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PAPER

Evaluation of a temporal trend heavy metals contamination in *Posidonia oceanica* (L.) Delile, (1813) along the western coastline of Sicily (Italy)

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The use of biological species in the monitoring of marine environmental quality allows the evaluation of biologically available levels of contaminants in the ecosystem and the effects of contaminants on living organisms. The seagrass *Posidonia oceanica* is a useful bioindicator because through the lepidochronology technique it is possible to obtain a historical contamination trend of a given area. This study aims to assess the temporal trend contamination by heavy metal investigations on dead sheaths of 100 samples of *P. oceanica* collected in the Protected Marine Area of “Plemmirio” (Sicily) and in the Siracusa bay. Important results were obtained because data show a significant negative temporal trend for the metals analysed especially for As, Co, Cr, Hg, Pb, Se, U and V that in the past had higher concentrations, with a stronger contamination in the Plemmirio area, the site much more exposed to the pollution of the nearby petrochemical complex. This study confirms the relevance of the use of *P. oceanica* as a biological indicator of metal contamination in coastal ecosystems. Thus the usefulness of *P. oceanica* as a tracer of spatial metal contamination and as a good tool for water quality evaluation is reinforced.

1. Introduction

Heavy metals contamination on the global scale is intensely studied because of their persistence in the environment, their toxicity and the risk of biomagnification through food chains.^{1–12}

The accumulation of toxic substances in hazardous concentrations in biota has led scientists in recent years to deepen the environmental monitoring programs using bioindicator organisms.^{5,6,12–20}

The species considered is the endemic seagrass *Posidonia oceanica* that holds a central position in the ecology of the Mediterranean Sea. The importance of this species lies in its extension,

in its high productivity and stability, and also because its meadows are functional as spawning area, hunting territory or permanent habitat for numerous plant and animal species.^{3,21–25}

This marine phanerogame has all the qualities required to be used as a biological indicator of chemical pollution as a benthic species widely spread throughout the Mediterranean basin, with a great power of concentration of chemical pollutants. For twenty years *P. oceanica* has been studied by some authors with the aim of assessing its usefulness as a trace metal biomonitor.^{14,22,26–35}

Since the early 50's, the southeast area of Sicily (Priolo) has become one of the largest and most complex petrochemical sites in Europe, containing several industrial plants, mainly consisting of oil refineries, petrochemical plants, power stations, military base and many other industrial installations. This has led to a growing concern about the possible health effects caused by environmental exposure of residents in neighboring towns until 1990, when the industrial area of Augusta-Priolo, in the Siracusa

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Environmental impact

This manuscript “Evaluation of a temporal trend heavy metals contamination in *Posidonia oceanica* (L.) Delile, (1813) along the western coastline of Sicily (Italy)” by analysing heavy metal concentrations in *Posidonia oceanica* from 1999 to 2009 contributes to improved knowledge of the quality of seawater, and the degree of metals contamination over a long period. This is the first report of coastal contamination of Sicily seas. Indeed, there are few published studies on this topic in the area of Sicily and in particular near one of the largest petrochemical complexes of Italy.

area, was declared an “area at high risk of environment crisis” in accordance with Italian law 349 of 8/7/86.

Sampling sites are located a few kilometres to the south from this industrial complex, and are examples of coastal environments with a relatively high influx of unregulated industrial and domestic effluents. In the petrochemical area there is a well documented bioavailability of contaminants such as heavy metals, PCBs and PAHs:^{36–38} a study conducted in this area on surveillance of congenital malformations and abortions revealed a frequency much higher than the national average in 2001, probably correlated with the petrochemical pollution source.^{38,39} Another study revealed high concentrations of Hg, PCB and HCB in milk and hair of pregnant women residents in Augusta with respect to the control area of Catania.³⁸

The aim of our study was to monitor the status of the marine environment in two stations of the southern Ionic coast of Sicily, the Protected Marine Area (P.M.A.) of “Plemmirio”, since 2003, and Siracusa bay. Matching the lepidochronological technique for the identification of yearly cycles of leaf production of *P. oceanica* rhizomes with a mass spectrometry technique for heavy metals analysis, we measured the temporal trend accumulation of heavy metals from 1999 to 2009.

2. Materials and methods

2.1 Sampling and sample preparation

In the summer of 2009 samples of *P. oceanica* were collected by scuba diving at a depth of 10 m in the Plemmirio P.M.A. station and in the Siracusa bay station (Fig. 1).

Each sample area was divided into 10 grids of 1 m² of about 1 m away from each other. At each grid point were collected 10

plants. In each plant, dead sheaths were separated by year with the lepidochronology technique, till 1999 and no more, due to the reduction of biomass to be analyzed.

Sheaths from each grid of the same year (sample group) were mixed and washed first in fresh water to have the mechanical removal of epiphyte and sediment, then in ultrapure water. Each sample group was dried and frozen at –20 °C before analysis.

2.2. Trace metals analysis

For the lab tests 1 g of each analytical sample group was weighed with an analytical scale (Mettler Toledo xs105) and mineralized in a microwave system Ethos Touch Control (MILESTONE S.r.l.—Italy) with a digestion solution prepared with 6 ml of 65% nitric acid (HNO₃) (Carlo Erba) and 2 ml of 30% peroxide hydrogen (H₂O₂) (Carlo Erba). After mineralization, ultrapure water (Merck) was added to the samples up to 30 ml.

The test for heavy metals analysis was performed with an ICP-MS Elan DRce (PERKIN ELMER-USA). Standards for the instrument calibration were prepared on the basis of multi-element certified reference solution ICP Standard (Merck).

The system was calibrated using a calibration line made up of 5 points of increasing concentration for each element, with a linear regression (*R*²) for the calibration line between 0.9931 and 0.9991.

Analytical blanks and recovery samples were run in the same way as the real samples and concentrations were determined using standard solutions prepared in the same acid matrix to validate the calibration.

Analytical quality control was performed with certified reference material BCR-670-Aquatic Plant, and with NIST-1946-Lake Superior Fish (Table 1).

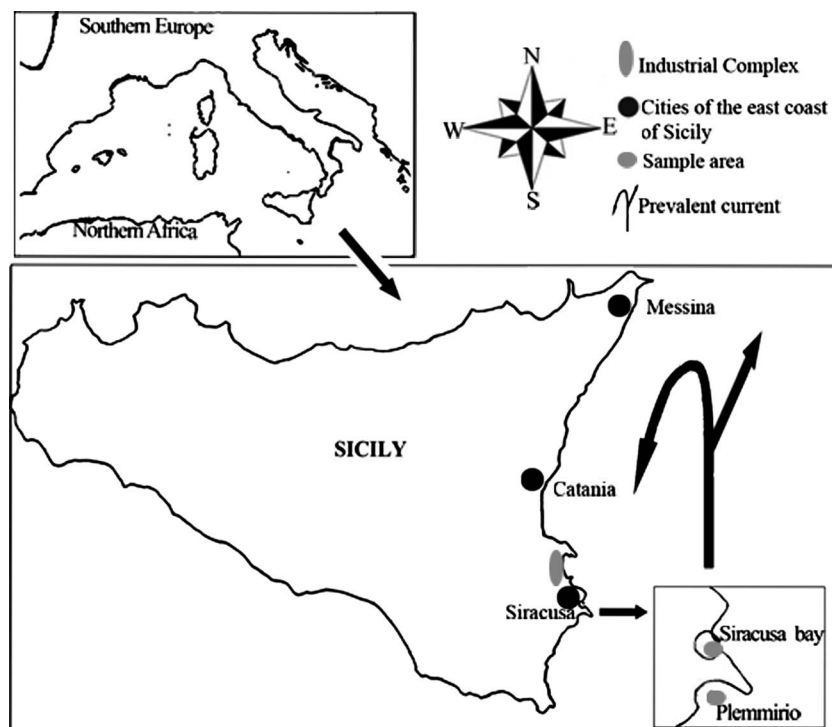


Fig. 1 Study area. Data on prevalent surface Ionian current are from Brandt *et al.* (1999).⁴⁰

Table 1 Analysis of trace metals with certified reference material BCR-670 (Aquatic Plant) ($\mu\text{g g}^{-1}$ dry wt) and with NIST-1946 (Lake Superior Fish) ($\mu\text{g g}^{-1}$ weight wt)

Metals	BCR-670	
	Certified	Found
As	1.98 ± 0.19	1.9 ± 0.07
Cd	0.075 ± 0.002	0.065 ± 0.001
Cr	2.05 ± 0.10	2.66 ± 0.32
Cu	1.82 ± 0.30	1.66 ± 0.05
Pb	2.06 ± 0.12	1.90 ± 0.06
Zn	24 ± 2.1	23.05 ± 2.05
Metals	Not certified	
		Found
Ni	2.12–2.58	2.63 ± 0.39
Se	0.149–0.273	0.238 ± 0.05
Metal	NIST-1946	
	Certified	Found
Hg	0.433 ± 0.009	0.442 ± 0.02

To compare the total metal content in different years and at different stations, the metal pollution index (MPI) defined by Usero *et al.* (2005)⁴¹ was used. It is obtained with the following equation: $\text{MPI} = (\text{Cf1} \times \text{Cf2} \times \text{Cfn})^{1/n}$, where Cfn is the concentration of the metal *n* in the sample.

2.3. Statistical analysis

The difference between metals concentrations of the stations was evaluated with a Student's *t*-test for the paired sample, applying $p = 0.001$ as the level of significance, using the software package SPSS 14.0.

3. Results

3.1. Comparison of the stations

The concentrations of most of the metals vary considerably depending on the location of the sampling stations: As, Cr, Cd, Cu, Ni ($P < 0.001$) and Hg ($P < 0.05$) concentrations in the Plemmirio station are significantly higher than in the Siracusa bay station; the highest Zn concentration is recorded in the Siracusa bay station ($P < 0.001$); Co, Pb, Se, U and V concentrations do not show significant differences (Fig. 2). Although there are differences in concentrations, for some metals there is a clear decrease of pollution load in both sites, as it can be seen from the value of the metal pollution index, MPI (Tables 2 and 3).

4. Discussions and conclusion

As shown in Fig. 2, for many metals such as As, Co, Hg, Pb, Se, V and U there is a clear downward trend over time in both stations analyzed. Cd, Ni and Zn show no significant temporal fluctuations, Cr and Cu do not show temporal oscillations in the Siracusa bay, while in Plemmirio there is a downward trend from 1999 to 2009.

The dead sheaths, although less sensitive to environmental changes than the living parts of the plant, are an index of

environmental contamination over long periods.³³ The persistence of metals from leaves to dead sheaths is not clear, and consequently it is unknown with certainty which are the metals that, found in the scale, reflect perfectly the concentrations in the living plant.

Some authors in fact noted that Cr, Ni, Cd, Pb and Zn accumulate preferably in leaves, as a result of a longer exposure to the metal load of the marine environment.^{14,34,42} Gosselin *et al.* (2006) instead observed concentrations of Cr and Pb major in dead sheaths, and Ni, Cd and Zn highest in the leaves. Kljakovic-Gaspic *et al.* (2004)⁴³ however showed no significant difference in Cd and Pb between the two parties, as Gosselin *et al.* (2006) reported for Cu.

A comparison with bibliographic data on bioaccumulation of heavy metals in *P. oceanica* revealed that concentrations we found are consistently lower than those found by other authors.

In particular the values of Cu found in the Siracusa bay (Table 2) are lower than those found in the literature, but those of Plemmirio (Table 3) are in the range of concentrations found by other authors in contaminated and non-contaminated sites, except for values found by Conti *et al.* (2007) in Ustica (Sicily, Italy) and Malea (1994)⁴⁴ in Antikyra Gulf (Greece), significantly higher than ours. The presence of Cu concentrations >16 ppm in some areas is linked to the use of this element as a fungicide in vineyards,³³ which are widespread in the province of Siracusa, or as antifouling agent in boat paints, as our study area has a heavy vessel traffic.

The data on the bioaccumulation of Cd, whose uptake occurs through a passive process that depends on the leaf surface exposed to a subsequent translocation in the roots and rhizomes,^{22,45} are among average values found in the literature.^{33,46}

With regard to Ni, only Gosselin *et al.* (2006) evaluated the concentrations in dead sheaths, finding values between 9 and 37 ppm in contaminated sites, and 111 ppm in sites contaminated by mining operation, values significantly higher than those detected by us in all sample groups, except one dated 1999 from Plemmirio, which corresponds to the year with greater accumulation of all metals in this area (Tables 2 and 3).

Zn is the only element found in higher concentrations in the dead sheaths of *P. oceanica* sampled in the Siracusa bay. The increased presence of Zn in this area could result from galvanizing operations of vessels, because within the bay there is a small marina and a small area for mooring vessels. The concentration range of Zn rated by other authors in dead sheaths (14–61 ppm)^{33,46} appears to be greater than ours (Tables 2 and 3). This indicates a lower contamination of Zn in our study area.

The temporal trend of Pb found in the sites is in agreement with the decreasing contamination of this element in the atmosphere and the sea surface, and is likely to be the result of progressive reduction in the use of lead-based additives in car fuels.⁴⁶ The concentration of lead found at both sites is among the lowest values reported in the literature on analysis made both in living leaves^{26,28,33,44,47,48} and in dead sheaths,^{33,46} except for the values reported by Conti *et al.* (2010)⁴⁹ in Linosa Island (Sicily, Italy), values that fall within our concentration range.

The concentrations of Cr in studies in different areas of the Mediterranean Sea in leaves of *P. oceanica* range from a minimum of 0.11 to a maximum of 5.73 ppm

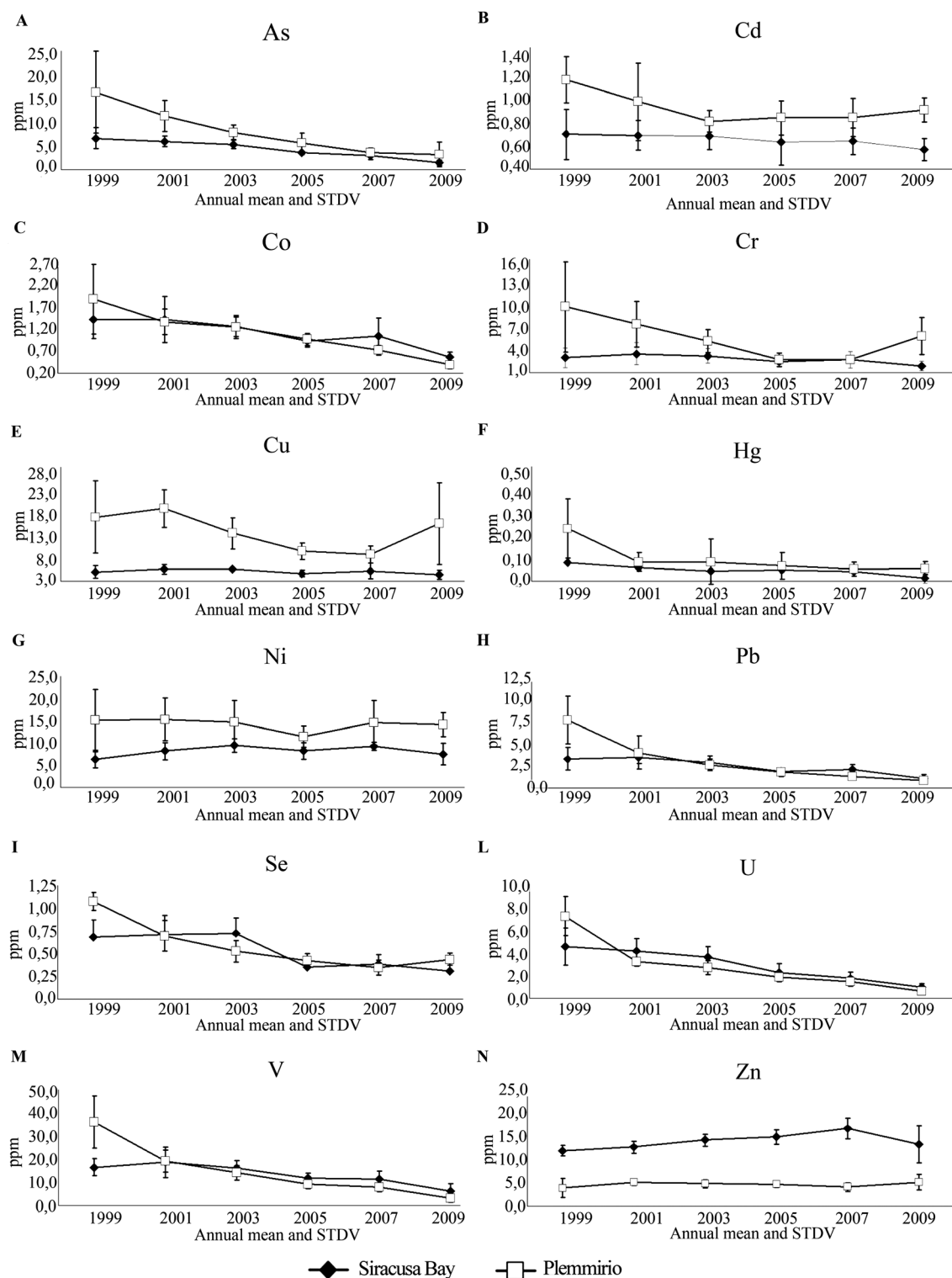


Fig. 2 Temporal trend contamination of heavy metals in *P. oceanica* dead sheaths of the Siracusa bay and Plemmirio ($\mu\text{g g}^{-1}$ dry wt).

(dry wt).^{27,33,34,42,47,50,51} The average concentrations of this element in studies of dead sheaths by Gosselin *et al.* (2006) at contaminated and non-contaminated sites in the north of Corsica are significantly higher than ours (Tables 2 and 3); instead Pergent in Pointe-Martini Chèvre (France) found an average concentration of 2.8 ppm (dry wt), comparable to our results.

The Hg concentration in the living sheaths remains stable throughout the year following the death of the dead sheaths, and it shows that no appreciable mercury flux (absorption or desorption) occurs in this tissue after the blade shedding.⁵² So the analysis of Hg in dead sheaths of *P. Oceanica* allows reconstruction of the trend contamination of coastal sites.⁵² Results obtained by other authors in sites considered contaminated have

Table 2 Mean concentrations ($\mu\text{g g}^{-1}$ dry wt), standard deviation and metal pollution index (MPI) of the analytical sample groups from dead sheaths of *P. oceanica* in the P.M.A. of "Plemmirio"

Year	Cd	Pb	Zn	Co	V	U	As	Cu	Ni	Cr	Hg	Se	MPI
1999	0.69 ± 0.21	3.43 ± 1.34	12.62 ± 1.23	1.42 ± 0.44	16.44 ± 3.77	4.54 ± 1.65	6.65 ± 2.21	5.10 ± 1.42	6.31 ± 1.80	2.94 ± 1.30	0.08 ± 0.01	0.68 ± 0.19	2.63
2001	0.68 ± 0.12	3.62 ± 0.75	13.39 ± 1.44	1.43 ± 0.53	18.60 ± 6.56	4.14 ± 1.08	5.99 ± 1.12	5.70 ± 1.04	8.18 ± 1.96	3.50 ± 1.48	0.06 ± 0.01	0.71 ± 0.21	2.69
2003	0.67 ± 0.11	3.02 ± 0.78	14.99 ± 1.38	1.26 ± 0.23	16.15 ± 3.19	3.62 ± 0.94	5.33 ± 0.86	5.74 ± 0.10	9.49 ± 1.54	3.21 ± 0.92	0.04 ± 0.01	0.73 ± 0.17	2.52
2005	0.63 ± 0.07	2.00 ± 0.22	15.73 ± 1.69	0.94 ± 0.14	11.67 ± 2.11	2.25 ± 0.80	3.59 ± 0.36	4.77 ± 0.75	8.26 ± 1.90	2.42 ± 0.43	0.05 ± 0.01	0.35 ± 0.08	1.93
2007	0.63 ± 0.11	2.17 ± 0.61	17.68 ± 2.36	1.05 ± 0.41	11.30 ± 3.45	1.75 ± 0.53	2.93 ± 0.80	5.35 ± 1.73	9.29 ± 0.94	2.68 ± 1.10	0.04 ± 0.01	0.39 ± 0.11	1.95
2009	0.56 ± 0.09	1.14 ± 0.46	14.02 ± 4.21	0.57 ± 0.12	6.23 ± 3.21	0.97 ± 0.32	1.43 ± 0.55	4.46 ± 1.02	7.57 ± 2.47	1.88 ± 0.63	0.01 ± 0.02	0.31 ± 0.10	1.22

Table 3 Mean concentrations ($\mu\text{g g}^{-1}$ dry wt), standard deviation and metal pollution index (MPI) of the analytical sample groups from dead sheaths of *P. oceanica* in the Siracusa bay

Year	Cd	Pb	Zn	Co	V	U	As	Cu	Ni	Cr	Hg	Se	MPI
1999	1.13 ± 0.19	8.05 ± 2.79	4.11 ± 2.22	1.89 ± 0.80	36.20 ± 11.39	7.21 ± 1.70	16.38 ± 8.68	17.68 ± 8.20	15.23 ± 6.90	9.70 ± 5.98	0.23 ± 0.13	1.08 ± 0.10	4.90
2001	0.96 ± 0.32	4.18 ± 1.94	5.34 ± 0.64	1.37 ± 0.29	19.11 ± 4.67	3.18 ± 0.37	11.34 ± 3.21	19.61 ± 4.35	15.39 ± 4.86	7.42 ± 3.02	0.09 ± 0.04	0.70 ± 0.17	3.43
2003	0.79 ± 0.09	2.72 ± 0.65	5.07 ± 0.92	1.26 ± 0.26	14.05 ± 3.23	2.68 ± 0.61	7.89 ± 1.54	13.89 ± 3.55	14.69 ± 4.99	5.13 ± 1.52	0.09 ± 0.10	0.53 ± 0.12	2.75
2005	0.83 ± 0.14	1.86 ± 0.48	5.00 ± 0.71	0.97 ± 0.14	9.12 ± 2.00	1.86 ± 0.31	5.69 ± 2.00	9.89 ± 1.90	11.46 ± 1.92	2.68 ± 0.88	0.07 ± 0.06	0.42 ± 0.08	2.06
2007	0.83 ± 0.16	1.37 ± 0.35	4.34 ± 0.81	0.72 ± 0.11	7.69 ± 1.89	1.44 ± 0.37	3.65 ± 0.88	9.08 ± 1.92	14.66 ± 4.95	2.72 ± 0.28	0.05 ± 0.03	0.35 ± 0.08	1.77
2009	0.89 ± 0.10	0.91 ± 0.40	5.40 ± 1.80	0.38 ± 0.08	3.16 ± 1.77	0.59 ± 0.06	3.26 ± 2.59	16.18 ± 9.33	14.16 ± 2.69	5.82 ± 2.43	0.06 ± 0.03	0.44 ± 0.07	1.62

average concentrations ranging between 0.09 and 0.51 ppm^{25,30,53,54} comparable to those found in Plemmirio from 1999 to 2003 (Table 3). Although the Mediterranean Sea is rich in natural emissions of this element,^{37,55} our data attest to a past contamination of industrial origin, but at the same time seem to reflect a decrease.

About the accumulation of As in dead sheaths Gosselin *et al.* (2006) found average values ranging between 14 and 20 ppm, in the years between 1999 and 2003. The values of Siracusa bay are lower (Table 2) than those found in the bibliography. Plemmirio values (Table 3) are within the range reported by other authors. The presence of this metal in the area is due both to natural events, such as volcanic activity, and to human activities, such as the use of herbicides and antifouling agents.

The results obtained for U, V, Co and Se show a similar decreasing trend over time with concentrations of metals overlapping in both sites, as well as for Pb.

The first consideration is certainly the greatest occurrence of metal load in Plemmirio, probably due to greater movement of water with respect to the Siracusa bay, the latter being a half-closed area with low hydrodynamic movement of water, despite being closer to the industrial area. The increased exposure of Plemmirio to pollutants is also underlined by the values of MPI when compared year to year with that of the Siracusa bay (Tables 2 and 3).

Certainly, important results were obtained because data show a negative temporal trend for the metals analysed especially for As, Co, Cr, Hg, Pb, Se, U and V that in the past had higher concentrations.

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