



Cite this: *RSC Adv.*, 2017, 7, 31611

Selective and colorimetric detection of Ba²⁺ ions in aqueous solutions using 11-mercaptoundecylphosphonic acid functionalized gold nanoparticles†

Blanca A. García Grajeda,^a Samuel G. Soto Acosta,^a Sergio A. Aguila,^b Héctor Peinado Guevara,^a Marta E. Díaz-García,^c Adriana Cruz Enríquez ^{*a} and José J. Campos-Gaxiola ^{*a}

In this paper we describe a simple colorimetric method for the sensitive and selective detection of Ba²⁺ ions, using 11-mercaptoundecylphosphonic acid functionalized gold nanoparticles (AuNPs-MPA) in aqueous solution. With increasing concentration, Ba²⁺ ions induced aggregation of AuNPs-MPA, causing a change in color of the colloidal solution from pink to blue. The detection result was monitored using the naked eye and UV-Vis spectroscopy. The maximum absorption was located at 625 nm. The stability of AuNPs-MPA and detection of Ba²⁺ by AuNPs-MPA was investigated at different pH values (pH = 3, 4, 5, 6, 7, 8, 9, 10, 11, 12). Results indicated that AuNPs-MPA at pH 6 were the optimal choice for detecting Ba²⁺ ions. The detection limit for Ba²⁺ was found to be 43.27 μmol L⁻¹ (R² = 0.9901) within a range of 20–120 μmol L⁻¹. In addition, the selectivity of this method was investigated using other metal ions. The AuNPs-MPA was selective for Ba²⁺ compared with other metal ions (Li⁺, Na⁺, K⁺, Mg²⁺, Ca²⁺, Sr²⁺, Ba²⁺, Mn²⁺, Fe²⁺, Co²⁺, Ni²⁺, Cu²⁺, Zn²⁺, Cd²⁺, Hg²⁺, Sb³⁺, Bi³⁺) leading to a remarkable color change. Moreover, the developed cost-effective probe has been successfully applied to real drinking and tap water samples which demonstrated the potential for field applications.

Received 4th April 2017
Accepted 8th June 2017

DOI: 10.1039/c7ra03861e

rsc.li/rsc-advances

1. Introduction

Heavy metal pollution in aquatic systems and soil has become a serious threat to human health and the environment.¹ Barium (Ba) is considered as a heavy metal and is widely used in

a variety of industrial processes, such as electroplating, the glass industry, the petroleum industry, alloying and fixing pigments, and in corrosion protection films. To date pollution of water by Ba²⁺ ions is still the most common.^{2,3} Barium compounds are also used in paint, bricks, ceramics, glass and rubber.⁴ Ba²⁺ contaminates soil and water, and violently reacts with the latter to form salts. All barium salts, especially the water and acid-soluble compounds are highly toxic (for example, barium sulphide).^{5,6} High concentrations of Ba in water used for irrigation have toxic effects on plant life.^{7,8} The soluble form of Ba can be accumulated by plants; this may result in toxicity for other organisms because accumulated Ba can be transferred to animals.^{7,9} The selective and sensitive barium is important because, in low doses, barium acts as a muscle stimulant, while in higher doses it affects the nervous system, causing cardiac irregularities, increased or decreased blood pressure, tremors, weakness, anxiety, dyspnea, changes in heart rhythm or paralysis and possibly death.⁹ At present, water pollution by Ba²⁺ ions are still very common. In this context, several methods have been used for detection of heavy metal ions, including Ba²⁺; such as, fluorescent chemosensors,^{10,11} inductively coupled plasma mass spectrometry (ICP-MS),¹² voltammetry¹³ and spectrophotometry,¹⁴ proving much more sensitivity and reliability than colorimetric

^aFacultad de Ingeniería Mochis, Universidad Autónoma de Sinaloa, Fuente de Poseidón y Prol. A. Flores S/N, C.P. 81223, C.U. Los Mochis, Sinaloa, México. E-mail: gaxiolajose@uas.edu.mx; cruzadriana@uas.edu.mx; Fax: +52 668 8127641; Tel: +52 668 8127641

^bCentro de Nanociencias y Nanotecnología, Universidad Nacional Autónoma de México (CNYN-UNAM), km. 107 Carretera Tijuana-Ensenada, Apartado Postal 14, C.P. 22800, Ensenada, Baja California, México

^cDepartment of Physical and Analytical Chemistry, University of Oviedo, Julián Clavería no. 8 primera planta, C.P. 33006, Oviedo, Spain

† Electronic supplementary information (ESI) available: Fig. S1–S8 [(S1) Absorbance ratio (A₆₂₅/A₅₃₀) of AuNPs-MPA in the presence of metal ions. Blue bars represent the addition of single metal ion (100 μM); red bars represent the mixture of Ba²⁺ (100 μM) with another metal ion (100 μM). (S2) IR spectra functionalized AuNPs-MPA and AuNPs-MPA detecting Ba²⁺ ions, (S3) UV-Vis spectra of AuNPs-citrate and AuNPs-citrate-Ba²⁺, (S4) TEM image of AuNPs-MPA evaluated at pH 3, (S5) effect of reaction time on the absorption ratio A₆₂₅/A₅₃₀ for AuNPs-MPA system in the presence of various concentrations of Ba²⁺ (S6) UV-Vis spectra of AuNPs-MPA detecting Ba²⁺ ions in drinking water and tap water, (S7) colorimetric response of drinking water and tap water, (S8) the linearity curve of the ratio of A₆₂₅/A₅₃₀ versus Ba²⁺ concentration drinking water and tap water]. See DOI: 10.1039/c7ra03861e



methods and other procedures. However, these methods require expensive instruments, tedious sample preparation, more time for analysis and these are not suitable for on-site assays. Therefore, the development of simple and rapid colorimetric method, a lower cost and less response time, that can be used in metal ions assays in environmental samples is necessary.

In recent years, gold nanoparticles (AuNPs) have been amply studied, and used in a wide range of applications including: optical sensors, electronics, surface-enhanced Raman spectroscopy and colorimetric probes. Colorimetric sensors based on gold nanoparticles (AuNPs) are of particular interest because they are simple and convenient; no sophisticated instruments are required and the chemical interaction can easily be monitored by color changes. Therefore, AuNPs are widely used for analyte detection. For example, AuNPs are red in color; however, AuNPs aggregates generated by different metal ions show changes to other colors such as yellow, purple or blue.^{15–20}

AuNPs display strong surface plasmon resonance (SPR) absorption properties,^{21,22} which are extremely sensitive to size, shape and interparticle plasmon coupling on analyte-induced aggregation of AuNPs. These results in a color change from red to blue that may be observed by the naked eye.^{23,24}

Colorimetric assays have been widely used to detect molecules such as oligonucleotides,²⁵ proteins,²⁶ small molecules²⁷ and metal ions.^{28–30} These assays are simple, rapid and, most importantly, cost effective. On the other hand, Ba²⁺ ion detection by the colorimetric method using specially functionalized gold and silver nanoparticles has been scarcely reported.^{3,31,32} To date, there are no reports detecting metals ions using phosphonic groups, more specifically 11-mercaptopoundecylphosphonic acid (MPA) functionalized gold nanoparticles. One of the advantages of using this ligand (MPA) is that it provides a wide range on pH sensitivity due to their dibasic nature, providing excellent stability in aqueous solutions under extremes pH.³³

In this work, 11-mercaptopoundecylphosphonic acid (MPA)-functionalized gold nanoparticles (AuNPs-MPA) were synthesized and used as a colorimetric sensor to detect Ba²⁺. The MPA was attached to the surface of AuNPs using the thiol group *via* the Au-S bond,^{34–36} then, AuNPs-MPA were used for metal ion detection. Other metal ions such as Li⁺, Na⁺, K⁺, Mg²⁺, Ca²⁺, Sr²⁺, Ba²⁺, Mn²⁺, Fe²⁺, Co²⁺, Ni²⁺, Cu²⁺, Zn²⁺, Cd²⁺, Hg²⁺, Sb³⁺, Bi³⁺ were tested for metal ion selectivity. Only the Ba²⁺ ion caused aggregation of AuNPs-MPA.

2. Experimental section

2.1 Materials

Hydrogen tetrachloroauric(III) trihydrate (HAuCl₄·3H₂O), trisodium citrate, 11-mercaptopoundecylphosphonic acid (MPA), NaOH, polyvinyl alcohol, LiCl, NaCl, KCl, MgCl₂, CaCl₂, SrCl₂, BaCl₂, MnCl₂, FeCl₂, CoCl₂, NiCl₂, CuCl₂, ZnCl₂, CdCl₂, HgCl₂, SbCl₃, BiCl₃, were obtained from Sigma-Aldrich. All reagents were of analytical grade and distilled water was used in all experiments.

2.2 Methods

UV-Vis spectra were recorded on a Shimadzu 1800 UV-Vis spectrophotometer at room temperature. IR data were obtained on an Alpha-Bruker Fourier-Transform Infrared Spectrometer. TEM images were obtained from a JEM-2010 transmission electron microscope.

2.3 Gold nanoparticle (AuNPs-MPA) preparation

AuNPs were produced by reduction of HAuCl₄·3H₂O with trisodium citrate in aqueous solution.^{29,33,34} Briefly, HAuCl₄·3H₂O (100 mg, 0.25 mmol) was dissolved in 200 mL water and the solution was heated (temperature ≤ 80 °C) while stirring. A trisodium citrate (40 mmol L⁻¹) solution was rapidly added with intense stirring. After the color had changed from pale yellow to wine red, an aqueous solution of 11-mercaptopoundecylphosphonic acid (81.7 mg, 0.3 mmol), containing an equivalent amount (0.0022 mg, 0.3 mmol) of sodium hydroxide, was added to the colloidal gold suspension. The solution was stirred for 1 h and then diluted to 250 mL with distilled water. In order to remove excess trisodium citrate and MPA, the colloidal suspension was dialyzed in an aqueous medium. Unwanted large aggregated particles were removed from the suspension by filtration through a 0.45 μm nylon filter. Finally, AuNPs-MPA were collected and stored at room temperature for later use. UV-Vis spectra were obtained at 530 nm to AuNPs-MPA.

2.4 Colorimetric detection of Ba²⁺

To test the sensitivity of AuNPs-MPA, one mL aliquots of solution containing AuNPs-MPA were taken to prepare 10 mL aqueous solution samples, using PVA 1% as a stabilizing agent, then increasing concentrations of Ba²⁺ (20–300 μmol L⁻¹) were added using the stock solution. The pH was adjusted to 3–12 using acetic acid/sodium acetate (0.1 M), as buffer. We investigated the selectivity of AuNPs-MPA for Ba²⁺ with other metal ions: Li⁺, Na⁺, K⁺, Mg²⁺, Ca²⁺, Sr²⁺, Ba²⁺, Mn²⁺, Fe²⁺, Co²⁺, Ni²⁺, Cu²⁺, Zn²⁺, Cd²⁺, Hg²⁺, Sb³⁺, Bi³⁺ under the same conditions.

Subsequently, 1.0 mL samples of solution containing AuNPs-MPA were separately mixed with the different metal ions (100 μmol L⁻¹). The mixtures were prepared at room temperature and then transferred separately into 1.5 mL quartz. Their SPR absorption bands were recorded by UV-Vis spectrophotometer.

2.5 Colorimetric detection of Ba²⁺ in real samples

Drinking water and tap water were collected in Los Mochis, Sinaloa, Mexico and filtered through a 0.45 μmol L⁻¹ membrane. A 0.0001 mol L⁻¹ standard solution of Ba²⁺ was prepared. A series of dispersions was prepared using 1.0 mL of AuNPs-MPA and 20–300 μmol L⁻¹ of Ba²⁺. The pH was adjusted to 6 using acetic acid/sodium acetate (0.1 M) as buffer. Subsequently, SPR absorption bands were recorded by UV-Vis spectrophotometer.



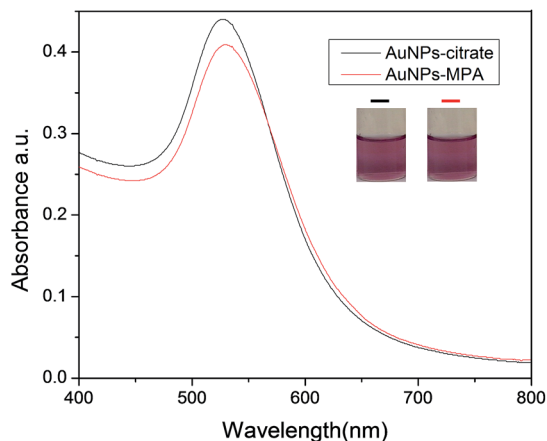


Fig. 1 UV-Vis spectra of (–) AuNPs–citrate and (–) AuNPs–MPA.

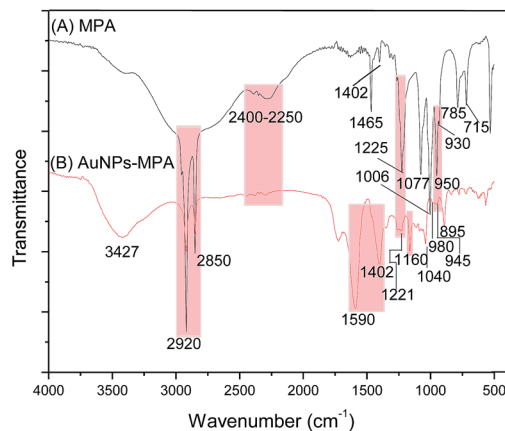


Fig. 3 FT-IR spectra of (A) MPA and (B) AuNPs–MPA.

3. Results and discussion

3.1 Characterization of AuNPs–MPA

AuNPs were prepared by reducing HAuCl_4 , with citrate as reducing agent. MPA adsorbed on the gold nanoparticles by thiol groups. Fig. 1 shows the optical absorption spectra of colloidal gold before and after modification with MPA. The surface plasmon band exhibits a slight red shift from 527 nm to 530 nm associated with MPA modification. This small red shift is due to the increase in refractive index of the medium that surrounds the nanoparticles.³⁷ Maximum absorption of the AuNPs surface plasmon resonance (SPR) was located at 530 nm,

determined by UV-Vis spectrophotometer. TEM images showed well-monodispersed AuNPs–MPA with a size ranging from 14 to 30 nm, average size of 20.8 nm (Fig. 2).

The AuNPs–MPA were also characterized using infrared spectroscopy. IR spectra of MPA and AuNPs–MPA are compared in Fig. 3. Note that the characteristic –SH stretching absorption peaks at $2400\text{--}2250\text{ cm}^{-1}$ corresponding to MPA have disappeared in the AuNPs–MPA IR spectrum, suggesting that MPA is coordinated with the gold atoms on the surface of AuNPs through the Au–S bond.

The MPA IR bands are assigned as follows: $2920\text{--}2850\text{ cm}^{-1}$ for symmetric (ν_s) and antisymmetric (ν_{as}) C–H stretching, respectively, at 1465 and 1402 cm^{-1} for the scissoring mode (δ), at 1225 cm^{-1} P=O (ν), 1077 and 1006 cm^{-1} (ν_s and ν_{as} of PO_3), at 950 and 930 cm^{-1} (ν , P–OH), at 785 and 715 cm^{-1} are IR bands assigned to CH_2 out of the plane (oop) and the P–C, respectively. All these bands are consistent with previous studies of phosphonic acids.^{33,38,39}

A broad band of the IR of AuNPs–MPA which appears at 3427 cm^{-1} corresponds to OH groups of the citrate and P–OH of MPA groups. The intensity of the P=O band at frequency $\sim 1220\text{ cm}^{-1}$ is clearly diminished, and a new band appears at a lower wavenumber (1160 cm^{-1}), associated with phosphonate groups; in addition, the small peaks at 980 and 945 cm^{-1} assigned to P–OH group moieties, concluding that these groups are deprotonated.^{33,39} Two vibrational bands at 1590 and 1402 cm^{-1} are assigned to the symmetric (ν_s) and antisymmetric (ν_{as}) stretching of COO^- groups of citrate.^{40,41} Data from complementary techniques such as XPS and NMR suitable for study the binding sites of phosphonic and thiol ligands to AuNPs, compare favorably with our results.^{42–44}

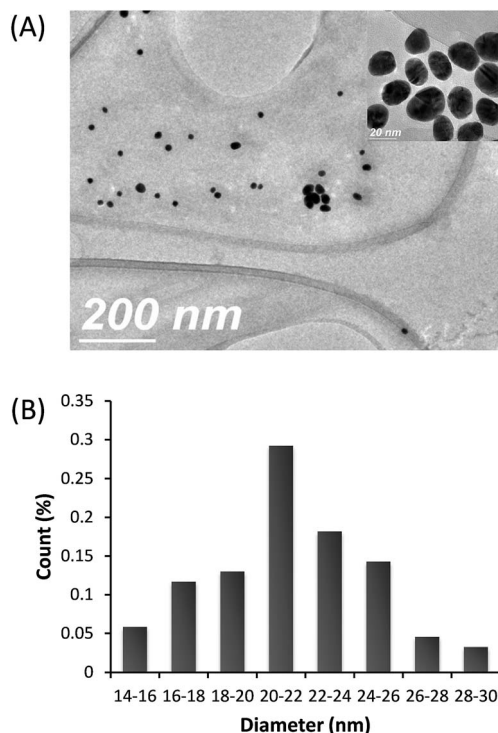


Fig. 2 (A) Typical TEM image of the AuNPs–MPA (scale bar: 200 nm and 20 nm) (B) histogram of AuNPs–MPA particle size distribution.

3.2 Interactions of AuNPs–MPA with various metal ions

The interactions of the AuNPs–MPA with various metal ions was monitored by UV-Vis spectrophotometer and also by color change visualized with the naked eye.

To evaluate the selectivity of AuNPs–MPA, various metal ions were used: Li^+ , Na^+ , K^+ , Mg^{2+} , Ca^{2+} , Sr^{2+} , Ba^{2+} , Mn^{2+} , Fe^{2+} , Co^{2+} , Ni^{2+} , Cu^{2+} , Zn^{2+} , Cd^{2+} , Hg^{2+} , Sb^{3+} and Bi^{3+} ($100\text{ }\mu\text{mol L}^{-1}$) under



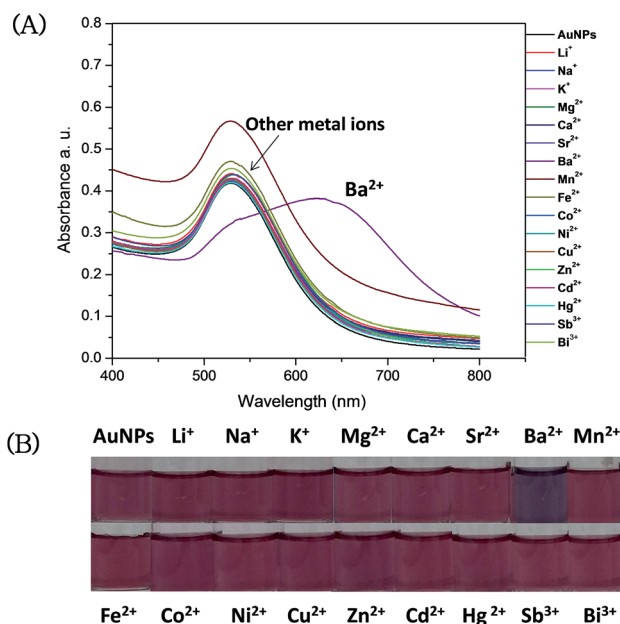


Fig. 4 (A) UV-Vis spectra of AuNPs-MPA in presence of different metal ions ($100 \mu\text{mol L}^{-1}$). (B) Images of AuNPs-MPA in presence of various metal ions ($100 \mu\text{mol L}^{-1}$).

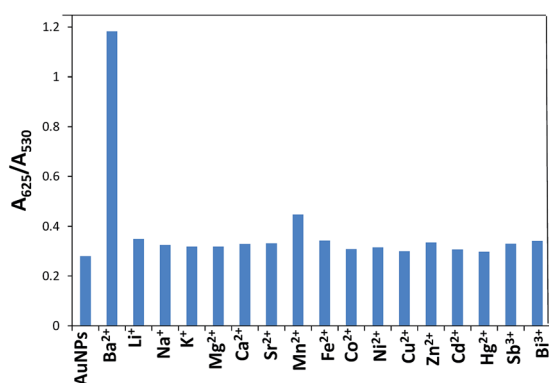


Fig. 5 Absorbance ratio (A_{625}/A_{530}) of AuNPs-MPA in the presence of $100 \mu\text{mol L}^{-1}$ Ba^{2+} and other metals ions.

the same conditions (pH 6). Fig. 4 shows that only Ba^{2+} causes a shift in the plasmon absorption peak from 530 nm to 625 nm. This red shift was observed as a color change from pink to blue.

The A_{625}/A_{530} data of AuNPs-MPA show excellent selectivity of Ba^{2+} over other metal ions (Fig. 5 and S1†). Ba^{2+} induced remarkable aggregation of the AuNPs-MPA, and other metal ions did not influence the absorption spectrum of AuNPs-MPA indicating that no aggregation occurred. These results could be evidence for the aggregation of AuNPs-MPA by Ba^{2+} seen in the TEM images shown in Fig. 6. Thus, the selectivity of AuNPs-MPA for the detection of Ba^{2+} might be suitable for the detection of Ba^{2+} in environmental samples.

The sensing mechanism was explained in terms of the Lewis acid-base theory that Ba^{2+} might coordinate with oxygen-donating chelates and could form a stable Ba^{2+} citrate and phosphonate complex. To gain a clearer understanding of the

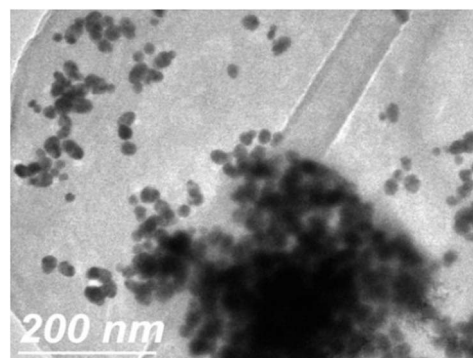


Fig. 6 TEM image of AuNPs-MPA in the presence of Ba^{2+} ($100 \mu\text{M}$).

interaction between AuNPs-MPA and Ba^{2+} , we used IR spectroscopy. The vibrational bands for AuNPs-MPA that were observed in the IR spectrum were: at $2920\text{--}2850 \text{ cm}^{-1}$ C-H stretching; at 1160 cm^{-1} phosphonate groups; at 1553 and 1392 cm^{-1} COO^- from the citrate compound. These appear with a lower wavenumber in the vibrational bands obtained for AuNPs-MPA + Ba^{2+} (see Fig. S2 in ESI†). These shifts show that AuNPs-MPA and citrate interact with Ba^{2+} in aqueous solution.^{45–47}

As shown in Fig. 7, aggregation of AuNPs-MPA in the presence of Ba^{2+} is due to the binding of Ba^{2+} ions through bonds formed between phosphonic and COO^- groups that can be captured by Ba^{2+} , producing a substantial shift in the plasmon band to a longer wavelength and the mentioned color change from pink to blue.

In addition, to investigate the metal-ligand interaction of citrate ions with Ba^{2+} , Ba^{2+} was added to unfunctionalized AuNPs (AuNPs-citrate). This did not influence the surface plasmon band shown by the UV-Vis results (see Fig. S3 in ESI†). The color of the dispersions remained pink, implying that citrate ions did not interact with Ba^{2+} ; therefore, functionalization of AuNPs is necessary to detect Ba^{2+} .

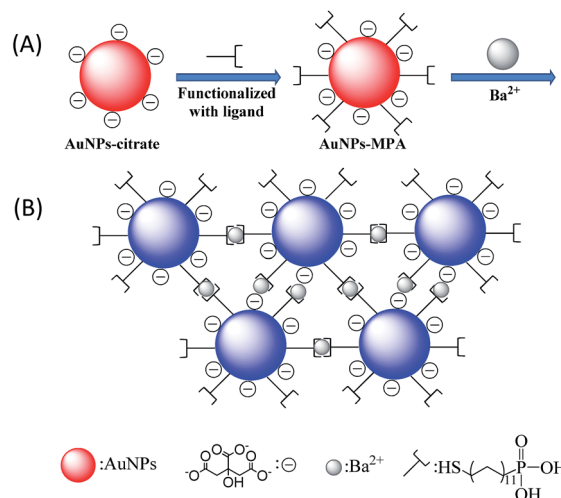


Fig. 7 Illustrative description of (A) functionalized of AuNPs-MPA and (B) Ba^{2+} induced aggregation of AuNPs-MPA.



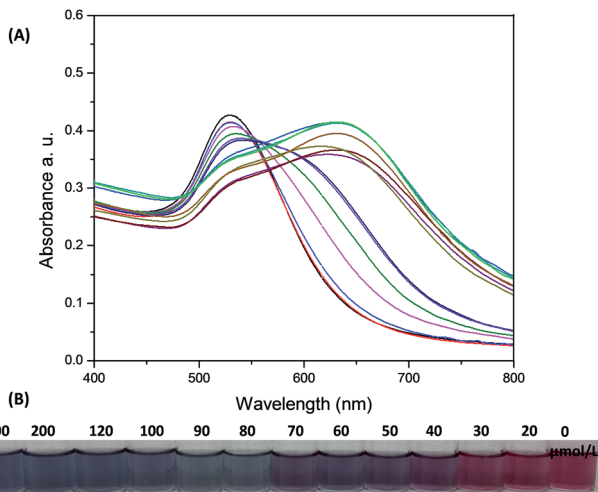


Fig. 8 (A) UV-Vis spectra of AuNPs-MPA after the addition of different concentrations of Ba^{2+} , (B) photographic image of AuNPs-MPA of detection.

On the other hand, to study the sensitivity (detectable minimum concentration of Ba^{2+}) and the linear range in more detail, different concentrations of the standard solution of barium in a range of 20–120 $\mu\text{mol L}^{-1}$ were taken, using the same volume of AuNPs-MPA at pH 6, and analyzed by UV-Vis spectrophotometry.

Fig. 8 shows a red shift in wavelength and a broadening of the SPR peaks of AuNPs-MPA. Simultaneously, the intensity of the SPR signal decreased at 530 nm and a new band appeared at around 625 nm with increasing concentration of barium, as a result of the induced aggregation of AuNPs-MPA. The color of the AuNPs-MPA is proportional to the amount of barium added to the samples, changing from pink to blue, and is accompanied by a bathochromic shift in the UV-Vis spectra. A linear relationship of Ba^{2+} was obtained from the absorbance ratio (A_{625}/A_{530}) and concentration of Ba^{2+} in the 0–120 $\mu\text{mol L}^{-1}$ range, with a correlation coefficient (R^2) of 0.9901, as shown in Fig. 9. The limit of detection (LOD) for Ba^{2+} was found at 43.27 $\mu\text{mol L}^{-1}$.

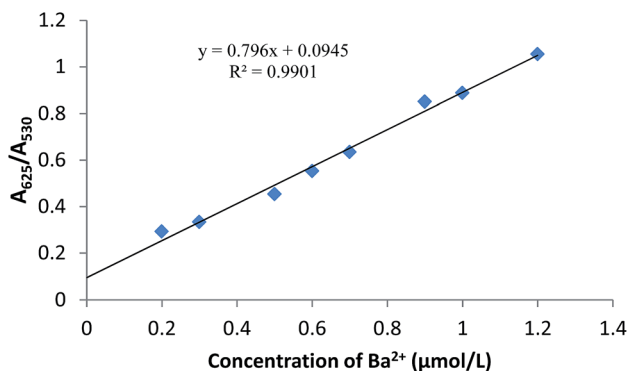


Fig. 9 The linearity curve of the ratio of A_{625}/A_{530} versus Ba^{2+} concentration.

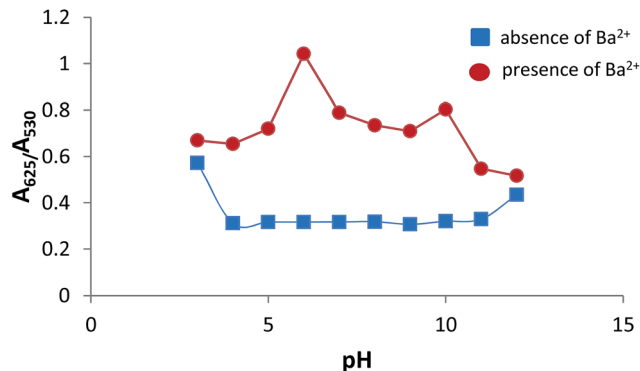


Fig. 10 Influence of pH on absorbance ratio (A_{625}/A_{530}) of AuNPs-MPA in the absence and presence of Ba^{2+} (100 $\mu\text{mol L}^{-1}$).

3.3 Influence of pH on the stability of AuNPs-MPA and aggregation induced by Ba^{2+}

The pH value of the system also plays an important part in the sensing system. The stability of AuNPs-MPA at different pH values was investigated. The pH was adjusted from 3–12 using an acetic acid/sodium acetate (0.1 M) buffer solution. Fig. 10 (blue line) shows that the absorbance ratio (A_{625}/A_{530}) of AuNPs-MPA increased at pH 3; thus, under very acidic conditions the absorbance ratio (A_{625}/A_{530}) is slightly higher than without Ba^{2+} . This may be explained because phosphonate groups, as mentioned above (Fig. 2), are deprotonated species, and protonation of these groups induces the formation of intermolecular hydrogen bonds between phosphonic groups. This causes the AuNPs-MPA to aggregate by interparticle plasmon coupling, which is unsuitable for monitoring Ba^{2+} ions under these conditions (see Fig. S4 in ESI†).^{33,48} In the pH range between 4 and 12, the absorbance ratio (A_{625}/A_{530}) was constant, indicating that AuNPs-MPA at these values were stable. The influence of pH on the Ba^{2+} -induced aggregation of AuNPs-MPA (Fig. 10, red line) shows a high absorbance ratio (A_{625}/A_{530}) at the pH range 4–10; an even higher aggregation was observed at pH 6.

At pH > 10, the A_{625}/A_{530} decreased because of the formation of colloidal BaOH_2 in the alkaline environment of the solutions. It is worth to mention the fact that the ionic strength of the medium may result in the AuNPs-MPA aggregation.^{49–51} In order to avoid such influence, a 0.1 M ionic strength was kept constant using the acetic acid/sodium acetate 0.1 M thoroughly. All assays were performed at room temperature in order to avoid changes in the hydrodynamic dimensions of the MPA-AuNPs and their aggregation.^{52,53}

3.4 Effect of reaction time

The influence of reaction time upon the interaction between AuNPs-MPA and Ba^{2+} was investigated and the results are recorded in Fig. 11. It was observed that after the addition of 100 μM Ba^{2+} to the AuNPs-MPA solution, the absorption ratio (A_{625}/A_{530}) for the reaction time increased gradually from 1 to 10 minutes and was then consistent from 10 to 20 minutes. This indicated that aggregation of the AuNPs solution was nearly complete at the 10 minutes reaction time. Thus it was



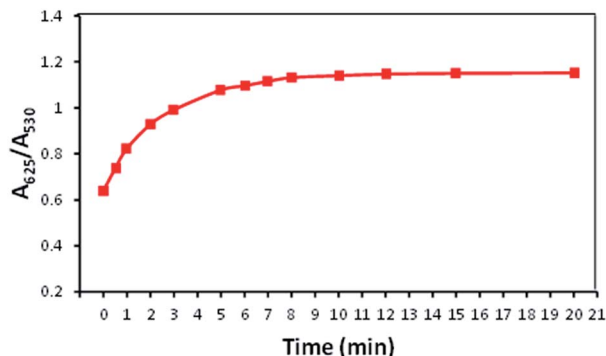


Fig. 11 Effect of reaction time on the absorption ratio (A_{625}/A_{530}) of AuNPs-MPA solution containing $100 \mu\text{M}$ Ba^{2+} .

determined that 10 minutes was the optimum reaction time and all further colorimetric measurements were performed for 10 minutes.

A sensor providing a fast response will be beneficial to the on-site and real time detection of Ba^{2+} . Therefore, the aggregation reaction kinetics of this assay with different concentrations of Ba^{2+} were investigated. As shown in Fig. S5,† in the concentration range ($80\text{--}150 \mu\text{M}$), the A_{625}/A_{530} increased with the increase in the incubation time and kept constant after 10 min. When the concentration of Ba^{2+} was $100 \mu\text{M}$, the

absorbance ratio increased very rapidly within 1 min and then reached equilibrium with time. To make the recognition reaction complete, the A_{625}/A_{530} was measured after incubation for 10 min. Therefore, 10 min time course was introduced to record the coordination reaction in the sensing system.

3.5 Analysis of Ba^{2+} ions in real water samples

To confirm the practical application of this method, we collected two different water samples (drinking water and tap water), and filtered them through a $0.45 \mu\text{m}$ membrane. Ba^{2+} ions were not detected in drinking water samples or in tap water, according with the UV-Vis results (see Fig. S6 in ESI†). We then prepared a series of samples by spiking the drinking and tap water with a standard solution of Ba^{2+} in a range of $20\text{--}300 \mu\text{M}$. Results of the detection of Ba^{2+} in these samples are shown in Fig. 12.

The SPR bands of AuNPs-MPA are observed at different concentrations ($20\text{--}300 \mu\text{mol L}^{-1}$) of Ba^{2+} . The intensity of absorbance of AuNPs-MPA decreased with increasing concentrations of Ba^{2+} , accompanied by a colour change of the colloidal dispersions from pink to blue for the drinking water, and from pink to purple for the tap water samples (see Fig. S7 ESI†).

The minimum detectable concentration of Ba^{2+} was found at $56.94 \mu\text{M}$ ($R^2 = 0.9577$) and $176.78 \mu\text{M}$ ($R^2 = 0.9738$) in drinking water and tap water, respectively (see Fig. S8 in ESI†).

These results demonstrate that this method is applicable for the detection of Ba^{2+} in both water samples. Higher sensitivity was observed for samples of drinking water, perhaps because these samples have less impurity.

4. Conclusions

In this work, we developed a colorimetric system of AuNPs-MPA that can be used to detect Ba^{2+} ions. In the presence of high concentration of potential interfering ions, Ba^{2+} was the only metal that induced aggregation of AuNPs-MPA, also showing a color change of colloidal solutions from pink to blue, accompanied by a plasmon absorption band shift from 530 nm to 625 nm . The colorimetric system is a method that allows monitoring Ba^{2+} ions at concentrations as low as $43.27 \mu\text{mol L}^{-1}$ at low cost and rapidly (the detection time within 10 min). This method was successfully applied to detect of Ba^{2+} in drinking and tap water samples. Furthermore, we believe that this approach has great potential for the routine detection of Ba^{2+} in environmental samples.

Acknowledgements

The authors gratefully acknowledge access to the analytical facilities (Transmission Electronic Microscopy Laboratory) at the Centro de Nanociencias y Nanotecnología, Universidad Nacional Autónoma de México (CNYN-UNAM). B. A. G. G. thanks the support from Consejo Nacional de Ciencia y Tecnología (CONACYT) in the form of a graduate scholarship (261564). This work was supported by Secretaría de Educación Pública México

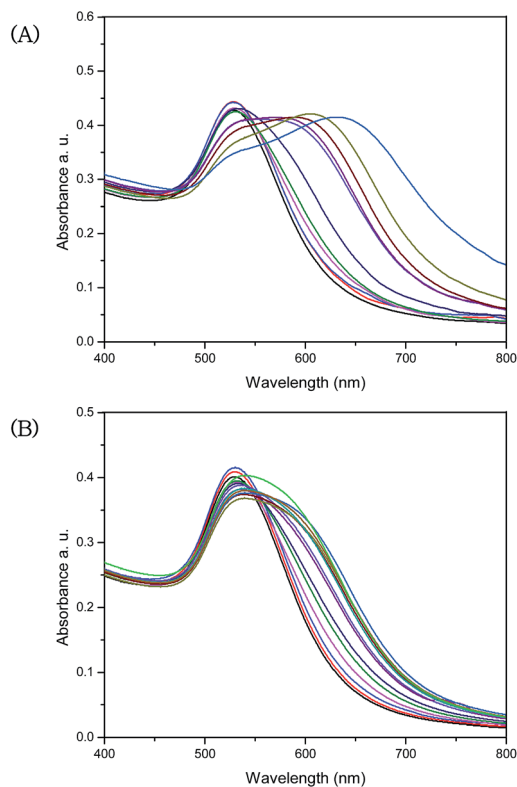


Fig. 12 UV-Vis absorption spectra of (A) drinking water and (B) tap water with different concentrations of Ba^{2+} ($20\text{--}300 \mu\text{mol L}^{-1}$), separately for both samples.



(UAS-CA-281). The authors are indebted to Prof. Francisco Ruiz Medina and Ph.D. María Isabel Pérez Montfort of Centro de Nanociencias y Nanotecnología, Universidad Nacional Autónoma de México (CNYN-UNAM) for analytical support and his valuable contributions to this work.

References

- (a) J. O. Nriagu and J. M. Pacyna, *Nature*, 1988, **333**, 134–139; (b) O. Kristofova, L. Trnkova, V. Adam, J. Zehnalek, J. Hubalek, P. Babula and R. Kizek, *Sensors*, 2010, **10**, 5308–5328.
- J. Kravchenko, T. H. Darrah, R. K. Miller, H. K. Lyerly and A. Vengosh, *Environ. Geochem. Health*, 2014, **36**, 797–814.
- R. P. Modi, V. N. Mehta and S. K. Kailasa, *Sens. Actuators, B*, 2014, **195**, 562–571.
- S. Martin and W. Griswold, *Environ. Sci. Technol.*, 2009, **15**, 1–6.
- R. Suwa, K. Jayachandran, N. T. Nguyen, K. Fujita and H. Saneoka, *Arch. Environ. Contam. Toxicol.*, 2008, **55**, 397–403.
- Canadian Council of Ministers of the environment-CCME, *Barium Canadian soil quality guidelines for the protection of environmental and human health, Canadian environmental quality guidelines*, 1999.
- W. de Oliveira, M. F. Batista de Carvalho, E. de Almeida, A. A. Menegário, R. Naves Domingos, A. L. Brossi-Garcia, V. F. do Nascimento Filho and R. Erthal Santelli, *Talanta*, 2012, **100**, 425–431.
- M. Llugany, C. Poschenrieder and J. Barcelo, *Arch. Environ. Contam. Toxicol.*, 2000, **39**, 440–444.
- (a) G. Machata, *Handbook on Toxicity of Inorganic Compounds*, ed. H. G. Seiler and H. Sigel, Marcel Dekker Inc, New York, 1988, p. 97; (b) P. Patnaik, *Handbook of Inorganic Chemical Compounds*, McGraw-Hill, New York, 2003, p. 77; (c) B. Kukiattrakoon, C. Hengtrakool and U. J. Kedjarune-Leggat, *Dental Science*, 2010, **5**, 189–200.
- (a) M. Licchelli, A. O. Biroli and A. Poggi, *Org. Lett.*, 2006, **8**, 915–918; (b) P. N. Basa, A. Bhowmick, M. M. Schulz and A. G. Sykes, *J. Org. Chem.*, 2011, **76**, 7866–7871.
- (a) Z. Jian-Min, Z. Qian-Shou and C. Chun-Feng, *J. Org. Chem.*, 2010, **75**, 5092–5098; (b) D. Yun, E. Cho, S. D. Dindulkar and S. Jung, *Macromol. Mater. Eng.*, 2016, **301**, 805–811.
- (a) A. Shailini, P. Barrie, B. Graeme, H. Kimberly and R. Malcolm, *Anal. Chim. Acta*, 2009, **653**, 191–199; (b) G. Huelga-Suarez, M. Moldovan, A. Garcia-Valiente, E. Garcia-Vazquez and J. I. Garcia Alonso, *Anal. Chem.*, 2012, **84**, 127–133; (c) G. Huelga-Suarez, B. Fernández, M. Moldovan and J. I. Garcia Alonso, *Anal. Bioanal. Chem.*, 2013, **405**, 2901–2909.
- S. V. Chikineva, V. P. Kovaleva and N. V. Gladyshev, *J. Anal. Chem.*, 2001, **56**, 449–452.
- P. Hooshang and F. Abbas, *Anal. Sci.*, 2000, **16**, 575–577.
- (a) A. Majzik, L. Füllöp, E. Csapó, D. Sebök, T. Martinek, F. Bogár, B. Penke and I. Dékány, *Colloids Surf., B*, 2010, **81**, 235–241; (b) E. Priyadarshini and N. Pradhan, *Sens. Actuators, B*, 2017, **238**, 888–902.
- A. Szalai, A. Sipos, E. Csapó, L. Tóth, M. Csete and I. Dékány, *Plasmonics*, 2013, **8**, 53–62.
- M. S. Han, A. K. R. Lytton-Jean, B. K. Oh, J. Heo and C. A. Mirkin, *Angew. Chem.*, 2006, **45**, 1807–1810.
- T. Lou, Z. Chen, Y. Wang and L. Chen, *ACS Appl. Mater. Interfaces*, 2011, **3**, 1568–1573.
- L. J. Miao, J. W. Xin, Z. Y. Shen, Y. J. Zhang, H. Y. Wang and A. G. Wu, *Sens. Actuators, B*, 2013, **176**, 906–912.
- F. Zhang, L. Zeng, Y. Zhang, H. Wang and A. Wu, *Nanoscale*, 2011, **3**, 2150–2154.
- M. S. Yavuz, G. C. Jensen, D. P. Penaloza, T. A. P. Seery, S. A. Pendergraph, J. F. Rusling and G. A. Sotzing, *Langmuir*, 2009, **25**, 13120–13124.
- I. Abdulhalim, M. Zourob and A. Lakhtakia, *Electromagnetism*, 2008, **28**, 214–242.
- T. Lou, L. Chen, Z. Chen, Y. Wang, L. Chen and J. Li, *ACS Appl. Mater. Interfaces*, 2011, **3**, 4215–4220.
- J. Du, L. Jiang, Q. Shao, X. Liu, R. S. Marks, J. Ma and X. Chen, *Small*, 2013, **9**, 1467–1481.
- H. Li and L. Rothberg, *Proc. Natl. Acad. Sci. U. S. A.*, 2004, **101**, 14036–14039.
- C. S. Tsai, T. B. Yu and C. T. Chen, *Chem. Commun.*, 2005, 4273–4275.
- Z. Sun, Z. Cui and H. Li, *Sens. Actuators, B*, 2013, **183**, 297–302.
- Y. M. Sung and S. P. Wu, *Sens. Actuators, B*, 2014, **201**, 86–91.
- M. C. Daniel and D. Astruc, *Chem. Rev.*, 2004, **104**, 293–346.
- S. P. Wu, Y. P. Chen and Y. M. Sung, *Analyst*, 2011, **136**, 1887–1891.
- (a) Z. Jia, W. Yong, X. Xiaowen and Y. Xiurong, *Analyst*, 2011, **136**, 3865–3868; (b) B. Lian-Yang, Z. Yu-Ping, C. Na, C. Jun, Z. Xia-Mao and H. Lin-Feng, *Micro Nano Lett.*, 2011, **6**, 337–341; (c) K. Shrivastava, P. Maji and K. Dewangan, *Spectrochim. Acta, Part A*, 2017, **173**, 630–636.
- (a) H. Li, L. Zhang, Y. Yao, C. Han and S. Jin, *Supramol. Chem.*, 2010, **22**, 544–547; (b) X. Wu, W. Tang, C. Hou, C. Zhang and N. Zhu, *Microchim. Acta*, 2014, **181**, 991–998.
- F. Zhang, Y. Zhou, Y. Chen, Z. Shi, Y. Tang and T. Lu, *J. Colloid Interface Sci.*, 2010, **351**, 421–426.
- I. Ojea-Jiménez and J. M. Campanera, *J. Phys. Chem. C*, 2012, **116**, 23682–23691.
- J. Kimling, M. Maier, B. Okenve, V. Kotaidis, H. Ballot and A. Plech, *J. Phys. Chem. B*, 2006, **110**, 15700–15707.
- S. C. D'Andre and A. Y. Fadeev, *Langmuir*, 2003, **19**, 7904–7910.
- E. S. Gawalt, G. Lu, S. L. Bernasek and J. Schwartz, *Langmuir*, 1999, **15**, 8929–8933.
- W. E. Ford, F. Abraham, F. Scholz, G. Nelles, G. Sandford and F. Von Wrochem, *J. Phys. Chem. C*, 2017, **121**, 1690–1703.
- S. A. Paniagua, P. J. Hotchkiss, S. C. Jones, S. R. Marder, A. Mudalige, F. S. Marrikar, J. E. Pemberton and N. R. Armstrong, *J. Phys. Chem. C*, 2008, **112**, 7809–7817.
- M. Matzapetakis, N. Karligiano, A. Bino, M. Dakanali, C. P. Raptopoulou, V. Tangoulis, A. Terzis, J. Giapintzakis and A. Salifoglou, *Inorg. Chem.*, 2000, **39**, 4044–4051.



- 41 (a) S. Aryal, B. K. C. Remant, N. Dharmaraj, N. Bhattarai, C. H. Kim and H. Y. Kim, *Spectrochim. Acta, Part A*, 2006, **63**, 160–163; (b) V. Lupusoru, D. A. Pricpo, M. Andries and D. Creanga, *J. Mol. Struct.*, 2016, **1126**, 192–199.
- 42 P. Fiurasek and L. Reven, *Langmuir*, 2007, **23**, 2857–2866.
- 43 S. D. Techane, L. J. Gamble and D. G. Castner, *J. Phys. Chem.*, 2011, **115**, 9432–9441.
- 44 M. R. Ivanov and A. J. Haes, *Anal. Chem.*, 2012, **84**, 1320–1326.
- 45 Y. Guo, Z. Wang, H. Shao and X. Jiang, *Carbon*, 2013, **52**, 583–589.
- 46 W. W. Yang, Y. Y. Di, Y. X. Kong, Y. F. Zhu and Z. C. Tan, *J. Chem. Eng. Data*, 2009, **54**, 2038–2044.
- 47 (a) K. Latham, K. F. White, K. B. Szpakolski, C. J. Rix and J. M. White, *Inorg. Chim. Acta*, 2009, **362**, 1872–1886; (b) D. Saha, T. Maity, R. Sen and S. Koner, *Polyhedron*, 2012, **43**, 63–70; (c) Z. P. Wang, B. Hu, X. H. Qi, N. N. Shen and X. Y. Huang, *Dalton Trans.*, 2016, **45**, 8745–8752.
- 48 D. Su, X. Yang, Q. Xia, F. Chai, C. Wang and F. Qu, *RSC Adv.*, 2013, **3**, 24618–24624.
- 49 X. Han, J. Goebl, Z. Lu and Y. Yin, *Langmuir*, 2011, **27**, 5282–5289.
- 50 R. Pamies, J. G. Hernández Cifre, V. Fernández Espín, M. Collado-González, F. G. Díaz Baños and J. García de la Torre, *J. Nanopart. Res.*, 2014, **16**, 2376.
- 51 M. A. Escudero Francos, R. Badial Laiño and M. E. Díaz-García, *Microchim. Acta*, 2015, **182**, 1591–1598.
- 52 M. V. Zyuzin, T. Honold, S. Carregal-Romero, K. Kantner, M. Karg and W. J. Parak, *Small*, 2016, **12**, 1723–1731.
- 53 S. Volden, A. L. Kjøniksen, K. Zhu, J. Genzer, B. Nyström and V. R. Glomm, *ACS Nano*, 2010, **4**, 1187–1201.

