub>/SBA-17h-450 at pH=3 has been shown in Fig. 6B and simulated by Langmuir and Freundlich equations:

<sup>45</sup> Langmuir isotherm:  $C_e/Q_e = 1/(Q_mK_L) + C_e/Q_m$ Freundlich isotherm:  $Log(Q_e) = Log(K_F) + (1/n)Log(C_e)$ where  $Q_e$  (mg·g<sup>-1</sup>) is the equilibrium adsorption capacity for AO7,  $C_e$  (mg·L<sup>-1</sup>) is the equilibrium concentration of AO7;  $Q_m$  (mg·g<sup>-1</sup>) and  $K_L$  (L·g<sup>-1</sup>) are Langmuir constants related to adsorption

 $_{\rm 50}$  capacity and energy of adsorption, respectively.  $K_{\rm F}~({\rm mg}\cdot{\rm g}^{-1})$  and 1/n are the Freundlich constants, indicating the capacity and intensity of adsorption, respectively.

AO7 uptake by FeO<sub>x</sub>/SBA-17h-450 was better fitted to Langmuir model than Freundlich model with  $r^2 = 0.99355$  and the

fitted Langmuir equations is  $C_e/Q_e = 0.11235 + 0.00541C_e$ . The excellent correlations assumed that all surface sites on the adsorbents have the same affinity for the adsorbate, the single 115 adsorbate binds to a single site on the adsorbent to form a

monolayer adsorption and no interactions between the adsorbates. The degradation of AO7 by FeO<sub>x</sub>/SBA-15 was further investigated. Fig. S2 shows UV–vis spectra changes of AO7 solution during Fenton reaction. It's obvious that the specific peak of azo bonds of AO7 at 484 nm disappeared in 10 min, indicating the complete degradation of AO7.



**Fig. 7** FTIR spectra of FeO<sub>x</sub>/SBA-17h-450 (a), AO7-adsorbed FeO<sub>x</sub>/SBA-80 17h-450 before Fenton reaction (b), AO7-adsorbed FeO<sub>x</sub>/SBA-17h-450 after Fenton reaction (c), and AO7 (d).

The FTIR spectra (Fig. 7) were used to characterize the structural change of AO7 molecules absorbed on FeO<sub>x</sub>/SBA-15 85 before and after degradation. The intense adsorption of AO7 at 1514 cm<sup>-1</sup> can be attributed to the vibrati ons of -N=N- bond or aromatic ring sensitive to the interaction with the azo bond, or to the N-H bending vibration mode (N-H) of the hydrazone form of the azo dye <sup>31</sup>. While most of the peaks of AO7 adsorbed on 90 FeO<sub>x</sub>/SBA-15 were overlapped by the strong and broad band of  $FeO_x/SBA-15$  in the range of 1000–1200 cm<sup>-1</sup>, the specific vibration peak at 1514 cm<sup>-1</sup> of AO7 at FeO<sub>x</sub>/SBA-15 can still be identified. The disappearance of the pea k at  $1514 \text{ cm}^{-1}$  after Fenton reaction indicates the degradation of AO7 absorbed on 95 FeOx/SBA-15. All these indicate AO7, both absorbed on the catalyst and in the solution, has been degraded completely. The mineralization degree of AO7 was examined by TOC analysis of the solution, as shown in Fig. S3. The TOC removal of the AO7 solution in 10 min reached above 60%, implying FeO<sub>x</sub>/SBA-15 100 can induce an extremely fast mineralization of AO7 by Fenton reaction. As a comparative, commercial Fe<sub>3</sub>O<sub>4</sub> was used to degrade AO7 at the same conditions. Fig. S4 clearly showed that it can only degrade 27% in 10 min and 66% in 2 h, demonstrating the huge advantage of FeO<sub>x</sub>/SBA-15 in heterogeneous Fenton 105 reaction.





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**Fig. 8** Influence of PVI duration on the catalytic activity of FeO<sub>x</sub>/SBA-15 (calcination temperature = 450 °C, [FeO<sub>x</sub>/SBA-15] = 0.6 g·L<sup>-1</sup>, [H<sub>2</sub>O<sub>2</sub>] = 10 mM, pH<sub>0</sub> = 3, C<sub>0</sub> = 100 mg·L<sup>-1</sup>).

- <sup>5</sup> The prolonged PVI duration, can lead to the increased  $FeO_x$  loading amount in SBA-15 and the decreased surface area, pore size/volume of  $FeO_x/SBA-15$ , which will noticeably affect their catalytic activity in Fenton reaction. Fig. 8 gives the concentration variations of AO7 during the adsorption to <sup>10</sup> equilibrium and the following degradation reaction initiated by
- $H_2O_2$  addition. FeO<sub>x</sub>/SBA-17h-450 demonstrates the best adsorption capacity and catalytic degradation activity for AO7. FeO<sub>x</sub>/SBA-9h-450 has larger BET surface area, but has lower adsorption capacity and catalytic activity than FeO<sub>x</sub>/SBA-17h-
- <sup>15</sup> 450 due to its smaller amount of active sites. FeO<sub>x</sub>/SBA-15 obtained at PVI durations above 17 h have higher FeO<sub>x</sub> contents as shown in Table S1, but show decreases in the adsorption capacity and degradation efficiency of AO7, which can be attributed to decreases of their BET surface areas and pore <sup>20</sup> sizes/volumes.

Nevertheless, for all  $FeO_x/SBA-nh-450$  samples, regardless of  $FeO_x$  amounts loaded, one hour is mostly enough to completely degrade AO7. Sample  $FeO_x/SBA-17h-450$  can completely eliminate the dye in as short as 10 min by Fenton reaction.

#### 25 Effect of calcination temperature on AO7 degradation



Fig. 9 Influence of calcination temperature on the catalytic activity of <sup>40</sup> FeO<sub>x</sub>/SBA-15 (PVI duration time = 17 h, [FeO<sub>x</sub>/SBA-15] = 0.6 g·L<sup>-1</sup>,  $[H_2O_2] = 10 \text{ mM}, pH_0 = 3, C_0 = 100 \text{ mg·L}^{-1}$ ).

Firstly, the adsorption capacity of the catalysts obtained at different calcination temperatures can be related not only to their <sup>45</sup> BET surface areas but also to the carbon residue after calcination in the catalysts (as shown in Fig. S5). As the decomposition temperature of ferrocece is 400 °C, the higher the calcination temperature was, the less the content of carbon in the catalyst is. FeO<sub>x</sub>/SBA-17h-450 shows the best adsorption capacity, which <sup>50</sup> should be attributed to the best combination of carbon content and the BET surface area. As reported in literatures, the catalytic activity of heterogeneous catalysts in Fenton reaction is dependent on their calcination temperature and amorphous oxides obtained at low calcination temperature are inclined to show high

<sup>55</sup> catalytic activity <sup>32, 33</sup>. Here, the degradation of AO7 by the heterogeneous Fenton reaction on FeO<sub>x</sub>/SBA-15 obtained at different calcination temperatures was investigated. Limited effects of calcination temperature of FeO<sub>x</sub>/SBA-15 on catalytic

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degradation activity can be found as shown in Fig. 9. This could <sup>60</sup> be attributed to the confinement effect of the mesopore channels of SBA-15 (Fig. 3 and 4), which lead to the compactness rather than particle growth of  $FeO_x$  coating on the mesopore surface during calcination at high temperature.

Effect of Fenton reaction conditions on AO7 degradation. <sup>65</sup> Varied catalyst amounts from 0 to 1 g·L<sup>-1</sup> were used to investigate the effect of FeO<sub>x</sub>/SBA-17h-450 dosage on AO7 degradation efficiency. As can be seen in Fig. S6A, negligible degradation of AO7 can be found without FeO<sub>x</sub>/SBA-17h-450, indicating that  $H_2O_2$  itself cannot generate HO• to initiate Fenton reaction. With <sup>70</sup> the increase of catalyst dosage, AO7 degradation was accelerated by the increased HO• concentration. The optimal concentration of FeO<sub>x</sub>/SBA-17h-450 is 0.6 g·L<sup>-1</sup>, above which no significant enhancement in degradation efficiency can be obtained.

It is well established that  $H_2O_2$  shows significant effect on the <sup>75</sup> degradation of dyes involving Fenton reaction. Consequently, effect of  $H_2O_2$  dosage on the degradation of AO7 was investigated. As shown in Fig. S6B, no degradation of AO7 can be observed without  $H_2O_2$ . The degradation efficiency was improved by the increasing concentration of  $H_2O_2$  until it reached <sup>80</sup> 15 mM. At over-low initial concentration of  $H_2O_2$ , only limited amount of HO• can be generated, which is too little to initiate quick dye degradation. However, the excessive  $H_2O_2$  with concentration above 15 mM, according to the following equations:  $H_2O_2 + HO• \rightarrow HO_2• + H_2O$  eq. (1) <sup>85</sup>  $HO_2• \rightarrow O_2•^- + H^+$  eq. (2)

the scavenging effect of HO• will happen and produce less active  $HO_2$ • and superoxide anion  $O_2$ •<sup>-34</sup> and thus retard the Fenton reaction.

Surface charge generally shows great influence on the <sup>90</sup> catalytic activity of Fenton catalysts <sup>35</sup> since the adsorption of the dye molecules on catalyst surface and their catalytic degradation by HO• are synergetic 36. As shown in Fig. S6C, with the decrease of pH from 9 to 2, both the adsorption performance and degradation efficiency can be promoted. Under alkaline 95 conditions, only a small amount of dye could be adsorbed and no significant degradation of the dye was observed. Comparatively, in strong acid environment, the catalyst showed higher adsorption capacity and degradation efficiency for AO7 because the catalyst surface was nearly-positively (The isoelectric point was pH = 2.5-100 2.7) charged and easily adsorbed negatively-charged AO7 molecules by electrostatic interaction. Furthermore, it has been reported that an acid environment, in which the maximum amount of HO• can be produced, are usually optimum for Fenton reaction  $^{37}$ . However, at pH = 2, little increase of degradation 105 efficiency can be observed since excessive Cl<sup>-</sup> can serve as HO• scavenger <sup>38</sup>. It is worth noting that the degradation efficiency could reach 89% in 120 min at pH = 5, indicating the catalyst are still effective in weak acid environment.

Stability and recycling performance of the heterogeneous 110 Fenton catalyst FeO<sub>x</sub>/SBA-15



Fig. 10 Recycling performance of the FeO<sub>x</sub>/SBA-17h-450 catalyst in degrading AO7 ([FeO<sub>x</sub>/SBA-15] = 0.6 g·L<sup>-1</sup>, [H<sub>2</sub>O<sub>2</sub>] = 15 mM, pH<sub>0</sub> = 3, 15 C<sub>0</sub> = 100 mg·L<sup>-1</sup>).

In order to investigate the stability of the  $FeO_x/SBA-15$  catalysts, the dissolved iron concentration in the aqueous solution were tested by ICP-OES. At pH = 3, the concentration of total iron ions dissolved in the solution in 30 min after the adding of

<sup>20</sup> H<sub>2</sub>O<sub>2</sub> was 5.77 ppm. To clarify whether it was a homogeneous reaction or heterogeneous reaction, the supernatant collected by centrifuging to remove FeO<sub>x</sub>/SBA-15 solid during Fenton reaction was used to degrade AO7 and a degradation efficiency of as low as 20% was obtained after 2 h. The catalysts were <sup>25</sup> recovered by centrifuging and dried in oven at 60 °C for 12 h and <sup>25</sup> the effect of 25 was as a solution.

then reused for AO7 degradation. The degradation efficiency of AO7 on  $\text{FeO}_x/\text{SBA-15}$  only showed a slight decrease from 98% to 95% after five recycles (Fig. 10). These facts confirm it is a heterogeneous Fenton reaction that dominates the dye <sup>30</sup> degradation on  $\text{FeO}_x/\text{SBA-15}$ .

**Fenton catalytic mechanism on FeO<sub>x</sub>/SBA-15.** Scavenging study was used to further identify the responsibility of HO• in this Fenton reaction. The degradation of AO7 was significantly inhibited and decreased from 99% to 77% by adding 2-propanol <sup>35</sup> (molar ratio 2-propanol/AO7 = 400/1), which is a typical HO•

scavenger, to the reaction solution (Fig. S7), Taking account into the 73% removal of AO7 by adsorption on  $FeO_x/SBA-15$ , the degradation of AO7 by Fenton catalyst was negligible in the presence of 2-propanol, that is, without HO•.



Fig. 11 Schematic diagram of the  $FeO_x/SBA-15$  synthesis and application as heterogeneous Fenton catalyst for efficient degradation of AO7.

A synergistic catalytic effect, which is featured with the successional functions of the catalyst components in activating dye molecules by adsorption on the pore surface of the catalyst, and generating the active HO• species by  $Fe^{2+}/Fe^{3+}$  cycling after the dye adsorption, was assumed as illustrated in Fig. 11 <sup>36</sup>. The

- processes of the dye activation and HO• generation are <sup>60</sup> independent of each other. Therefore, theoretically, the two processes can be regarded as in parallel. Nevertheless, in the practical operation of the Fenton reactions, dye molecules were first adsorbed and activated on the pore surface of FeO<sub>x</sub>/SBA-15, then the HO• generated by the reaction between Fe<sup>n+</sup> and later-
- $^{\rm 65}$  added  $\rm H_2O_2$  oxidizes the activated dye molecules to small molecules or  $\rm CO_2$  and  $\rm H_2O.$

#### Conclusions

- A simple physical vapor infiltration process has been developed 70 to load extraordinarily large amount of ferric oxide into ordered mesoporous silica SBA-15. The FeOx, which is demonstrated to be deposited as a thin coating on the mesopore channel surface of SBA-15, are amorphous and enriched with Fe<sup>2+</sup> ions (58%, relative to 42% Fe<sup>3+</sup>). The high dispersion of a large amount of  $_{75}$  Fe<sup>2+</sup>-rich FeO<sub>x</sub> species and the well-opened mesoporous structure of the composite catalysts are believed to the main contributors to the high performance in AO7 dye degradation (~ 99% AO7 degradation in 10 min at pH = 3,  $[H_2O_2] = 15$  mM, and  $[FeO_x/SBA-15 \text{ catalyst}] = 0.6 \text{ g}\cdot\text{L}^{-1}$ ). The Fenton reaction is 80 proposed to progress by a synergetic effect featured with the successional/parallel functions of catalyst components in dye adsorption-activation on the pore surface of the catalyst and the generation of active HO• species by Fe<sup>2+</sup>/Fe<sup>3+</sup> cycling, which are responsible for the high catalytic activity in the dye degradation.
- 85 The catalyst also showed good stability with negligible decrease of degradation efficiency after 5 recycles. These mesoporous FeO<sub>x</sub>/SBA-15 composites obtained by the low-cost convenient PVI approach are promising candidates as high-efficiency heterogeneous Fenton catalysts for water treatment.

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#### 95 Notes and references

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- † Electronic Supplementary Information (ESI) available: [Pore structure parameters; UV–vis spectra changes of AO7solution with reaction time; TOC; Degradation efficiency of AO7; Influence of catalyst dosage, H<sub>2</sub>O<sub>2</sub>
   <sup>105</sup> dosage, initial pH value on the catalytic activity; Effect of HO• scavenger

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High dispersion of large amount of  $Fe^{2+}$ -rich  $FeO_x$  and unblocked pore channels of catalysts obtained by physical-vapor-infiltration facilitate Fenton reaction.